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## Research Letters

# High genetic diversity and strong spatial genetic structure in *Cabralea canjerana* (Vell.) Mart. (Meliaceae): implications to Brazilian Atlantic Forest tree conservation



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## ABSTRACT

The Atlantic Forest is the biome most severely affected by deforestation in Brazil. *Cabralea canjerana* ssp. *canjerana* is a dioecious tree species that due to its wide distribution and reproductive biology traits can be considered a model for Atlantic Forest conservation studies. Using microsatellite loci, the genetic diversity and its structure were evaluated in natural subpopulations of this species located in forest fragments in Southern Minas Gerais State, Brazil. High levels of genetic diversity and a strong genetic structure among these subpopulations were found. Genotype assignment analyses suggested that the eight subpopulations can be grouped into two clusters, probably as a consequence of environmental landscape elements affecting gene flow patterns. High levels of spatial genetic structure were found among subpopulations, suggesting that pollen mediated gene flow and seed dispersal occur mainly within small distances.

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## Introduction

Brazil is recognized as one of the most important repositories of worldwide biological diversity. The Atlantic Forest has high levels of biodiversity and endemism but it is the biome

most severely affected by deforestation in Brazil. Habitat fragmentation alters the patterns of random genetic distribution within and between populations and changes the patterns of pollen and seed dispersal (Ghanzoul, 2005). This process has a direct influence on the genetic composition of populations and can produce negative genetic effects like inbreeding and

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**Table 1 – Geographical and genetic traits of *C. canjerana* subpopulations.**

Subpopulations	Latitude S	Longitude W	Elevation (m)	Population density	Isolation degree (km)	Fragment area (ha)	N	$A_R$	$H_O$	$\bar{h}$	f
Spp. 1	22°41'11"	45°54'28"	1681	26	4.08	0.90	27	5.560	0.716	0.667	-0.076
Spp. 2	22°40'59"	45°53'58"	1667	76	4.10	1.40	33	6.370	0.718	0.737	0.025
Spp. 3	22°41'12"	45°54'11"	1630	86	3.89	12.00	31	6.666	0.773	0.749	-0.033
Spp. 4	22°41'18"	45°54'06"	1603	86	3.51	7.50	33	5.486	0.733	0.683	-0.075
Spp. 5	22°41'35"	45°53'32"	1598	187	2.92	15.00	17	6.651	0.615	0.727	0.156
Spp. 6	22°41'43"	45°53'28"	1600	67	2.73	1.95	11	6.238	0.675	0.713	0.055
Spp. 7	22°42'19"	45°52'52"	1810	221	2.03	4000.00	39	7.272	0.675	0.750	0.101
Spp. 8	22°42'47"	45°53'52"	1928	298	0.00	4000.00	30	7.391	0.579	0.779	0.259

N = number of sampled individuals;  $A_R$  = average allelic richness;  $H_O$  = observed heterozygosity;  $\bar{h}$  = Nei's gene diversity; f = intrapopulation fixation index. The isolation degree is the distance (km) between all subpopulations and the subpopulation 8, taken as reference. Population density refers to the number of adult *C. canjerana* trees per hectare.

genetic drift. Also, habitat fragmentation may decrease reproductive fitness, disease resistance and population adaptation to new environments increasing susceptibility to local extinction (Fahrig, 2003). These considerations explain the inclusion of genetic diversity levels by IUCN as an important parameter to be used in the characterization of endangered species (IUCN, 2000). The understanding of the spatial distribution of genetic diversity within and among populations is considered fundamental to the design of efficient conservation strategies (Caballero et al., 2010). The knowledge about the distribution patterns of genotypes in the landscape is extremely important for the conservation of natural resources because it helps to define in situ conservation strategies for endangered species.

*Cabralea canjerana* belongs to the family Meliaceae and comprises three subspecies. One of them, *C. canjerana* ssp. *canjerana*, is a dioecious tree species that occurs in Neotropical forests. The genus *Cabralea* has a range of distribution from Costa Rica to north of Argentina. In Brazil, *C. canjerana* ssp. *canjerana* is abundant from southern Minas Gerais to the Rio Grande do Sul State, predominantly in mountain ranges and Ombrophilous Dense Forests, being absent only in the Araucaria Forest (Barreiros and Souza, 1986).

