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# Methane cracking as a bridge technology to the hydrogen economy

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## ABSTRACT

Shifting the fossil fuel dominated energy system to a sustainable hydrogen economy could mitigate climate change through reduction of greenhouse gas emissions. Because it is estimated that fossil fuels will remain a significant part of our energy system until midcentury, bridge technologies which use fossil fuels in an environmentally cleaner way offer an opportunity to reduce the warming impact of continued fossil fuel utilization. Methane cracking is a potential bridge technology during the transition to a sustainable hydrogen economy since it produces hydrogen with zero emissions of carbon dioxide. However, methane feedstock obtained from natural gas releases fugitive emissions of methane, a potent greenhouse gas that may offset methane cracking benefits. In this work, a model exploring the impact of methane cracking implementation in a hydrogen economy is presented, and the impact on global emissions of carbon dioxide and methane is explored. The results indicate that the hydrogen economy has the potential to reduce global carbon dioxide equivalent emissions between 0 and 27%, when methane leakage from natural gas is relatively low, methane cracking is employed to produce hydrogen, and a hydrogen fuel cell is applied. This wide range is a result of differences between the scenarios and the CH4 leakage rates used in the scenarios. On the other hand, when methane leakage from natural gas is relatively high, methane steam reforming is employed to produce hydrogen and an internal combustion engine is applied, the hydrogen economy leads to a net increase in global carbon dioxide equivalent emissions between 19 and 27%. © 2016 The Authors. Published by Elsevier Ltd on behalf of Hydrogen Energy Publications LLC. This is an open access article under the CC BY-NC-ND license (http://

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# Introduction

Transformation of the global energy system is recognized as one of the most consequential factors to successful global warming mitigation, as the energy sector is responsible for about two-thirds of all anthropogenic greenhouse gas (GHG) emissions today [1]. Reduction of GHG emissions is a technological and societal challenge of vast dimensions, requiring a deep change of our energy system which is currently

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dominated by fossil fuels [2,3]. There is a general consensus about the need to shift from a fossil fuel-based to a sustainable, low-carbon society. According to the Intergovernmental Panel on Climate Change (IPCC), GHG emissions must be near or below zero by the end of the 21st century to limit warming to 2 °C relative to pre-industrial levels [4]. Such a transformation of our energy system will be a challenge, requiring new technological breakthroughs and renewable energy investments, and could take decades or generations to be carried out. In spite of the urgent need to reduce GHG emissions to limit the increase in warming, fossil fuels are expected to remain an overwhelming share of the worldwide energy demand at least until 2040 [2,3].

The hydrogen (H<sub>2</sub>) economy has been proposed as a method for effecting this transformation, in which H<sub>2</sub> serves as one of the main global energy carriers [5]. H<sub>2</sub> can provide energy for transportation, buildings and industry, and can serve as a way to store energy [5]. Furthermore, it can be used as input to the H<sub>2</sub> fuel cell, which operates with a relatively high efficiency [5,6]. Importantly, the H<sub>2</sub> economy offers a way to mitigate global warming because H2 oxidation is carbonfree. However, H<sub>2</sub> can be produced by a wide array of fossil fuel and sustainable energy sources, meaning that GHG emissions can be released depending on the type of production process [5,7-9]. Today, the majority of H<sub>2</sub> is derived from fossil fuels by steam reforming and gasification techniques [10,11]. Since these fossil fuel-based conventional technologies generate GHG emissions [10], they have the potential to offset H<sub>2</sub>'s environmental benefits. The low production costs of technologies utilizing fossil fuels compared to the relatively high costs of renewable alternatives suggests that this trend will not change in the foreseeable future [7].

In this context, the deployment of technologies that utilize fossil fuel resources while generating low or zero GHG emissions may be required in the meantime. These technologies may constitute a bridge between the current unsustainable energy system and a future sustainable society, as they satisfy world energy demand through available fossil fuel resources in an environmentally cleaner way during the development of a renewable-based system so as to keep GHG emissions under control.

Methane cracking may be considered a bridge technology. In this process,  $CH_4$  is separated under high temperatures and in the absence of oxygen to produce elemental carbon and  $H_2$ , i.e., the reaction itself generates zero emissions of  $CO_2$  [12–14]. Methane cracking has the potential to cross the bridge from our current carbon-intensive energy system, based on fossil fuels, to a low-carbon energy future. As fossil fuels will likely be required in the energy transition [14], a bridging solution is needed in which fossil fuels may be used with low or zero emissions to mitigate climate change until a sustainable system is developed. Another competing bridge technology is carbon capture and sequestration (CCS), which avoids  $CO_2$  emissions to the atmosphere by storing  $CO_2$  underground [15].

During the methane cracking process, unreacted  $CH_4$  is separated from  $H_2$ , and is recirculated to the reactor. While the gas feedstock for methane cracking is mainly composed of  $CH_4$ , other hydrocarbons present are cracked in the same way as  $CH_4$  by thermal splitting of the C–H bonds. In the absence of oxygen,  $H_2$  produced by methane cracking is free of CO and CO<sub>2</sub>, making methane cracking a suitable method to produce  $H_2$  for fuel cells which require pure  $H_2$ . Another benefit of methane cracking is that the elemental carbon produced may possess an economic value because it is essential in the production of carbon fiber, which can be used in a variety of manufacturing applications. Nevertheless, developing a viable industrial implementation of the produced carbon for the economic benefits to be realized will be challenging [13,16].

Recent developments indicate that a large-scale, practical and viable application of this technology may be possible, by the process of bubbling methane into a liquid metal bath [17]. Lab-scale tests have overcome one of the main technological issues, i.e., carbon deposition leading to clogging of the reactor, which previously had prevented the development of methane cracking processes on the industrial-scale. For a practical and massive application of methane cracking, facilities to provide up to ~100–500 ton/day of hydrogen are required [18]. Work is in progress to advance the construction of a pilot plant to confirm the scalability of that technology.

The feedstock for methane cracking technology, natural gas, has been rapidly expanding in the United States due to significant advances in hydraulic fracturing and horizontal drilling in unconventional gas extraction [19,20]. While the upward trend of unconventional gas extraction has been mostly confined to the US, it is expected that this trend will continue globally [20]. Natural gas is often promoted as a bridging fuel on the road to a decarbonized energy system because it emits less CO<sub>2</sub> per unit of energy than oil or coal [21]. However, during the extraction, processing, and use of natural gas, CH<sub>4</sub> can escape to the atmosphere. This poses a problem to global warming mitigation because CH4 is a potent GHG, with a global warming potential (GWP) of 86 over 20 years and 34 over 100 years [4], meaning that even small CH<sub>4</sub> leaks in the natural gas supply chain can have a large impact on global warming. The leakage rate of methane is thus critical in determining the overall climate impact of natural gas.

