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13 **Gene transcription and higher-level effects of multigenerational Zn**  
14 **exposure in *Daphnia magna***

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35 **Abstract**

36 Zn exposure of *Daphnia magna* during one generation has been shown to modulate gene  
37 transcription differently in Zn exposed organisms compared to their non-exposed offspring.  
38 Here we studied the transcriptional gene regulation with a cDNA microarray in *D. magna*  
39 exposed to Zn for three generations (F<sub>0</sub>-F<sub>2</sub>). For the first time molecular effects of  
40 multigeneration toxicant exposure in *D. magna* are described. Out of 73 differentially  
41 transcribed genes in the F<sub>1</sub> Zn exposed generation (compared to the F<sub>1</sub> control), only 7 genes  
42 were also differentially transcribed in the same direction in the F<sub>0</sub> Zn exposed daphnids (up  
43 or down, compared to the F<sub>0</sub> control). The majority of the differentially transcribed unigenes  
44 in F<sub>1</sub> Zn exposed daphnids (78 %) were not differentially transcribed in the F<sub>0</sub> Zn exposed  
45 organisms. This indicates that Zn exposure affected other molecular pathways in the second  
46 exposed generation, although a reduced reproduction and a reduction in juvenile growth  
47 were observed in both Zn exposed generations, compared to the respective controls. In the  
48 third Zn exposed generation (F<sub>2</sub>), no reduction in growth or reproduction compared to the  
49 control was observed. This acclimation was reflected in a significantly lower number of  
50 differentially transcribed genes, compared to the Zn exposed F<sub>0</sub> and F<sub>1</sub> generations.

51

52 **Keywords**

53 Acclimation, microarray, ecotoxicology, stress, ecotoxicogenomics

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## 56 1. Introduction

57 In the young and rapidly growing research field of ecotoxicogenomics, genomic tools are  
58 used to detect the molecular responses an organism experiences when exposed to  
59 pollutants, providing clues to the toxic effects in the organism and the compensatory  
60 mechanisms that are induced (Poynton and Vulpe, 2009). With DNA microarrays,  
61 ecotoxicological effects of exposure can be linked with transcription profiles of large  
62 numbers of genes. The transcriptional patterns obtained provide a means to identify  
63 complex pathways and strategies that are altered or induced in an organism when it is  
64 exposed to environmental stressors (Steinberg et al., 2008). In recent years, a number of  
65 studies has investigated the transcriptional responses of *Daphnia* sp. exposed to different  
66 types of environmental stress, using *Daphnia* microarrays. This way, molecular effects  
67 induced by exposure of daphnids to e.g. Cd, dietary Zn, fenarimol, Ni and even binary metal  
68 mixtures or munitions constituents have been discovered and elucidated (Soetaert et al.,  
69 2007; Cannon et al., 2008; De Schamphelaere et al., 2008; Garcia-Reyero et al., 2009;  
70 Vandenbrouck et al., 2009).

71 Under continuous, multigenerational exposure to certain metals, *Daphnia magna* is known  
72 to develop tolerance to this stress. This was demonstrated in experiments with Cd, Cu and  
73 Zn (Bossuyt and Janssen, 2004; Muyssen and Janssen, 2004; 2005). Molecular analyses can  
74 reveal insights into the underlying mechanisms of tolerance development during the  
75 acclimation period. This knowledge may be useful for screening or monitoring potential  
76 tolerance development in response to chemical exposure, or for investigating other  
77 environmental factors that could affect this tolerance. Except for an investigation of  
78 metallothionein induction, related to multigenerational Cd acclimation (Guan and Wang,

79 2006), no molecular studies related to tolerance development in metal acclimated *D. magna*  
80 are available in the literature.

81 In a recent study, transcriptional patterns of *D. magna* exposed to Zn for one generation and  
82 cultured under non-exposed standard conditions for two subsequent generations were  
83 analyzed using a custom cDNA microarray (Vandeghechuchte *et al.*, 2010b). This revealed  
84 transcriptional regulation of several genes, both in the exposed daphnids and in the two  
85 subsequent non-exposed generations. An interesting observation was that the differentially  
86 transcribed genes of the F<sub>0</sub> Zn exposed daphnids (compared to the F<sub>0</sub> control organisms)  
87 were different from those in their non-exposed F<sub>1</sub> and F<sub>2</sub> offspring (compared to F<sub>1</sub> and F<sub>2</sub>  
88 control daphnids).

89 In parallel with these two generations of non-exposed offspring, two generations of  
90 offspring were cultured under continuous Zn exposure. In the present study, gene  
91 transcription as well as higher-level effects in three generations of Zn exposed daphnids  
92 were studied to evaluate transcriptional effects of continuous multigeneration Zn exposure  
93 and to elucidate the acclimation process at a transcriptional level.