The aim of our study was to evaluate the levels and structure of genetic diversity in subpopulations of *C. canjerana* ssp. *canjerana* from forest fragments of different sizes. To test the hypothesis that habitat fragmentation does affect the pattern of genetic diversity distribution, the genetic diversity of eight subpopulations of *C. canjerana* was evaluated at a small spatial scale in an area of approximately 10 km<sup>2</sup> in high-altitudes of Atlantic Forest (montane forest) fragments. The level of spatial genetic structure among these subpopulations was also evaluated to study the effects of habitat fragmentation on pollen and seed dispersion.

## Materials and methods

### Sampled sites and molecular procedures

The sampling sites are located in the Fernão Dias Environmental Protection Area (EPA), situated in a mountain range in southern Minas Gerais State, Brazil and occupy an area of 180,073 ha. Samples were taken in six small fragments (area  $\leq$  15 ha), of which two are also corridor fragments, and in two

large fragments located in a continuous forest area, totaling eight sampled sites. Each sampled site was considered as a subpopulation (Table 1).

The Atlantic Forest was continuous in the study area until 1910, when human activity increased its footprint in the landscape. Intense and inappropriate logging along the past 60 years has produced severe habitat fragmentation. The deforestation became more intense after 1950, due to timber exploitation. These fragments decreased in size until 1985, when timber exploitation and the expansion of agriculture became less intense in the area. Currently, several fragments are well conserved, but others are still subject to selective logging and cattle grazing (Aguari, 2001).

Leaves of *C. canjerana* adult plants were collected randomly in the eight sampled sites. On average, 28 individuals were sampled from each subpopulation, totaling 221 individuals. Genotyping was carried out for six microsatellite loci described by Pereira et al. (2011).

### Statistical analyses

Allelic richness and unbiased expected heterozygosity ( $H_E$ ) (Nei, 1973) were estimated for each subpopulation. The parameters intrapopulation fixation index (f), overall inbreeding coefficient (F) and the measure of genetic structure among subpopulations ( $\theta$ ) were estimated according to Weir and Cockerham (1984). These parameters are analogous to Wright's F-statistics,  $F_{IS}$ ,  $F_{IT}$  and  $F_{ST}$ , respectively. These estimates were also produced using FSTAT 2.9.3.2 (Goudet, 2002) software. Slatkin's  $R_{ST}$  genetic structure parameter was estimated using  $R_{ST}$ -calc (Goodman, 1997). Recent population bottleneck was evaluated for each subpopulation using Luikart and Cornuet (1998) empirical test available in Bottleneck software (Cornuet and Luikart, 1996). The presence of null allele frequencies was checked for each loci with 10<sup>4</sup> interactions in Gibbs sampling as implemented in INEst software (Chybicki and Burczyk, 2009).

To identify the number of clusters that best explained the genetic structure of the data set, an analysis was performed using Structure 2.3.3 (Falush et al., 2003), with default parameters. K values ranging from 1 to 10 were evaluated using 20 independent runs for each K and 10<sup>6</sup> MCMC steps, assuming a burn-in period of 50,000 steps. The suggested K values were evaluated using the  $\Delta K$  statistic developed by Evanno et al.

**Table 2 – Genetic structure parameters estimates obtained for six microsatellite loci, for eight *C. canjerana* subpopulations.**

Locus	N	A	$f^a$	$F^b$	$\theta^c$	$R_{ST}$
Ccan17	219	27	0.064 <sup>d</sup>	0.164 <sup>d</sup>	0.106 <sup>d</sup>	0.346
Ccan04	205	32	0.344	0.376	0.049	0.038
Ccan31	215	32	0.055 <sup>d</sup>	0.131 <sup>d</sup>	0.083 <sup>d</sup>	0.402
Ccan28	218	14	-0.259	-0.150	0.087 <sup>d</sup>	0.005
Ccan32	171	24	0.283	0.482	0.278	0.167
Ccan10	179	18	-0.184	0.039 <sup>d</sup>	0.189 <sup>d</sup>	0.178
Average	-	24.5	0.061	0.184	0.131 <sup>d</sup>	0.189

<sup>a</sup> Intrapopulation fixation index ( $F_{IS}$ ).<sup>b</sup> Total fixation index ( $F_{IT}$ ).<sup>c</sup> Population differentiation coefficient (analogous to  $F_{ST}$ ).<sup>d</sup> Statistically different from zero ( $p < 0.05$ ).

