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Bree, L.G.J. van, Rijpstra, W.I.C., Al-Dhabi, N.A., Verschuren, D., Sinninghe Damsté, J.S. & Leeuw, J.W. de (2016). *Des*-A-lupane in an East African lake sedimentary record as a new proxy for the stable carbon isotopic composition of C₃ plants *Organic Geochemistry*, 101, 132–139

Published version: dx.doi.org/10.1016/j.orggeochem.2016.09.003

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1	Des-A-lupane in an East African lake sedimentary record as a new proxy for C3-plant
2	stable carbon isotopic composition
3	
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20	To be submitted to Organic Geochemistry

22 Abstract

We studied the high-resolution and well-dated 25,000 year sedimentary record of Lake 23 Challa, a deep tropical crater lake in equatorial East Africa, to explore new proxies for 24 paleoenvironmental and paleohydrological change. Sedimentary biomarker analysis revealed 25 the presence of *des*-A-triterpenoids with oleanane, ursane and lupane carbon skeletons, 26 microbial degradation products of angiosperm plant triterpenoids. Their increased influx from 27 16,000 years ago corresponds with previously documented changes in the terrestrial 28 29 vegetation of the Lake Challa basin during postglacial warming, in particular the relative increase in C₃/C₄ plant ratio inferred from the stable carbon isotopic signature (δ^{13} C) of 30 sedimentary *n*-alkanes derived from plant leaf waxes. In contrast to this *n*-alkane δ^{13} C, the 31 δ^{13} C of des-A-lupane maintains a constant value of -27.4±1.1‰ across the glacial–interglacial 32 transition. Since *des*-A-lupane is derived from C₃ plants, its δ^{13} C signature is here proposed to 33 34 represent a novel and independent proxy for the time-variable carbon isotopic composition of local terrestrial C₃ plants, which can improve estimates of the C₃/C₄ plant ratio based on two-35 end member mixing models of *n*-alkane δ^{13} C values. 36

37

38 Keywords

39 *Des*-A-triterpenoids; *des*-A-lupane; δ¹³C of *des*-A-lupane; lacustrine sediments; East Africa;
40 C₃/C₄ plant ratio; vegetation reconstruction.

42 **1. Introduction**

43 Non-hopanoid pentacyclic triterpenoids preserved in lake and marine sediments are relatively well studied biomarkers, and used in organic geochemistry as a proxy for the input of 44 45 terrestrial angiosperm plants (e.g. Rullkötter et al., 1982; Freeman et al., 1994; Sabel et al., 2005). These wax components are synthesized almost exclusively by higher plants as a 46 defense mechanism against insects, pathogens and herbivores (Langenheim, 1994). Pentacylic 47 48 triterpenoids with, for example, lupane, oleanane and ursane skeletons, are widely accepted as general biomarkers for angiosperms, yet most of these triterpenoids are not species-specific 49 (e.g. Ohmoto et al., 1970; ten Haven and Rullkötter, 1988; Regnery et al., 2013). Des-A-50 triterpenoids, in turn, are diagenetic products of triterpenoid A-ring degradation (Trendel et 51 al., 1989), a process probably mediated by microorganisms under anoxic conditions 52 (Lohmann et al., 1990). 53

54 Plant wax lipids, such as *n*-alkanes, can be used as indicators of the composition of local vegetation, as their compound-specific stable carbon isotopic value (δ^{13} C) depends on 55 the dominant biochemical pathway used by plants for photosynthesis. Higher plants usually 56 fix CO₂ by either the C₃ (Calvin-Benson) or C₄ (Hatch-Slack) cycle. In C₄ photosynthesis, 57 atmospheric CO₂ entering the plant stomata is pre-concentrated, resulting in less carbon 58 59 isotopic depletion. In tropical Africa, lowland vegetation consists mostly of C₄ grasses and C₃ trees and shrubs, and (seasonal) moisture availability is the dominant control on the 60 distribution and abundance of these C4 grasses across the modern-day landscape. On glacial-61 interglacial time scales, variation in atmospheric pCO_2 exerted important control on the 62 composition of savanna vegetation of East Africa (Sinninghe Damsté et al., 2011). When 63 pCO_2 varies little, as within the Holocene, the distribution of C₃ and C₄ vegetation can be 64 65 used as a measure of past water availability (Castañeda et al., 2007).

During an earlier high-resolution stratigraphic study of *n*-alk-1-enes in the continuous and well-dated 25,000-year (25 kyr) sedimentary record of Lake Challa, a deep crater lake in equatorial East Africa (van Bree et al., 2014), we also encountered a wide range of *des*-Atriterpenoids. In this paper, we report our findings regarding the origin and potential paleoenvironmental application of the pentacyclic triterpenoid degradation products derived from higher plants. A study of *des*-A-arborenes/fernenes also present in Lake Challa sediments, but of microbial origin, will be published separately.

73

74 2. Material and Methods

75 <u>2.1 Study area</u>

Lake Challa (3°19'S, 37°42'E) is a relatively unproductive, tropical freshwater lake, situated 76 ~880 m above sea level on the lower south-eastern slope of Mt. Kilimanjaro (Fig. 1). Shared 77 by Kenya and Tanzania, it has a surface area of 4.5 km², a steep-sided crater catchment of 78 \sim 3.4 km² and a maximum depth of \sim 90 m. Rainfall seasonality is mainly determined by the 79 80 semi-annual passing of the Inter-Tropical Convergence Zone (ITCZ), resulting in moderate 81 'long rains' (March-May) and more intense 'short rains' (October-December) separated by a long dry season during Southern Hemisphere winter (June-September). Since the local 82 83 balance between precipitation and evaporation is negative by three to one, the lake's water budget must be maintained by substantial groundwater inflow, derived from rainfall on the 84 forested mid-elevation slopes of Mt. Kilimanjaro. Also temporary creek discharge breaching 85 the crater's northwestern rim can occur during very heavy rains (Buckles et al., 2014). Daily 86 wind-driven mixing of the water column is limited to the uppermost 15-20 m year-round. 87 Seasonal deeper mixing (June-September) extends to 40-60 m, implying a permanently 88 89 stratified lower water column. Given limited density stratification, the possibility of complete mixing cannot be excluded but its recurrence frequency is likely decadal or longer, rather than 90

inter-annual (Wolff et al., 2011, 2014). Wind-blown (aeolian) particles, supplemented by soil
and litter input from the steep-sided inner slopes of the crater, contribute allochthonous
organic matter to Lake Challa sediments. The vegetation outside the crater and on top of the
rim is a woodland-savannah with shrubs, trees and C4 grasses, while inside the crater rim a
more varied vegetation occurs, including CAM plants in a dry succulent forest occupying the
middle slopes, and a fringe of evergreen forest around the lake shore (Hemp, 2006; Sinninghe
Damsté et al., 2011).

98

99 <u>2.2 Study material</u>

A 21.65 m composite core of mostly finely-laminated organic muds was retrieved from the 100 center of Lake Challa (Fig. 1) during coring activities in 2003 and 2005 (Verschuren et 101 al.,2009). This yielded, after excision of five turbidites, a 20.82 m long master sequence 102 103 covering the last 25 kyr of continuous offshore lacustrine sedimentation. The age-depth model is based on a smoothed spline through INTCAL04-calibrated AMS ¹⁴C ages of 164 bulk 104 105 organic carbon samples (Blaauw et al., 2011). For this study, a total of 148 sediment samples with 4-cm thickness were extracted and processed for biomarker analysis, generally at ~200-106 year intervals, with a higher resolution of ~50 year intervals in the youngest 3.2 kyr. In 107 addition, we analyzed the compound-specific δ^{13} C of selected biomarkers in 52 sediment 108 109 samples distributed throughout the sequence.

In order to trace the origin of the biomarkers found in Lake Challa sediments, we
further analyzed soil and litter samples from around the lake, suspended particulate matter
(SPM) and settling particles collected by a sediment trap (Sinninghe Damsté et al. 2009;
Buckles et al. 2014). This sediment trap was deployed at 35 m water depth in a most often
suboxic part of the water column, and samples were retrieved at ca. 4-week intervals between
November 2006 and December 2007 (Sinninghe Damsté et al., 2009). The SPM sampling was

conducted on 10–11 September 2006 along a vertical profile in the center of the lake, every 5 116 m between 0 and 30 m, and every 10 m between 30 and 90 m water depth. The water samples 117 (4 to 9 L) were filtered through GF/F filters and stored frozen until processing. Eight soil 118 samples were collected from within the catchment area of Lake Challa in 2005 (Sinninghe 119 Damsté et al., 2009; Fig. 1). Sampling of litter in September 2012 included three near-shore 120 samples representing leaf and fruit remains from forest trees and shrubs, and two samples 121 from just below the crater rim consisting of leaf and twig remains and small non-diagnostic 122 organic debris. These samples were also stored frozen until analysis. 123

124

125 2.3 Lipid extraction

The freeze-dried and powdered sediment samples were extracted with a DionexTM 126 Accelerated Solvent Extractor (ASE), using a dichloromethane (DCM)/methanol (9:1, v/v) 127 128 mixture at high temperature (100°C) and pressure (7.6 x 106 Pa) (Sinninghe Damsté et al., 2011). The total extracts were separated over an activated Al₂O₃ column into an apolar and a 129 polar fraction with hexane/DCM (9:1, v/v) and DCM/MeOH (1:1, v/v), respectively. 130 Accumulation rates of apolar compounds (in mg m⁻² yr⁻¹) were calculated based on their 131 concentration, the wet weight and water content of the sediment samples, and the sediment's 132 133 age-depth profile. For analysis of stable carbon isotopic composition, a subset of the apolar fractions were separated into a saturated and an unsaturated hydrocarbon fraction using a 134 small Ag⁺-impregnated silica column with hexane and ethyl acetate as eluents, respectively. 135 SPM and sediment trap samples were extracted previously (Sinninghe Damsté et al., 136 2009). Fluxes of apolar compounds (in mg m⁻² yr⁻¹) in the sediment trap samples were 137 calculated using the concentration of each of these components relative to total particle flux. 138 Litter and soil samples were extracted ultrasonically with DCM/MeOH (2:1, v/v), after 139 cutting the larger leaf, stem or fruit remains into small pieces. The extracts were evaporated to 140

dryness, methylated with diazomethane in diethyl ether and separated over an activated Al₂O₃ 141

column into apolar, ketone and polar fractions, using hexane/DCM (9:1, v/v), hexane/DCM 142

(1:1, v/v) and DCM/MeOH (1:1, v/v) as eluents, respectively. 143

144

2.4 Lipid identification and quantification 145

The apolar fractions of sediments, SPM, sediment trap, soil and litter extracts were analyzed 146 by gas chromatography (GC) and GC-mass spectrometry (GC/MS), after addition of a known 147 amount of internal standard. A few polar fractions of sediment and SPM samples were 148 methylated, silvlated and screened for functionalized triterpenoids. 149

GC was performed using a Hewlett-Packard (HP6890) instrument equipped with an 150 on-column injector and a flame ionization detector (FID). A fused silica capillary column (25 151

m x 0.32 mm) coated with CP Sil-5 CB (film thickness 0.12 µm) was used with helium as 152

153 carrier gas. The samples were injected at 70°C and the oven temperature was programmed to

rise at 20°C/min to 130°C, and then at 4°C/min to 320°C, at which it was held for 20 min. 154

155 GC-MS was performed on a Finnigan Trace DSQ mass spectrometer operated at 70 156 eV with a mass range of m/z 40 to 800 and a cycle time of 1.7 s. The gas chromatograph was equipped with a fused silica capillary column as described above. The carrier gas was helium 157 158 and the same oven temperature program as for GC was used. Identification of the des-A-

159 triterpenoids and other triterpenoid hydrocarbons is based on relative retention times,

published mass spectra (including the NIST98 spectral library), and interpretation of observed 160

fragmentation patterns. Quantification of compounds was performed by peak area integration

of appropriate peaks (including that of the internal standard) in the FID chromatograms. 162

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164 2.5 Compound-specific carbon isotope analyses

We subjected 52 saturated hydrocarbon fractions to compound-specific δ^{13} C analysis using an 165 166 Agilent 6800 GC coupled to a ThermoFisher Delta V isotope-ratio monitoring mass spectrometer. The isotope values were measured with reference to a calibrated external 167 168 reference gas, and performance of the instrument was monitored daily by injections of a mixture of a C₂₀ and a C₂₄ perdeuterated *n*-alkane with known isotopic composition. The δ^{13} C 169 values are reported in standard delta notation against the Vienna Pee Dee Belemnite (VPDB). 170 All samples were run at least in duplicate, allowing estimation of the standard deviation of 171 these measurements. δ^{13} C values of the *n*-alkanes in these same samples have been published 172 previously (Sinninghe Damsté et al., 2011). 173

174

175 **3. Results**

176 <u>3.1 The sedimentary record of *des*-A-triterpenoids in Lake Challa</u>

177 Our analysis of the apolar fractions of 148 biomarker extracts from Lake Challa sediments

178 covering the last 25 kyr showed the presence of *n*-alkanes, *n*-alk-1-enes, phytadienes,

aromatic triterpenoids, hopenes, and a suite of *des*-A-triterpenoid hydrocarbons with lupane,

180 oleanane, ursane, and fernane or arborane skeletons. The stratigraphic distribution, origin and

palaeoenvironmental significance of the *n*-alkanes and *n*-alk-1-enes have been discussed by

182 Sinninghe Damsté et al. (2011) and van Bree et al. (2014), respectively.

We detected 11 distinct *des*-A-triterpenoid hydrocarbon compounds on the basis of on their relative retention times, published mass spectra, and interpretation of mass spectrometric fragmentation patterns. The mass spectra of ten *des*-A-triterpenes and one *des*-A-triterpane exhibit molecular ions at m/z 324, 326, 328 or 330. In this paper we focus on the distribution, origin and potential paleoenvironmental application of the five *des*-A-triterpenoids with lupane, oleanane and ursane skeletons (Fig. 2A). The identification and geochemical significance of the other, arborane- or fernane-derived *des*-A-triterpenoids will be discussed elsewhere. Next to these *des*-A-triterpenoids with arborane or fernane skeleton, several other,
non-identified *des*-A-triterpenoids co-elute, occur infrequently or in low relative abundance
and are not further discussed.

193	Compound 1 (M ⁺⁺ at m/z 328; Fig. 2B) was identified as <i>des</i> -A-olean-13(18)-ene
194	(Corbet, 1980; Logan and Eglinton, 1994; Jacob et al., 2007; Huang et al., 2008), and
195	compound 2 (M ^{+•} at m/z 328; Fig. 2B) was tentatively assigned to des-A-olean-12-ene
196	(Corbet, 1980; Logan and Eglinton, 1994; Jacob et al., 2007). Des-A-triterpenoid 3 (M ^{+•} at
197	<i>m</i> /z 328; Fig. 2B) was identified as <i>des</i> -A-urs-13(18)-ene (Corbet, 1980; Logan and Eglinton,
198	1994; Jacob et al., 2007). Compound 4 (M ⁺⁺ at m/z 328; Fig. 2B) represents a mixture of des-
199	A-ursenes, which could not be further identified. <i>Des</i> -A-triterpenoid 5 exhibiting an M^{+} at
200	<i>m</i> / <i>z</i> 330 (Fig. 2B) was identified as <i>des</i> -A-lupane (Corbet, 1980; Schmitter et al., 1981;
201	Trendel et al., 1989; Boreham et al., 1994; Jacob et al., 2007; Huang et al., 2008).
202	The accumulation rates (in mg m ⁻² yr ⁻¹) of these <i>des</i> -A-triterpenoid hydrocarbons in
203	the sedimentary record vary considerably over time (Fig. 3). The stratigraphic distributions of
204	des-A-triterpenoids with lupane (5) , ursane (3) and oleanane $(1+2)$ skeletons are all similar
205	and significantly inter-correlated: des-A-lupane with des-A-oleanenes (R ² =0.53), des-A-
206	oleanenes with <i>des</i> -A-ursene (R^2 =0.62) and <i>des</i> -A-lupane with <i>des</i> -A-ursene (R^2 =0.81).
207	Throughout the record, des-A-lupane is more abundant than des-A-ursene, whereas des-A-
208	oleanenes concentration is lowest (Fig. 3). They all exhibit low accumulation rates from
209	25,000 to 15,000 years ago (the glacial and early post-glacial period) and much higher (on
210	average 5 to 11 times) accumulation rates in last 12,000 years (the Holocene).
211	The δ^{13} C values of <i>des</i> -A-lupane (Fig. 4) vary between -25.1‰ and -29.4‰ (-
212	27.4‰ \pm 1.1‰ on average) with no significant trend over time (R ² =0.042; <i>n</i> =38; p=0.22).
213	Determinations of δ^{13} C of other <i>des</i> -A-triterpenoids of the oleanane and ursane type in the
214	unsaturated hydrocarbon fraction were unsuccessful due to co-elution or low abundances.

216	3.2 Triterpenoids and <i>des</i> -A-triterpenoids in soil and litter surrounding the lake
217	No des-A-triterpenoids were identified in the hydrocarbon fraction of extracts from soils and
218	litter. Functionalized triterpenoids in soils along the crater rim (Fig. 1) were identified as urs-
219	12-en-3 β -ol (α -amyrin), olean-12-en-3 β -ol (β -amyrin), olean-13(18)-en-3 β -ol, ursa-9(11), 12-
220	dien-3-one, taraxerone, lup-20(29)-en-3-one, lup-20(29)-en-3-ol acetate and friedelan-3-one
221	and, tentatively, methyl-ursa-2,12-dien-28-oate. Also three pentacyclic triterpene methyl
222	ethers (PTMEs) with oleanane and taraxerane skeletons were detected: 3β -methoxy-olean-12-
223	ene (iso-sawamilletin; β -amyrin ME), 3 β -methoxy-olean-18-ene ME (miliacin; germanicol
224	ME) and 3β -methoxy-taraxer-14-ene (sawamilletin, crusgallin or taraxerol ME). Litter
225	samples contained the oleanane-type triterpenoids β -amyrenone, olean-18-en-3 β -one
226	(germanicone), olean-18-en-3 β -ol (germanicol) and β -amyrin acetate. The functionalized
227	ursane-type triterpenoids present were α -amyrinone and α -amyrin acetate. One litter sample
228	contained a lupeol acetate.
229	
230	3.3 Des-A-triterpenoids in sedimenting particles and suspended particulate matter
231	In the extracts of sedimenting particles, only low concentrations of <i>des</i> -A-urs-13(18)-ene and
232	des-A-lupane were detected, next to traces of functionalized triterpenoids such as α -amyrin.
233	Hydrocarbon concentrations in the SPM extracts were too low for identification of specific
234	triterpenoid compounds.
235	
236	4. Discussion
237	The strong correlation between accumulation rates of des-A-oleanenes, des-A-ursene and des-
238	A-lupane (Fig. 3) suggest a common origin for these des-A-triterpenoids. Higher plant

239 material originates predominantly from washed-in debris of local terrestrial vegetation from

within the crater catchment, because of the absence of riverine input, the limited amount of 240 241 leaf waxes deposited from the air, and the near-absence of submerged or emergent aquatic macrophytes due to the steep crater walls both above and below the lake surface (Sinninghe 242 243 Damsté et al., 2011). While various des-A-triterpenoid hydrocarbons are present both in the sediment record and to some extent in settling particles, these compounds are lacking in the 244 collected litter and soil. Des-A-triterpenoids are thought to result from microbial degradation 245 246 of functionalized triterpenoids, under anoxic or reducing conditions (e.g. Lohmann et al., 247 1990; Jacob et al., 2007; Huang et al., 2008); their presence in the sediment trap material indicates that microbial degradation is not restricted to the anoxic lower water column in Lake 248 249 Challa. One other mechanism for des-A-triterpenoids to enter the system, namely the washing-in of microbial degradation products from (anoxic) soils, is unlikely in Lake Challa, 250 251 as the studied litter and soils did not contain *des*-A-triterpenoids. 252 The principal degradation processes affecting pentacyclic triterpenoids have been described by various authors (e.g. Trendel et al., 1989; Hauke et al., 1992a, 1992b; Jacob et 253 254 al., 2007), although not all transformation routes are completely established. Early diagenetic degradation of C-3 oxygenated triterpenoids involves either A to D-ring aromatization or the 255 loss of the A-ring, followed by progressive aromatization from ring B to D (Trendel et al., 256 257 1989; Lohmann et al., 1990). The dominant transformation pathway of non-hopanoid triterpenoids in Lake Challa sediments is loss of the A-ring. This loss can be initiated by the 258 formation of A-seco-intermediates, a process that can already occur within the vegetation by 259 photochemical or photomimetic influences (Corbet, 1980; Baas, 1985), but in Lake Challa 260 sediments such intermediates were not identified. 261

262

263 <u>4.1 Des-A-oleanenes and des-A-ursenes</u>

Oleanane- and ursane-type triterpenoids are generally considered to be biomarkers of terrestrial higher plants, specifically angiosperms (e.g. Diefendorf et al., 2012). Smetanina et al. (2001) identified miliacin (3 β -methoxy-olean-18-ene) in a marine fungus, which would imply that oleanoid-type triterpenoids are not exclusively produced by terrestrial higher plants. However, a new study on this fungal species did not yield any miliacin (Bossard et al., 2013). We therefore consider all oleanane- and ursane-type triterpenoids as angiosperms biomarkers.

Functionalized triterpenoids with ursane and oleanane skeletons occur in the local 271 vegetation, soils and litter of Lake Challa, often as α - or β -amyrin. The accumulation of their 272 des-A-counterparts in the sediments increased markedly from 15 kyr BP onwards (Fig. 3). A 273 comparatively low accumulation of higher plant des-A-triterpenes occurred during the C4-274 plant dominated glacial and early late-glacial periods (Sinninghe Damsté et al., 2011; Fig. 4), 275 276 suggesting that the C₄ grasses which dominated the vegetation inside the crater at that time 277 did not produce (much of these) triterpenoids. The increase in higher plant des-A-triterpenoid 278 accumulation after 15 kyr BP (Fig. 3) broadly coincides with the shift towards a mixed C_3/C_4 279 vegetation as inferred from δ^{13} C signature of *n*-C₃₁ alkanes (Fig. 4; Sinninghe Damsté et al., 2011) and palynological data (van Geel et al., 2011), following the post-glacial intensification 280 281 of the region's monsoon rainfall (Verschuren et al., 2009).

Concentrations of *des*-A-oleanenes in Lake Challa sediments are relatively low compared to the other *des*-A-triterpenoids, especially considering the predominance of oleanane-type functionalized triterpenoids in the investigated soils and litter. Future studies on the hydrocarbons in local Lake Challa vegetation might shed light on this apparent discrepancy. Possibly, aromatization of some triterpenoid types is favored over the loss of the A-ring, although the low and infrequent occurrence of aromatic triterpenoids in the sedimentary record does not seem to fit this scenario.

290 <u>4.2 *Des*-A-lupane: a tracer of C₃-plant vegetation composition</u>

291	The only saturated des-A-triterpenoid present in Lake Challa sediments is des-A-lupane, as
292	has also been reported from other late-Quaternary lake-sediment records (see e.g. Jacob et al.,
293	2007; Huang et al., 2008). This saturation may be a result of its specific precursors such as
294	lupeol, lupanol or lupanone, or because des-A-lupane is more resistant to diagenesis
295	compared to other saturated <i>des</i> -A-triterpenoids (Jacob et al., 2007; Huang et al., 2008).
296	Des-A-lupane records have been interpreted in different ways. Huang et al. (2008),
297	who studied the Dajiuhu peat deposit in China, used des-A-lupane as a proxy for the
298	depositional environment at the time of burial, linking sections of the biomarker record with
299	more des-A-lupane to episodes of limited degradation (i.e. better preservation). This
300	mechanism is less likely to apply to Lake Challa, because of its relatively constant
301	sedimentation rate and continuously anoxic bottom-water conditions. In a study on the
302	Brazilian Lake Caçó, des-A-lupane was thought to be derived from a belt of spike-rush
303	(Eleocharis sp.), an emergent aquatic macrophyte (Jacob et al., 2007). As already mentioned,
304	aquatic macrophytes are not a likely source of organic matter input in Lake Challa, as the
305	shoreline of Lake Challa is rocky and near-vertical (Fig. 1), which largely prevents aquatic
306	macrophyte growth. Even substantial lake-level lowering, which occurred during the early
307	late-glacial period (Moernaut et al. 2010), is not expected to have created more favorable
308	conditions for development of aquatic macrophytes (Sinninghe Damsté et al., 2011). Given
309	the absence of functionalized triterpenoids in the sedimentary record of Lake Challa, we
310	consider the sedimentary des-A-lupane record (Fig. 3) to reflect the input of its precursors in
311	the lake, with a constant transformation of these precursors into des-A-lupane, and minor or
312	no degradation of <i>des</i> -A-lupane over the last 25-kyr.

313	Most non-hopanoid triterpenoids are not species-specific, and the lupane-type
314	triterpenoids are no exception. Although these compounds occur in both C_3 and C_4 plants (e.g.
315	Misra et al., 1988; Macías-Rubalcava et al., 2007; Saleem, 2009; Singariya et al., 2012, 2014),
316	the specific δ^{13} C value of <i>des</i> -A-lupane (-27.4‰ on average, SD=±1.1‰, <i>n</i> =38; Fig. 4)
317	clearly indicates a C ₃ -plant origin (<i>cf.</i> Castañeda et al., 2009; Diefendorf et al., 2012).
318	Regnery et al. (2013) reported a comparable δ^{13} C signature (ranging from -28‰ to -30.6‰)
319	of des-A-lupane in lake sediments from the Holsteinian interglacial (cf. Marine Isotope Stage
320	11c) at Dethlingen in Germany, also similar to des-A-lupane isotope values in Tertiary brown
321	coal from China (-28.1±0.6‰; Schoell et al., 1994). Some genera of the Betulaceae (birch
322	family) biosynthesize <i>des</i> -A-lupane precursors and, therefore, this C ₃ plant family is regularly
323	designated as an important biological source of des-A-lupane (Regnery et al., 2013; Schnell et
324	al., 2014). However, Betulaceae do not naturally occur in tropical Africa and no Betulaceae
325	vegetation or pollen are found in Lake Challa (van Geel et al., 2011; Sinninghe Damsté et al.,
326	2011), so in this setting, <i>des</i> -A-lupane must originate from other C ₃ -plant species.
327	In line with Regnery et al. (2013), and in contrast to Jacob et al. (2007), Huang et al.
328	(2008) and Diefendorf et al. (2012), the concentration of des-A-lupane in Lake Challa
329	sediments correlates well with oleanane- and ursane-type des-A-triterpenoids. We propose
330	that this dichotomy can be explained by one or more different sources of des-A-lupane in
331	these latter studies.
332	An interesting feature of the <i>des</i> -A-lupane δ^{13} C record is the apparent lack of trend
333	over longer timescales (Fig. 4). This record does show some variability over time, but,
334	unexpectedly, no trend, even though large changes in pCO_2 (and its carbon isotopic
335	composition), rainfall and vegetation occur over the glacial-interglacial transition. Recently it
336	was shown that carbon isotope discrimination in C ₃ land plants is independent of natural

variations in *p*CO2 (e.g. Kohn, 2016). Furthermore, the δ^{13} C value of CO₂ varies by only

³³⁸ ~1‰ (between -7‰ and -6‰) over the LGM and Holocene, a variability that is hard to ³³⁹ differentiate within the analytical error margins of compound specific δ^{13} C analysis. Lastly, ³⁴⁰ the modest variability in the *des*-A-lupane δ^{13} C record in Lake Challa that does occur shows ³⁴¹ no clear connection to the regional vegetation changes and the alternation of wetter and drier ³⁴² periods documented from other proxies. This absence of a clear trend in δ^{13} C values of *des*-A-³⁴³ lupane was also noted by Regnery et al. (2013) in their interglacial record from Dethlingen ³⁴⁴ paleolake.

Recent studies have indicated the need for more direct proxies of C₃ and C₄ higher plants 345 (Diefendorf et al., 2010; Castañeda and Schouten, 2011), and our findings may be an 346 important step in this development. Firstly, the concentration of des-A-triterpenoids with 347 ursane/oleanane/lupane skeletons in the Lake Challa record is high when C₃ plants (trees and 348 shrubs) are common in local/regional vegetation (van Geel et al., 2011; Sinninghe Damsté et 349 350 al., 2011). Therefore, we might be able to use these compounds as an absolute measure of C₃ plant abundance, instead of the usual estimate of C₃ or C₄ percentage based on variation in *n*-351 alkane δ^{13} C values. Secondly, we introduce a new way of estimating the average δ^{13} C value of 352 C_3 vegetation at any point in time, which can be important for carbon cycle studies, as $\delta^{13}C$ 353 records of C_3 plants could be used by modelers to assess the possible pCO_2 -effect of plant 354 δ^{13} C (Kohn, 2016). Thirdly, we can use this C₃-plant specific δ^{13} C record for improving 355 356 estimates of the C_3/C_4 ratio in past vegetation. Recently, the relative contribution of C_3 and C_4 plants in terrestrial vegetation has been modeled using the δ^{13} C value of long-chain *n*-alkanes 357 (e.g. Castañeda et al., 2007; Sinninghe Damsté et al., 2011; Berke et al., 2012). In the 358 Holocene section of our Lake Challa record, *des*-A-lupane δ^{13} C values are on average less 359 depleted (-27.6‰) than those off the *n*-C₃₁ alkane (-29.1‰; Sinninghe Damsté et al., 2011); 360 the situation is reversed during the glacial period, where average des-A-lupane δ^{13} C values (-361 27.2‰) are more depleted than those of n-C₃₁ alkanes (-24.3‰; Sinninghe Damsté et al., 362

2011). During the Holocene period, C_3 plants are estimated to contribute ca. 50% of the *n*- C_{31} 363 alkanes (Sinninghe Damsté et al., 2011). This value was calculated using a simple binary box 364 model, in which a 50/50 mixture of C₃- and C₄-derived alkanes (with mean values of 365 366 respectively -35.2‰ and -21.7‰ in modern plants; Castañeda et al., 2009) yields an average δ^{13} C value of -28.5%. During the glacial period, vegetation in the Lake Challa region was 367 dominated by C₄ grasses; the mixed C_3/C_4 composition developed only from 16.5 kyr BP 368 onwards (Sinninghe Damsté et al., 2011), when a riparian forest of C₃ trees and shrubs started 369 370 growing inside the crater (van Geel et al., 2011). This local growth of C₃ vegetation is also clearly evident in the record of des-A-triterpenoid hydrocarbons with lupane, ursane and 371 372 oleanane skeletons (Fig. 3). The carbon isotopic fractionation between acetyl CoA-based compounds (e.g. n-373

alkanes) and the isoprene-based isoprenoids essentially depends on the biosynthetic pathways 374 375 by which they are produced (e.g. MVA for isoprenoids in angiosperms). In their study of 376 North American C₃ plants, Diefendorf et al. (2012) showed that this biosynthetic offset 377 between terpenoids and n-alkanes is ~4-6‰. Correspondingly, also in sedimentary records, 378 the ${}^{13}C$ of terpenoids is typically reported to be enriched by 5-6‰ compared to *n*-alkanes. We do not expect this biosynthetic offset between triterpenoids and *n*-alkanes to be different in 379 East African vegetation. Temperate C₃ plants may have significantly lower 'overall' lipid 380 fractionation values compared to tropical species (Diefendorf et al., 2011), but this will 381 influence the *n*-alkanes and the terpenoids in the same way. Hence, we can use the δ^{13} C of 382 des-A-lupane as a temporal (i.e. time-specific) C_3 endmember in modeling the C_3/C_4 ratio of 383 local vegetation. To this end we subtracted an average terpenoid enrichment of 5.5% from the 384 individual *des*-A-lupane δ^{13} C values of to estimate the δ^{13} C values of *n*-C₃₁ alkanes derived 385 386 from local C₃-plants at each time interval. We use the average offset as recorded in sediments (Diefendorf et al., 2012) for this down-core correction exercise, as sediments contain more 387

388	integrated, averaged signals compared to fresh plant material. Introduction of these corrected
389	<i>n</i> -C ₃₁ alkane δ^{13} C values in the binary box model for C ₃ /C ₄ ratio calculation, using a constant
390	<i>n</i> -C ₃₁ alkane δ^{13} C value of -21.7‰ derived from C ₄ plants (<i>cf.</i> Castañeda et al., 2009), results
391	in an estimated increase in he relative proportion of C_3 plants by 0 and 19% (Fig. 5) as
392	compared to values obtained by Sinninghe Damsté et al. (2011). Our %C3 estimates are
393	different from the previous estimates (t-test; p<0.0001, n=35), and the difference remains
394	significant throughout the terpenoid to <i>n</i> -alkane offset range of 4 to 6‰ (i.e. the range that
395	exists in modern plants; Diefendorf et al., 2012). From this exercise we conclude that
396	especially in the Holocene part of the Lake Challa record, when local vegetation had a mixed
397	C_3/C_4 composition, a reconstruction using fixed C_3 -plant $\delta^{13}C$ values underestimates the
398	fraction of C ₃ vegetation, while there is less discrepancy during the glacial and early late-
399	glacial periods when C ₄ plants (here mostly grasses) were dominant. This result is especially
400	valuable for carbon-cycle modeling studies, where correct assessment of climate-vegetation
401	feedbacks strongly depends on correct estimates of past C_3/C_4 (and hence biome) distribution.
402	Our method for time-specific correction of the fraction of C ₃ plants in local vegetation
403	has the potential to enhance the accuracy of C_3/C_4 vegetation reconstructions in all situations
404	where the local C ₃ -plant signal is as strong as during the Holocene around Lake Challa (Fig.
405	5). Using sample-specific <i>des</i> -A-triterpenoid δ^{13} C values for estimating this local C ₃
406	vegetation component is convenient, as for example des-A-lupane is present in the saturated
407	aliphatic lipid fraction, just like the <i>n</i> -alkanes, and can therefore be measured in the same
408	analysis.

5. Conclusion

411 We investigated the possibility to use the degradation products of terrestrial higher plant
412 pentacyclic triterpenoids as a proxy for local vegetation reconstructions. The accumulation of

413 des-A-triterpenoids with oleanane/ursane/lupane skeletons serves as a proxy record for the 414 local abundance of C₃ vegetation. The δ^{13} C signature of des-A-lupane can be used as a proxy 415 for the stable carbon isotopic composition of local C₃ plants if, as in our study site of Lake 416 Challa, des-A-lupane is exclusively of C₃-plant origin. Therefore, it can be applied as a 417 temporally variable C₃-plant end member representing the local C₃ vegetation component in 418 reconstructions of the C₃/C₄ ratio through time.

419

420 Acknowledgements

Sample collection for this study was carried out with permission of the Permanent Secretary 421 of the Ministry of Education, Science and Technology of Kenya under research permit 422 13/001/11C to D.V. This work was performed as contribution to the ESF EuroClimate project 423 CHALLACEA and the ICDP project DeepCHALLA, financially supported mainly by grants 424 425 from the Dutch Organization for Scientific Research (NWO) and FWO-Vlaanderen (Belgium) to J.S.S.D. and D.V., respectively. Part of the work was carried out under the program of the 426 427 Netherlands Earth System Science Centre (NESSC), financially supported by the Ministry of 428 Education, Culture and Science (OCW). We thank C.M. Oluseno for field assistance, and A. Mets, M. Verweij, J. Ossebaar and M. van der Meer for technical assistance. We further thank 429 430 J. Jacob and an anonymous reviewer for their valuable comments, which greatly improved

- 431 this manuscript.
- 432

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587

588 Figure legends

Figure 1: [A] Location of Lake Challa ($3^{\circ}19$ 'S, $37^{\circ}42$ 'E) in equatorial East Africa. [B] Lake Challa and its crater basin with bathymetry at 10-m intervals and sampling locations of the suspended particulate matter (SPM) profile (Δ), the sediment trap (\Box) and the 25,000-year sediment record (\blacksquare), as well as catchment soils (\bigcirc) and litter (\odot). The outer bold line indicates the crater rim, which defines the lake's catchment area. Modified after Moernaut et al. (2010) and van Bree et al. (2014).

595

Figure 2: Higher plant derived *des*-A-triterpenoids in the Lake Challa sediments. Panel [A]
shows a partial summed mass chromatogram (*m*/*z* 163+177+189+203+218+309) of the
distribution of *des*-A-triterpenoids in the apolar fraction of the lipid biomarker extract from
996-1000 cm core depth, deposited ~12.1 kyr BP. Panel [B] shows the electron-impact mass
spectra of *des*-A-triterpenoids in the Lake Challa sediment record: *des*-A-olean-13(18)-ene
(compound 1); *des*-A-olean-12-ene (2); *des*-A-urs-13(18)-ene (3); mixture of *des*-A-ursenes
(4); *des*-A-lupane (5). Additional information on these compounds is presented in Table 1.

Figure 3: Variation through time in the accumulation rates (in mg m⁻² yr⁻¹, not cumulative) of
the *des*-A-triterpenoid compounds for *des*-A-lupane, *des*-A-ursenes and *des*-A-oleanenes in
Lake Challa sediments. Shaded areas represent the LGM (26.5-19 kyr BP), Heinrich event H1
(16.8-15.4 kyr BP) and YD (13-11.5 kyr BP).

Figure 4: Variation through time in the stable carbon isotopic composition (δ¹³C, in ‰ vs.
VPDB) of *n*-C₃₁ alkanes (Sinninge Damsté et al., 2011) and *des*-A-lupane (this study).

612	Figure 5: Percentage of C ₃ vegetation, based on <i>n</i> -C ₃₁ alkane δ^{13} C values (data from
613	Sinninghe Damsté et al., 2011), calculated in two different ways: first (solid line) using des-
614	A-lupane as the local and temporally variable C ₃ -plant endmember and an average terpenoid
615	fractionation of 5.5‰ relative to the <i>n</i> -alkanes, and second (dashed line) based on fixed δ^{13} C
616	values for the C ₃ -plant (-35.2‰) and C ₄ -plant (-21.7‰) endmembers taken from the literature
617	(Sinninghe Damsté et al., 2011). The range of C ₃ estimates when the modern terpenoid
618	fractionation range of 4 to 6‰ is taken into account is plotted as a grey band. Boxplots show
619	the differences between the original and <i>des</i> -A-lupane (grey, left-hand side) corrected C_3
620	estimate in the glacial and early late-glacial (25 to 16 kyr BP) and in the interglacial period
621	(Holocene, 12.3 to 0 kyr BP) (thus excluding the glacial-interglacial transition), with median,
622	first and third quartiles, and whiskers depicting the minimum/maximum values.
623	
624	Table 1: List of <i>des</i> -A-triterpenoids in the Lake Challa sediment record with retention times,
625	mass spectral data and (tentative) identifications. References: A: Corbet (1980); B: Logan and
626	Eglinton (1994); C: Trendel et al. (1989); D: Jacob et al. (2007); E: Huang et al. (2008); F:

627 Schmitter et al. (1981); G: Boreham et al. (1994)







B







Peak number	Retention time (GCMS)	Most significant ions (in order of decreasing abundance)	\mathbf{M}^+	Formula	Tentative identification	References
1	18.16	109, 313, 189, 328 [M ⁺], 205, 204, 218, 161	328	$C_{24}H_{40}$	Des-A-olean-13(18)ene	ABCDE
2	18.30	203, 218, 313, 189, 328 [M ⁺], 231, 243	328	$C_{24}H_{40}$	Des-A-olean-12-ene	A B D
3	18.52	313, 177, 189, 121, 175, 328 [M ⁺], 218	328	$C_{24}H_{40}$	Des-A-urs-13(18)ene	A B D
4	19.01	218, 203, 313, 189, 133, 328 [M ⁺], 231	328	$(C_{24}H_{40})$	Mixture, <i>i.a. des</i> -A-ursenes	
5	19.60	123, 149, 163, 191, 287, 206, 330 [M ⁺], 315	330	$C_{24}H_{42}$	Des-A-lupane	A C D E F G