## 94 **2. Materials and methods**

### 95 **2.1 *Daphnia* cultures and experimental design**

96 *D. magna* Straus (clone K6) used in our experiments was originally collected from a pond in  
97 Kiel (Antwerp, Belgium) and has been successfully cultured under controlled laboratory  
98 conditions for more than 10 years in aerated carbon filtered tap-water, enriched with  
99 selenium (1 µg/L) and vitamins (7.5 mg/L thiamin, 100 µg/L cyanocobalamin and 75 µg/L  
100 biotin).

101 Daphnids were cultured in 10 mL medium per surviving daphnid during the first week and in  
102 20 mL medium per surviving daphnid from the second week onwards, maintaining a  
103 constant density of organisms and food, as described by Vandegehuchte et al. (2010b).  
104 Culture media were renewed three times per week and juveniles were removed at these  
105 occasions. The experimental design used in the current study is as follows. A set of neonates  
106 (0-24h) taken from the laboratory culture was divided into two batches. One batch was  
107 transferred to modified standard M4 medium (Elendt and Bias, 1990) and cultured in this  
108 control medium for three generations ( $F_0C-F_2C$ ). A second batch of neonates was transferred  
109 into the same medium, but with the Zn concentration adjusted to 388  $\mu\text{g/L}$  and cultured in  
110 this Zn contaminated medium for three generations ( $F_0Zn-F_1Zn$ ). Based on previous studies,  
111 the higher Zn concentration was estimated to be sublethal, with a significant effect on  
112 reproduction (Heijerick *et al.*, 2005; Muysen and Janssen, 2005). Each combination of  
113 generation and exposure (control or Zn contaminated medium) is termed a 'treatment'  
114 throughout this paper (Fig. 1). The standard M4 medium was modified by replacing EDTA  
115 and Fe by 4 mg/L of natural Dissolved Organic Carbon (DOC) to avoid the use of excessively  
116 high metal concentrations due to EDTA complexation and to increase the environmental  
117 relevance of the medium. The dissolved organic matter was collected from a small unpolluted  
118 creek (Ruisseau de St. Martin, Bihain, Belgium) using a portable reverse osmosis system (PROS/2)  
119 (Sun *et al.*, 1995). It was stored in the dark at 4 °C in a 50 L barrel, at a concentration of  
120 approximately 400 mg/L DOC. This DOC stock was thoroughly mixed each time before the  
121 preparation of new medium. The same batch of DOC was used for all treatments and media  
122 renewals. The Zn concentration in the control medium was adjusted to 19  $\mu\text{g/L}$  Zn, i.e. within  
123 the optimal concentration range of this essential element for daphnids (Muysen and  
124 Janssen, 2004).

125   Reproduction as total number of living juveniles per surviving adult after 21 days was  
126   measured by counting the number of juveniles per organism three times per week for each  
127   individual daphnid. Ten individual daphnids were kept in plastic cages (fitted with 200  $\mu\text{m}$   
128   mesh size gauze) which were suspended in the same aquaria as the treatment cultures. The  
129   length from the top of the head until the base of the spine was measured for ten different  
130   individual organisms per treatment by analyzing a microscopic image with UTHSCSA Image  
131   Tool 3.0 (San Antonio, TX, USA). This was done on day 6, day 13 and one to three days after  
132   the fifth brood was observed in the aquarium, when sufficient 0-24h offspring were available  
133   to start the next generation treatment. Internal Zn concentrations were determined as  
134   described in Vandegehuchte et al. (2010b). All Zn concentrations were measured by atomic  
135   absorption spectrometry (SpectrAA-100, Varian, Mulgrave, Australia).

## 136   **2.2 Statistical analysis**

137   All statistics were performed with Statistica (Statistica, Tulsa, USA). Differences between the  
138   Zn exposed and the control daphnids in reproduction (total number of juveniles per surviving  
139   female), length or internal Zn concentration were assessed using t-tests. For the comparison  
140   of the internal Zn concentrations in daphnids from the three Zn exposed generations, a one-  
141   way ANOVA was used. Assumptions of normality and homoscedasticity were tested with  
142   Shapiro-Wilk's test and Bartlett's test, respectively. When one of these assumptions was not  
143   met, non-parametric Mann-Whitney U tests were performed to assess differences between  
144   exposed and control treatments (USEPA, 2000). In all tests, the limit of significance was set  
145   at  $p = 0.05$ .