(2005). The level of genetic structure among the K clusters suggested by this analysis was also estimated.

The level of spatial genetic structure among the eight subpopulations was evaluated by the Mantel test with  $10^4$  permutations using pairwise  $R_{ST}$  values and the Euclidean geographical distances matrix. The geographical distances matrix was calculated using geographical coordinates data in UTM. Geographical distance matrix, Mantel test and dispersion plots were made in R software (R Core Team, 2013).

## Results

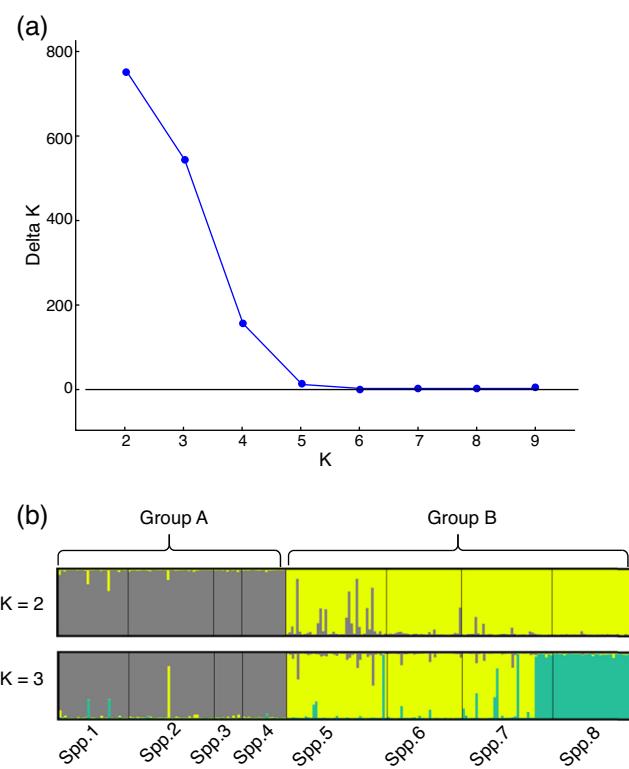
One hundred forty seven alleles were found in the eight *C. canjerana* subpopulations for the six loci evaluated, an average of 24.5 alleles per loci. The average genetic diversity in the eight subpopulations was high ( $\bar{h} = 0.725$ ), as was the average allelic richness ( $A_R = 6.454$ ). Subpopulations 7 and 8 showed the highest levels of genetic diversity and allelic richness (Table 1). No bottleneck effect was detected in the eight subpopulations, except for subpopulation 2, in which an excess of heterozygosity was found according to the Wilcoxon test ( $p = 0.015$ ). The frequency of null alleles for each locus was very low.

The genetic structure parameters estimates were  $f = 0.061$  [ $CI_{95\%} = -0.174$  to  $0.278$ ] and  $\theta = 0.131$  [ $CI_{95\%} = 0.066$ – $0.222$ ] showing a low level of within subpopulation inbreeding and a moderate genetic differentiation among subpopulations (Table 2). The  $\theta$  estimate reveals that 0.869 of the total genetic diversity is within subpopulations. The  $R_{ST} = 0.198$  ( $p < 0.01$ ) corroborates the evidence of a moderate genetic structure among all subpopulations. This pattern of genetic structure is typical of outcrossing plants species (Loveless and Hamrick, 1984). The assignment analysis showed that the eight subpopulations were structured into two clusters (Group A and Group B) according to the Evanno et al. (2005) criteria (Fig. 1). A high  $\Delta K$  value was also found for  $K=3$ . This third group refers mainly to individuals from subpopulation 8 and can be regarded as a subdivision of Group B.

