



Blood serum concentrations of perfluorinated compounds in men from Greenlandic Inuit and European populations

Christian H. Lindh^{a,*}, Lars Rylander^a, Gunnar Toft^b, Anna Axmon^a, Anna Rignell-Hydbom^a, Aleksander Giwercman^c, Henning S. Pedersen^d, Katarzyna Góalczyk^e, Jan K. Ludwicki^e, Valentyna Zvyezday^f, Roel Vermeulen^g, Virissa Lenters^g, Dick Heederik^g, Jens Peter Bonde^b, Bo A.G. Jönsson^a

^a Division of Occupational and Environmental Medicine, Department of Laboratory Medicine, Lund University, 221 85 Lund, Sweden

^b Department of Occupational Medicine, Aarhus University Hospital, Aarhus, Denmark

^c Reproductive Medicine Centre, Skane University Hospital Malmö, Lund University, Malmö, Sweden

^d Institute of Natural Resources, Nuuk, Greenland

^e Department of Environmental Toxicology, National Institute of Public Health – National Institute of Hygiene, Warsaw, Poland

^f Laboratory of Human Reproduction, Kharkiv State Medical University, Kharkiv, Ukraine

^g Institute for Risk Assessment Sciences, Utrecht University, Utrecht, The Netherlands

ARTICLE INFO

Article history:

Received 3 November 2011

Received in revised form 13 March 2012

Accepted 15 March 2012

Available online 10 April 2012

Keywords:

Perfluorinated compounds

PFOS

PFOA

Europeans

Greenlandic Inuits

Seafood

ABSTRACT

Perfluorinated compounds (PFCs), such as perfluorooctane sulfonate (PFOS) and perfluorooctanoic acid (PFOA), are used in large quantities. They are persistent and found in measurable levels in human serum around the world. They have been associated with developmental, hepatic, and carcinogenic effects in animal studies. The aim of the present study was to describe levels of PFCs in serum among Inuits from Greenland and inhabitants from Warsaw, Poland and Kharkiv, Ukraine. Furthermore, the aim was to define social- and lifestyle related determinants of exposure for these compounds. Serum levels of seven PFCs were analyzed by liquid chromatography/tandem mass spectrometry (LC/MS/MS). The concentrations of PFOS and PFOA were the highest of all PFCs in all three populations with a total amount of almost 90% of the PFCs. The mean levels of PFOS and PFOA were in the Greenlandic Inuits 52 and 4.8 ng mL⁻¹, in Poland 19 and 5.2 ng mL⁻¹, and in Ukraine 8.1 and 1.9 ng mL⁻¹, respectively. Thus, levels of PFCs in the serum of Inuits on Greenland were among the highest described in a general population whereas the levels in Poland were similar to other industrialized countries. The exposure in Ukraine was rather low. In the Greenlandic Inuit population, intake of seafood, tea, age and area of living were significant determinants of PFOS concentrations and explained about 22% of the variation. For the other populations no strong determinants were found.

© 2012 Elsevier Ltd. Open access under [CC BY-NC-ND license](http://creativecommons.org/licenses/by-nc-nd/3.0/).

1. Introduction

Perfluorinated compounds (PFCs), such as perfluorooctane sulfonate (PFOS) and perfluorooctanoic acid (PFOA), have since the 1950s been used in large quantities in a number of applications such as in surface coatings, making products water- and oil resistant, and fire fighting foams. In addition, several of the compounds are approved for use in food containers and cooking pans (e.g. Teflon®) and may thus be transferred to food (Fromme et al., 2009). The annual production of PFCs is in the range of several hundred metric tons per year (Lau et al., 2004).

The use of PFOS and PFOA has decreased in the last years and as a result, levels in human matrices are declining (Harada et al., 2004, 2007; Olsen et al., 2005, 2007, 2008; Calafat et al., 2007;

Jin et al., 2007; Haug et al., 2009; Wilhelm et al., 2009; Sundström et al., 2011; Axmon et al., in preparation). However, PFOS and PFOA have been substituted by other PFCs, such as perfluorononanoic acid (PFNA), perfluorodecanoic acid (PFDA) and perfluoroundecanoic acid (PFUnDA). Consequently, human levels of the latter compounds have been found to be increasing (Calafat et al., 2007; Haug et al., 2009; Axmon et al., in preparation).

In animal studies, PFCs have been found to be associated with several outcomes, such as cancer, neurobehavior and reproductive health problems (Luebker et al., 2005; Lau et al., 2007; Chang et al., 2008; Jensen and Leffers, 2008; Johansson et al., 2008; DeWitt et al., 2009; Sonne, 2010). Furthermore, the evidence of effects on humans in the general population is increasing. *In utero* exposure to PFOS and PFOA has been suggested to decrease birth weight (Apelberg et al., 2007; Fei et al., 2007; Stein et al., 2009; Washino et al., 2009) as well as weight in early childhood (Andersen et al., 2010). Moreover, associations have been found between body

* Corresponding author. Tel.: +46 46173819; fax: +46 46143702.

E-mail address: Christian.Lindh@med.lu.se (C.H. Lindh).

levels of PFCs and sperm count (Joensen et al., 2009), glucose homeostasis and metabolic syndrome (Lin et al., 2009), attention deficit/hyperactivity disorder (ADHD) (Hoffman et al., 2010) and thyroid function (Dallaire et al., 2009a).

The PFCs are biopersistent compounds with half-lives in humans that have been estimated to be between 3 and 9 years (Olsen et al., 2007, 2008; Bartell et al., 2010; Brede et al., 2010; Nilsson et al., 2010; Seals et al., 2011). Thus, they will accumulate in humans. In recent years there have been many publications describing serum levels in different populations worldwide (Kannan et al., 2004; Kukulnyik et al., 2004; Guruge et al., 2005; Falandysz et al., 2006; Kärrman et al., 2006, 2007; Calafat et al., 2007; Fromme et al., 2007; Weihe et al., 2008; Harada et al., 2009, 2010; Rylander et al., 2009a,b; Hanssen et al., 2010; Hemat et al., 2010; Ingelido et al., 2010; Roosens et al., 2010; Vassiliadou et al., 2010; Zhang et al., 2010). Generally, levels in industrialized countries are higher (10–30 ng PFOS mL⁻¹) than in developing countries (few ng PFOS mL⁻¹), although levels in the general population seem to be much lower than those of occupational exposed workers (µg PFOS mL⁻¹) (Fromme et al., 2009). However, measurements in many populations are still missing or insufficient. One such population is the Greenlandic Inuits where only recently levels in breast cancer cases and controls have been described (Bonefeld-Jorgensen et al., 2011). The levels in the controls were similar to those found in other industrialized countries. This is in accordance with results from Canadian Inuits (Dallaire et al., 2009b; Château-Degat et al., 2010). However, the levels may be expected to be high in certain areas of Greenland with high intake of local animals, as animals living in polar areas have elevated levels of PFC (Jensen and Leffers, 2008).