## 146   **2.3 Microarrays**

147 Three *D. magna* cDNA libraries enriched with genes related to energy metabolism, molting  
148 and life stage specific processes have been developed by Soetaert et al. (2006; 2007) using  
149 the suppression subtractive hybridization technique. Next to these cDNA libraries, two extra  
150 cDNA fragments, corresponding to expressed sequence tags (ESTs) from genes that are  
151 reported to be sensitive to Zn were spotted on the array: ESTs with homology to (1) ferritin  
152 (AJ292556) and (2) retinol dehydratase (DV437801) gene fragments (Poynton et al., 2007).  
153 Finally, also two ESTs with homology to putative MTs (metallothioneins) (DV437799 and  
154 DV437826) were spotted because MTs have been shown to be induced by Zn (Fan et al.,  
155 2009). The preparation and spotting of the sequences are reported by Vandegehuchte et al.  
156 (2010b).

157

#### 158 **2.4 Microarray preparation**

159 Three replicates of ten adult daphnids per treatment ('treatment' = combination of  
160 generation and exposure type, see Fig. 1) were sampled for mRNA analysis on the day the  
161 next generation was started (see above). The methods for RNA-extraction, conversion into  
162 cDNA, labeling and hybridization following a universal reference design can be found in  
163 Vandegehuchte et al. (2010b).

#### 164 **2.5 Bioinformatic analysis of microarray data**

165 The microarrays were scanned using a Genepix personal 4100 Scanner (Axon instruments,  
166 USA). Scanned images were analyzed using Genepix Pro Software 4.0 (Axon Instruments) for  
167 spot identification and for quantification of the fluorescent signal intensities. Subsequently,  
168 data were further evaluated using the Bioarray Software Environment database (BASE



169 1.2.17, <http://www.islab.ua.ac.be/base/>), i.e. a MIAME based microarray analysis package  
170 developed by the Intelligent Systems Laboratory (University of Antwerp, Belgium). Spots  
171 were background corrected by local background subtraction. Spots with saturated intensities  
172 were filtered out by visual inspection. The Cy5/Cy3 ratio was calculated for each spot,  $\log_2$   
173 transformed, and normalized between arrays using variance stabilization normalization  
174 (Huber et al., 2002). Analysis of significant differences in transcription between treatments  
175 was performed by using Limma (linear models for microarray data) (Smyth, 2004; Smyth et  
176 al., 2005). Fragments for which the p-value, adjusted for false discovery rate, was lower than  
177 0.05, were retained as significantly up- or downregulated (Benjamini and Hochberg, 1995).  
178 Only those fragments for which the  $\log_2$  ratio was outside the interval [-0.75, 0.75] were  
179 retained for further analysis. Sequence descriptions and annotations were obtained through  
180 Blast2GO (Conesa et al., 2005)([www.blast2go.de](http://www.blast2go.de)), which allowed genes to be classified into  
181 functional groups (Fig. 2). A heat plot was created with MultiExperiment Viewer (MeV) 4.5.1  
182 (Saeed *et al.*, 2006).

## 183 **Results and discussion**

184 Differences between exposed and control treatments will only be mentioned when they are  
185 statistically significant ( $p < 0.05$ ).

186 An effect of Zn exposure on growth (vs. the respective controls) was noted in 6-day old  
187 daphnids of the F<sub>0</sub>Zn and F<sub>1</sub>Zn treatments (Fig. 3A, Table 1). Growth reduction in juvenile  
188 daphnids is not uncommon and has been observed in toxicity tests with  
189 cetyltrimethylammonium bromide and 5-azacytidine (Knops *et al.*, 2001; Vandegheuchte *et*  
190 *al.*, 2010a). Like in the F<sub>0</sub> generation, a Zn induced reduction in juvenile growth (compared to  
191 the respective control) was also observed in their F<sub>1</sub>Zn offspring. However, no growth  
192 reduction was noted in the F<sub>2</sub> generation (compared to the F<sub>2</sub> control). The absence of  
193 growth reduction in the F<sub>2</sub>Zn daphnids can be interpreted as acclimation to Zn in the third  
194 exposed generation. This acclimation in the F<sub>2</sub>Zn organisms is also suggested by the fact that  
195 their reproduction is not affected (compared to the F<sub>2</sub> control daphnids), although  
196 reproduction results in F<sub>2</sub> should be interpreted with care, considering the decreased control  
197 reproduction in F<sub>2</sub>C. In the first and second generation of Zn exposed daphnids a reduction in  
198 reproduction was observed (compared to the control of the same generation, Fig. 3B, Table  
199 1). Muysen *et al* (2005) showed that exposure to Zn for six generations can increase or  
200 decrease the reproductive output, depending on the acclimation concentration and the test  
201 concentration to which the sixth-generation daphnids were exposed. These authors  
202 reported a significantly higher reproduction in daphnids of the sixth versus the first  
203 generation acclimated to 45 µg/L Zn<sup>2+</sup> (which is higher than the optimal concentration  
204 range), when exposed to an optimal test concentration of 22 µg/L Zn<sup>2+</sup>. In that study,  
205 reproduction in the actual acclimation treatments was not reported. Tolerance  
206 development/acclimation to a metal can occur even after two generations of exposure, as

207 demonstrated for net reproduction in *D. magna* exposed to 5 to 35 µg/L of Cu (Bossuyt and  
208 Janssen, 2003). This is in accordance with our results on reproduction.