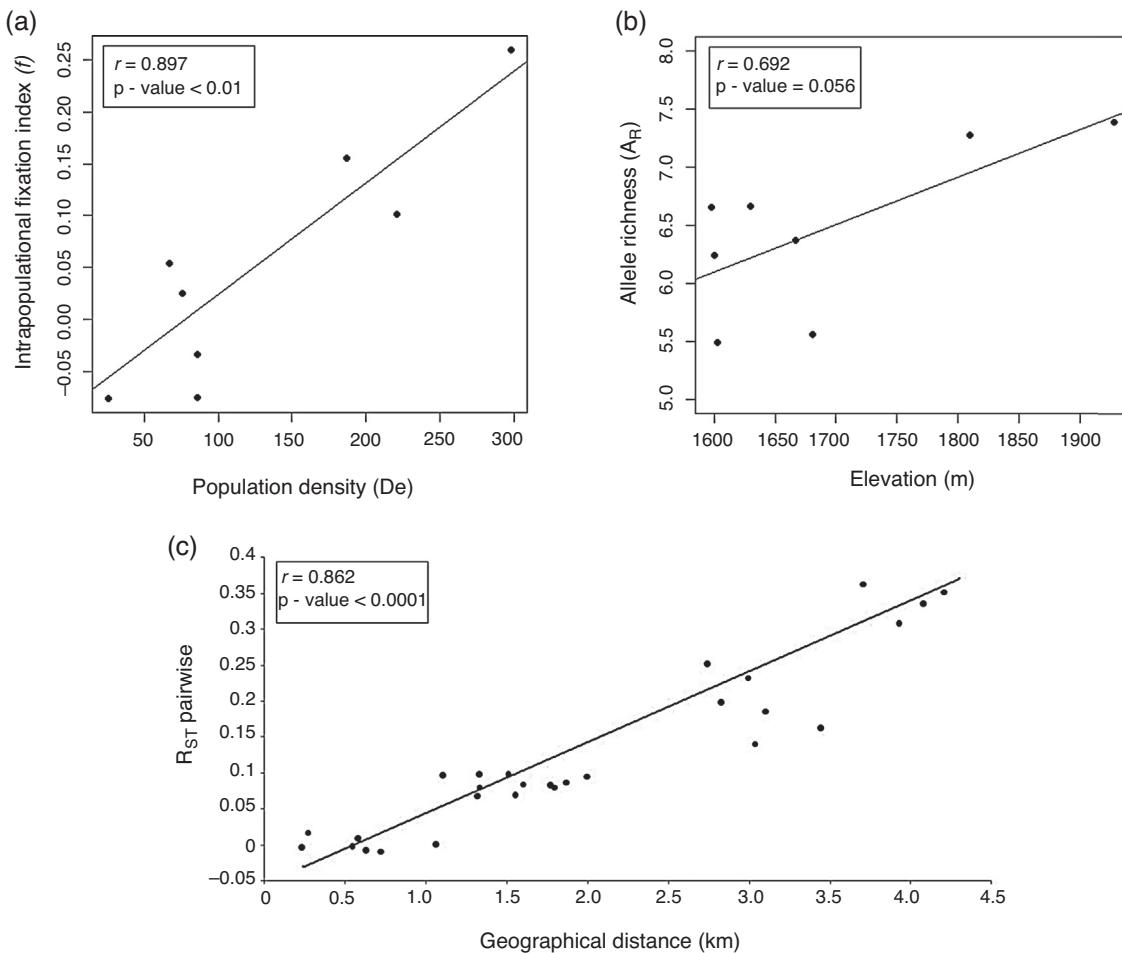
The genetic structure analysis including only the four subpopulations of Group A showed a low but significant genetic structure ( $\theta = 0.029$  [ $CI_{95\%} = 0.017$ – $0.041$ ]) and high genetic diversity ( $\bar{h} = 0.712$ ). The analysis of Group B showed a moderate and significant genetic structure ( $\theta = 0.074$  [ $CI_{95\%} = 0.04$ – $0.118$ ]) and a genetic diversity ( $\bar{h} = 0.755$ ) higher than that of Group A. The estimated overall apparent gene

flow among all subpopulations was 1.65 migrant individuals per generation. The apparent gene flow rates within Group A, Group B and between these groups were estimated at 8.370, 2.875 and 1.548 migrant individuals per generation, respectively.

The Pearson's correlation coefficient between the subpopulations pairwise  $R_{ST}$  values and geographical distances was high ( $r = 0.836$ ) and significant ( $p < 0.0001$ ). These results suggest that the pattern of genetic divergence among subpopulations is structured in the geographical landscape due to isolation by distance (Fig. 2).



