



# A CASE OF ECOLOGICAL RENATURATION IN A DRAINED MEDITERRANEAN PEATLAND: THE CASE STUDY OF THE MASSACIUCCOLI LAKE BASIN (TUSCANY, IT)

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The research was carried out with funding from the Consorzio di Bonifica 1 Toscana Nord, 55049 - Viareggio (Lu)

The Massaciuccoli Lake floodplain is located in the Natural Park of San Rossore, Migliarino and Massaciuccoli (Figure 1), which is one of the most important residual coastal marshy areas of the Tuscany (Italy). Since the 1930s, a large part of the Massaciuccoli floodplain has been drained for agricultural purposes. To ensure a water table depth suitable for cultivation, a complex network of artificial drains and pumping stations has been used to drain the superficial aquifer and recharge the aquifer AA.VV., 1997). In the drained area, cultivated peat soils (autri-sapric and endo-salic hydrosols, with values of organic matter reaching up to 50% in some cases) are present (Pistocchi et al. 2012). Land use is characterised by conventional agriculture (covers 80% of the area) and perurban infrastructures, such as a wastewater treatment plant. In the peatland area, cropping peats are based on continuous production of maize, (Zea mays L.), sunflower (*Helianthus annuus* L.) or maize-wheat rotations, while winter cereals are mainly cultivated in the remaining part of the basin. As a consequence of land use, several environmental concerns arose in the last 50 years. The most important concerns are those related to: I. eutrophication of the lake due to nutrient enrichment (N, P) in the surface- and groundwater. Indeed, from the 1970s, the lake, from an initial oligotrophic status, progressively turned into an eutrophic/hypereutrophic system; II. the subsidence rate (2-3 m in 70 years) due to compaction and increased mineralization of peat. This process, started since land reclamation, left the lake perched above the drained area, which is now 0 to 4 m below the sea level.



Fig.1 Geographical location of study area

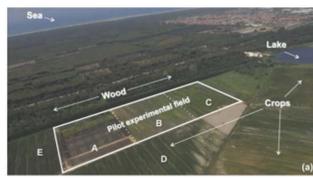


Fig.2 The area of study in the basin landscape

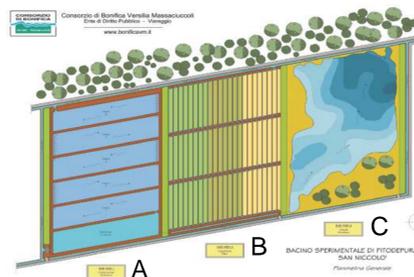


Fig.3 The pilot experimental field



Fig.4 Construction works

Fig.5 Overview of the subarea C after construction works



The project RestoMedPeatland (<https://sites.google.com/site/restomedpeatland/>) started in 2011, identified rewetting and setting-up a phyto-treatment system as the solution to improve water quality, and reduce soil organic matter (SOM) mineralisation, and, therefore, a method to restore the ecological functions of this site (Fig. 2). A pilot experimental field of 15 ha was set-up and three different management systems, with increasing anthropogenic impact, has been tested (Fig. 3): constructed wetland (A), paludicultural system (B) and natural wetland (C). This implies a gradient in regulation of water regime (from a regulated wetland, controlled system to a "quasi" natural rewetting), plant communities (from cultivated to native communities) and harvesting strategies. The soil-plant continuum systems are expected to reduce nutrient load. In addition, a conventionally cultivated (D) and an uncultivated drained (E) peat soil (the latter characterized by a natural vegetation succession), were used as controls (Fig. 4, 5, 6).



Fig.6 The pilot experimental field in 2012

The water capacity of the area C is about 8500-9000 m<sup>3</sup>, with a maximum depth of 50 cm. The flow is between 13 - 15 L / s.

Before the beginning of the construction works, in the abandoned fields existing on the sub-area C, a floristic survey was carried out. In the natural wetland (C), after top soil removal, excavation and rewetting with drainage water, the vegetation has evolved naturally. Each year thereafter, a detection of the flora was performed in spring / summer.

In the summers of 2014 and 2015 phytosociological surveys, by Braun-Blanquet method (Tabb. 2,3,4,5)

The floristic survey highlighted the development of a spontaneous hydro-gyrophilous flora, partially similar to that Massaciuccoli banks, which strongly replaced the preexisting weeds (Tabb. 1 a, b).