209 The average Zn body burdens of the exposed F<sub>1</sub>Zn and F<sub>2</sub>Zn treatments (resp. 165 and 157  
210 µg Zn/g dry weight) were higher than those of the F<sub>1</sub>C and F<sub>2</sub>C controls (resp. 49 and 51 µg  
211 Zn/g dry weight). This is in accordance with the previously reported internal Zn  
212 concentrations of F<sub>0</sub>Zn and F<sub>0</sub>C (resp. 229 and 69 µg Zn/g dry weight, Vandeghechuchte *et al.*,  
213 2010b). There was no significant difference between the internal Zn concentrations of the  
214 three Zn exposed treatments.

215 When the gene transcription patterns of control treatments were compared (i.e. F<sub>0</sub>C vs F<sub>1</sub>C, F<sub>1</sub>C vs  
216 F<sub>2</sub>C or F<sub>0</sub>C vs F<sub>2</sub>C), a large number of genes were found to be differentially transcribed, as reported by  
217 Vandeghechuchte *et al.* (2010b). This concerned more than 15% of the unigenes on the array. The  
218 differential transcription of these genes is likely due to differences in the molting phases and  
219 reproductive cycles of the daphnids in the different generations and is as such not specific to the Zn  
220 exposure. Therefore, those genes that significantly varied in transcription between different control  
221 generations, were removed from the list of differentially transcribed genes between Zn treated  
222 organisms and controls obtained with the microarray analysis. Thus, 38 to 46 % of the differentially  
223 transcribed unigenes between treatments and controls were retained for further analysis. In the  
224 following section of the manuscript, differential transcription will always be related to the  
225 control of the same generation. Differentially transcribed genes for which a sequence  
226 description could be obtained are listed in Fig. 2. Genes for which no homology was found  
227 are summarized in the supplementary online material. Redundant fragments on the array  
228 were grouped into contigs. The resulting 1207 unique identified fragments on the array are  
229 termed unigenes (Vandeghechuchte *et al.*, 2010b).

230 In the F<sub>1</sub>Zn daphnids, 73 differentially transcribed unigenes were found (Table 1). This  
231 number is comparable to the 71 regulated unigenes in the F<sub>0</sub>Zn treatment, where also a  
232 reduction in reproduction and in body length at day 6 were observed. Seven genes were  
233 regulated in the same direction in F<sub>0</sub>Zn and in F<sub>1</sub>Zn. However, another set of seven common  
234 genes were differentially transcribed in opposite directions in F<sub>0</sub>Zn and F<sub>1</sub>Zn (Fig. 2 and  
235 supplementary online table). Although some of the remaining 59 differentially transcribed  
236 unigenes in F<sub>1</sub>Zn may belong to the same gene as fragments that were differentially  
237 transcribed in F<sub>0</sub>Zn (such as genes with homology to *D. magna* vitellogenin or to a  
238 hemoglobin subunit), for most of these fragments this is not the case. Zn exposure in the  
239 second generation daphnids clearly elicited different effects at the transcriptional level  
240 compared to the first generation. Some differentially transcribed genes in F<sub>1</sub>Zn for which a  
241 sequence description could be obtained through Blast will be discussed in the next  
242 paragraphs.

243 General trends per functional group of genes differ between F<sub>0</sub>Zn and F<sub>1</sub>Zn organisms. While  
244 in F<sub>0</sub>Zn all affected transcription and translation related genes were downregulated, four out  
245 of five transcription and translation related genes are upregulated in F<sub>1</sub>Zn. All five of these  
246 regulated unigenes are different from those in F<sub>0</sub>Zn. The potential stress-induced energy-  
247 saving mechanism of decreasing ribosomal protein synthesis (Brown-Peterson et al., 2005),  
248 which was suggested based on the downregulation of ribosomal protein coding genes in  
249 F<sub>0</sub>Zn, is not present in the second generation of Zn exposed daphnids anymore. Similarly, the  
250 oxidative stress response related genes peroxiredoxin 6 and glutathione S-transferase, which  
251 were upregulated in F<sub>0</sub>Zn, were not differentially regulated in F<sub>1</sub>Zn.