There is some information about determinants of PFCs in human samples. Diet has been found to be a major determinant (Fromme et al., 2009; Vestergren and Cousins, 2009; Egeghy and Lorber, 2011; Haug et al., 2011). Among foods suggested to account for this is seafood (Falandysz et al., 2006; Dallaire et al., 2009b; Rylander et al., 2009a; Haug et al., 2010a), meat (Halldorsson et al., 2008; Haug et al., 2010a), poultry, frozen entrées, fast food, and microwave popcorn (Tittlemier et al., 2007) and peppers, cakes and cookies, lunchmeats, and green vegetables (Ostertag et al., 2009). While food seems to be the most important source of exposure in the general population, some studies have indicated a substantial contribution also from dust and air (Fromme et al., 2009; Gewurtz et al., 2009; Harrad et al., 2010; Egeghy and Lorber, 2011; Haug et al., 2011).

Even though some exposure routes for PFCs are well known, this only explains a small proportion of the differences in the human PFC levels. Studies on possible determinants such as age (Harada et al., 2004; Kannan et al., 2004; Calafat et al., 2006; Fromme et al., 2007; Jin et al., 2007; Olsen et al., 2008; Rylander et al., 2009a,b; Wilhelm et al., 2009; Haug et al., 2010a) and gender (Fromme et al., 2009) have so far been inconclusive. Thus, further studies are needed.

In this study, we report serum levels of seven PFCs from Inuit men on Greenland and male inhabitants from Warsaw, Poland and Kharkiv, Ukraine. Furthermore, we investigate determinants of exposure to these PFCs. Such knowledge is important e.g. for risk evaluations and measures for decreasing exposure. The large variation in PFC levels between the different populations may constitute a valuable basis for studies of dose–response studies of different health effects.

2. Materials and methods

2.1. Subjects and data collection

The study population has been described in detail previously (Jönsson et al., 2005). In this study, male subjects delivering semen

samples from Greenland, Poland and Ukraine were investigated. Serum samples were obtained from 196 Greenlandic Inuit males, 190 men from Poland and 203 men from Ukraine in the years 2002–03. Demographic data are shown in Table 1. The study was approved by the local ethical committees representing all participating populations and all subjects signed an informed consent.

2.2. Interview data

Information on lifestyle, medical and reproductive history was collected through interviews. In the present study we used information on age, height and weight (for calculation of BMI), smoking habits, intake of seafood and tea consumption at baseline. For the population from Greenland we also used information regarding area of living, due to different life-styles between different areas, exemplified by different nutrition habits, which might affect PFC levels.

2.3. Collection of blood samples

Blood samples were drawn from a cubital vein into 10 mL vacuum tubes for serum collection without additives (Becton Dickinson, Maylan, France). After cooling to room temperature the tubes were centrifuged at 4000g for 15 min. Serum was transferred with ethanol rinsed Pasteur pipettes to ethanol rinsed brown glass bottles (Termometerfabriken, Gothenburgh, Sweden). A piece of aluminum foil was placed on top of the bottles which were then sealed. Sera were stored at –20 °C until shipment, but refrigeration was accepted for up to 4 d. After the arrival at the analyzing laboratory the samples were stored at –80 °C until analyzed.

2.4. Analysis of PFCs

The analyses of PFHxS, PFOS, PFOA, PFNA, PFDA, PFUnDA and perfluorododecanoic acid (PFDoDA) were performed by LC/MS/MS. Aliquots of 100 µL serum were added with 10 µL glucuronidase, 10 µL ammonium acetate buffer, 10 µL methylumbelliferyl-β-D-glucuronide and 25 µL of a water:acetonitrile (50:50) solution containing ¹³C- or ¹⁸O-labeled internal standards for all evaluated compounds and digested at 37 °C for 90 min. The proteins were precipitated with 175 µL acetonitrile and vigorously shaking for 30 min. The samples were thereafter centrifuged at 4200g and 3 µL of the supernatant was injected on a LC (UFLC^{XR}, SHIMADZU Corporation, Kyoto, Japan; LC/MS/MS) using hybrid triple quadrupole linear ion trap mass spectrometry equipped with a Turbolon-Spray source (QTRAP 5500, Applied Biosystems). The column used was a C₈ column (4 µm, 2.1 mm i.d. × 20 mm GENESIS; Grace Vydac, Hesperia, CA, USA). The mobile phase was A: 0.1% ammonia (NH₃) in water and B: 0.1% NH₃ in acetonitrile. The mobile phase was kept at 25% B for 1 min after injection. A gradient was then applied in 3 min to 95% B where it was kept for 1 min. The column was then conditioned at 25% B for 3 min. Air was used as nebulizer

Table 1

Background characteristics for 196 Inuit men (Greenland), 190 men from Warsaw (Poland), and 203 men from Kharkiv (Ukraine).

	Greenland	Poland	Ukraine
Age (year)	30	30	25
Median (5th, 95th perc)	(20, 43)	(25, 37)	(19, 39)
BMI (kg m ⁻²)	25.5	25.4	24.0
Median (5th, 95th perc)	(20.4, 32.1)	(20.4, 31.8)	(19.6, 29.5)
Tea (cups d ⁻¹)	2	2	2
Median (5th, 95th perc)	(0, 6)	(0, 5)	(1, 5)
Seafood (d week ⁻¹)	2	1	4
Median (5th, 95th perc)	(0, 6)	(0, 3)	(2, 5)

and auxiliary gas while pure nitrogen was used as curtain gas and collision gas. The temperature of the ion source was set at 500 °C and the declustering potential was –80 V. The MS analyses were carried out using selected reaction monitoring (SRM) in the negative ion mode. SRM transitions and collision energies (CEs) used in the analysis are described in Table 2.

About 100 samples were analyzed in each batch. The results reported is the average of the two measurements from the same sample worked-up and analyzed on different days. These results were also used to calculate the reproducibility of the method (Table 2), determined as the coefficient of variation (CV) of the duplicate samples. The limits of detection (LOD), determined as the concentrations corresponding to three times the standard deviation of the responses in chemical blanks, are also shown in Table 2. In all sample batches, the quality of the measurements was controlled by analyzing chemical blanks and in-house quality control samples, prepared from a large volume of serum spiked with small amounts of different PFCs. The analyses of PFOS and PFOA are part of the Round Robin inter-comparison program (Professor Dr. med. Hans Drexler, Institute and Out-Patient Clinic for Occupational, Social and Environmental Medicine, University of Erlangen-Nuremberg, Germany) with results within the tolerance limits, determined as three times the standard deviation of the results from special reference laboratories.

2.5. Statistics

For levels below the LOD the numbers obtained in the evaluation of the MS data were used assuming that these values also have some meaning. Mann–Whitney tests were used for pairwise comparisons of the PFC exposure concentrations between countries (Greenland vs. Poland, Greenland vs. Ukraine, and Poland vs. Ukraine). Separate analyses for each population were performed regarding correlations between PFOS, PFOA and the other PFCs as well as evaluation of the impact of different determinants. The correlations between PFOS, PFOA and the other PFCs were evaluated by Spearman's correlation coefficients. The impact of potential determinants on inter-individual variations in PFOS and PFOA serum concentrations were assessed by linear regression models. As PFOS and PFOA were essentially normally distributed no transformation was performed. Model assumptions were checked by analysis of residuals. The determinants evaluated were age (continuous), BMI (continuous), smoking (categorical: never, ex-smoker and current smoker), seafood (days per week) and tea consumption (cups d⁻¹). In addition, the impact of area of resi-

dence was assessed for the Greenlandic Inuits divided into three categories: (Mid: (Nuuk, Sisimiut, Maniitsoq and Paamiut), South/East: (Narsaq, Qaqortoq, Nanortalik, Tasiilaq, Kulusuk and Kuummiit) and North: (Kangaatsiaq, Aasiaat, Ilulissat, Qeqertarsuaq, Qasigiannnguit, Uummannaq, Qaanaaq, Illorsuit, Saattut and Ukkusissat)).