Tab 1 a Pre-processing floristic inventory (sub area C)

1. <i>Abutilon theophrasti</i> Medik.	Malvaceae
2. <i>Amaranthus retrofractus</i> L.	Amaranthaceae
3. <i>Arctium lappa</i> L.	Asteraceae
4. <i>Artemisia verborum</i> Lam.	Asteraceae
5. <i>Bidens tripartita</i> L.	Asteraceae
6. <i>Bromus tectorum</i> L.	Poaceae
7. <i>Calystegia sepium</i> (L.) R.Br.	Convolvulaceae
8. <i>Datura stramonium</i> L.	Solanaceae
9. <i>Echinocloa crus-galli</i> (L.) P. Beauv.	Poaceae
10. <i>Linaria vulgaris</i> Mill.	Plantaginaceae
11. <i>Phragmites australis</i> (Cav.) Trin.	Poaceae
12. <i>Typha latifolia</i> L.	Typhaceae
13. <i>Lythrum salicaria</i> L.	Lythraceae
14. <i>Phytolacca americana</i> L.	Phytolaccaceae
15. <i>Rumex crispus</i> L.	Polygonaceae
16. <i>Silene alba</i> (Miller) Krause	Caryophyllaceae
17. <i>Xanthium strumarium</i> L.	Asteraceae

The most abundant plant species were *C. sepium*, *P. australis*, *A. lappa* and *B. tectorum*

Tab 1 b. Three years after floristic inventory (sub area C)

1. <i>Alisma plantago-aquatica</i> L.	Alismataceae
2. <i>Apium nodiflorum</i> (L.) Lag	Apiaceae
3. <i>Aster squamatus</i> Spreng.	Asteraceae
4. <i>Bidens tripartita</i> L.	Convolvulaceae
5. <i>Calystegia sepium</i> (L.) R.Br.	Asteraceae
6. <i>Carex otrubae</i> Podp.	Cyperaceae
7. <i>Cirsium arvense</i> (L.) Scop.	Asteraceae
8. <i>Cyperus glomeratus</i> L.	Cyperaceae
9. <i>Datura stramonium</i> L.	Solanaceae
10. <i>Echinocloa crus-galli</i> (L.) P. Beauv.	Poaceae
11. <i>Epilobium hirsutum</i> L.	Onagraceae
12. <i>Holcus lanatus</i> L.	Poaceae
13. <i>Juncus articulatus</i> L.	Juncaceae
14. <i>Juncus bufonius</i> L.	Juncaceae
15. <i>Juncus effusus</i> L.	Juncaceae
16. <i>Lemma minor</i> L.	Lemnaceae
17. <i>Lycopus europaeus</i> L.	Hydrangeaceae
18. <i>Lythrum salicaria</i> L.	Lythraceae
19. <i>Matricaria camomilla</i> L.	Asteraceae
20. <i>Miriophyllum aquaticum</i> (Vell.) Verdc.	Haloragaceae
21. <i>Paspalum dilatatum</i> Poir.	Poaceae
22. <i>Phragmites australis</i> (Cav.) Trin.	Poaceae
23. <i>Poa trivialis</i> L.	Poaceae
24. <i>Polygonum monspeliensis</i> (L.) Desf.	Poaceae
25. <i>Polygopus viridis</i> (Gouan) Breistr.	Poaceae
26. <i>Ranunculus sceleratus</i> L.	Ranunculaceae
27. <i>Ranunculus sardous</i> Crantz	Ranunculaceae
28. <i>Samolus valerandi</i> L.	Samoebaceae
29. <i>Schoenoplectus tabernaemontani</i> (Gmel.) Palla	Cyperaceae
30. <i>Typha latifolia</i> L.	Typhaceae

The area C, was also periodically monitored through the use of a drone that, at average heights of 15 meters, has carried out photographic strips (Fig. 7) through which, once assembled and imported into a GIS environment Map Info, it was possible to realize a dynamics map of the vegetation (Fig. 8).