In 2011, Howarth and colleagues estimated CH<sub>4</sub> leakage rates for conventional natural gas production at 3.8% and unconventional at 5.8%, which factors in upstream (well site and gas processing) and downstream (transmission, storage, and distribution) emissions over the lifetime production of a well (i.e., full life cycle-based emissions estimate) [22]. These values were much higher than the United States Environmental Protection Agency's (US EPA) estimate for conventional gas at the time, which was 1.1% (no separate estimate existed then for unconventional natural gas) [23]. Since then, many reports have been published on CH<sub>4</sub> leakage estimates from natural gas production, giving a range of 0.47-6% for conventional and 0.67-7.9% for unconventional natural gas production over the lifetime production of a well [24]. Several top-down atmospheric measurement campaigns have also been performed which quantify CH<sub>4</sub> leakage rates from natural gas system activity at specific regions. CH<sub>4</sub> leakage rates published by these studies range from lower to significantly higher compared with EPA estimates, with values ranging from 0.18 to 17.3% [25-33].

In a review on 20 years of literature on  $CH_4$  emissions from natural gas systems, Brandt et al. 2014 found that official inventories such as the US EPA frequently underestimate  $CH_4$  emissions; nevertheless, he found that very high values found in some regional atmospheric studies are unlikely to be representative of typical emissions from natural gas systems. Therefore, considerable uncertainty still remains on the actual extent of  $CH_4$  leakage from natural gas production. This is an important consideration when examining natural gas's viability as a bridging fuel, and in its promotion with technologies such as methane cracking to reduce the global warming impact. Other important considerations remain in addition to  $CH_4$  leakage when promoting technologies that utilize natural gas, such as the potential impacts on water, environmental and air pollution, but these considerations are not addressed in this work, which does not have the aim of being a life cycle assessment (LCA) following the ISO 14040/ 14044 Standards and methodologies [34].

In this paper, a model for the evaluation of a hydrogen economy with the practical implementation of methane cracking is presented, and used to quantify the impacts on global emissions of  $CO_2$  and  $CH_4$ . Scenarios are developed to illustrate the impact of replacing current industrial hydrogen production with methane cracking, and transport fuel with methane cracking-produced hydrogen, on emissions. The sensitivity of the results to  $CH_4$  leakage rates from natural gas production is evaluated. Furthermore, other important parameters such as elemental carbon production from methane cracking, fugitive  $H_2$  emissions, and the required efficiency of the methane cracking process are also analyzed.

# Methods

# Scenarios

Scenarios were designed to study different aspects of the methane cracking – hydrogen (MC-H<sub>2</sub>) economy to evaluate the range of effects on global CO<sub>2</sub> and CH<sub>4</sub> emissions. The MC-H<sub>2</sub> economy is defined here as a H<sub>2</sub> economy facilitated by methane cracking. The scenarios were constructed as a projected snapshot of the present-day situation, in which the proposed changes are enacted immediately. Sectors of special interest to the MC-H<sub>2</sub> economy and therefore of focus in the scenarios are industrial H<sub>2</sub> production and road transportation. An overview of the scenarios is provided in Table 1.

# Baseline

A baseline scenario was developed to represent the present situation regarding industrial  $H_2$  production and road transportation and their impact on emissions. This scenario was made as a means to compare alternative scenarios.

# Industrial H<sub>2</sub>

 $H_2$  is a vital feedstock in the global chemical industry. Current values for global  $H_2$  production are varied in the literature. For example, in the International Energy Agency's (IEA) 2007 report it estimates that 65 Mton  $H_2$  are produced annually for industrial end-use applications [7], while Bond et al. 2011 estimate 47 Mton [35], and the US Department of Energy estimates >50 Mton [36]. In this work, the IEA value was assumed for industrial  $H_2$  production since other data were also obtained from the IEA inventory (see Section Activity data).

| Table 1 — Scenario overview.            |  |  |  |  |
|---|--|--|--|--|
| Scenario                                | Full name                                      | Description  |  |  |
| Baseline                                | Baseline                                       | Scenario representing actual<br>situation and emissions<br>regarding H <sub>2</sub> production<br>and road transportation.   |  |  |
| Industrial H <sub>2</sub>               | Industrial<br>hydrogen                         | Methane cracking is fully<br>implemented for production of<br>industrial H <sub>2</sub> .  |  |  |
| RoadTrans H <sub>2</sub>                | Road<br>transportation<br>hydrogen             | $H_2$ fuel is fully implemented to<br>cover the energy needs of<br>the road transportation sector.<br>The $H_2$ implemented is<br>produced by methane cracking.  |  |  |
| Pessimistic<br>RoadTrans H <sub>2</sub> | Pessimistic road<br>transportation<br>hydrogen | $H_2$ fuel is fully implemented to<br>cover the energy needs of<br>the road transportation sector,<br>in the context of pessimistic<br>(high-emission) assumptions.  |  |  |
| H <sub>2</sub> Econ                     | Hydrogen<br>economy                            | Contains implementations of<br>both Industrial $H_2$ and<br>RoadTrans $H_2$ combined in one<br>scenario. This scenario is used<br>to investigate various aspects<br>of the $H_2$ economy, i.e., the<br>upper limit of $CH_4$ leakage in<br>natural gas production, and the<br>methane cracking efficiency. |  |  |

Around 96% of global  $H_2$  is derived from fossil fuels by conventional technologies, i.e., methane steam reforming, oil/ naphtha reforming and coal gasification, which generate considerable GHG emissions (Table 2) [10]. Therefore, in the *Industrial hydrogen full penetration* scenario (*Industrial*  $H_2$ ), methane cracking supplies current industrial  $H_2$  demand, thereby replacing methane steam reforming, oil/naphtha reforming and coal gasification. Methane cracking does not replace electrolysis or the share of technologies producing  $H_2$ for methanol synthesis (because the methanol reaction relies on syngas- $H_2$  and CO/CO<sub>2</sub>, which is not produced by methane cracking).

 $CO_2$  emission factors (EF) associated with  $H_2$  production technologies are much greater in magnitude compared to those of  $CH_4$ , e.g., for coal gasification 107,000 kg  $CO_2$  and 4 kg  $CH_4$  are emitted per TJ (Table 2). This is because the  $CO_2$  EFs are based on the carbon content of the fuel, and here reflect the assumption that 100% of the fuel's carbon gets oxidized during

| Table 2 – Percent shares of global hydrogen production,  |
|--|
| and the corresponding emission factors (EF) for hydrogen |
| production technologies [10,21].                         |

| H <sub>2</sub> production | Product  | EF [kg/TJ] <sup>a</sup> |                 |        |
|---------------------------|----------|-------------------------|-----------------|--------|
| technology                | Baseline | Industrial $H_2$        | CO <sub>2</sub> | $CH_4$ |
| Methane steam reforming   | 48%      | 0%                      | 54,500          | 2      |
| Oil/naphtha reforming     | 30%      | 0%                      | 71,100          | 4      |
| Coal gasification         | 18%      | 0%                      | 107,000         | 4      |
| Electrolysis              | 4%       | 4%                      | 0               | 0      |
| Methane cracking          | 0%       | 96%                     | 0               | 2      |

<sup>a</sup> These emission factors were calibrated [16] and selected from a range provided by the IPCC 2006 Guidelines.

combustion. On the other hand, the  $CH_4$  EFs represent fugitive emissions released from fuels, which are comparatively low in magnitude to  $CO_2$  emissions in the energy sector [21].