The objective of the multivariate model building was to evaluate the impact of each of the determinants. In the first step we evaluated the associations for one determinant at a time with the PFOS and PFOA concentrations in serum, respectively. In the second step we included in the multivariate model all determinants with a *p*-value below 0.05 in the univariate analyses. However, if a determinant was statistically significant in one population but not in the others we present the univariate estimates also for the other countries. In the results we also give the explained variances (adjusted *r*² obtained from SPSS™ version 18 for the final models. To avoid ecological fallacies in these analyses, the three populations were analyzed separately.

3. Results

Concentrations of the analyzed PFCs are described in Table 3. With the exception of PFOA, the highest mean concentration for each PFC was found in the Greenlandic Inuit population, with the overall highest concentration displayed by PFOS (mean 52 ng mL⁻¹). The mean PFOS concentration in serum for the three regions in Greenland were: Mid: 44.7 ng mL⁻¹, South/East: 59.3 ng mL⁻¹, and North: 56.3 ng mL⁻¹. For PFOA, the Polish population had slightly higher concentrations than the Greenlandic Inuits. The dominating PFC in all three populations was PFOS with a fraction of 80%, 69% and 68%, of the total PFCs on weight basis, in Greenland, Poland and Ukraine, respectively. The second most dominating PFC was PFOA with a fraction of 7%, 19% and 15%, respectively, for the three populations. Thus, PFOS and PFOA made up almost 90% of the PFCs in all three populations.

The numbers of samples below the LOD in the three populations are included in Table 3. For PFHxS, PFOS, PFOA and PFNA there were no or just a few samples below the LOD, for PFDA about half of the Ukraine samples were below the LOD, and for PFUnA and PFDoA the majority of the samples from Poland and Ukraine were below the LOD.

Correlations between PFOS and PFOA and the other PFCs in the three populations are included in Table 4. In the Greenlandic Inuit population there were high correlations between PFOS and most PFCs (*r*_s between 0.65 and 0.79) except for PFOA where the corre-

Table 2
Parameters for analysis, limit of detection levels and reproducibility^a for the analyzed compounds.

PFC	SRM transition	Collision energy (V)	Limit of detection (ng mL ⁻¹)	Reproducibility (%)	Concentration at reproducibility ^b (ng mL ⁻¹)
PFHxS	399/99	–80	0.06	8	1.5
PFOS	499/99	–85	0.2	5	26
PFOA	413/169	–25	0.6	6	3.9
PFNA	463/419	–15	0.2	9	1.6
PFDA	513/469	–16	0.2	9	0.6
PFUnA	563/519	–19	0.3	10	0.7
PFDoA	613/569	–25	0.07	22	0.08
¹⁸ O ₂ -PFHxS	403/103	–80	–	–	–
¹³ C ₄ -PFOS	503/99	–110	–	–	–
¹³ C ₄ -PFOA	417/169	–30	–	–	–
¹³ C ₅ -PFNA	468/423	–15	–	–	–
¹³ C ₃ -PFDA	515/470	–16	–	–	–
¹³ C ₂ -PFUnA	565/520	–19	–	–	–
¹³ C ₂ -PFDoA	615/570	–25	–	–	–

^a Determined as the coefficient of variation (CV) of duplicate samples worked-up and analyzed on different days.

^b The mean concentration for the samples from which the reproducibility was determined.

Table 3

Concentrations (ng mL⁻¹) of the analyzed PFCs in Inuits from Greenland and inhabitants from Warsaw, Poland and Kharkiv, Ukraine and the number of samples below the limit of detection.

	Greenland n = 196	Poland n = 190	Ukraine n = 203
PFHxS^{a,b,c}			
Mean (SD)	2.88 (2.54)	1.22 (0.44)	0.40 (0.30)
Median (min, max)	2.18 (0.91, 20.5)	1.18 (0.43, 3.78)	0.34 (<LOD ^d , 3.42)
Samples below LOD	0	0	1
PFOS^{a,b,c}			
Mean (SD)	51.9 (24.4)	18.6 (5.67)	8.08 (3.98)
Median (min, max)	44.7 (12.3, 161)	18.5 (8.20, 40.2)	7.60 (2.77, 29.9)
Samples below LOD	0	0	0
PFOA^{b,c}			
Mean (SD)	4.83 (1.60)	5.25 (2.14)	1.79 (2.75)
Median (min, max)	4.54 (1.52, 13.7)	4.84 (1.48, 16.0)	1.29 (<LOD, 35.0)
Samples below LOD	0	0	16
PFNA^{a,b,c}			
Mean (SD)	2.26 (1.74)	1.32 (0.69)	1.13 (0.54)
Median (min, max)	1.74 (0.53, 11.6)	1.19 (0.48, 6.11)	1.02 (0.25, 4.19)
Numbers below LOD	0	0	0
PFDA^{a,b,c}			
Mean (SD)	1.08 (0.81)	0.41 (0.17)	0.23 (0.14)
Median (min, max)	0.87 (<LOD, 5.92)	0.38 (<LOD, 1.37)	0.21 (<LOD, 0.96)
Samples below LOD	2	11	99
PFUnDA^{a,b}			
Mean (SD)	1.76 (1.74)		
Median (min, max)	1.28 (<LOD, 13.4)	<LOD (<LOD, 0.70)	<LOD (<LOD, 0.53)
Samples below LOD	11	168	178
PFDoDA^{a,b,c}			
Mean (SD)	0.19 (0.20)		
Median (min, max)	0.14 (<LOD, 1.68)	<LOD (<LOD, 0.23)	<LOD (<LOD, 0.10)
Samples below LOD	38	174	197

^a Significant difference ($p < 0.05$) between Greenland and Poland (Mann–Whitney test).

^b Significant difference ($p < 0.05$) between Greenland and Ukraine (Mann–Whitney test).

^c Significant difference ($p < 0.05$) between Poland and Ukraine (Mann–Whitney test).

^d Below the limit of detection.

lation was moderate ($r_s = 0.42$). The correlations between PFOA and the other PFCs were lower ($r_s < 0.30$). In the Ukrainian population there were moderate to high correlations ($r_s = 0.39–0.72$) between PFCs, while the correlations between PFCs in the Polish population were low to moderate ($r_s = 0.10–0.53$).