252 While most metabolism-related differentially transcribed genes were upregulated in F<sub>0</sub>Zn,  
253 this was the case for only four out of nine metabolism-related differentially transcribed  
254 genes in F<sub>1</sub>Zn. A gene coding for a serine threonine protein phosphatase, which was  
255 upregulated in F<sub>0</sub>Zn, was downregulated in F<sub>1</sub>Zn. In the presence of Fe<sup>2+</sup>, Zn<sup>2+</sup> is known to  
256 influence the activity of these phosphatases (Chu et al., 1996). It is hypothesized that in the  
257 F<sub>0</sub>Zn daphnids, the internally available Zn<sup>2+</sup> concentration may have been high enough to  
258 reduce the phosphatase activity compared to the control daphnids. A transcriptional  
259 upregulation could compensate for this. Still following this hypothesis, the internally  
260 available Zn<sup>2+</sup> concentration may have changed in the F<sub>1</sub>Zn daphnids, due to Zn induced  
261 defense mechanisms, resulting in a phosphatase activity which is near the optimum and  
262 higher than in the control, thus explaining the lower transcription. The upregulation of a  
263 serine protease, as seen in the F<sub>1</sub>Zn treatment, was also observed specifically after Zn  
264 exposure in a study of transcriptional responses in *Daphnia magna* exposed to munitions  
265 constituents, such as metals and nitroaromatic compounds (Garcia-Reyero et al., 2009).  
266 Similarly, the observed downregulation of a chitinase is consistent with previous studies with  
267 Zn exposed *D. magna*, where Zn toxicity was suggested to be associated with molting and  
268 exoskeleton maintenance (Poynton et al., 2007; Garcia-Reyero et al., 2009).

269 The upregulation of a gene coding for the heat shock protein Hsp90 can be a stress response  
270 leading to elevated levels of Hsp90 in Zn exposed daphnids, as observed in earthworms  
271 exposed to Zn and Pb contaminated soils (Marino et al., 1999). Another likely stress  
272 response, which was already noted in the F<sub>0</sub>Zn treatment, is the upregulation of a gene  
273 related to glutathione S-transferase, which is involved in oxidative stress abatement  
274 (Newman and Clements, 2008). Also similar to F<sub>0</sub>Zn, all differentially transcribed genes with  
275 homology to *D. magna* vitellogenin, which is fused with a superoxide dismutase module

276 (Kato et al., 2004), were upregulated. These genes are involved with vitellogenesis, the  
277 production of yolk proteins in the oocytes. Their differential transcription is likely due to  
278 random differences in reproductive cycle phases and associated vitellogenesis between the  
279 F<sub>1</sub>Zn and F<sub>1</sub>C daphnids, as indicated by the differential transcription between two control  
280 treatments of a unigene with the same homology (Vandeghechuchte *et al.*, 2010b). Stibor  
281 (2002) has demonstrated large differences in yolk protein levels at different times between  
282 the deposition of two consecutive broods into the brood pouch.

283 The transcriptional downregulation of genes coding for a hemoglobin protein subunit was  
284 already noted in F<sub>0</sub>Zn. Martinez-Tabche et al. (2000) reported that Zn exposure decreased  
285 the hemoglobin level in the oligochaete worm *Limnodrilus hoffmeisteri*. These authors  
286 suggested that this was caused by a Zn induced inhibition of heme synthesis. If Zn inhibits  
287 heme synthesis, it can be speculated that transcription of hemoglobin related genes would  
288 not lead to the formation of hemoglobin protein and transcriptional downregulation could  
289 be an energy-saving mechanism. Zn exposure is indeed known to decrease the hemoglobin  
290 content in *D. magna* (Berglind, 1986). A last remarkable upregulated gene in the F<sub>1</sub>Zn  
291 treatment showed homology to cytochrome p450. P450s are proteins involved with phase I  
292 detoxification, lipid metabolism and hormone synthesis/breakdown (Baldwin et al., 2009).  
293 Transcriptional upregulation of a P450 coding gene in *D. magna* was also observed after Cd  
294 exposure (Connon et al., 2008). Zn exposure, as well as Cu exposure, induced P450 activity in  
295 earthworms (Lukkari et al., 2004).

296 It is striking that in the third generation of Zn exposed daphnids (F<sub>2</sub>Zn) a much lower number  
297 of genes than in the previous generations are differentially transcribed: only 23 of which 11  
298 were upregulated (Table 1). Daphnids from this treatment seem to be acclimated to the Zn

299 exposure in the sense that no negative effects on reproduction or body length were  
300 observed, although the internal Zn concentration of 157  $\mu\text{g Zn/g}$  dry weight in body tissue  
301 was still elevated and not significantly different from the previous Zn exposed generations.  
302 Roelofs et al. (2009) also reported a smaller number of Cd-induced differentially transcribed  
303 genes in a Cd tolerant versus a reference population of the springtail *Orchesella cincta*.  
304 Additionally, these authors suggested that the absence of inhibitory effects on translation  
305 and digestive enzyme related genes could explain the smaller growth reduction upon Cd  
306 exposure in tolerant *Orchesella* populations (Posthuma et al., 1992). Our results for Zn are in  
307 line with this suggestion. No growth reduction was observed in the  $\text{ZnF}_2$  daphnids, for which  
308 only one translation and two metabolism related genes were differentially regulated,  
309 compared to six to seven and nine genes, respectively, in the previous generations with  
310 juvenile growth reduction. Two notable differences between the present study and that of  
311 Roelofs et al. (2009) can be remarked. First, springtails, unlike daphnids, are not  
312 parthenogenetic and thus genetic variation was present in their populations. Second, the  
313 springtails were selected from populations in different field sites, of which one had a long  
314 history of metal pollution, whereas the daphnids in the present study originated from the  
315 same parental generation and only differ in their three-generation exposure history. As such,  
316 no genetic selection can have acted on the daphnids in this study.

317 The genes for hydroxyisourate hydrolase (HIU hydrolase, involved in purine metabolism) and  
318 for obstructor d, involved in chitin metabolism, were downregulated in  $\text{F}_2\text{Zn}$ . Genes involved  
319 in chitin metabolism have been observed to be both up- and downregulated in several  
320 studies with metal exposed *D. magna* (Poynton et al., 2007; De Schamphelaere et al., 2008;  
321 Vandenbrouck et al., 2009). As in the other Zn exposed treatments, a gene coding for  
322 vitellogenin was upregulated and genes coding for hemoglobin subunits were

323 downregulated. Next to these, genes coding for a wd repeat protein and for a small  
324 nucleolar ribonucleoprotein involved in mRNA splicing or its regulation as well as genes with  
325 homology to chromosome 3 open reading frame 23 and to an inorganic pyrophosphatase  
326 were downregulated. Transcriptional upregulation was observed for genes coding for two  
327 proteins: one with homology to a hypothetical protein of the body louse *Pediculus humanus*  
328 *corporis* and another one with homology to a midline fasciclin, which mediates cell adhesion  
329 and signaling (Hu et al., 1998).

330 In conclusion, continuous Zn exposure resulted in acclimated *D. magna* in the third exposed  
331 generation, which exhibited no adverse effect on reproduction or growth. At the  
332 transcriptional level, few unigenes were regulated in the same direction in the three  
333 generations of Zn exposed daphnids: two genes with no homology, a vitellogenin coding  
334 gene and a hemoglobin chain coding gene. In the second Zn exposed generation (F<sub>1</sub>Zn), a  
335 large number of the differentially transcribed genes were different from those in F<sub>0</sub>Zn,  
336 although a reduction in reproduction and juvenile growth was observed in both treatments.  
337 Multigenerational exposure to Zn elicits different molecular effects in the different  
338 generations. The acclimation in the third exposed generation was reflected in a considerably  
339 smaller number of differentially transcribed genes. No direct molecular acclimation  
340 mechanisms could be deduced from the transcriptional results obtained with this custom  
341 cDNA microarray, on which a limited, although ecotoxicologically relevant, set of genes is  
342 represented. Currently, the *D. magna* genome is being sequenced by the *Daphnia* Genomics  
343 Consortium, coordinated at Indiana University. When this genome becomes available, wider  
344 transcriptome studies can be undertaken to elucidate the molecular mechanisms of metal  
345 acclimation in *D. magna*.



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351

## 352 References

- 353 Baldwin, W., Marko, P., Nelson, D., 2009. The cytochrome P450 (CYP) gene superfamily in *Daphnia*  
354 *pulex*. BMC Genomics 10, 169.
- 355 Benjamini, Y., Hochberg, Y., 1995. Controlling the False Discovery Rate: A Practical and Powerful  
356 Approach to Multiple Testing. Journal of the Royal Statistical Society. Series B (Methodological) 57,  
357 289-300.
- 358 Berglind, R., 1986. Combined and separate effects of cadmium, lead and zinc on ALA-D activity,  
359 growth and hemoglobin content in *Daphnia magna*. Environmental Toxicology and Chemistry 5, 989-  
360 995.
- 361 Bossuyt, B.T.A., Janssen, C.R., 2003. Acclimation of *Daphnia magna* to environmentally realistic  
362 copper concentrations. Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology  
363 136, 253-264.
- 364 Bossuyt, B.T.A., Janssen, C.R., 2004. Influence of multigeneration acclimation to copper on tolerance,  
365 energy reserves, and homeostasis of *Daphnia magna* Straus. Environmental Toxicology and  
366 Chemistry 23, 2029-2037.
- 367 Brown-Peterson, N.J., Larkin, P., Denslow, N., King, C., Manning, S., Brouwer, M., 2005. Molecular  
368 indicators of hypoxia in the blue crab *Callinectes sapidus*. Marine Ecology-Progress Series 286, 203-  
369 215.
- 370 Chu, Y., Lee, E.Y.C., Schlender, K.K., 1996. Activation of Protein Phosphatase 1. Journal of Biological  
371 Chemistry 271, 2574-2577.
- 372 Conesa, A., Gotz, S., Garcia-Gomez, J.M., Terol, J., Talon, M., Robles, M., 2005. Blast2GO: a universal  
373 tool for annotation, visualization and analysis in functional genomics research. Bioinformatics 21,  
374 3674-3676.
- 375 Cannon, R., Hooper, H.L., Sibly, R.M., Lim, F.L., Heckmann, L.H., Moore, D.J., Watanabe, H., Soetaert,  
376 A., Cook, K., Maund, S.J., Hutchinson, T.H., Moggs, J., De Coen, W., Iguchi, T., Callaghan, A., 2008.  
377 Linking molecular and population stress responses in *Daphnia magna* exposed to cadmium. Environ.  
378 Sci. Technol. 42, 2181-2188.
- 379 De Schampelaere, K.A.C., Vandenbrouck, T., Muysen, B.T.A., Soetaert, A., Blust, R., De Coen, W.,  
380 Janssen, C.R., 2008. Integration of molecular with higher-level effects of dietary zinc exposure in  
381 *Daphnia magna*. Comparative Biochemistry and Physiology Part D: Genomics and Proteomics 3, 307-  
382 314.
- 383 Elendt, B.-P., Bias, W.-R., 1990. Trace nutrient deficiency in *Daphnia magna* cultured in standard  
384 medium for toxicity testing. Effects of the optimization of culture conditions on life history  
385 parameters of *D. magna*. Water Res. 24, 1157-1167.
- 386 Fan, W.-H., Tang, G., Zhao, C.-M., Duan, Y., Zhang, R., 2009. Metal accumulation and biomarker  
387 responses in *Daphnia magna* following cadmium and zinc exposure. Environmental Toxicology and  
388 Chemistry 28, 305-310.
- 389 Garcia-Reyero, N., Poynton, H.C., Kennedy, A.J., Guan, X., Escalon, B.L., Chang, B., Varshavsky, J.,  
390 Loguinov, A.V., Vulpe, C.D., Perkins, E.J., 2009. Biomarker Discovery and Transcriptomic Responses in  
391 *Daphnia magna* exposed to Munitions Constituents. Environ. Sci. Technol. 43, 4188-4193.
- 392 Guan, R., Wang, W.-X., 2006. Multigenerational cadmium acclimation and biokinetics in *Daphnia*  
393 *magna*. Environmental Pollution 141, 343-352.
- 394 Heijerick, D.G., De Schampelaere, K.A.C., Van Sprang, P.A., Janssen, C.R., 2005. Development of a  
395 chronic zinc biotic ligand model for *Daphnia magna*. Ecotoxicology and Environmental Safety 62, 1-  
396 10.
- 397 Hu, S., Sonnenfeld, M., Stahl, S., Crews, S.T., 1998. Midline fasciclin: A *Drosophila* fasciclin-I-related  
398 membrane protein localized to the CNS midline cells and trachea. Journal of Neurobiology 35, 77-93.
- 399 Huber, W., von Heydebreck, A., Sultmann, H., Poustka, A., Vingron, M., 2002. Variance stabilization  
400 applied to microarray data calibration and to the quantification of differential expression.  
401 Bioinformatics 18, S96-104.

402 Kato, Y., Tokishita, S., Ohta, T., Yamagata, H., 2004. A vitellogenin chain containing a superoxide  
403 dismutase-like domain is the major component of yolk proteins in cladoceran crustacean *Daphnia*  
404 *magna*. *Gene* 334, 157-165.

405 Knops, M., Altenburger, R., Segner, H., 2001. Alterations of physiological energetics, growth and  
406 reproduction of *Daphnia magna* under toxicant stress. *Aquat. Toxicol.* 53, 79-90.

407 Lukkari, T., Taavitsainen, M., Soimasuo, M., Oikari, A., Haimi, J., 2004. Biomarker responses of the  
408 earthworm *Aporrectodea tuberculata* to copper and zinc exposure: differences between populations  
409 with and without earlier metal exposure. *Environmental Pollution* 129, 377-386.

410 Marino, F., Winters, C., Morgan, A.J., 1999. Heat shock protein (hsp60, hsp70, hsp90) expression in  
411 earthworms exposed to metal stressors in the field and laboratory. *Pedobiologia* 43, 615-624.

412 Martinez-Tabche, L., Cabrera, I.G., Olivan, L.G., Martinez, M.G., Faz, C.G., 2000. Toxic effects of zinc  
413 from trout farm sediments on ATP, protein, and hemoglobin concentrations of *Limnodrilus*  
414 *hoffmeisteri*. *J. Toxicol. Env. Health Pt A* 59, 575-583.

415 Muysen, B.T.A., Janssen, C.R., 2004. Multi-generation cadmium acclimation and tolerance in  
416 *Daphnia magna* Straus. *Environmental Pollution* 130, 309-316.

417 Muysen, B.T.A., Janssen, C.R., 2005. Importance of acclimation to environmentally relevant zinc  
418 concentrations on the sensitivity of *Daphnia magna* toward zinc. *Environmental Toxicology and*  
419 *Chemistry* 24, 895-901.

420 Newman, M.C., Clements, W.H., 2008. *Ecotoxicology: a comprehensive treatment*. CRC Press, Boca  
421 Raton, FL

422 Posthuma, L., Hogervorst, R.F., Vanstraelen, N.M., 1992. Adaptation to soil pollution by cadmium  
423 excretion in natural populations of *Orchesella cincta* (L.) (Collembola). *Archives of Environmental*  
424 *Contamination and Toxicology* 22, 146-156.

425 Poynton, H.C., Varshavsky, J.R., Chang, B., Cavigliolo, G., Chan, S., Holman, P.S., Loguinov, A.V., Bauer,  
426 D.J., Komachi, K., Theil, E.C., Perkins, E.J., Hughes, O., Vulpe, C.D., 2007. *Daphnia magna*  
427 ecotoxicogenomics provides mechanistic insights into metal toxicity. *Environ. Sci. Technol.* 41, 1044-  
428 1050.

429 Poynton, H.C., Vulpe, C.D., 2009. Ecotoxicogenomics: emerging technologies for emerging  
430 contaminants. *Journal of the American Water Resources Association* 45, 83-96.

431 Roelofs, D., Janssens, T.K.S., Timmermans, M., Nota, B., Marien, J., Bochdanovits, Z., Ylstra, B., Van  
432 Straalen, N.M., 2009. Adaptive differences in gene expression associated with heavy metal tolerance  
433 in the soil arthropod *Orchesella cincta*. *Molecular Ecology* 18, 3227-3239.

434 Saeed, A.I., Bhagabati, N.K., Braisted, J.C., Liang, W., Sharov, V., Howe, E.A., Li, J., Thiagarajan, M.,  
435 White, J.A., Quackenbush, J., Alan, K., Brian, O., 2006. *TM4 Microarray Software Suite*. Meth.  
436 *Enzymol.* Academic Press, pp. 134-193.

437 Smyth, G.K., 2004. Linear Models and Empirical Bayes Methods for Assessing Differential Expression  
438 in Microarray Experiments. *Statistical Applications in Genetics and Molecular Biology* 3, Article 3.

439 Smyth, G.K., Michaud, J., Scott, H.S., 2005. Use of within-array replicate spots for assessing  
440 differential expression in microarray experiments. *Bioinformatics* 21, 2067-2075.

441 Soetaert, A., Moens, L.N., Van der Ven, K., Van Leemput, K., Naudts, B., Blust, R., De Coen, W.M.,  
442 2006. Molecular impact of propiconazole on *Daphnia magna* using a reproduction-related cDNA  
443 array. *Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology* 142, 66-76.

444 Soetaert, A., van der Ven, K., Moens, L.N., Vandenbrouck, T., van Remortel, P., De Coen, W.M., 2007.  
445 *Daphnia magna* and ecotoxicogenomics: Gene expression profiles of the anti-ecdysteroidal fungicide  
446 fenarimol using energy-, molting- and life stage-related cDNA libraries. *Chemosphere* 67, 60-71.

447 Steinberg, C.E.W., Stürzenbaum, S.R., Menzel, R., 2008. Genes and environment - Striking the fine  
448 balance between sophisticated biomonitoring and true functional environmental genomics. *Science*  
449 *of The Total Environment* 400, 142-161.

450 Stibor, H., 2002. The role of yolk protein dynamics and predator kairomones for the life history of  
451 *Daphnia magna*. *Ecology* 83, 362-369.

452 Sun, L., Perdue, E.M., McCarthy, J.F., 1995. Using reverse osmosis to obtain organic matter from  
453 surface and ground waters. *Water Res.* 29, 1471-1477.

454 USEPA, 2000. Methods for Measuring the Toxicity and Bioaccumulation of Sediment-associated  
455 Contaminants with Freshwater Invertebrates. