ELSEVIER

Contents lists available at ScienceDirect

Journal of Environmental Radioactivity

journal homepage: www.elsevier.com/locate/jenvrad



Depth profile of ²³⁶U/²³⁸U in soil samples in La Palma, Canary Islands

M. Srncik a,*, P. Steier b, G. Wallner a

ARTICLEINFO

Article history: Received 15 November 2010 Received in revised form 27 January 2011 Accepted 18 March 2011 Available online 9 April 2011

Keywords: Soil samples Global fallout ²³⁶U/²³⁸U ²⁴⁰Pu/²³⁹Pu ²³⁶U/²³⁹Pu

ABSTRACT

The vertical distribution of the $^{236}\text{U}/^{238}\text{U}$ isotopic ratio was investigated in soil samples from three different locations on La Palma (one of the seven Canary Islands, Spain). Additionally the $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratio, as it is a well establish tool for the source identification, was determined. The radiochemical procedure consisted of a U separation step by extraction chromatography using UTEVA® Resin (Eichrom Technologies, Inc.). Afterwards Pu was separated from Th and Np by anion exchange using Dowex 1x2 (Dow Chemical Co.). Furthermore a new chemical procedure with tandem columns to separate Pu and U from the matrix was tested. For the determination of the uranium and plutonium isotopes by alpha spectrometry thin sources were prepared by microprecipitation techniques. Additionally these fractions separated from the soil samples were measured by Accelerator Mass Spectrometry (AMS) to get information on the isotopic ratios $^{236}\text{U}/^{238}\text{U}$, $^{240}\text{Pu}/^{239}\text{Pu}$ and $^{236}\text{U}/^{239}\text{Pu}$, respectively. The ^{236}U concentrations [atoms/g] in each surface layer (~ 2 cm) were surprisingly high compared to deeper layers where values around two orders of magnitude smaller were found. Since the isotopic ratio $^{240}\text{Pu}/^{239}\text{Pu}$ indicated a global fallout signature we assume the same origin as the probable source for ^{236}U . Our measured $^{236}\text{U}/^{239}\text{Pu}$ value of around 0.2 is within the expected range for this contamination source.

© 2011 Elsevier Ltd. Open access under CC BY-NC-ND license.

1. Introduction

The major source of actinides (as e.g. plutonium) in the environment is global fallout which arose from about 541 atmospheric nuclear explosions (UNSCEAR Report, 1993). Other sources are nuclear reactor accidents (Chernobyl), discharges of radioactive waste (Sellafield, La Hague, Mayak) and accidents with nuclear devices (e.g. Palomares, Thule) (Salbu, 2001). Soil samples from three different locations on La Palma, Canary Islands, (see Fig. 1) were collected in November 2007 and in February 2009. The emphasis on this work was the vertical distribution of $^{236}\mathrm{U}/^{238}\mathrm{U}$ in the soil and the measurement of Pu. In addition the determination of the $^{236}\mathrm{U}/^{239}\mathrm{Pu}$ isotopic ratio in these samples was carried out.

 236 U with a half life of 2.3×10^7 years is continuously produced by thermal neutron capture on 235 U which is omnipresent in the environment. The natural 236 U production results from neutrons produced by (α, n) -reactions on lighter nuclides (e.g. Na and Mg), spontaneous fission of 238 U, induced fission of 235 U and at the earth's surface from cosmic rays (Wilcken et al., 2008). A 236 U/ 238 U isotopic ratio between 1×10^{-14} and 5×10^{-14} can be expected for typical crustal rocks with a uranium and thorium content of a few

ppm (Steier et al., 2008). On the other hand, significant releases to the environment can be expected as 236 U is a by-product in nuclear power plants. This potentially enhances the very low natural isotopic ratio, but the environmental distribution of anthropogenic ²³⁶U is not well investigated yet. The present work is part of a larger effort to assess the dispersion of anthropogenic ²³⁶U in different compartments of the environment. The summit of La Palma Island is chosen since it allows relatively easy access to a site with clean tropospheric air, far away from anthropogenic emissions. It is thus well suited for studying fallout at minimum risk of local contamination. The climate of the Canary Islands (Herrera et al., 2001) is dominated by the direct influence of the trade wind belt, which blows mainly against the northern of the islands. La Palma (2312 m a.s.l.) is the second highest island of the Canary archipelago, and the most humid (~ 1000 mm/year at the sampling sites). The exceptional steep topography of La Palma leads to condensation directly from high-altitude clean air. Since the deposition of global fallout is controlled by precipitation, a sufficient amount of radionuclides is expected. On the other hand, the high mountain ranges of La Palma are easily accessible on good roads, since they host several European astronomical observatories.

Measurements of ²³⁶U became possible only recently by AMS on very few facilities, as e.g. VERA (Vienna Environmental Research Accelerator); plutonium isotopes, on the contrary, are comparably well studied as tracer for anthropogenic actinides with alpha

^a Department of Inorganic Chemistry, University of Vienna, Währinger Straße 42, A-1090 Vienna, Austria

^b VERA Institute, Faculty of Physics – Isotope Research, University of Vienna, Währinger Straße 17, A-1090 Vienna, Austria

^{*} Corresponding author. Tel.: +43 1 4277 52624; fax: +43 1 4277 52620. E-mail address: michaela.srncik@univie.ac.at (M. Srncik).

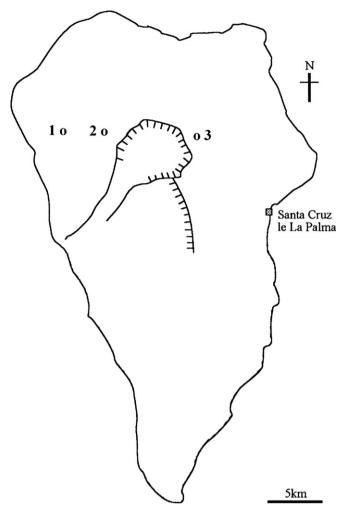


Fig. 1. Modified map from La Palma (Moss et al., 1999). The numbers with the respective cycles indicate the sampling sites (1 represents IIC, IID and LP_B1, LP_B2, number 2 stands for IA, IB and 3 for LP_A).

spectrometry and ICP-MS (Ketterer and Szechenyi, 2008) but also by AMS (Oughton et al., 2001). Although the environmental behaviour of U and Pu differs significantly, we consider plutonium as the best candidate for a methodical comparison. Also for Pu, AMS seems to have the best overall sensitivity (Fifield, 2008) but it can also be measured by other mass spectrometric techniques with partially higher detection efficiency (Steier et al., 2010). Recently soil samples from a site which was solely influenced by global fallout could be investigated for their 236 U concentration. The amount ranged from 10^8-10^9 atoms/g dry soil (Sakaguchi et al., 2009). A constant 236 U/ 239 Pu ratio of \sim 0.2 was found, which was interpreted as the original fingerprint of global fallout. In the general environment, however, differences in the mobility of Pu and U will lead to deviations from this ratio.

In this paper the combined procedure for the determination of U and Pu in soil samples is presented. Two column chromatography methods were applied: at the beginning the sample solution was passed through a column filled with UTEVA® Resin (Eichrom Technologies, Inc.) to separate U, and then the Pu fraction was purified with a second anion exchange step by using Dowex 1x2 (100–200 mesh, Dow Chemical Co.). The second method consisted of a double column system. The first column was filled with Dowex 1x8 and the second one with UTEVA® Resin. The UTEVA column was placed directly below the anion exchange column.

The advantage of this method is that no further purification (i.e. additional column chromatography steps) of the Pu fraction was necessary.

We present first 236 U/ 238 U ratios from a clean air area as well as 240 Pu/ 239 Pu and 236 U/ 239 Pu atomic ratios, from which the contamination source (namely global fallout) can be derived.

2. Materials and methods

The samples were collected at three different sites on the island La Palma in the Atlantic Ocean. Despite samples were taken close to roads or trails, we looked out for an intact and site-typical vegetation cover as indication for the absence of anthropogenic disturbance. From the first sampling location (N 28° 45.233′/W 17° 53,421', 2330 m a.s.l.) on the northern slope of Roque de los Muchachos two cores (IA and IB) were available which were separated into 2 cm layers and stored in plastic bags. The full length was 10 cm and 12 cm for IA and IB, respectively. At the second site several 100 m downhill (N 28° 46.191'/W 17° 54.224', 1970 m a.s.l.) again two cores with 8 cm length were collected (IIC and IID). IIC was separated into 2 cm layers whereas IID was taken as a bulk sample. Two years later two additional bulk samples (LP_B1 and LP_B2) were collected from the same site as IIC and IID: the depth of LP_B1 ranged from 0 to 5 cm. LP_B2 was sampled from an outcrop apparently exposed by road construction between 2007 and 2009. Approximately 10 cm of the vertical surface of the outcrop were removed, and then the material was taken from 15 to 40 cm below the horizontal surface. Additionally a third bulk sample (LP A) was collected on Pico de La Nieve (N 28° 43.717'/W 17° 49.636'. ~2000 m a.s.l.) with a depth down to 5 cm. In the time period between the two collection dates the sample procedure was improved. While the samples picked up in 2007 were prepared according to the chemical separation procedure A, the samples taken in 2009 were processed following the faster and more comfortable sample preparation B. The main difference of the second procedure is the usage of UTEVA® Resin coupled with Dowex 1x8 but few other changes are in the detail and therefore both sample treatments are described separately below. For spiking a stock solution of ²³⁶Pu was obtained from the University of Helsinki and calibrated against SRM 4334G (242Pu solution) by alpha spectrometry. Furthermore a stock solution of SRM 4324B (²³²U solution) was used. One reference sample (IAEA-135, Radionuclides in Irish Sea Sediment) was investigated with respect to Pu and U in order to check the validity of our results.

2.1. Chemical separation procedure A

After drying the samples by air, they were ashed in porcelain crucibles in an electric muffle furnace at 420 °C for 24 h. The radiochemical procedure applied to the analysis of U and Pu by alpha spectrometry followed the procedures given in (Eichrom Technologies, 2005; Hrnecek et al., 2002) with slight modifications which are discussed briefly below. Between 5 and 20 g soil were leached in 100 mL 65% HNO₃ and 25 mL 37% HCl, 0.5 g NaNO₂, $^{236}\mbox{Pu}$ and $^{232}\mbox{U}$ tracer solutions were added. After boiling the mixture for 3 h, the sample was centrifuged for 30 min at 4000 rpm (relative centrifugal force (RCF) is 1646). The soil residue was leached twice with 20 mL 65% HNO₃ and 15 mL 48% HF for 1 h and was then rejected. The supernatant solutions were combined and evaporated to near dryness. The residue was fumed twice with 10 mL boric acid (c = 5 g/100 mL), three times with 10 mL 65% HNO₃ and 4 mL 32% H₂O₂. The residue was taken up in 20 mL 3 M $HNO_3 - 1$ M $Al(NO_3)_3$ and insoluble particles were removed by centrifugation.

2.1.1. Separation of uranium

0.5 g UTEVA $^{\$}$ resin was pre-conditioned with 3 M HNO₃, the sample solution was brought onto the column which was then washed with 35 mL 3 M HNO₃. Pu was eluted with 10 mL 9 M HCl and 30 mL 5 M HCl - 0.05 M oxalic acid, and finally U was eluted with 30 mL 1 M HCl.

2.1.2. Purification of plutonium

The solution containing Pu was taken to dryness and fumed three times with 5 mL 65% HNO $_3$ and 2 mL 32% H $_2$ O $_2$. The residue was taken up with 20 mL 1 M HNO $_3$ and 100 mg Mohr's salt [(NH $_4$) $_2$ Fe(SO $_4$)x6H $_2$ O] were added to reduce Pu to trivalent state. Afterwards the solution was adjusted to 8 M HNO $_3$ by adding 20 mL 65% HNO $_3$. Finally Pu was oxidized to tetravalent oxidation state by addition of 0.5 g NaNO $_2$. The excess of nitrous acid was destroyed by gently boiling (NO $_3$ fumes). 6 g Dowex 1x2 were pre-conditioned with 8 M HNO $_3$ and then the sample solution was loaded onto the column which was subsequently washed with 50 mL 8 M HNO $_3$. Th was removed with 30 mL 37% HCl and Pu was eluted with 50 mL 0.36 M HCl - 0.014 M HF.

The U and Pu fractions were evaporated until dryness and fumed three times with 5 mL 65% HNO $_3$ and 2 mL 32% H $_2$ O $_2$. The U fraction was additionally fumed with 37% HCl (3x). Afterwards the Pu was dissolved in 20 mL 1 M HNO $_3$ and U in 20 mL 1 M HCl to carry out the microprecipitation (description see Section 2.3) for the alpha source preparation.

2.2. Chemical separation procedure B

The procedure described below (modified from Eichrom Technologies, 2005; Moreno et al., 1997; Warneke et al., 2002) was used for the bulk samples collected in 2009. Samples of 20 g were dried in an electric oven at 100 °C to constant weight and afterwards ashed in porcelain crucibles in an electric muffle furnace at 420 °C for 24 h. The sample was transferred to a Teflon beaker and was leached according to 2.1 "Chemical separation procedure A". Afterwards the combined solutions were evaporated till dryness, fumed three times with 10 mL boric acid (c = 5 g/100 mL), with 10 mL 37% HCl (3x) and with 10 mL 65% HNO₃ (3x). The residue was dissolved in 8 M HNO3 and a redox adjustment of Pu was carried out as described in Section 2.1. The insoluble part was removed by centrifugation (5 min at 4000 rpm). Two columns were prepared; the first one was filled with 6 g Dowex 1x8 (100-200 mesh, Dow Chemical Co.) and the second column with 0.5 g UTEVA® (Eichrom Technologies, Inc.). The UTEVA® column was placed directly under the anion exchange column and both were pre-conditioned with 8 M HNO₃. The sample solution was transferred to the Dowex 1x8 column and the eluent passed directly onto the UTEVA® column. After washing with 60 mL 8 M HNO₃ the columns were separated. The Dowex 1x8 column was washed with 50 mL 37% HCl to remove Th, and Pu was finally eluted with 50 mL $0.1 \text{ M NH}_4\text{I} - 9 \text{ M HCl}$. The iodide reduced Pu(IV) to Pu(III) which does not form chloride complexes and was no longer retained by the column. The UTEVA® resin was washed with 40 mL 3 M HNO₃, traces of Th and Np were eluted with 10 mL 9 M HCl and 30 mL 5 M HCl - 0.05 M oxalic acid. Finally U was eluted with 30 mL 1 M HCl. The U and Pu fractions were treated according the last paragraph in Section 2.1.2.

2.3. Microprecipitation

To the plutonium fraction a tip of a spatula of Mohr's salt, 0.5 mL of a 25% NaNO2 solution and 50 μL of a Nd $^{3+}$ -solution (c=1 mg/mL) were added.

To the U fraction 50 μ L of a Nd³⁺-solution (c=1 mg/mL) were added and U was reduced to tetravalent oxidation state by addition of 100 μ L 15% TiCl₃-solution. Afterwards 5 mL 40% HF were added to both of the solutions. After 1 h the samples were filtered through a cellulose nitrate membrane filter (Whatman®, 0.1 μ m pore size), washed three times with 2 mL 4% HF and twice with 2 mL Millipore water (18.2 M Ω cm).

2.4. Alpha measurement

Alpha spectrometry was performed with a PIPS (Passivated Implanted Planar Silicon) Detector, Model 7401 VR, Canberra/Packard with an active area of 450 mm². The counting time was 252 000 s and 604 800 s for U and Pu, respectively. For these measurements, the detection limits calculated according to Currie (1968) were 1.12 mBq/sample for ²³⁸U and 0.80 mBq/sample for ²³⁹⁽⁴⁰⁾Pu. The software Genie 2.1 (Canberra, USA) was used for the evaluation of the spectra.

2.5. AMS measurement

The filters already analysed by alpha spectrometry were reprocessed for the AMS measurement according to Srncik et al. (2008). The filters were dissolved in HNO₃, the nuclides of interest were co-precipitated with iron hydroxide, and combusted to oxides which are pressed into an aluminium sample holder either with or without silver as binder. The VERA setup as used for actinide measurements is described in Vockenhuber et al. (2003) and Steier et al. (2010).

PuO⁻ at 50 keV was produced in a Cs-sputter source, mass analysed, accelerated to 3 MeV and stripped to positive charge state in a gas cell. By these means, all interfering molecular isobars are destroyed. The 5+ charge state is selected in a second, high energy mass spectrometer, after acceleration to 18 MeV. A combination of an ionization chamber and a time-of-flight measurement allow a unique identification of the isotope of interest. Significant background exists for ²⁴⁰Pu⁵⁺, apparently from stable ¹⁴⁴Nd³⁺, which shares the same mass-to-charge ratio and thus cannot be suppressed by the mass spectrometer. In the present measurement, the residual Nd concentration in the sputter samples was, however, low enough and could therefore be separated in the energy-resolving detector.

The machine was tuned with $^{238}U^{16}O^{-}$ as a pilot beam, and the parameters were than scaled to the different rare radioisotopes which do not provide sufficient beam current for tuning. Alternating measurements of ²³⁹Pu and ²⁴⁰Pu were performed, with duration of several 100 s each. By these means, one tries to minimize the impact of source output variations on the measured count rate results. For 236 U, 238 U 16 O $^{-}$ can be measured once per second as a current before the accelerator, which renders source output variations negligible. Our in-house standard Vienna-KkU, 236 U/U = $(6.98 \pm 0.32) \times 10^{-11}$ (Steier et al., 2008) was used to normalize the ²³⁶U measurements. However, no standard is available for Pu; this introduces an additional uncertainty, which is however expected to be in the range of a few percent. Measurement times per sample ranged from 30 min to several hours, depending mainly on the machine time availability. Generally, the prepared samples last for several hours, and an efficiency of $\sim 10^{-4}$ (counts in the detector per atom in sputter sample) is achieved if the samples are completely sputtered.

3. Results and discussion

The chemical separation A has been already applied in our laboratory but for sample preparation B the validity was checked by analysing an IAEA-135 reference material. The obtained result was

 (225 ± 7) Bq/kg $^{239(40)}$ Pu; this value is in good agreement with the recommended value of (205–226) Bq/kg for $^{239(40)}$ Pu. The IAEA-135 standard was also measured for its 236 U/ 238 U ratio by AMS. Our value of $(1.48\pm0.04)\times10^{-6}$ agrees with Hotchkis et al. (2000) who determined $(1.48\pm0.37)\times10^{-6}$ via AMS. Our result indicated a very good accuracy of the proposed method.

In Table 1 the results of the $^{236}\text{U}/^{238}\text{U}$ isotopic ratios and ^{236}U [atoms/g] for all core layers and also for the bulk samples are shown. In each core (IA, IB and IID) the main amount of ²³⁶U was found in the surface layer (between 10⁹ and 10¹⁰ atoms/g dry soil). With increasing depth the number of atoms per gram measured was in the range of 10⁷, except for the depth profile of IA. Here again an increase of two orders of magnitude was obtained for the deepest layer. As this value is so high and therefore implausible, we assume that a mistake during the sample collection occurred. According to the results that a relatively high value of ²³⁶U atoms/g in each surface layer was found in contrast to the deeper layers (decrease by two orders of magnitude), it seems that the anthropogenic ²³⁶U was not moving downwards. The layers below the "surface layer" show at all sites (IA, IB and IID) at least the same order of magnitude of ²³⁶U atoms/g. The result of the respective sub-samples of LP B1 and LP B2 are consistent with the exception of LP_B2_2 which was below the blank value. The result of sample LP_A_2 is higher but agrees with LP_A_1 and LP_A_3 within 2σ. Both bulk samples LP_B1 (0-5 cm) and LP_B2 (15-40 cm) show a ²³⁶U concentration in the range of 10⁸ atoms/g dry soil. For the LP A bulk sample (0-5 cm depth) the results are one order of magnitude higher which reflects inhomogeneities in the ²³⁶U distribution at different sample sites.

Table 1 Results of the isotopic ratio 236 U/ 238 U and 236 U [atoms/g] in soil samples from La Palma. The measurement uncertainties are given in \pm 1 σ .

²³⁶U/²³⁸U

Core, depth [cm]

Depth profile – IA

²³⁶U [atoms/g]

	le – IA			
LP01a	IA, 2-6	$(2.37 \pm 0.07) \times 10^{-7}$	$(9.69 \pm 1.62) \times 10^9$	
LP02a	IA, 6-8	$(2.96 \pm 0.20) \times 10^{-9}$	$(6.08 \pm 0.73) \times 10^7$	
LP03	IA, 8-10	$(1.23 \pm 0.04) \times 10^{-7}$	$(1.93 \pm 0.31) \times 10^9$	
	,	(, .	,	
Depth profil		7	0	
LP04	IB, 0–2	$(3.15 \pm 0.09) \times 10^{-7}$	$(8.76 \pm 1.62) \times 10^9$	
LP05a	IB, 2–4	$(3.01 \pm 0.09) \times 10^{-9}$	$(3.91 \pm 0.56) \times 10^{7}$	
LP06	IB, 4–6	$(3.90 \pm 0.21) \times 10^{-9}$	$(8.12 \pm 1.24) \times 10^{7}$	
LP07	IB, 6–8	$(2.10 \pm 0.23) \times 10^{-9}$	$(4.95 \pm 1.12) \times 10^{7}$	
LP08a	IB, 8–10	$(1.00 \pm 0.11) \times 10^{-9}$	$(1.43 \pm 0.33) \times 10^{7}$	
LP09	IB, 10-12	$(2.13 \pm 0.16) \times 10^{-9}$	$(5.46 \pm 0.89) \times 10^7$	
Depth profil	le — IID			
LP10	IID, 0–2	$(2.54\pm0.08){\times}10^{-7}$	$(2.82 \pm 0.48) \times 10^{9}$	
LP11	IID, 2–4	$(4.93 \pm 0.34) \times 10^{-9}$	$(2.61 \pm 0.48) \times 10^7$	
LP12a	IID, 4–6	$(3.25 \pm 0.16) \times 10^{-9}$	$(5.64 \pm 0.96) \times 10^7$	
LP12a LP13	IID, 4–0 IID. 6–8	$(3.54 \pm 0.33) \times 10^{-9}$	$(3.85 \pm 0.49) \times 10^7$	
LP13 LP14	IID, 8–8 IID, 8–10	$(3.54 \pm 0.33) \times 10^{-9}$	$(2.83 \pm 0.49) \times 10^{7}$ $(4.41 \pm 0.96) \times 10^{7}$	
LF 14	110, 6–10	$(3.32 \pm 0.29) \times 10$	$(4.41 \pm 0.30) \times 10$	
IIC bulk sample				
iic buik san	ipie			
LP15a	IIC, 0–8	$(4.35\pm0.24){\times}10^{-9}$	$(5.75\pm 1.18){\times}10^{7}$	
		$(4.35 \pm 0.24) \times 10^{-9}$	$(5.75 \pm 1.18) \times 10^{7}$	
LP15a Sample	IIC, 0–8 Depth [cm]	²³⁶ U/ ²³⁸ U		
LP15a Sample	IIC, 0–8	²³⁶ U/ ²³⁸ U	²³⁶ U [atoms/g]	
LP15a Sample LP_B1 bulk	IIC, 0–8 Depth [cm] sample, again from IIC	²³⁶ U/ ²³⁸ U /IID		
LP15a Sample LP_B1 bulk: LP_B1_1	IIC, 0–8 Depth [cm] sample, again from IIC 0–5	236 U/ 238 U /IIID $^{(2.36 \pm 0.07) \times 10^{-8}}$	236 U [atoms/g] $(4.44 \pm 0.89) \times 10^8$	
LP15a Sample LP_B1 bulk : LP_B1_1 LP_B1_2 LP_B1_3	Depth [cm] sample, again from IIC 0-5 0-5 0-5	236 U/ 238 U /IID 236 ± 0.07)×10 ⁻⁸ (2.45 ± 0.07)×10 ⁻⁸ (2.48 ± 0.07)×10 ⁻⁸	236 U [atoms/g] $(4.44 \pm 0.89) \times 10^{8}$ $(2.50 \pm 0.41) \times 10^{8}$	
Sample LP_B1 bulk: LP_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk:	Depth [cm] sample, again from IIC 0-5 0-5 0-5 sample, again from IIC	236 U/ 238 U /IID (2.36 ± 0.07)×10 ⁻⁸ (2.45 ± 0.07)×10 ⁻⁸ (2.48 ± 0.07)×10 ⁻⁸	$ \begin{array}{c} ^{236}\text{U [atoms/g]} \\ \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \end{array} $	
LP15a Sample LP_B1 bulk: LP_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk: LP_B2_1	Depth [cm] sample, again from IIC 0-5 0-5 0-5 sample, again from IIC 15-40	236 U/ 238 U /IID 236 ± 0.07)×10 ⁻⁸ (2.45 ± 0.07)×10 ⁻⁸ (2.48 ± 0.07)×10 ⁻⁸	236 U [atoms/g] $(4.44 \pm 0.89) \times 10^{8}$ $(2.50 \pm 0.41) \times 10^{8}$	
LP15a Sample LP_B1 bulk: LP_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk: LP_B2_1 LP_B2_1	Depth [cm] sample, again from IIC 0-5 0-5 0-5 sample, again from IIC 15-40 15-40	$\begin{array}{c} ^{236}\text{U}/^{238}\text{U} \\ \text{/IID} \\ (2.36 \pm 0.07) \times 10^{-8} \\ (2.45 \pm 0.07) \times 10^{-8} \\ (2.48 \pm 0.07) \times 10^{-8} \\ \text{/IID} \\ (4.85 \pm 1.69) \times 10^{-9} \\ - \end{array}$	$ \begin{array}{c} ^{236}\text{U [atoms/g]} \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \\ (2.49 \pm 0.92) \times 10^{8} \\ - \end{array} $	
LP15a Sample LP_B1 bulk: LP_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk: LP_B2_1	Depth [cm] sample, again from IIC 0-5 0-5 0-5 sample, again from IIC 15-40	236 U/ 238 U /IID (2.36 ± 0.07)×10 ⁻⁸ (2.45 ± 0.07)×10 ⁻⁸ (2.48 ± 0.07)×10 ⁻⁸	$ \begin{array}{c} ^{236}\text{U [atoms/g]} \\ \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \end{array} $	
LP15a Sample LP_B1 bulk: LP_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk: LP_B2_1 LP_B2_1	Depth [cm] sample, again from IIC 0-5 0-5 0-5 15-40 15-40 15-40	$\begin{array}{c} ^{236}\text{U}/^{238}\text{U} \\ \text{/IID} \\ (2.36 \pm 0.07) \times 10^{-8} \\ (2.45 \pm 0.07) \times 10^{-8} \\ (2.48 \pm 0.07) \times 10^{-8} \\ \text{/IID} \\ (4.85 \pm 1.69) \times 10^{-9} \\ - \end{array}$	$ \begin{array}{c} ^{236}\text{U [atoms/g]} \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \\ (2.49 \pm 0.92) \times 10^{8} \\ - \end{array} $	
LP15a Sample LP_B1 bulk 1P_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk 1P_B2_1 LP_B2_2 LP_B2_3	Depth [cm] sample, again from IIC 0-5 0-5 0-5 15-40 15-40 15-40	236 U/ 238 U /IIID $^{(2.36 \pm 0.07) \times 10^{-8}}$ $^{(2.45 \pm 0.07) \times 10^{-8}}$ $^{(2.48 \pm 0.07) \times 10^{-8}}$ $^{(110)}$ $^{(4.85 \pm 1.69) \times 10^{-9}}$ $^{(3.25 \pm 1.54) \times 10^{-9}}$ $^{(4.21 \pm 0.13) \times 10^{-8}}$	$\begin{array}{c} ^{236}\text{U [atoms/g]} \\ \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \\ (2.49 \pm 0.92) \times 10^{8} \\ \\ \\ - \\ (1.78 \pm 0.88) \times 10^{8} \\ \\ \\ (4.01 \pm 0.48) \times 10^{9} \end{array}$	
LP15a Sample LP_B1 bulk LP_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk LP_B2_1 LP_B2_1 LP_B2_3 LP_A bulk sa	Depth [cm] sample, again from IIC 0-5 0-5 0-5 sample, again from IIC 15-40 15-40 15-40 ample	236 U/ 238 U /IID $^{(2.36 \pm 0.07) \times 10^{-8}}$ $^{(2.45 \pm 0.07) \times 10^{-8}}$ $^{(2.48 \pm 0.07) \times 10^{-8}}$ $^{(2.48 \pm 0.07) \times 10^{-8}}$ /IID $^{(4.85 \pm 1.69) \times 10^{-9}}$ $^{(3.25 \pm 1.54) \times 10^{-9}}$ $^{(4.21 \pm 0.13) \times 10^{-8}}$ $^{(3.41 \pm 0.35) \times 10^{-8}}$	$\begin{array}{c} ^{236}\text{U [atoms/g]} \\ \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \\ (2.49 \pm 0.92) \times 10^{8} \\ \\ \\ - \\ (1.78 \pm 0.88) \times 10^{8} \end{array}$	
LP15a LP_Bl bulk: LP_B1_1 LP_B1_2 LP_B1_3 LP_B2 bulk: LP_B2_1 LP_B2_2 LP_B2_3 LP_A2_1 LP_A2_1	Depth [cm] sample, again from IIC 0-5 0-5 0-5 sample, again from IIC 15-40 15-40 15-40 ample 0-5	236 U/ 238 U /IIID $^{(2.36 \pm 0.07) \times 10^{-8}}$ $^{(2.45 \pm 0.07) \times 10^{-8}}$ $^{(2.48 \pm 0.07) \times 10^{-8}}$ $^{(110)}$ $^{(4.85 \pm 1.69) \times 10^{-9}}$ $^{(3.25 \pm 1.54) \times 10^{-9}}$ $^{(4.21 \pm 0.13) \times 10^{-8}}$	$\begin{array}{c} ^{236}\text{U [atoms/g]} \\ \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \\ (2.49 \pm 0.92) \times 10^{8} \\ \\ \\ - \\ (1.78 \pm 0.88) \times 10^{8} \\ \\ \\ (4.01 \pm 0.48) \times 10^{9} \end{array}$	
LP15a Sample LP_B1 bulk: LP_B1_1 LP_B1_2 LP_B1_3 LP_B2_1 LP_B2_1 LP_B2_3 LP_B2_3 LP_A2 bulk st. LP_B2_3 LP_A2 bulk st. LP_A2	Depth [cm] sample, again from IIC 0-5 0-5 0-5 sample, again from IIC 15-40 15-40 15-40 ample 0-5 0-5 0-5	236 U/ 238 U /IID $^{(2.36 \pm 0.07) \times 10^{-8}}$ $^{(2.45 \pm 0.07) \times 10^{-8}}$ $^{(2.48 \pm 0.07) \times 10^{-8}}$ $^{(2.48 \pm 0.07) \times 10^{-8}}$ /IID $^{(4.85 \pm 1.69) \times 10^{-9}}$ $^{(3.25 \pm 1.54) \times 10^{-9}}$ $^{(4.21 \pm 0.13) \times 10^{-8}}$ $^{(3.41 \pm 0.35) \times 10^{-8}}$	$\begin{array}{c} ^{236}\text{U [atoms/g]} \\ \\ (4.44 \pm 0.89) \times 10^{8} \\ (2.50 \pm 0.41) \times 10^{8} \\ (3.22 \pm 0.64) \times 10^{8} \\ \\ (2.49 \pm 0.92) \times 10^{8} \\ \\ - \\ (1.78 \pm 0.88) \times 10^{8} \\ \\ (4.01 \pm 0.48) \times 10^{9} \\ (2.20 \pm 0.43) \times 10^{9} \end{array}$	

The isotopic ratios 236 U/ 238 U ranged from 10^{-7} near the surface to 10^{-9} with increasing depth. The obtained ratios of 10^{-9} are not influenced by our blanks; this was proven by a blank correction. At the sample location LP_B1, (0-5 cm) the level of few parts in 10^{-8} was found and at LP_B2 (15-40 cm) the isotopic ratio is one order of magnitude lower. The value for LP_A is slightly higher than the bulk samples at the LP_B site.

Table 2 shows the vertical distribution of $^{239(40)}\text{Pu}$ [Bq/kg] for each of the three sample locations (cores IA, IB and IID). At the sample site IA the surface layer (0–2 cm) got lost. The specific activities of $^{239(40)}\text{Pu}$ in layer LP01a (2–6 cm) and LP02a (6–8 cm) at location IA are similar [(0.32 \pm 0.05) Bq/kg and (0.35 \pm 0.06) Bq/kg, respectively]; in the deepest layer a decrease of one order of magnitude occurred. The highest specific activity at location IB of (0.46 \pm 0.03) Bq/kg was found in the surface layer (0–2 cm). The values for LP05a (2–4 cm) and LP09 (10–12 cm) were below the limit of detection (<0.040 Bq/kg). At the sample location IID the highest value occurred in the layer between 2–4 cm and the decrease of the activity is less marked than at the other two locations. Also here two samples, LP12a (4–6 cm) and LP14 (8–10 cm), were below the detection limit (<0.080 Bq/kg).

The accumulated levels at site IB and IID were (36 \pm 11) Bq $^{239(40)}$ Pu m $^{-2}$ and (22 \pm 7) Bq $^{239(40)}$ Pu m $^{-2}$, respectively. These $^{239(40)}$ Pu results are in the range of the expected value [(36 \pm 3) Bq $^{239(40)}$ Pu m $^{-2}$] for global fallout in the Northern Hemisphere at latitude 20°–30° N (Hardy et al., 1973).

Buesseler (1997) reports for the global fallout 240 Pu/ 239 Pu isotopic ratio a typical interval of 0.175–0.19. According to Kelley et al. (1999) the average composition of fallout 240 Pu/ 239 Pu for the north equatorial region (latitude, 30°N-0) is 0.178 with a standard deviation of 0.019 (2 σ). The ratio for Chernobyl is 0.45–0.52 (Kutkov et al., 1995).

In the present work, for some samples the 240 Pu/ 239 Pu ratio could not be determined (indicated by "–" in Table 3) due to relatively small 239 Pu and 240 Pu count rates. One reason was the low Pu chemical yield (the average was 24.3 ± 0.7) determined by alpha spectrometry which is attributed to difficulties during the leaching and fuming steps before the column separation step. Additionally, sometimes problems during the Fe(OH) $_3$ co-precipitation occurred which would further explain losses of Pu and therefore the low count rate by AMS. From Table 3 it can be seen that at the sample location IA only the first layer could by measured and is in agreement with global fallout, although the value was slightly higher.

Table 2 Vertical distribution of $^{239(40)}$ Pu [Bq/kg] at the sample location IA, IB and IID and the measurement uncertainty is given $\pm 1\sigma$.

Sample	Depth [cm]	²³⁹⁽⁴⁰⁾ Pu [Bq/kg]
Depth profile – IA		
LP01a	IA, 2–6	0.322 ± 0.045
LP02a	IA, 6–8	0.350 ± 0.055
LP03	IA, 8–10	0.026 ± 0.007
Depth profile – IB		
LP04	IB, 0-2	0.458 ± 0.028
LP05a	IB, 2-4	_
LP06	IB, 4–6	0.088 ± 0.010
LP07	IB, 6–8	0.055 ± 0.011
LP08a	IB, 8-10	0.037 ± 0.013
LP09	IB, 10-12	_
Depth profile – IID		
LP10	IID, 0-2	0.296 ± 0.065
LP11	IID, 2-4	0.601 ± 0.100
LP12a	IID, 4–6	_
LP13	IID, 6–8	0.136 ± 0.025
LP14	IID, 8–10	_

The 240 Pu count rates of both LP02a (6–8 cm) and LP03 (8–10 cm) were too low to determine the 240 Pu/ 239 Pu isotopic ratio. The 240 Pu/ 239 Pu results in all layers of location IB and IID showed a ratio indicating global fallout. The average isotopic ratio of bulk sample LP_B1 was 0.221 \pm 0.029. For all three sub-samples of LP_B2 (15–40 cm below the surface) 239 Pu was far below the blank value and therefore no atomic ratio could be determined. The mean 240 Pu/ 239 Pu ratio of bulk sample LP_A was 0.287 \pm 0.025 which is slightly higher than the expected value for global fallout.

Table 4 shows the ²³⁶U/²³⁹Pu results for all layers and bulk samples, respectively. According to the isotopic ratios of 240 Pu/ 239 Pu found in the La Palma soil samples (Table 3), the source of contamination seems to be global fallout. Ketterer et al. (2007) reported ²³⁶U/²³⁹Pu atomic ratios between 0.05 and 0.5 for samples containing Pu of purely stratospheric fallout. The wide range was attributed to the evidently higher mobility of fallout ²³⁶U compared to Pu (Ketterer et al., 2007). However, this migration behaviour of U and Pu was not observed by Sakaguchi et al. (2009), who investigated at three sites 20 m apart from each other soil samples (0–30 cm) from Ishikawa Prefecture (Japan); here the ²³⁶U/²³⁹Pu isotopic ratios were in the much smaller range of 0.212-0.253. In our work (compare Table 1 and 2) the overwhelming part of ²³⁶U in the core samples was found in the surface layers, while the (unfortunately sparse) Pu data indicates a more even distribution. Consequently, the three topmost samples reveal an enhanced ²³⁶U/²³⁹Pu ratio higher than 3. All other values are in the range supposed by Ketterer et al. (2007) and for the bulk subsamples LP B1 (0–5 cm) a ²³⁶U/²³⁹Pu value of 0.2 similar to the bulk

Table 3 Results of 240 Pu/ 239 Pu by AMS in soil samples from La Palma. The measurement uncertainties are given in \pm 1σ .

Core, depth [cm]

Depth profile – IA

²⁴⁰Pu/²³⁹Pu

LP01a LP02a LP03	IA, 2–6 IA, 6–8 IA, 8–10	0.267 ± 0.056 $ -$
Depth profile — IB LP04 LP05a LP06 LP07 LP08a LP09	IB, 0-2 IB, 2-4 IB, 4-6 IB, 6-8 IB, 8-10 IB, 10-12	$\begin{array}{c} 0.149 \pm 0.006 \\ 0.178 \pm 0.015 \\ 0.198 \pm 0.050 \\ 0.149 \pm 0.051 \\ - \\ - \end{array}$
Depth profile — IID LP10 LP11 LP12a LP13 LP14	IID, 0-2 IID, 2-4 IID, 4-6 IID, 6-8 IID, 8-10	$\begin{array}{c} 0.150 \pm 0.018 \\ 0.195 \pm 0.089 \\ 0.189 \pm 0.035 \\ - \\ - \end{array}$
77.0.1 11 1		
IIC bulk sample LP15a	IIC, 0–8	0.131 ± 0.029
	IIC, 0–8 Depth [cm]	0.131 ± 0.029 $^{240} Pu/^{239} Pu$
LP15a	Depth [cm]	
LP15a Sample LP_B1 bulk sample, again f LP_B1_1 LP_B1_2	Depth [cm] rom IIC/IID 0-5 0-5 0-5	$\begin{array}{c} 240 \text{Pu} / ^{239} \text{Pu} \\ \\ 0.251 \pm 0.060 \\ 0.198 \pm 0.027 \end{array}$

samples of Sakaguchi et al. (2009) was found. We obtained for the bulk sample LP_A (0–5 cm) a higher average ratio (236 U/ 239 Pu = 0.5) but with a larger scatter between sub-samples.

4. Conclusion

Soil samples from a remote clean air site, the La Palma Island. were investigated with the emphasis on the $^{236}U/^{238}U$ ratio. A conventional and a new tandem column procedure were applied for the sequential and simultaneous separation of U and Pu. respectively. Samples were measured both by alpha spectrometry and AMS. The isotopic ratio of $^{236}\text{U}/^{238}\text{U}$ was between 10^{-7} for the surface samples and 10^{-9} for deeper layers. The 236 U [atoms/g] results at each surface layer were surprisingly high and the decrease to deeper layers were around two orders of magnitude; this phenomenon was not observed for Pu. The isotopic ratio 240 Pu/ 239 Pu is a well established indicator for the identification of the contamination source. In our samples the ²⁴⁰Pu/²³⁹Pu atom ratios of 0.149-0.267 for site IA, IB and IID indicate that global fallout is the probable source for the Pu isotopes similarly to site LP_B1 where the mean isotopic ratio is 0.221 \pm 0.029. In accordance with the ²⁴⁰Pu/²³⁹Pu results, most of our measured ²³⁶U/²³⁹Pu isotopic ratios were in the order of 0.04 and 0.78 and agree with the sparse set of already published values for the global fallout signature.

Table 4 Results of 236 U/ 239 Pu in La Palma soil samples. The measurement uncertainties are given in $+ 1\sigma$.

51ven in ± 10.		
Sample	Core, depth [cm]	²³⁶ U/ ²³⁹ Pu
Depth profile – IA		
LP01a	IA, 2-6	14 ± 3
LP02a	IA, 6-8	_
LP03	IA, 8–10	_
Depth profile – IB		
LP04	IB, 0-2	3.3 ± 0.5
LP05a	IB, 2-4	_
LP06	IB, 4-6	0.38 ± 0.07
LP07	IB, 6-8	0.27 ± 0.08
LP08a	IB, 8-10	_
LP09	IB, 10–12	_
Depth profile – IID		
LP10	IID, 0-2	4.6 ± 1.5
LP11	IID, 2–4	0.04 ± 0.01
LP12a	IID, 4–6	-
LP13	IID, 6–8	0.08 ± 0.02
LP14	IID, 8–10	- -
	,	
IIC bulk sample		
LP15a	IIC, 0–8	0.14 ± 0.03
Sample	Depth [cm]	²³⁶ U/ ²³⁹ Pu
LP_B1 bulk sample, ag	gain from IIC/IID	
LP_B1_1	0-5	0.23 ± 0.03
LP_B1_2	0-5	0.17 ± 0.02
LP_B1_3	0-5	0.21 ± 0.03
LP_B2 bulk sample, ag	gain from IIC/IID	
LP_B2_1	15-40	_
LP_B2_2	15-40	_
LP_B2_3	15-40	_
LP_A bulk sample		
LP_A_1	0-5	0.78 ± 0.14
LP_A_2	0-5	0.43 ± 0.08
LP_A_3	0-5	0.43 ± 0.08 0.23 ± 0.11
ri _i_ɔ	0-3	0.23 ± 0.11

Acknowledgement

Financial support by the FWF Austrian Science Fund (project number P21403-N19) is gratefully acknowledged.

References

- Buesseler, K.O., 1997. The isotopic signature of fallout plutonium in the North Pacific. J. Environ. Radioact. 36 (1), 69–83.
- Currie, L.A., 1968. Limits for qualitative detection and quantitative determination. Anal. Chem. 40 (3), 586–593.
- Eichrom Technologies, Inc., 10th of February 2005. Uranium in Soil (2 grams Samples). Analytical Procedures, ACS07, Rev. 1.5.
- Fifield, L.K., 2008. Accelerator mass spectrometry of the actinides. Quat. Geochronol. 3, 276–290.
- Hardy, E.P., Krey, P.W., Volchok, H.L., 1973. Global inventory and distribution of fallout plutonium. Nature 241, 444–445.
- Herrera, R.G., Puyol, D.G., Martín, E.H., 2001. Influence of the North Atlantic oscillation on the Canary Islands precipitation. J. Clim. 14 (19), 3889–3903.
- Hotchkis, M.A.C., Child, D., Fink, D., Jacobsen, G.E., Lee, P.J., Mino, N., Smith, A.M., Tuniz, C., 2000. Measurement of ²³⁶U in environmental media. Nucl. Instrum. Methods Phys. Res. B 172, 659–665.
- Hrnecek, E., Aldave de las Heras, L., Betti, M., 2002. Application of microprecipitation on membrane filters for alpha spectrometry, liquid scintillation counting and mass spectrometric determination of plutonium isotopes. Radiochim. Acta 90, 721–725.
- Kelley, J.M., Bond, L.A., Beasley, T.M., 1999. Global distribution of Pu isotopes and ²³⁷Np. Sci. Total Environ. 237–238, 483–500.
- ²³⁷Np. Sci. Total Environ. 237–238, 483–500.
 Ketterer, M.E., Groves, A.D., Strick, B.J., 2007. ²³⁶U inventories, ²³⁶U/²³⁸U, and ²³⁶U/²³⁹Pu: the stratospheric fallout signature. goldschmidt conference abstracts. Geochim. Cosmochim. Acta 71, A480.
- Ketterer, M.E., Szechenyi, S.C., 2008. Determination of plutonium and other transuranic elements by inductively coupled plasma mass spectrometry: a historical perspective and new frontiers in the environmental sciences. Spectrochim. Acta Part B 63. 719–737.
- Kutkov, V.A., Arefieva, Z.S., Murav'ev, Y.B., Komaritskaya, O.I., 1995. Unique form of airborne radioactivity: nuclear fuel "hot particles" released during the

- chernobyl accident. In: Environmental Impact of Radioactive Releases, Proceedings of an International Symposium on Environmental Impact of Radioactive Releases. IAEA, Vienna.
- Moreno, J., Vajda, N., Danesi, P.R., Larosa, J.J., Zeiller, E., Sinojmeri, M., 1997. Combined procedure for the determination of ⁹⁰Sr, ²⁴¹Am and Pu radionuclides in soil samples. J. Radioanal. Nucl. Chem. 226 (1–2), 279–284.
- Moss, J.L., McGuire, W.J., Page, D., 1999. Ground deformation monitoring of a potential landslide at La Palma, Canary Islands. J. Volcanol. Geoth. Res. 94, 251–264.
- Oughton, D., Day, P., Fifield, K., 2001. In: Kudo, A. (Ed.), Plutonium Measurement using Accelerator Mass Spectrometry: Methodology and Applications. Plutonium in the Environment, vol. 1. Elsevier Science Ltd, pp. 47–62.
- Sakaguchi, A., Kawai, K., Steier, P., Quinto, F., Mino, K., Tomita, J., Hoshi, M., Whitehead, N., Yamamoto, M., 2009. First results on ²³⁶U levels in global fallout. Sci. Total Environ. 407. 4238–4242
- Salbu, B., 2001. In: Kudo, A. (Ed.), Actinides Associated with Particles. Plutonium in the Environment, vol. 1. Elsevier Science Ltd, pp. 121–138.
- Steier, P., Bichler, M., Fifield, L.K., Golser, R., Kutschera, W., Priller, A., Quinto, F., Richter, S., Srncik, M., Terrasi, P., Wacker, L., Wallner, A., Wallner, G., Wilcken, K.M., Wild, E.M., 2008. Natural and anthropogenic ²³⁶U in environmental samples. Nucl. Instrum. Methods Phys. Res. B 266, 2246–2250.
- Steier, P., Dellinger, F., Forstner, O., Golser, R., Knie, K., Kutschera, W., Priller, A., Quinto, F., Srncik, M., Terrasi, F., Vockenhuber, C., Wallner, A., Wallner, G., Wild, E.M., 2010. Analysis and application of heavy isotopes in the environment. Nucl. Instrum. Methods Phys. Res. B 268, 1045—1049.
- Srncik, M., Hrnecek, E., Steier, P., Wallner, A., Wallner, G., Bossew, P., 2008. Vertical distribution of ²³⁸Pu, ²³⁹⁽⁴⁰⁾Pu, ²⁴¹Am, ⁹⁰Sr and ¹³⁷Cs in Austrian soil profiles. Radiochim. Acta 96, 733–738.
- UNSCEAR Report, 1993. Annex B: Exposure from Man-made Sources of Radiation. Sources and Effects of Ionizing Radiation, pp. 91–120.
- Vockenhuber, C., Ahmad, I., Golser, R., Kutschera, W., Liechtenstein, V., Priller, A., Steier, P., Winkler, S., 2003. Accelerator mass spectrometry of heavy long-lived radionuclides. Int. J. Mass Spectrom. 223–224, 713–732.
- Warneke, T., Croudace, I.W., Warwick, P.E., Taylor, R.N., 2002. A new ground-level fallout record of uranium and plutonium isotopes for northern temperate latitudes. Earth Planet. Sci. Lett. 203, 1047–1057.
- Wilcken, K.M., Fifield, L.K., Barrows, T.T., Tims, S.G., Gladkis, L.G., 2008. Nucleogenic ³⁶Cl, ²³⁶U and ²³⁹Pu in uranium ores. Nucl. Instrum. Methods Phys. Res. B 266, 3614–3624.