Among the Greenlandic Inuits, age, intake of tea and area of living were significantly associated with PFOS serum concentrations (Table 5). Together, they explained 19% of the variation. If we also included intake of seafood, the explained variance increased to 22%. Moreover, excluding one of the variables intake of seafood or area of living in the models did not change the estimate of the remaining variables, speaking in favor of these two variables as independent. In Poland, the only significant determinant was BMI

and in Ukraine there was no statistically significant determinant. Regarding PFOA, the only significant association was observed for tea consumption among the Greenlandic Inuits. An increase of one cup per day corresponded to an increase in PFOA of 0.17 ng mL⁻¹ (95% CI, 0.05, 0.29; $R^2 = 0.04$). Smoking was not associated with PFOS or PFOA.

4. Discussion

The present study provides data on levels of PFCs in populations for which such information is lacking or only scarce. The levels among Inuits in Greenland were among the highest described in the literature. Levels in Poland were similar to other industrialized populations while those in Ukraine were lower. Determinants found in most cases seemed to be similar to previous studies (Fromme et al., 2009; Haug et al., 2010a). However, the finding of different levels of PFOS in different areas of Greenland was novel. Also, the finding of tea consumption as a determinant for both PFOS and PFOA was not previously reported.

For the purpose of this study we developed a simple method for determination of several PFCs. The method uses only 0.1 mL serum and may therefore be suitable to apply on samples stored in serum banks. The coefficient of variation of duplicate samples worked-up and analyzed on different was generally below 10% for duplicate analysis of the same sample, analyzed in different sample batches on different days. Moreover, the detection limits were sufficient for determination of levels in most samples. Furthermore, a high accuracy of the method, at least for PFOS and PFOA, was proven by participation in an inter-laboratory control program. However, it should be emphasized that a quantitative comparison of the presented data, especially with earlier published data, must be done with caution since there have been many challenges in the development of analytical methods for the PFCs over time (Martin et al., 2004).

The levels of PFOS in the Inuits on Greenland indicate that these are among the highest exposed populations in the world and was higher than those at 22 ng mL⁻¹ reported by Bonfeld-Jorgensen et al. (2011) for the controls in their study. This is in accordance with the high levels found in animals living in the Arctic area. Moreover, the results show that the Greenlandic Inuit population is highly exposed to several other, more recently industrially introduced PFCs, such as PFNA, PFDA, PFUnA and PFDoA. This indicates a fast distribution of these compounds to the Arctic area. Interestingly, the levels of PFOA on Greenland were similar to the levels found in industrialized countries in Europe and the US.

Previous studies on Polish populations have found that people with a high consumption of fish from or residing by the Baltic Sea have considerably higher levels of PFCs than non-fish eating populations (Kannan et al., 2004; Falandysz et al., 2006). In the present study, the levels of PFCs in Poland were comparable to those found in Polish non-fish eating populations. This was to be expected, since the population was recruited from Warsaw, which

Table 4

Spearman's correlations coefficients^a (and p -values) between PFOS, PFOA and the other PFCs.

	Greenland n = 196		Warsaw (Poland) n = 190		Kharkiv (Ukraine) n = 203	
	PFOS	PFOA	PFOS	PFOA	PFOS	PFOA
PFHxS	0.65 (<0.001)	0.20 (0.01)	0.44 (<0.001)	0.23 (<0.001)	0.72 (<0.001)	0.49 (<0.001)
PFOA	0.42 (<0.001)		0.34 (<0.001)		0.56 (<0.001)	
PFNA	0.76 (<0.001)	0.18 (0.01)	0.43 (<0.001)	0.53 (<0.001)	0.57 (<0.001)	0.39 (<0.001)
PFDA	0.79 (<0.001)	0.16 (0.02)	0.37 (<0.001)	0.44 (<0.001)	–	–
PFUnDA	0.69 (<0.001)	0.05 (0.51)	–	–	–	–
PFDoDA	0.65 (<0.001)	0.04 (0.50)	–	–	–	–

^a Correlation coefficients were only calculated when we had data above LOD for at least 60% of the participants.

Table 5

Linear regressions (univariate and multivariate models) for determinants for PFOS serum concentrations (ng/mL) among Inuit men (Greenland), Warsaw men (Poland), and Kharkiv men (Ukraine). Significant associations are marked with bold text. In addition, explained variances (R^2) are shown.

Determinants	Greenland				Poland		Ukraine	
	Univariate		Multivariate		Univariate		Univariate	
	β	95% CI	β	95% CI	β	95% CI	β	95% CI
Age (yr)	0.87	(0.38, 1.4)	0.84	(0.36, 1.3)	0.062	(-0.14, 0.27)	-0.008	(-0.10, 0.087)
BMI (kg/m ²)	-0.37	(-1.2, 0.42)	-	-	-0.26	(-0.50, -0.014)	-0.003	(-0.18, 0.17)
Tea (cups/day)	2.6	(0.88, 4.2)	2.1	(0.52, 3.6)	0.38	(-0.21, 0.97)	-0.33	(-0.77, 0.10)
Seafood (days/week)	3.0	(1.1, 4.9)	2.1	(0.24, 3.9)	0.74	(-0.095, 1.6)	0.042	(-0.40, 0.48)
Area (North reference)								
South/East	3.0	(-6.0, 12)	0.67	(-8.0, 9.4)	-		-	
Mid	-12	(-20, -3.4)	-13	(-21, -5.8)	-		-	
Adjusted R ²			0.22		0.02 ^a		_b	

^a Including only BMI in the model.

^b No significant associations.

is in the Polish inland, and did not have a high consumption of Baltic Sea fish. Some earlier pilot studies in Warsaw showed even somewhat lower levels of PFCs confirming lower exposure of inlanders (Strucinski et al., 2006). The levels of PFCs in Poland were in the same range as for other industrialized countries (Fromme et al., 2009).

To the best of our knowledge, there are no previous reports of levels of PFCs in Ukraine. The levels found in this study indicate a rather low exposure compared to most industrialized countries. For PFOS the levels were similar to those at about 10 ng mL⁻¹ found in Brazil and Colombia (Kannan et al., 2004).

Although PFOS correlated well with all other PFCs, except with PFOA, the correlations between PFOA and the other PFCs were low. One interpretation is that the exposure routes for PFOS differ from those for PFOA. E.g. difference in PFOS and PFOA fish concentrations has been observed that could indicate a lower potential of PFOA to bioaccumulate in fish than PFOS (Fromme et al., 2009). Thus, it is not surprising that we found different determinants for PFOA and PFOS.

We found significant determinants for PFOS only in the Greenlandic Inuit population. Possible explanations for this may include higher levels of PFOS among the Greenlandic Inuits than in Poland or Ukraine, but also a difference in exposure levels from the same exposure routes, e.g. that seafood consumed on Greenland carries a higher exposure than the same amount consumed in Poland. Still, only 22% of the Greenlandic Inuit's PFOS levels were explained by the determinants investigated.

In the Greenlandic Inuit population, seafood was one of the determinants for PFOS. PFOS and - although to a lesser extent - PFOA have been found in both freshwater and marine fish species world wide (D'Hollander et al., 2010). Accordingly, studies relating fish and shellfish consumption to human levels of PFCs have generally found positive correlations (Falandysz et al., 2006; Rylander et al., 2009a,b; Haug et al., 2010a), although not always for PFOA (Rylander et al., 2009a,b). Thus, the results in the present study are in agreement with previous research.