## RoadTrans H<sub>2</sub>

The road transportation sector is responsible for generating a significant level of  $CO_2$  and  $CH_4$  emissions each year, primarily due to the use of oil and natural gas (Table 3). In fact, transportation accounts for nearly a quarter of global  $CO_2$  emissions [37]. Therefore, in the *Road transportation hydrogen full penetration* scenario (*RoadTrans H*<sub>2</sub>), oil and natural gas in road transportation are fully replaced with H<sub>2</sub> that is produced by methane cracking, while the relative shares of electricity and biofuels are kept the same. In this scenario it is assumed that vehicles are powered by hydrogen fuel cells, because fuel cells possess a relatively high tank-to-wheel efficiency (a measure of the drivetrain performance) compared to an internal combustion engine (ICE).

# Pessimistic RoadTrans H<sub>2</sub>

In the Pessimistic road transportation hydrogen scenario (Pessimistic RoadTrans H<sub>2</sub>), H<sub>2</sub> fuel covers the energy needs of the road transportation sector as done in the RoadTrans H<sub>2</sub> scenario, but in the context of pessimistic assumptions. Specifically, it is assumed that H<sub>2</sub> fuel is produced by methane steam reforming, which is a conventional, fossil fuel-based (i.e., natural gas) H<sub>2</sub> production technology that releases CO<sub>2</sub> during the reaction process. Furthermore, it is assumed that CH<sub>4</sub> leakage rates from natural gas production are on the upper-end of the EF range, and that a H<sub>2</sub> internal combustion engine (ICE) is employed because this is considerably less efficient in terms of tank-to-wheel efficiency than H<sub>2</sub> fuel cell vehicles.

#### $H_2Econ$

The  $H_2Econ$  scenario is not handled like the previous scenarios, and instead is used for other applications of the MC- $H_2$  model discussed in Section Other aspects of interest to the MC- $H_2$ economy. The  $H_2Econ$  scenario contains the implementations assumed in the Industrial  $H_2$  and RoadTrans  $H_2$  scenarios.

## Natural gas production methane leakage rates

The scenarios were calculated with three sets of  $CH_4$  leakage rates in natural gas production, which cover a broad range of

Table 3 – Road transportation energy share (based on energy activity per energy form for year 2012), and tankto-wheel efficiency (TTW) in road transportation based on fuel type and energy converter [2,46].

| Fuel        | Energy share [%] |                | Energy    | TTW            |
|-------------|------------------|----------------|-----------|----------------|
| type        | Baseline         | RoadTrans      | converter | efficiency [%] |
|             |                  | H <sub>2</sub> |           |                |
| Oil         | 92.8%            | 0%             | ICE       | 22             |
| Natural     | 3.8%             | 0%             | ICE       | 16             |
| gas         |                  |                |           |                |
| Biofuels    | 2.4%             | 2.4%           | ICE       | 22             |
| Electricity | 1.0%             | 1.0%           | Battery   | 82             |
| Hydrogen    | 0%               | 96.6%          | Fuel cell | 53             |
|             |                  |                | ICE       | 28             |

estimates in the literature that we consider reasonable. This was done to explore the impact of varying and potential CH<sub>4</sub> leakage rates in the natural gas system on total emissions in the scenarios. These sets were retrieved from EPA 1996 [23], Hultman 2011 [38] and Howarth 2011 [22] (see Table 4). The EPA 1996 values were used as lower-end rates, those from Howarth 2011 were used as higher-end rates, and those from Hultman 2011 were used as middle-value rates. However, for the RT H<sub>2</sub> Pessimistic scenario, upper-bound CH<sub>4</sub> leakage rates from Howarth 2011 were applied. Leakage rates are disaggregated into upstream and downstream processes from natural gas production. Upstream emissions occur at the well site and during gas processing, while downstream emissions occur during storage, transport and distribution of gas to customers. A distinction is made here between conventional and unconventional gas. Unconventional gas is obtained from relatively new sources that require unconventional methods for its extraction (e.g., hydraulic fracturing of shale gas), whereas conventional gas is obtained from traditional sources for which conventional methods can be used for its extraction. Aside from the extraction method, no difference is assumed between unconventional and conventional natural gas itself. Each calculation was performed twice, once assuming natural gas supply via 100% conventional natural gas, and once assuming natural gas supply via 100% unconventional natural gas. This was done to provide the full range of emissions estimates, because emissions are generally lower in conventional and higher in unconventional natural gas production.

Note that the downstream leakage rates displayed in Table 4 are representative of decentralized natural gas utilization. This is because these downstream leakage rates are higher than those that would result from centralized natural gas utilization. This is due to the fact that low-pressure urban distribution lines have a higher leakage rate than gas lines delivering natural gas to centralized power plants [20]. The decentralized downstream leakage rates were used to reduce complexity, so that the same rates could be employed for all segments of natural gas utilization in the model (i.e., for both industrial and private consumer end-use). While decentralized natural gas utilization leads to greater  $CH_4$  emissions, decentralized production also has its own benefits with respect to methane cracking such as sharply reducing delivery

| Table 4 – Methane emission factor estimates expressed     as a percentage of total natural gas produced. |          |          |            |  |
|--|----------|----------|------------|--|
| Source   | Upstream | Upstream | Downstream |  |
|  |          |          | 11         |  |
|  | gas      | gas      | _          |  |
| EPA 1996 <sup>a</sup> [23]   | 0.2%     | _        | 0.9%       |  |
| Hultman 2011 [38]  | 1.3%     | 2.8%     | 0.9%       |  |
| Howarth 2011 [22]  | 1.4%     | 3.3%     | 2.5%       |  |
| Upper-end  | 2.4%     | 4.3%     | 3.6%       |  |
| estimates  |          |          |            |  |
| of Howarth   |          |          |            |  |
| 2011 [22]  |          |          |            |  |
| <sup>a</sup> EPA 1996 [23] did not provide an EF for upstream unconventional                             |          |          |            |  |
| gas. Therefore upstream unconventional gas is assumed to have  |          |          |            |  |

gas. I herefore upstream unconventional gas is assumed to hav the same EF as upstream conventional gas for EPA 1996 data.

and storage costs of  $H_2$ , and providing more independence to the consumer [39].

Displayed in Table 5 is an emissions comparison between natural gas, coal and oil, factoring in the three EF sets for natural gas provided in Table 4. The 20-year global warming potential (GWP) of CH<sub>4</sub> was used to convert CH<sub>4</sub> emissions into  $CO_2$ -eq emissions; the reason for this is explained in Section Emissions and CH<sub>4</sub> global warming potential. Note that the  $CO_2$ -eq for natural gas production is different among the EPA, Hultman and Howarth, since each of these sources have different CH<sub>4</sub> EF estimates.

## Other aspects of interest to the MC-H<sub>2</sub> economy

## Elemental carbon production from methane cracking

A side effect of methane cracking employed in the *Industrial*  $H_2$  and *RoadTrans*  $H_2$  scenarios is that elemental carbon (carbon black) is produced as a by-product of methane cracking. Today, carbon black production amounts to 8.1 Mton per year [40]. The carbon produced by methane cracking in the *Industrial*  $H_2$  and *RoadTrans*  $H_2$  scenarios was quantified at 170 and 790 Mton carbon per year, some orders of magnitude higher that the current black carbon market. The amount of high quality carbon produced by methane cracking will make this material available at a very low cost, likely boosting the development of new carbon-based technologies, for instance, graphene applications.

#### H<sub>2</sub> emissions

Incomplete combustion of fossil fuels and leakage of industrial  $H_2$  are both sources of anthropogenic  $H_2$  emissions. In the  $H_2$  economy,  $H_2$  emissions would be emitted during the production, transport, storage, and use of  $H_2$ . Increased  $H_2$ 

| Table 5 — Aggregated emission factors for production of coal, oil and natural gas. |                 |          |                     |  |
|--|-----------------|----------|---------------------|--|
| Fossil fuel production   |                 | EF [kg/T | [[                  |  |
|  | CO <sub>2</sub> | $CH_4$   | CO <sub>2</sub> -eq |  |
| Coalª  | -               | 450      | 39 000              |  |
| Oil <sup>b</sup>   | 2900            | 280      | 27 000              |  |
| Natural gas <sup>c</sup> (EPA 1996 [23])   | 4100            | 240      | 25 000              |  |
| Natural gas <sup>c</sup> (Hultman 2011 [38])                                       | 4100            | 540      | 51 000              |  |
| Natural gas <sup>c</sup> (Howarth 2011 [22])                                       | 4100            | 930      | 84 000              |  |
|  |                 |          |                     |  |

<sup>a</sup> Coal: The CO<sub>2</sub> EF is not provided in the literature because it is still being developed; therefore it is not included in this model. Note that the CO<sub>2</sub> EF for coal production is estimated to be low in comparison to the CH<sub>4</sub> EF. The CH<sub>4</sub> EF covers mining and postmining emissions, averaged for underground and surface coal mines [21].

- <sup>b</sup> Oil: The CO<sub>2</sub> EF covers oil production, transport and refining [21]. The CH<sub>4</sub> EF covers emissions from oil production, transport and refining [21,47].
- <sup>c</sup> Natural gas: The CO<sub>2</sub> EF is based on the conventional and unconventional CO<sub>2</sub> EFs [22], calculated based on the year 2012 ratio of conventional to unconventional natural gas production [2]. The CH<sub>4</sub> EF covers upstream (conventional and unconventional) and downstream emissions from the EPA 1996, Howarth 2011 and Hultman 2011 publications, and is calculated based on the year 2012 ratio of conventional to unconventional natural gas production [2].

emissions may exacerbate global warming by reducing the atmospheric concentration of the hydroxyl radical (OH), which would increase the lifetime of CH<sub>4</sub> and other trace gases [37,41]. For this reason, exploring the effects of increased H<sub>2</sub> usage on H<sub>2</sub> emissions is useful to understand ancillary effects of the H<sub>2</sub> economy on the climate. Therefore, H<sub>2</sub> emissions from H<sub>2</sub> fuel use in road transportation under the RoadTrans H<sub>2</sub> scenario were quantified and compared to estimates on current H<sub>2</sub> emissions from incomplete combustion of oil and natural gas in road transportation by Wokaun and Wilhelm 2011 [42]. This provides the H<sub>2</sub> emissions balance that would occur in road transportation, i.e., the incurred H<sub>2</sub> emissions from introduced H2 fuel use compared to the eliminated H<sub>2</sub> emissions from phased out oil and natural gas use. H<sub>2</sub> fuel leakage rates in road transportation of 1 and 2% were selected from Wokaun and Wilhelm 2011. While higher values of up to 4% were provided, these rates were deemed unrealistic as they would result in a huge commodity and significant monetary loss. Furthermore, it was found that industrial H<sub>2</sub> leakage rates from the H<sub>2</sub> distribution grid in Germany are already as low as 0.1% [41,43,44].

#### Upper limit of CH<sub>4</sub> leakage in natural gas production

As discussed in Section Introduction, there is a high level of uncertainty surrounding CH<sub>4</sub> leakage rates in natural gas production. Therefore, it was determined what the upper limit of CH<sub>4</sub> leakage in natural gas production is that would lead to the same CO<sub>2</sub>-eq emissions in H<sub>2</sub>Econ as in the baseline. The CH<sub>4</sub> leakage limit was determined by performing an optimization, in which the same natural gas production CH<sub>4</sub> leakage rate was applied to the baseline and to H<sub>2</sub>Econ and adjusted until the CO<sub>2</sub>-eq emissions were the same in both scenarios.

#### Methane cracking efficiency

Throughout the previous scenarios, an energy conversion efficiency of 55% was assumed for the methane cracking process based on estimations from the literature [14]. The energy conversion efficiency for the methane cracking process is defined here as the ratio between the energy output in terms of H<sub>2</sub> and the energy input in terms of CH<sub>4</sub>, where CH<sub>4</sub> also supplies energy to overcome the reaction barrier ( $\Delta H^0 = 74.85 \text{ kJ/mol}$ ). However, methane cracking has not yet been implemented on a commercial scale and so the efficiency that will be realized in practice is not yet known. In order to further explore the MC-H<sub>2</sub> economy, the minimum required methane cracking efficiency was determined. This was done by performing an optimization, in which the same efficiency was applied to the baseline and H<sub>2</sub>Econ and adjusted until the CO<sub>2</sub>-eq emissions were the same in both scenarios.

#### Model

The methane cracking – hydrogen economy (MC-H<sub>2</sub>) model was developed in order to implement the scenarios described in Section Scenarios, quantifying the effect of changes in H<sub>2</sub> production and road transportation on  $CO_2$  and  $CH_4$  emissions. The model quantifies anthropogenic emissions on the global scale from the sectors of relevance to the industrial use of hydrogen and the road transportation sector. Namely, global  $CO_2$  emissions include emissions from production of

coal, oil, and natural gas, as well as their use in industry, transport, buildings (residential, services and non-specified other) and other (agriculture and non-energy use), but exclude electricity and heat generation emissions.  $CO_2$  emissions from electricity generation are only considered for electricity requirements for electrolysis in  $H_2$  production. Global  $CH_4$  emissions include stationary and combustion, oil and natural gas, coal mining, and biomass.

The MC-H<sub>2</sub> model is based on a series of input parameters and equations. The structure of the model is based on various aspects of the energy sector that are required to quantify changes in  $CO_2$  and  $CH_4$  emissions resulting from implementation of a H<sub>2</sub> economy. These aspects are referred to here as domains; in total, the model is made up of 17 domains. The model domains and the information flow in the model are shown in Fig. 1. A full description of the model is given in Weger 2015 [16].