**Fig. 1 – Genetic structure analysis results obtained by the Structure software. (a) Evanno et al. (2005)  $\Delta K$  statistic. (b) Bar plot with assignment probabilities of all *C. canjerana* individuals to the inferred clusters ( $K = 2$  and  $K = 3$ ).**



**Fig. 2 – Pearson correlation analysis results: (a) Population density and intrapopulation fixation index (*f*), (b) Elevation (m) and Allelic richness, and (c) Geographical distances (km) among eight *C. canjerana* subpopulations and  $R_{ST}$  pairwise values.**

## Discussion

### Level and distribution of genetic diversity in the landscape

*C. canjerana* subpopulations in the Fernão Dias EPA have a high level of genetic diversity. The genetic diversity estimates are in agreement with those found by Nybom (2004) in different plant tree species and are greater than those found for endemic species (Kang et al., 2008; Furches et al., 2009). The genetic diversity levels detected in Group A subpopulations were lower than those found in Group B, suggesting an effect of altitude (Fig. 2a) and that subpopulations from Group A have already suffered the effects of forest fragmentation. The intrapopulation fixation index estimates were strongly correlated to subpopulation densities (Fig. 2b). Interestingly, significant estimates of *f* were found in subpopulations 7 and 8 from a continuous forest area. Since *C. canjerana* is a dioecious species, this is probably a result of biparental inbreeding as a consequence of the spatial genetic structure within each subpopulation.

The estimate of  $\theta=0.131$  indicates a moderate level of genetic differentiation among the subpopulations. However,

due to the small-scale geographical distribution of these subpopulations, the detected genetic structure can be considered strong. The similarity between the  $\theta$  and  $R_{ST}$  estimates (Table 2) suggests that genetic drift is the major force shaping the genetic structure and that the dispersive effect of mutation is still of minor importance. The recent fragmentation scenario can explain the high levels of genetic diversity retained by the fragments. Franceschinelli et al. (2007), working in the same area with young plants of *Myrciaria floribunda*, a common Brazilian Atlantic Forest tree, also found high levels of genetic diversity.

Subpopulations from Group B are located at higher altitudes than those from Group A. The genetic differentiation within Group A was low, most likely because these subpopulations are very close to each other, favoring gene flow among them. In contrast, the genetic structure found in Group B can be related to the geographical distance and variation in altitude among these subpopulations. Although the apparent rate of gene flow in Group B was low, the genetic diversity measure in this group was high, most likely because subpopulations 7 and 8 are located in higher and better-preserved forest fragments.

The genetic structure between the two groups was high ( $\theta=0.139$ ), suggesting a low rate of gene flow between the subpopulations belonging to different groups ( $Nm=1.584$ ). Our results suggest that the gene flow among subpopulations is affected by distance (Fig. 2c) and the landscape of the study area. These two groups are separated by a deforested valley, which explains the higher apparent levels of gene flow within each group than between them. Our results also support the hypothesis of an additional genetic isolation between these two groups beyond that expected by isolation by distance. The presence of a deforested valley separating the two groups is probably limiting the agents of seed and pollen dispersal. Primarily birds, bats and small mammals perform seed dispersal in *C. canjerana* and there is evidence that pollen dispersal is performed exclusively by small moths (Pizo & Oliveira, 1998). The pattern of gene flow found may reflect the dispersal behavior of these animals.

### Implications to conservation

Two environmental scenarios clearly describe these eight *C. canjerana* subpopulations. The first scenario includes subpopulations located at lower altitudes with little landscape variation. These subpopulations are in the most severely degraded forest fragments and show reduced levels of genetic diversity. For this reason, these subpopulations deserve substantial attention by conservation programs. The second scenario includes two subpopulations located in a preserved continuous area at higher altitudes. These subpopulations are located in areas with less logging activity and maintain high levels of plant density and genetic diversity. Gene flow is still important evolutionary force acting on these subpopulations, and their survival depends on the preservation of the Fernão Dias EPA. These scenarios suggest that the high altitude subpopulations must be considered priority areas for conservation policy since they still retain the higher levels of genetic diversity. Once these areas are protected, conservation efforts that improve the gene flow mediated by pollen and seed dispersal must be stimulated, mainly on direction of subpopulations in small fragments.

### Conflicts of interest

The authors declare no conflicts of interest.

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