Unfortunately we have no information regarding the Greenlandic Inuit's seafood consumption was related to consumption of fish or marine mammals. In some studies, consumption of marine mammals has been associated with high levels of PFCs in serum; Weihe et al. (2008) found a relationship between PFCs in serum of consumers of Faroese whale meat. In the present study, we found higher levels of PFOS but not PFOA in the remote areas of Greenland, where the diet tend to consist of more traditional Greenlandic Inuit food with consumption of fish and marine mammals, than in Nuuk, the capital of Greenland. Thus, the traditional Greenlandic Inuit food is most likely related to a higher exposure. In agreement with our study, Dallaire et al. (2009a,b) did not find

that the remote geographical location and traditional lifestyle protected the Canadian Nunavut Inuits from PFOS exposure.

Previous studies on a possible relationship between age and PFCs are inconclusive. Whereas some have found increasing levels of PFOS with increasing age (Olsen et al., 2005; Kärrman et al., 2006; Fromme et al., 2007; Rylander et al., 2009a,b; Haug et al., 2010a), others have failed to show any age related changes for PFOS or PFOA (Harada et al., 2004; Kannan et al., 2004; Calafat et al., 2006; Jin et al., 2007; Olsen et al., 2008; Wilhelm et al., 2009). Similarly, in the present study, different results for age were found for the different PFCs and the different populations. The influence of age on the levels of PFCs should be further investigated, especially considering that since environmental levels of PFCs have not been at a steady state during the past decades, age is not necessarily a proxy for exposure time.

In Poland there was an inverse relationship between PFOS and BMI. The effect estimate was similar in the Greenlandic population, although not statistically significant. A similar relationship has been found in the Danish general population (Eriksen et al., 2011) and a German population exposed to PFCs through contaminated drinking water (Hölzer et al., 2008). However, other studies have not shown such an association (Halldórsson et al., 2008; Rylander et al., 2009a,b). There was no expectation of finding an association with BMI since PFCs are mainly bound to proteins and not fat. We can therefore not rule out that in the Polish population, BMI is a proxy for an unmeasured determinant, such as diet.

It is interesting to note that tea seemed to be a determinant for PFOS and PFOA in the Greenlandic Inuit population. However, tea consumption was not a determinant of PFOS and PFOA level in Poland or Ukraine, with similar tea consumption as the Greenlandic Inuits. High levels of PFCs in tea could be a result of either high levels in the water with which the tea was brewed, from the tea leaves themselves or from the container (e.g. tea bag or tea infuser for home brewed tea or bottle for ready made tea, such as iced tea). Drinking water is generally not thought of as a major exposure source for PFCs. In most studies, its contribution to the daily intake of PFOS and PFOA has been found to be below 3% (Thompson et al., 2011). However, in some reports, it has been estimated to be as high as 55% (Noorlander et al., 2011; Thompson et al., 2011). We have found no reports on PFC levels in Greenlandic drinking water, and it is therefore difficult to draw any conclusions about the possibility that the water is the contributing factor the results found for tea consumption. In a study from Sri Lanka, workers from a rural conventional tea cultivation area had levels of PFOS and PFOA that were in the same range as people from the county's capital (Guruge et al., 2005). However, workers involved in organic tea cultivation had considerably lower levels of both PFOS and PFOA. Thus, there is the possibility that something in the conventional

tea cultivation might cause PFC contamination. Indeed, Haug et al. (2010a,b) measured PFCs in a small sample of brewed tea, and found that the levels of PFOA but not PFOS were considerably higher in the prepared drink than in the water it was made from. Thus, it is possible that both water and tea leaves can contribute to the results found in this study.

5. Conclusion

The levels of PFCs in serum of Inuits on Greenland were among the highest described in a general population, whereas the levels in a population from Warsaw, Poland were similar to other industrialized countries. The levels in Ukraine were rather low. Determinants for PFOS on Greenland were seafood and tea intake, age and area of living. Tea was also a determinant of PFOA at Greenland. For the other populations no strong determinants were observed.

Acknowledgements

The collection of the serum samples use in this study was part of the project “INUENDO – Biopersistent organochlorines in diet and human fertility. Epidemiological studies of time to pregnancy and semen quality in Inuit and European populations”, supported by The European Commission to the 5th Framework Programme Quality of Life and Management of Living Resources, Key Action 4 on Environment and Health (Contract No. QLK4-CT-2001-00202), <http://www.inuendo.dk>. The present project was funded by the project “CLEAR – Climate change, environmental contaminants and reproductive health (Grant Agreement No. 226217)” supported by The European Commission to the 7th Framework Programme: FP7-ENV-2008-1 Environment (including Climate Change), <http://www.inuendo.dk/clear>. The work has also been funded by the Swedish Research Council and the Swedish Research Council for Environment, Agricultural sciences and Spatial Planning. Authors thank Åsa Amilon for performing chemical analyses in a skilful way.