# Model evaluation

The MC-H<sub>2</sub> model was evaluated by a sensitivity analysis and subsequently calibrated. This evaluation is described in detail in Weger 2015 [16]. The parameters shown to have the greatest sensitivity are production of both oil and natural gas. The upper-end CH<sub>4</sub> EF value for oil production led to a 200% increase in CO<sub>2</sub>-eq emissions, whereas the upper-end CH<sub>4</sub> EF value for upstream unconventional natural gas production led to a 35% increase in emissions. The considerable sensitivity observed here is due to wide uncertainty ranges in the EFs for production of both oil and natural gas. Calibration of the model is described in detail in Weger 2015 [16].

#### Data

#### Activity data

Activity data represent the amount of fuel consumption associated with a specific activity. The activity data used in this model were retrieved from the IEA. The data represent activity from the year 2012. The IEA activity data used in MC-H<sub>2</sub> represent total final consumption (TFC) energy, excluding electricity and heat. TFC includes industry, transport, buildings (residential, services and non-specified other) and other (agriculture and non-energy use) [2]. The demand for saleable fuel for combustion is treated as equivalent to the fuel preproduction to simplify calculations.

#### **Emission** factors

EF's represent the average emissions released per unit of given activity, and are typically provided as a range. The EFs used in the model were primarily retrieved from the 2006 IPCC Guidelines [21]. When data could not be retrieved from the IPCC, the data were obtained from alternative sources, as listed in Table 6.

# Emissions and CH<sub>4</sub> global warming potential

CO2 and CH4 emissions are presented in units of CO2 equivalent (CO<sub>2</sub>-eq). CO<sub>2</sub>-eq emissions are calculated by multiplying CH<sub>4</sub> emissions by the GWP of CH<sub>4</sub>, and adding this value to the CO<sub>2</sub> emissions. The GWP measures the relative heat-trapping ability of CH<sub>4</sub> compared to CO<sub>2</sub>, and is a function of the time frame considered after a pulse emission of CH4. The high warming potential of CH4 is primarily due to its strong absorption of infrared radiation emitted from the Earth's surface. Common time intervals to discuss the GWP of CH<sub>4</sub> are 20 and 100 years; CH<sub>4</sub> has a much larger 20-year GWP, because its atmospheric lifetime is about 12 years, while CO<sub>2</sub> has an effective influence on the atmosphere for about a century [4,20]. The 20-year GWP of CH<sub>4</sub> is 86, while the GWP drops to 34 over 100 years. The shorter, 20-year time scale for the GWP is more relevant and appropriate because of the urgent need to reduce  $CH_4$  emissions in the next decades, so as to prevent climate tipping points such as the melting of permafrost, and to slow the rate of global warming. Therefore it is the 20-year GWP that was used here to convert CH<sub>4</sub> to CO<sub>2</sub>-eq emissions.

#### Methodological approach

The Tier 1 approach from the 2006 *IPCC Guidelines* was employed to calculate emissions [21]. In this approach, emissions are estimated based on the quantities of fuel combusted (activity data, AD) and globally-averaged EFs for a particular energy



Fig. 1 – Schematic of the MC-H<sub>2</sub> model domains. The arrows follow the flow of energy demand from end-use applications to the point of energy production. The main component groups in MC-H<sub>2</sub> are H<sub>2</sub> (purple), fossil fuels (yellow), and transportation (green)

| Table 6 – Emission factor data sources used in MC-H <sub>2</sub> . |                    |                 |  |  |
|--|--------------------|-----------------|--|--|
| Domain   | Parameter          | Source          |  |  |
| Fossil fuel  | Coal               | IPCC [21]       |  |  |
| economy  | Oil                | IPCC [21]       |  |  |
|  | Natural gas        | IPCC [21]       |  |  |
| Road   | Oil                | IPCC [21]       |  |  |
| transportation   | Natural gas        | IPCC [21]       |  |  |
|  | Electricity        | -               |  |  |
|  | Biofuels           | IPCC [21]       |  |  |
|  | Hydrogen           | Wokaun and      |  |  |
|  |                    | Wilhelm [42]    |  |  |
| Coal   | Underground mining | IPCC [21]       |  |  |
| production   | Underground        | IPCC [21]       |  |  |
|  | post-mining        | IDCC [21]       |  |  |
|  | Surface maining    | IPCC [21]       |  |  |
| Cool   | Process            | IPCC [21]       |  |  |
| gasification   | FIOCESS            |                 |  |  |
| Oil production   | Oil well           | IPCC [21]       |  |  |
|  | Oil production     | IPCC [21];      |  |  |
|  | -                  | Cai [47]        |  |  |
|  | Oil transport      | IPCC [21]       |  |  |
|  | Oil refining       | IPCC [21]       |  |  |
| Oil/naphtha  | Process            | IPCC [21]       |  |  |
| reforming  |                    |                 |  |  |
| Natural gas  | Upstream           | EPA [23];       |  |  |
| production   | conventional       | Hultman [38];   |  |  |
|  |                    | Howarth [22]    |  |  |
|  | Upstream           | EPA [23];       |  |  |
|  | unconventional     | Hultman [38];   |  |  |
|  |                    | Howarth [22]    |  |  |
|  | Downstream         | EPA [23];       |  |  |
|  |                    | Hultman [38];   |  |  |
|  | _                  | Howarth [22]    |  |  |
| Methane  | Process            | IPCC [21]       |  |  |
| cracking   | <b>D</b>           | IDGG [04]       |  |  |
| steam<br>reforming   | Process            | IPCC [21]       |  |  |
| Biofuel  | Process            | Mortimer [48]   |  |  |
| production   |                    | Cai [47]        |  |  |
| Electricity  | Process            | IEA [49]        |  |  |
| production   | 100000             | Ecometrica [50] |  |  |
| Electrolysis   | Process            | -               |  |  |

form/application, where emissions =  $AD_{fuel} \times EF_{GHG, fuel}$ . CO<sub>2</sub> EF's are primarily dependent on the carbon content of the fuel because the majority of the carbon will be oxidized to CO<sub>2</sub>, meaning that globally-averaged CO<sub>2</sub> EFs do not introduce considerable uncertainty. On the other hand, CH<sub>4</sub> EF's are dependent on factors subject to variability, i.e., combustion technology and operating conditions, meaning that globally-averaged CH<sub>4</sub> EFs lead to relatively high uncertainty (for more information, refer to the 2006 IPCC Guidelines).

# **Results and discussion**

# Industrial H<sub>2</sub>

In this scenario set, *Industrial*  $H_2$ , conventional fossil fuelbased technologies are replaced with methane cracking for industrial  $H_2$  production as explained in Section Scenarios. This generally leads to a decrease in  $CO_2$ -eq emissions; however, the achieved emissions decrease is low, within the range of 0–3.8%. The results are displayed in Table 7 and Fig. 2.