Fig.7 The aerophotographic mosaic of area C

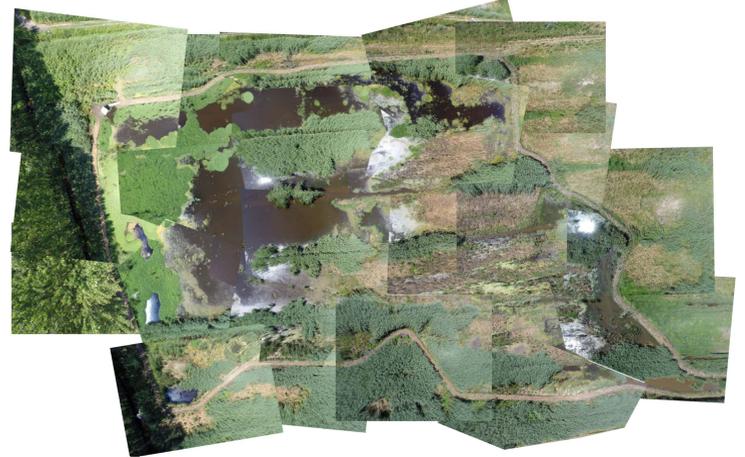


Fig.8 Vegetation map (spring 2015)

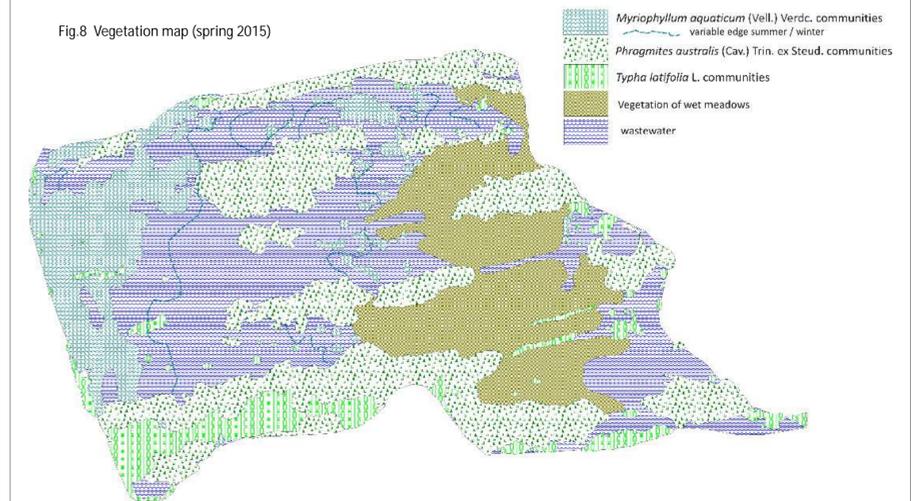


Fig.9 *Phragmites australis* (Cav.) Trin. ex Steud. communities



Fig.10 *Typha latifolia* L. communities



Fig.11 *Myriophyllum aquaticum* (Vell.) Verdc. communities



Fig.12 Vegetation of wet meadows

Tab. 2 *Phragmites australis* (Cav.) Trin. ex Steud. communities (Fig. 9)

Rel. no.	2	6	8	10	12	13	14
Surface (m <sup>2</sup> )	25	25	25	25	25	25	25
Coverage (%)	100	100	100	100	100	100	100
n° species	6	3	6	6	7	5	5

<i>Phragmites australis</i> (Cav.) Trin. ex Steud.	4	5	5	4	4	4	4
<i>Calystegia sepium</i> L.	+	+	+	+	+	+	+
<i>Eupatorium cannabinum</i> L.	-	-	-	-	-	-	-
<i>Stachys palustris</i> L.	-	-	r	r	r	-	-
<i>Lythrum salicaria</i> L.	+	-	-	+	+	-	r
<i>Typha latifolia</i> L.	-	r	r	-	-	-	-
<i>Schoenoplectus tabernaemontani</i> (Gmel.) Palla	r	-	-	r	-	-	-
<i>Mentha aquatica</i> L.	-	-	r	-	-	-	-
<i>Iris pseudacorus</i> L.	1	-	-	+	+	-	-
<i>Oenanthe aquatica</i> L.	-	-	+	-	-	-	+
<i>Alisma plantago-aquatica</i> L.	r	-	-	-	-	-	r
<i>Polygonum monspeliensis</i> (L.) Desf.	-	-	-	-	-	-	r

Tab. 3 *Typha latifolia* L. communities (Fig. 10)

Rel. no.	25	26	27
Surface (m <sup>2</sup> )	9	9	4
Coverage (%)	80	80	100
n° species	4	6	3

<i>Typha latifolia</i> L.	3	2	4
<i>Apium nodiflorum</i> (L.) Lag	+	1	+
<i>Calystegia sepium</i> L.	+	+	1
<i>Phragmites australis</i> (Cav.) Trin. ex Steud.	1	1	-
<i>Schoenoplectus tabernaemontani</i> (Gmel.) Palla	-	-	-
<i>Lythrum salicaria</i> L.	+	1	-

Tab. 4 *Myriophyllum aquaticum* (Vell.) Verdc. communities (Fig. 11)

Rel. no.	28	29	30
Surface (m <sup>2</sup> )	60	100	100
Coverage (%)	80	80	100
n° species	4	3	1

<i>Myriophyllum aquaticum</i> (Vell.) Verdc.	5	5	5
<i>Juncus articulatus</i> L.	+	-	-
<i>Lythrum salicaria</i> L.	+	+	-
<i>Lemma minor</i> L.	+	+	-

Tab. 5 Vegetation of wet meadow (Fig. 12)

Rel. no.	31	32	33	34
Surface (m <sup>2</sup> )	25	25	25	25
Coverage (%)	50	70	80	100
n° species	10	12	11	7

<i>Phragmites australis</i> (Cav.) Trin. ex Steud.	1	+	2	1
<i>Echinocloa crus-galli</i> (L.) P. Beauv.	2	1	+	3
<i>Calystegia sepium</i> L.	+	+	+	+
<i>Eupatorium cannabinum</i> L.	-	r	-	-
<i>Poa trivialis</i> L.	+	-	+	1
<i>Juncus bufonius</i> L.	-	-	+	-
<i>Juncus effusus</i> L.	-	+	+	-
<i>Carex otrubae</i> Podp.	r	r	-	-
<i>Lythrum salicaria</i> L.	+	+	2	+
<i>Schoenoplectus tabernaemontani</i> (Gmel.) Palla	r	r	-	-
<i>Iris pseudacorus</i> L.	-	r	-	r
<i>Paspalum dilatatum</i> Poir.	+	+	+	-
<i>Ranunculus sceleratus</i> L.	+	+	+	-
<i>Ranunculus sardous</i> Crantz	+	+	+	-
<i>Samolus valerandi</i> L.	+	-	-	-
<i>Epilobium hirsutum</i> L.	-	-	+	+

In four years from the flooding, the area C has become a wetland with spontaneous hydrophytic communities. These, in part, were the associations of helophytes of Lake Massaciuccoli (Tomei et al., 1997), partly dissimilar and yet others still in evolution. The populations of *Phragmites australis* appear floristically similar to *Phragmites australis* Gams 1927 of banks of Lake Massaciuccoli, while the populations of *Typha latifolia* here were differentiated from those of the lake instead made only by *Typha angustifolia*. This is probably due to a percolation of propagules and / or seeds from the ditches south of the lake where *T. angustifolia* is always present, as well as ecological conditions selective between the two species. The extensive population in *Myriophyllum aquaticum*, N-American exotic species, is only present in the reservoir and is, fortunately, absent from the lake and its massive coverage is likely linked to the particular conditions of low depth and high availability of nutrients. This area shows great potential for the study and the identification of hydro-gyrophilous species and has become a preferred site for nesting and source of food for birds (Fig. 13)

During 2015, drainage waters of this system presented a reduction of about 50% of total phosphorus and of about 43% of total nitrogen compared to the coming drainage water, mainly characterized by the following characteristics: 5.9 mg/L of nitrogen and 0.10 mg/L of phosphorus (Giannini et al., 2015).

In the hypothesis of using this strategy for the phyto-treatment of the drainage waters of the Massaciuccoli Lake, 50 ha will be needed.

The efficiency of this system is related also to the higher volume of drainage water treated per surface unit, making it one of the more promising phyto-treatment solution for the low cost construction and the high level of biodiversity provided.

In the perspective of an integration of this system in the existing agricultural district, a series of investigations on the potential biomass produced and on the quality that it has will be performed.

Studying the maps created on the basis of photographic strips, biomass samplings on areas characterized of homogeneous vegetation (e.g. monophytic stands of *Phragmites*) will be carried out and qualitative analyses of biomass will be performed to verify the suitability to different valorization chains (e.g. combustion, biogas).

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(Fig. 13) The wetland has quickly developed into a refuge for nesting and as a source of food for birds

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