References

- Andersen, C.S., Fei, C., Gamborg, M., Nohr, E.A., Sørensen, T.I., Olsen, J., 2010. Prenatal exposures to perfluorinated chemicals and anthropometric measures in infancy. *Am. J. Epidemiol.* 172, 1230–1237.
- Apelberg, B.J., Witter, F.R., Herbstman, J.B., Calafat, A.M., Halden, R.U., Needham, L.L., Goldman, L.R., 2007. Cord serum concentrations of perfluorooctane sulfonate (PFOS) and perfluorooctanoate (PFOA) in relation to weight and size at birth. *Environ. Health Perspect.* 115, 1670–1676.
- Axmon, A., Axelsson, J., Jakobsson, K., Lindh, C.H., Jönsson, B.A.G., in preparation. Time trends between 1987 and 2007 for perfluorinated chemicals in plasma from Swedish women.
- Bartell, S.M., Calafat, A.M., Lyu, C., Kato, K., Ryan, P.B., Steenland, K., 2010. Rate of decline in serum PFOA concentrations after granular activated carbon filtration at two public water systems in Ohio and West Virginia. *Environ. Health Perspect.* 118, 222–228.
- Bonefeld-Jørgensen, E.C., Long, M., Bossi, R., Ayotte, P., Asmund, G., Kruger, T., Ghisari, M., Mulvad, G., Kern, P., Nzulumiki, P., Dewailly, E., 2011. Perfluorinated compounds are related to breast cancer risk in Greenlandic Inuit: a case control study. *Environ. Health* 10, 88.
- Brede, E., Wilhelm, M., Goen, T., Muller, J., Rauchfuss, K., Kraft, M., Holzer, J., 2010. Two-year follow-up biomonitoring pilot study of residents' and controls' PFC plasma levels after PFOA reduction in public water system in Arnsberg, Germany. *Int. J. Hyg. Environ. Health* 213, 217–223.
- Calafat, A.M., Kuklenyik, Z., Caudill, S.P., Reidy, J.A., Needham, L.L., 2006. Perfluorochemicals in pooled serum samples from United States residents in 2001 and 2002. *Environ. Sci. Technol.* 40, 2128–2134.
- Calafat, A.M., Wong, L.Y., Kuklenyik, Z., Reidy, J.A., Needham, L.L., 2007. Polyfluoroalkyl chemicals in the US population: data from the National Health and Nutrition Examination Survey (NHANES) 2003–2004 and comparisons with NHANES 1999–2000. *Environ. Health Perspect.* 115, 1596–1602.
- Chang, S.C., Thibodeaux, J.R., Eastvold, M.L., Ehresman, D.J., Bjork, J.A., Froehlich, J.W., Lau, C., Singh, R.J., Wallace, K.B., Butenhoff, J.L., 2008. Thyroid hormone status and pituitary function in adult rats given oral doses of perfluorooctanesulfonate (PFOS). *Toxicology* 243, 330–339.
- Château-Degat, M.L., Pereg, D., Dallaire, R., Ayotte, P., Dery, S., Dewailly, E., 2010. Effects of perfluorooctanesulfonate exposure on plasma lipid levels in the Inuit population of Nunavik (Northern Quebec). *Environ. Res.* 110, 710–717.
- Dallaire, R., Dewailly, E., Pereg, D., Dery, S., Ayotte, P., 2009a. Thyroid function and plasma concentrations of polyhalogenated compounds in Inuit adults. *Environ. Health Perspect.* 117, 1380–1386.
- Dallaire, R., Ayotte, P., Pereg, D., Dery, S., Dumas, P., Langlois, E., Dewailly, E., 2009b. Determinants of plasma concentrations of perfluorooctanesulfonate and brominated organic compounds in Nunavik Inuit adults (Canada). *Environ. Sci. Technol.* 43, 5130–5136.
- DeWitt, J.C., Shnyra, A., Badr, M.Z., Loveless, S.E., Hoban, D., Frame, S.R., Cunard, R., Anderson, S.E., Meade, B.J., Peden-Adams, M.M., Luebke, R.W., Luster, M.I., 2009. Immunotoxicity of perfluorooctanoic acid and perfluorooctane sulfonate and the role of peroxisome proliferator-activated receptor alpha. *Crit. Rev. Toxicol.* 39, 76–94.
- D'Hollander, W., de Voogt, P., De Coen, W., Bervoets, L., 2010. Perfluorinated substances in human food and other sources of human exposure. *Rev. Environ. Contam. Toxicol.* 208, 179–215.
- Egeghy, P.P., Lorber, M., 2011. An assessment of the exposure of Americans to perfluorooctane sulfonate: a comparison of estimated intake with values inferred from NHANES data. *J. Expo. Sci. Environ. Epidemiol.* 21, 150–168.
- Eriksen, K.T., Sørensen, M., McLaughlin, J.K., Tjønneland, A., Overvad, K., Raaschou-Nielsen, O., 2011. Determinants of plasma PFOA and PFOS levels among 652 Danish Men. *Environ. Sci. Technol.* 45, 8137–8143.
- Falandysz, J., Taniyasu, S., Gulkowska, A., Yamashita, N., Schulte-Oehlmann, U., 2006. Is fish a major source of fluorinated surfactants and repellents in humans living on the Baltic Coast? *Environ. Sci. Technol.* 40, 748–751.
- Fei, C., McLaughlin, J.K., Tarone, R.E., Olsen, J., 2007. Perfluorinated chemicals and fetal growth: a study within the Danish National Birth Cohort. *Environ. Health Perspect.* 115, 1677–1682.
- Fromme, H., Schlummer, M., Möller, A., Gruber, L., Wolz, G., Ungewiss, J., Böhmer, S., Dekant, W., Mayer, R., Liebl, B., Twardella, D., 2007. Exposure of an adult population to perfluorinated substances using duplicate diet portions and biomonitoring data. *Environ. Sci. Technol.* 41, 7928–7933.
- Fromme, H., Tittlemier, S.A., Völkel, W., Wilhelm, M., Twardella, D., 2009. Perfluorinated compounds – exposure assessment for the general population in Western countries. *Int. J. Hyg. Environ. Health* 212, 239–270.
- Gewurtz, S.B., Bhavsar, S.P., Crozier, P.W., Diamond, M.L., Helm, P.A., Marvin, C.H., Reiner, E.J., 2009. Perfluoroalkyl contaminants in window film: indoor/outdoor, urban/rural, and winter/summer contamination and assessment of carpet as a possible source. *Environ. Sci. Technol.* 43, 7317–7323.
- Guruge, K.S., Taniyasu, S., Yamashita, N., Wijeratna, S., Mohotti, K.M., Seneviratne, H.R., Kannan, K., Yamanaka, N., Miyazaki, S., 2005. Perfluorinated organic compounds in human blood serum and seminal plasma: a study of urban and rural tea worker populations in Sri Lanka. *J. Environ. Monit.* 7, 371–377.
- Halldorsson, T.I., Fei, C.Y., Olsen, J., Lipworth, L., McLaughlin, J.K., Olsen, S.F., 2008. Dietary predictors of perfluorinated chemicals: a study from the Danish National Birth Cohort. *Environ. Sci. Technol.* 42, 8971–8977.
- Hanssen, L., Röllin, H., Odland, J.Ø., Moe, M.K., Sandanger, T.M., 2010. Perfluorinated compounds in maternal serum and cord blood from selected areas of South Africa: results of a pilot study. *J. Environ. Monit.* 12, 1355–1361.
- Harada, K., Saito, N., Inoue, K., Yoshinaga, T., Watanabe, T., Sasaki, S., Kamiyama, S., Koizumi, A., 2004. The influence of time, sex and geographic factors on levels of perfluorooctane sulfonate and perfluorooctanoate in human serum over the last 25 years. *J. Occup. Health* 46, 141–147.
- Harada, K., Koizumi, A., Saito, N., Inoue, K., Yoshinaga, T., Date, C., Fujii, S., Hachiya, N., Hirokawa, I., Koda, S., Kusaka, Y., Murata, K., Omae, K., Shimbo, S., Takenaka, K., Takeshita, T., Todoriki, H., Wada, Y., Watanabe, T., Ikeda, M., 2007. Historical and geographical aspects of the increasing perfluorooctanoate and perfluorooctane sulfonate contamination in human serum in Japan. *Chemosphere* 66, 293–301.
- Harada, K., Saito, N., Inoue, K., Yoshinaga, T., Watanabe, T., Sasaki, S., Kamiyama, S., Harada, K.H., Koizumi, A., 2009. Environmental and biological monitoring of persistent fluorinated compounds in Japan and their toxicities. *Environ. Health Perspect. Med.* 14, 7–19.
- Harada, K.H., Yang, H.R., Moon, C.S., Hung, N.N., Hitomi, T., Inoue, K., Niisoe, T., Watanabe, T., Kamiyama, S., Takenaka, K., Kim, M.Y., Watanabe, K., Takasuga, T., Koizumi, A., 2010. Levels of perfluorooctane sulfonate and perfluorooctanoic acid in female serum samples from Japan in 2008, Korea in 1994–2008 and Vietnam in 2007–2008. *Chemosphere* 79, 314–319.
- Harrad, S., de Wit, C.A., Abdallah, M.A., Bergh, C., Björklund, J.A., Covaci, A., Darnerud, P.O., de Boer, J., Diamond, M., Huber, S., Leonards, P., Mandalakis, M., Ostman, C., Haug, L.S., Thomsen, C., Webster, T.F., 2010. Indoor contamination with hexabromocyclododecanes, polybrominated diphenyl ethers, and perfluoroalkyl compounds: an important exposure pathway for people? *Environ. Sci. Technol.* 44, 3221–3231.
- Haug, L.S., Thomsen, C., Becher, G., 2009. Time trends and the influence of age and gender on serum concentrations of perfluorinated compounds in archived human samples. *Environ. Sci. Technol.* 43, 2131–2136.
- Haug, L.S., Thomsen, C., Brantsæter, A.L., Kvale, H.E., Haugen, M., Becher, G., Alexander, J., Meltzer, H.M., Knutsen, H.K., 2010a. Diet and particularly seafood are major sources of perfluorinated compounds in humans. *Environ. Int.* 36, 772–778.
- Haug, L.S., Salihovic, S., Jogsten, I.E., Thomsen, C., van Bavel, B., Lindström, G., Becher, G., 2010b. Levels in food and beverages and daily intake of perfluorinated compounds in Norway. *Chemosphere* 80, 1137–1143.