The main reason for the low emissions decrease is that the combined  $CO_2$  and  $CH_4$  EFs from natural gas production are higher than from coal and oil production (see Table 5), which curtails the emissions reductions achieved through methane cracking (see Table 2) from its zero-CO<sub>2</sub> emissions. The emissions decrease is greatest when the EPA 1996 EFs are applied and lowest when the Howarth 2011 EFs are applied. This is due to the combined CO<sub>2</sub> and CH<sub>4</sub> EFs being lowest with EPA 1996 and the highest with Howarth 2011. Furthermore, the decrease in emissions is greater when the natural gas supply used in the methane cracking process is 100% conventional, and lower when the natural gas supply is 100% unconventional. This is because the EFs for CO<sub>2</sub> and CH<sub>4</sub> combined are greater during unconventional natural gas production than conventional natural gas production. Furthermore, in the Howarth 2011 Industrial H<sub>2</sub> scenario supplied with 100% unconventional gas, no change in CO2-eq emissions is observed. In this scenario, the Howarth CH4 leakage rate from unconventional natural gas production is high enough that it completely offsets the decrease in CO<sub>2</sub> emissions achieved through methane cracking. On the other hand, the EPA 1996 and Hultman 2011 EFs for CH<sub>4</sub> leakage from unconventional natural gas production are low enough so that they still lead to emissions reductions in their corresponding Ind H<sub>2</sub> Full Pen scenarios. This underlines that the effectiveness of methane cracking in reducing emissions from industrial H<sub>2</sub> production is dependent on the leakage rate of  $CH_4$  in natural gas production.

# RoadTrans H<sub>2</sub>

In the next scenario set,  $RoadTrans H_2$ , oil and natural gas are fully replaced with  $H_2$  produced by methane cracking to cover the energy needs in road transportation as explained in

Table 7 – Total global emissions of carbon dioxide and methane (Mton CO<sub>2</sub>-eq) from the baseline and from the Industrial  $H_2$  scenario. Emissions are calculated for supply with 100% conventional and for 100% unconventional natural gas.

| Data source                             | Scenario                  | Total emissions [Mton CO <sub>2</sub> -eq] <sup>a</sup> |                               |  |
|---|---------------------------|---|-------------------------------|--|
| for natural<br>gas<br>production<br>EFs |                           | Conventional<br>natural gas                             | Unconventional<br>natural gas |  |
| EPA 1996                                | Baseline                  | 26 000  | 26 000                        |  |
|   | Industrial H <sub>2</sub> | 25 000  | 25 000                        |  |
|   | Change %                  | -3.8%   | -3.8%                         |  |
| Hultman 2011                            | Baseline                  | 27 000  | 29 000                        |  |
|   | Industrial H <sub>2</sub> | 26 000  | 28 000                        |  |
|   | Change %                  | -3.7%   | -3.4%                         |  |
| Howarth 2011                            | Baseline                  | 29 000  | 31 000                        |  |
|   | Industrial H <sub>2</sub> | 28 000  | 31 000                        |  |
|   | Change %                  | -3.4%   | 0%                            |  |

The italics represent the % change in emissions from the baseline to the Industrial  $H_2$  scenario.

<sup>a</sup> CO<sub>2</sub>-eq for CH<sub>4</sub> over 20 years, with a GWP value of 86 [4].



Fig. 2 – Total global emissions of carbon dioxide and methane (Mton  $CO_2$ -eq). 100% conventional natural gas (in blue), represents the total emissions incurred from 100% conventional natural gas supply. 100% unconventional natural gas (in red), represents the additional emissions incurred from 100% unconventional natural gas supply.

Section Scenarios. This consistently leads to a decrease in  $CO_2$ -eq emissions, and in some scenarios substantially so, within the range of 6.5–27%. This wide range results from the different  $CH_4$  leakage rates applied in the scenarios, and will be explained in greater detail below. The results are displayed below in Table 8 and Fig. 3.

The trends observed in the *Industrial*  $H_2$  scenarios were generally observed in the *RoadTrans*  $H_2$  scenarios as well. Namely, the emissions decrease is greatest when applying the EPA 1996 EFs and least when applying the Howarth 2011 EFs, and the emissions decrease is greater when natural gas supply is 100% conventional, and less when it is 100% unconventional. The explanations for these observations are discussed in the previous section. However, differences are also observed between the scenario sets. First, the magnitude of the emissions decrease is substantially greater among the *RoadTrans*  $H_2$  scenarios compared to the *Industrial*  $H_2$  scenarios (6.5–27% CO<sub>2</sub>-eq emissions reductions in the

Table 8 – Total global emissions of carbon dioxide and methane (Mton CO<sub>2</sub>-eq) from the baseline and from the RoadTrans  $H_2$  scenario. Emissions are calculated for supply with 100% conventional and for 100% unconventional natural gas.

| Data source  | Scenario        | Total emissions [Mton CO <sub>2</sub> -eq] <sup>a</sup> |                            |
|--------------|-----------------|---|----------------------------|
|              |                 | Conventional natural gas                                | Unconventional natural gas |
| EPA 1996     | Baseline        | 26 000  | 26 000                     |
|              | RoadTrans $H_2$ | 19 000  | 20 000                     |
|              | Change %        | -27%  | -23%                       |
| Hultman 2011 | Baseline        | 27 000  | 29 000                     |
|              | RoadTrans $H_2$ | 22 000  | 25 000                     |
|              | Change %        | -18%  | -14%                       |
| Howarth 2011 | Baseline        | 29 000  | 31 000                     |
|              | RoadTrans $H_2$ | 25 000  | 29 000                     |
|              | Change %        | -14%  | -6.5%                      |

The italics represent the % change in emissions from the baseline to the RoadTrans  $H_2$  scenario.

<sup>a</sup>  $CO_2$ -eq for  $CH_4$  over 20 years, with a GWP value of 86 [4].



Fig. 3 – Total global emissions of carbon dioxide and methane (Mton  $CO_2$ -eq). 100% conventional natural gas (in blue), represents the total emissions incurred from 100% conventional natural gas supply. 100% unconventional natural gas (in red), represents the additional emissions incurred from 100% unconventional natural gas supply.

RoadTrans H<sub>2</sub> scenarios compared with 0-3.8% Industrial H<sub>2</sub> scenarios). The main reason for the is that the RoadTrans H<sub>2</sub> scenarios require substantially less fuel in road transportation compared to the baseline and Industrial H2 scenarios. This is because the efficiency of the H<sub>2</sub> fuel cell is more than twice as high as that of the petrol/diesel ICE, which essentially reduces the road transportation fuel demand by half. Second, the range of emissions decrease is greater in the RoadTrans H<sub>2</sub> scenarios compared to the Industrial H<sub>2</sub> scenarios. The wide range of emissions decrease is due to the varying CH<sub>4</sub> leakage rates used, which have a more pronounced impact on total emissions in the RoadTrans H<sub>2</sub> scenarios than in the Industrial H<sub>2</sub> scenarios. This is because more natural gas is needed in total in the RoadTrans H<sub>2</sub> scenarios to provide H<sub>2</sub> fuel by methane cracking for road transportation. Third, the Howarth Industrial H<sub>2</sub> scenario utilizing 100% unconventional natural gas leads to no change in emissions, while the Howarth RoadTrans H<sub>2</sub> scenario utilizing 100% unconventional natural gas leads to a net decrease in emissions. One of the main reasons an emissions decrease was calculated in the latter scenario is due to the significant decrease in road transportation fuel demand. Nevertheless, the emissions decrease in this scenario is low, at 6.5%, which indicates that the very high CH<sub>4</sub> leakage rate from unconventional natural gas production provided by Howarth 2011 is close to the limit at which the increase in CH<sub>4</sub> emissions due to increased gas production cannot be compensated by the decrease in CO2 emissions achieved through methane cracking. This emphasizes the importance of CH4 emissions in natural gas production on climate benefits through emissions reductions achieved in the RoadTrans H<sub>2</sub> scenarios.

# Pessimistic RoadTrans H<sub>2</sub>

In the next scenario, *Pessimistic RoadTrans*  $H_2$ ,  $H_2$  fuel replaces oil and natural gas in the road transportation sector, under pessimistic assumptions as explained in Section Scenarios. With 100% conventional natural gas supply,  $CO_2$ -eq emissions

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Mton CO<sub>2</sub>-eq

increase by 19% from the baseline to the Pessimistic RoadTrans  $H_2$  (from 31,000 to 37,000 Mton CO<sub>2</sub>-eq), and with 100% unconventional natural gas supply CO<sub>2</sub>-eq emissions increase by as much as 27% (from 33,000 to 42,000 Mton CO<sub>2</sub>-eq). The results are shown in Fig. 4.

The results underline the critical role of assumptions in whether the H<sub>2</sub> economy has a beneficial or detrimental impact on the climate through emissions. Namely, pessimistic assumptions as applied here, including the production of H<sub>2</sub> by high-emissions-generating methane steam reforming, relatively high CH4 leakage rates from natural gas production, and a low tank-to-wheel efficiency of H<sub>2</sub> through use of a H<sub>2</sub> internal combustion engine, result in a significant and unfavorable impact on the climate through emissions increase. On the other hand, optimistic assumptions such as the production of H<sub>2</sub> by low-emissions-generating methane cracking, relatively low CH4 leakage rates from natural gas production, and a high tank-to-wheel efficiency of H<sub>2</sub> by way of a H<sub>2</sub> fuel cell enable climate benefits through emissions reduction. Thus, the H<sub>2</sub> economy can lead to net reduction of GHG emissions, but only if a low-emissions-generating H<sub>2</sub> production technology like methane cracking is applied, if CH<sub>4</sub> emissions from natural gas production are low, and a high tank-to-wheel efficiency of H<sub>2</sub> (i.e., application of the H<sub>2</sub> fuel cell) is possible.

# Elemental carbon production from methane cracking

In this section, the elemental carbon produced by methane cracking is quantified in the *Industrial*  $H_2$  and *RoadTrans*  $H_2$  scenarios as explained in Section Scenarios. In the *Industrial*  $H_2$  scenario, 170 Mton carbon is produced by methane cracking. On the other hand, 790 Mton carbon is produced by methane cracking in the *RoadTrans*  $H_2$  scenario, which amounts to 4.5 times more carbon production than in the *Industrial*  $H_2$  scenario. This is due to the fact that much more  $H_2$  is required to be produced by methane cracking in the road transportation sector than in the industrial  $H_2$  scenario. If  $H_2$  scenario are carbon production than in the road transportation sector than in the industrial  $H_2$  scenario.



Fig. 4 – Total global emissions of carbon dioxide and methane (shown in Mton  $CO_2$ -eq). 100% conventional natural gas (in blue), represents the total emissions incurred from 100% conventional natural gas supply. 100% unconventional natural gas (in red), represents the additional emissions incurred from 100% unconventional natural gas supply.

the carbon produced from methane cracking can be later used as a commodity in commercial applications, this would reduce life cycle emissions incurred from carbon production [45].

#### Hydrogen emissions

In this section,  $H_2$  emissions from  $H_2$  fuel use in road transportation under the RoadTrans H<sub>2</sub> scenario are quantified and compared to estimates on current H<sub>2</sub> emissions from incomplete combustion of oil and natural gas in road transportation, as explained in Section H<sub>2</sub> emissions. The results are displayed in Fig. 5. In the RoadTrans H<sub>2</sub> scenario, a H<sub>2</sub> leakage rate of 1% results in 2.6 Mton H<sub>2</sub> emissions per year, while a H<sub>2</sub> leakage rate of 2% results in 5.4 Mton H<sub>2</sub> emissions per year (explained in Section H<sub>2</sub> emissions). These values are comparable to current emissions in the literature [2,42], from which extrapolated H<sub>2</sub> emissions based on the increase of oil and natural gas in road transportation for year 2012 are 4.75 Mton [2]. Therefore, with full penetration of H<sub>2</sub> fuel in road transportation, H<sub>2</sub> emissions stay roughly the same, changing by -2.15 to +0.65 Mton H<sub>2</sub> per year. Based on the modest change in H<sub>2</sub> emissions resulting from replacing oil and natural gas with H<sub>2</sub> fuel, it is unlikely that H<sub>2</sub> fuel use under the RoadTrans H<sub>2</sub> scenario would lead to a considerable direct effect on OH. Because OH controls the atmospheric lifetime of GHGs and pollutants through oxidation, it is therefore unlikely that the lifetimes of these species would change to an appreciable extent. However, global-scale replacement of the ICE with the H<sub>2</sub> fuel cell may significantly decrease NO<sub>x</sub> emissions [41]. This is because  $NO_x$  is generated from the nitrogen and oxygen present in ambient air during fuel combustion due to the high temperatures reached. Decreased NO<sub>x</sub> emissions would have an indirect effect of reducing OH, which in turn would reduce OH's global oxidizing capacity of trace gases. Nevertheless, a reduction in NO<sub>x</sub> emissions would improve human and environmental health through reduction of tropospheric  $O_3$  formation, since tropospheric  $O_3$  is a powerful air pollutant.



Fig. 5 – Hydrogen emissions (Mton H<sub>2</sub>) from road transportation. "H<sub>2</sub>%" refers to the specific hydrogen leakage rate applied from hydrogen fuel use in the RoadTrans H<sub>2</sub> scenarios.

# CH4 leakage limit in natural gas production

In this section, the upper limit of CH<sub>4</sub> leakage in natural gas production is determined as explained in Section Scenarios. It was found that with 100% conventional natural gas supply, the upper limit of CH<sub>4</sub> leakage is 7.1% combined for upstream and downstream CH<sub>4</sub> emissions. With 100% unconventional natural gas supply, the CH<sub>4</sub> leakage limit is 7.0% combined for upstream and downstream CH4 emissions. The limit is slightly higher for 100% conventional natural gas supply because CO<sub>2</sub> emissions from conventional natural gas production are less than from unconventional natural gas production. While the EPA 1996, Howarth 2011 and Hultman 2011 natural gas production leakage rates are well under the CH4 leakage limits presented here, higher natural gas leakage rates have been reported in the literature [20,25,27-30,33]. In fact, the upper-end CH<sub>4</sub> leakage rate for unconventional natural gas production for upstream and downstream combined, from Howarth et al. 2011 [22], is as high as 7.9%. It is also noteworthy that these EFs were measured for natural gas production in the US, and EFs may be higher in countries with less stringent regulations, perhaps substantially so. Ensuring that the globally averaged CH<sub>4</sub> leakage rate from natural gas production is below the CH<sub>4</sub> leakage limits presented here is decisive in the MC-H<sub>2</sub> economy providing climate benefits.

# Required methane cracking efficiency

In this section, the minimum required methane cracking efficiency above which the  $MC-H_2$  economy provides net climate benefits is determined as explained in Section Scenarios. The results are displayed below in Table 9. The analysis reveals that the minimum required methane cracking efficiency strongly varies based on the  $CH_4$  leakage rate from natural gas production. Most notably, the minimum required methane cracking efficiency for each  $CH_4$  EF used here, and for both 100% conventional and 100% unconventional natural gas supply, are all well under the 55% energy efficiency mark postulated in the literature as an achievable value on the commercial scale. Of course, with higher  $CH_4$  leakage rates from natural gas production, an even higher methane cracking efficiency would be required than the ones

Table 9 – Minimum required efficiency of methane cracking so that  $CO_2$ -eq emissions from the  $H_2Econ$ scenario are equal to those from the baseline scenario. Emissions are calculated for 100% conventional natural gas supply and for 100% unconventional natural supply.

| Data source <sup>a</sup>       | Required methane cracking<br>efficiency <sup>b</sup> |                            |
|--------------------------------|--|----------------------------|
|                                | Conventional natural gas                             | Unconventional natural gas |
| EPA 1996 [23]                  | 11%  | 12%                        |
| Hultman 2011 <mark>[38]</mark> | 20%  | 32%                        |
| Howarth 2011 [22]              | 33%  | 47%                        |

<sup>a</sup> Source of EF data set for CH<sub>4</sub> leakage in natural gas production.
<sup>b</sup> Efficiency from providing energy for the reaction to proceed, and pressurizing the reaction vessel.

displayed in Table 9. Nevertheless, this result is promising for the MC-H<sub>2</sub> economy, provided that  $CH_4$  leakage rates from natural gas production do not greatly exceed those of Howarth 2011.

# **Conclusion and outlook**

The results presented here support the proposition that a fossil-fuel-enabled bridge to a fully renewable-based H<sub>2</sub> economy can bring benefits to the climate through reduction of CO<sub>2</sub>-eq emissions. However, the impact of the MC-H<sub>2</sub> economy on emissions is highly dependent on the factors facilitating it. Optimistic assumptions, including the production of H<sub>2</sub> by methane cracking, relatively low CH<sub>4</sub> leakage rates from natural gas production, and a high tank-to-wheel efficiency of H<sub>2</sub> by way of a H<sub>2</sub> fuel cell enable climate benefits through emissions reduction. On the other hand, pessimistic assumptions such as the production of  $H_2$  by conventional, fossil-based, high-emission technologies like methane steam reforming, relatively high CH<sub>4</sub> leakage rates from natural gas production, and a high tank-to-wheel efficiency of  $H_2$  through use of a  $H_2$  internal combustion engine, result in an unfavorable climate impact through considerable emissions increase.

In order to achieve net  $CO_2$ -eq emissions decrease with the MC-H<sub>2</sub> economy, it is important that the globally-averaged CH<sub>4</sub> leakage rates from natural gas production are below 7%, and even lower for more substantial emissions reductions to be realized (see Section CH<sub>4</sub> leakage limit in natural gas production, Industrial H<sub>2</sub> and RoadTrans H<sub>2</sub>). However, higher CH<sub>4</sub> leakage rates have been reported in the literature [20,25,27–30,33]. Nevertheless, some of these very high CH<sub>4</sub> leakage rates were observed in areas where high CH<sub>4</sub> fluxes were expected, and are not necessarily representative of typical CH<sub>4</sub> leakage rates on a large scale [20]. In any case, due to the high degree of uncertainty surrounding the CH<sub>4</sub> leakage rates, further research is required in order to form more robust and consistent CH<sub>4</sub> EF estimates for natural gas production.

Methane cracking and the  $H_2$  fuel cell are likewise needed to achieve net  $CO_2$ -eq emissions decrease with the MC- $H_2$ economy. Both of these technologies require further research and development in order to become realized on the global scale. Additionally, it is important to determine the energy conversion efficiency of methane cracking as well as the tankto-wheel efficiency of a commercialized hydrogen fuel cell that could be realistically achieved, so as to determine the impact of the MC- $H_2$  economy on emissions. It would also be interesting to explore the effect of centralized  $H_2$  production on the MC- $H_2$  economy because this would avoid downstream  $CH_4$  emissions, which in turn may lead to lower  $CH_4$ emissions.

Based on the sensitivity results, more study is needed to better understand  $CH_4$  leakage from oil production (see Section Model evaluation). Due to the high uncertainty in  $CH_4$ emissions from oil production, the potential impact of this sector on global  $CO_2$ -eq emissions is considerable. It is important that the uncertainty in  $CH_4$  leakage from natural gas production does not overshadow the considerable

uncertainty of the  $CH_4$  leakage rate from oil production. More research is required to understand oil's true impact on the climate, which may be far worse than what is currently perceived.

If the MC-H<sub>2</sub> economy is implemented, caution is advised to prevent technological lock-in into a natural gas fossil fuel economy, and to prevent a delay in the shift of the energy system to renewables. Instead, the MC-H<sub>2</sub> economy can potentially serve as a bridge to a renewable H<sub>2</sub> economy, in which it facilitates development of the H<sub>2</sub> economy infrastructure, if in the process it can provide climate benefits through reduced  $CO_2$ -eq emissions.

An important aspect missing from this work is the cost competitiveness of the MC-H<sub>2</sub> economy. While the MC-H<sub>2</sub> economy may have the ability to significantly mitigate emissions, it would never be realized if it is not economically feasible. In this context it would be interesting to consider the economic potential of the elemental carbon produced by the methane cracking process. Therefore, more research is needed to explore the financial considerations of MC-H<sub>2</sub>.

Finally, while the results presented here support the MC-H<sub>2</sub> economy's potential in benefitting the climate through reduction of CO<sub>2</sub>-eq emissions, there are other aspects not considered in this work that may have important environmental consequences. For instance, increased shale gas production without effective environmental regulations on a global scale may negatively impact human and environmental health by degrading air quality and contaminating drinking water. It is important that future research into the MC-H<sub>2</sub> economy consider comprehensive potential environmental consequences of increased shale gas production, lest certain environmental goals be achieved while others are sacrificed.

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