- Haug, L.S., Huber, S., Becher, G., Thomsen, C., 2011. Characterisation of human exposure pathways to perfluorinated compounds – comparing exposure estimates with biomarkers of exposure. *Environ. Int.* 37, 687–693.
- Hemat, H., Wilhelm, M., Völkel, W., Mosch, C., Fromme, H., Wittsiepe, J., 2010. Low serum levels of perfluorooctanoic acid (PFOA), perfluorooctane sulfonate (PFOS) and perfluorohexane sulfonate (PFHxS) in children and adults from Afghanistan. *Sci. Total Environ.* 408, 3493–3495.
- Hoffman, K., Webster, T.F., Weisskopf, M.G., Weinberg, J., Vieira, V.M., 2010. Exposure to polyfluoroalkyl chemicals and attention deficit/hyperactivity disorder in US children 12–15 years of age. *Environ. Health Perspect.* 118, 1762–1767.
- Hölzer, J., Midasch, O., Rauchfuss, K., Kraft, M., Reupert, R., Angerer, J., Kleeschulte, P., Marschall, N., Wilhelm, M., 2008. Biomonitoring of perfluorinated compounds in children and adults exposed to perfluorooctanoate-contaminated drinking water. *Environ. Health Perspect.* 116, 651–657.
- Ingelido, A.M., Marra, V., Abballe, A., Valentini, S., Iacovella, N., Barbieri, P., Porpora, M.G., Domenico, A., Felip, E.D., 2010. Perfluorooctanesulfonate and perfluorooctanoic acid exposures of the Italian general population. *Chemosphere* 80, 1125–1130.
- Jensen, A.A., Leffers, H., 2008. Emerging endocrine disruptors: perfluoroalkylated substances. *Int. J. Androl.* 31, 161–169.
- Jin, Y., Saito, N., Harada, K.H., Inoue, K., Koizumi, A., 2007. Historical trends in human serum levels of perfluorooctanoate and perfluorooctane sulfonate in Shenyang, China. *Tohoku J. Exp. Med.* 212, 63–70.
- Joensen, U.N., Bossi, R., Leffers, H., Jensen, A.A., Skakkebaek, N.E., Jørgensen, N., 2009. Do perfluoroalkyl compounds impair human semen quality? *Environ. Health Perspect.* 117, 923–927.
- Johansson, N., Fredriksson, A., Eriksson, P., 2008. Neonatal exposure to perfluorooctane sulfonate (PFOS) and perfluorooctanoic acid (PFOA) causes neurobehavioural defects in adult mice. *Neurotoxicology* 29, 160–169.
- Jönsson, B.A.G., Rylander, L., Lindh, C., Rignell-Hydbom, A., Giwercman, A., Toft, G., Pedersen, H.S., Ludwicki, J.K., Góralczyk, K., Zvezday, V., Spanò, M., Bizzaro, D., Bonefeld-Jørgensen, E.C., Manicardi, G.C., Bonde, J.P., Hagmar, L., Inuendo, 2005. Inter-population variations in concentrations, determinants of and correlations between 2,2',4,4',5,5'-hexachlorobiphenyl (CB-153) and 1,1-dichloro-2,2-bis (p-chlorophenyl)-ethylene (p, p'-DDE): a cross-sectional study of 3161 men and women from Inuit and European populations. *Environ. Health* 4, 27.
- Kannan, K., Corsolini, S., Falandysz, J., Fillmann, G., Kumar, K.S., Loganathan, B.G., Mohd, M.A., Olivero, J., Van Wouwe, N., Yang, J.H., Aldoust, K.M., 2004. Perfluorooctanesulfonate and related fluorochemicals in human blood from several countries. *Environ. Sci. Technol.* 38, 4489–4495.
- Kärman, A., Mueller, J.F., van Bavel, B., Hardten, F., Toms, L.M., Lindström, G., 2006. Levels of 12 perfluorinated chemicals in pooled Australian serum, collected 2002–2003, in relation to age, gender, and region. *Environ. Sci. Technol.* 40, 3742–3748.
- Kärman, A., Ericson, I., van Bavel, B., Darnerud, P.O., Aune, M., Glynn, A., Lignell, S., Lindström, G., 2007. Exposure of perfluorinated chemicals through lactation: levels of matched human milk and serum and a temporal trend, 1996–2004, in Sweden. *Environ. Health Perspect.* 115, 226–230.
- Kuklanyik, Z., Reich, J.A., Tully, J.S., Needham, L.L., Calafat, A.M., 2004. Automated solid-phase extraction and measurement of perfluorinated organic acids and amides in human serum and milk. *Environ. Sci. Technol.* 38, 3698–3704.
- Lau, C., Butenhoff, J.L., Rogers, J.M., 2004. The developmental toxicity of perfluoroalkyl acids and their derivatives. *Toxicol. Appl. Pharmacol.* 198, 231–241.
- Lau, C., Anitole, K., Hodes, C., Lai, D., Pfahles-Hutchens, A., Seed, J., 2007. Perfluoroalkyl acids: a review of monitoring and toxicological findings. *Toxicol. Sci.* 99, 366–394.
- Lin, C.Y., Chen, P.C., Lin, Y.C., Lin, L.Y., 2009. Association among serum perfluoroalkyl chemicals, glucose homeostasis, and metabolic syndrome in adolescents and adults. *Diabetes Care* 32, 702–707.
- Luebker, D.J., York, R.G., Hansen, K.J., Moore, J.A., Butenhoff, J.L., 2005. Neonatal mortality from in utero exposure to perfluorooctanesulfonate (PFOS) in Sprague-Dawley rats: dose-response, and biochemical and pharmacokinetic parameters. *Toxicology* 215, 149–169.
- Martin, J.W., Kannan, K., Berger, U., de Voogt, P., Field, J., Franklin, J., Giesy, J.P., Harner, T., Muir, D.C., Scott, B., Kaiser, M., Järnberg, U., Jones, K.C., Mabury, S.A., Schroeder, H., Simcik, M., Sottani, C., van Bavel, B., Kärman, A., Lindström, G., van Leeuwen, S., 2004. Analytical challenges hamper perfluoroalkyl research. *Environ. Sci. Technol.* 38, 248A–255A.
- Nilsson, H., Kärman, A., Westberg, H., Rotander, A., van Bavel, B., Lindström, G., 2010. A time trend study of significantly elevated perfluorocarboxylate levels in humans after using fluorinated ski wax. *Environ. Sci. Technol.* 44, 2150–2155.
- Noorlander, C.W., van Leeuwen, S.P.J., te Biesebeek, J.D., Mengelers, M.J.B., Zeilmaker, M.J., 2011. Levels of perfluorinated compounds in food and dietary intake of PFOS and PFOA in The Netherlands. *J. Agric. Food Chem.* 59, 7496–7505.
- Olsen, G.W., Huang, H.Y., Helzlsouer, K.J., Hansen, K.J., Butenhoff, J.L., Mandel, J.H., 2005. Historical comparison of perfluorooctanesulfonate, perfluorooctanoate, and other fluorochemicals in human blood. *Environ. Health Perspect.* 113, 539–545.
- Olsen, G.W., Burris, J.M., Ehresman, D.J., Froehlich, J.W., Seacat, A.M., Butenhoff, J.L., Zobel, L.R., 2007. Half-life of serum elimination of perfluorooctanesulfonate, perfluorohexanesulfonate, and perfluorooctanoate in retired fluorochemical production workers. *Environ. Health Perspect.* 115, 1298–1305.
- Olsen, G.W., Mair, D.C., Church, T.R., Ellefson, M.E., Reagen, W.K., Boyd, T.M., Herron, R.M., Medhizadehkashi, Z., Nobiletti, J.B., Rios, J.A., Butenhoff, J.L., Zobel, L.R., 2008. Decline in perfluorooctanesulfonate and other polyfluoroalkyl chemicals in American Red Cross adult blood donors, 2000–2006. *Environ. Sci. Technol.* 42, 4989–4995.
- Ostertag, S.K., Chan, H.M., Moisey, J., Dabeka, R., Tittlemier, S.A., 2009. Historic dietary exposure to perfluorooctane sulfonate, perfluorinated carboxylates, and fluorotelomer unsaturated carboxylates from the consumption of store-bought and restaurant foods for the Canadian population. *J. Agric. Food Chem.* 57, 8534–8544.
- Roosens, L., D'Hollander, W., Bervoets, L., Reynders, H., Van Campenhout, K., Cornelis, C., Van Den Heuvel, R., Koppen, G., Covaci, A., 2010. Brominated flame retardants and perfluorinated chemicals, two groups of persistent contaminants in Belgian human blood and milk. *Environ. Pollut.* 158, 2546–2552.
- Rylander, C., Brustad, M., Falk, H., Sandanger, T.M., 2009a. Dietary predictors and plasma concentrations of perfluorinated compounds in a coastal population from northern Norway. *J. Environ. Public Health* 2009, 268219.
- Rylander, C., Phi, D.T., Odland, J.O., Sandanger, T.M., 2009b. Perfluorinated compounds in delivering women from south central Vietnam. *J. Environ. Monit.* 11, 2002–2008.
- Seals, R., Bartell, S.M., Steenland, K., 2011. Accumulation and clearance of perfluorooctanoic acid (PFOA) in current and former residents of an exposed community. *Environ. Health Perspect.* 119, 119–124.
- Sonne, C., 2010. Health effects from long-range transported contaminants in Arctic top predators: an integrated review based on studies of polar bears and relevant model species. *Environ. Int.* 36, 461–491.
- Stein, C.R., Savitz, D.A., Dougan, M., 2009. Serum levels of perfluorooctanoic acid and perfluorooctane sulfonate and pregnancy outcome. *Am. J. Epidemiol.* 170, 837–846.
- Struciński, P., Góralczyk, K., Ludwicki, J.K., Czaja, K., Hernik, A., Korcz, W., 2006. Levels of selected organochlorine insecticides, polychlorinated biphenyls, phthalates and perfluorinated aliphatic substances in blood – polish WWF study. *Roczn. PZH* 57, 99–112 (in Polish).
- Sundström, M., Ehresman, D.J., Bignert, A., Butenhoff, J.L., Olsen, G.W., Chang, S.C., Bergman, A., 2011. A temporal trend study (1972–2008) of perfluorooctanesulfonate, perfluorohexanesulfonate, and perfluorooctanoate in pooled human milk samples from Stockholm, Sweden. *Environ. Int.* 37, 178–183.
- Thompson, J., Eaglesham, G., Mueller, J., 2011. Concentrations of PFOS, PFOA and other perfluorinated alkyl acids in Australian drinking water. *Chemosphere* 83, 1320–1325.
- Tittlemier, S.A., Pepper, K., Seymour, C., Moisey, J., Bronson, R., Cao, X.L., Dabeka, R.W., 2007. Dietary exposure of Canadians to perfluorinated carboxylates and perfluorooctane sulfonate via consumption of meat, fish, fast foods, and food items prepared in their packaging. *J. Agric. Food Chem.* 55, 3203–3210.
- Vassiliadou, I., Costopoulou, D., Ferderigou, A., Leondiadis, L., 2010. Levels of perfluorooctanesulfonate (PFOS) and perfluorooctanoate (PFOA) in blood samples from different groups of adults living in Greece. *Chemosphere* 80, 1199–1206.
- Vestergren, R., Cousins, I.T., 2009. Tracking the pathways of human exposure to perfluorocarboxylates. *Environ. Sci. Technol.* 43, 5565–5575.
- Washino, N., Saijo, Y., Sasaki, S., Kato, S., Ban, S., Konishi, K., Ito, R., Nakata, A., Iwasaki, Y., Saito, K., Nakazawa, H., Kishi, R., 2009. Correlations between prenatal exposure to perfluorinated chemicals and reduced fetal growth. *Environ. Health Perspect.* 117, 660–667.
- Weihe, P., Kato, K., Calafat, A.M., Nielsen, F., Wanigatunga, A.A., Needham, L.L., Grandjean, P., 2008. Serum concentrations of polyfluoroalkyl compounds in Faroese whale meat consumers. *Environ. Sci. Technol.* 42, 6291–6295.
- Wilhelm, M., Holzer, J., Dobler, L., Rauchfuss, K., Midasch, O., Kraft, M., Angerer, J., Wiesmuller, G., 2009. Preliminary observations on perfluorinated compounds in plasma samples (1977–2004) of young German adults from an area with perfluorooctanoate-contaminated drinking water. *Int. J. Hyg. Environ. Health* 212, 142–145.
- Zhang, T., Wu, Q., Sun, H.W., Zhang, X.Z., Yun, S.H., Kannan, K., 2010. Perfluorinated compounds in whole blood samples from infants, children, and adults in China. *Environ. Sci. Technol.* 44, 4341–4347.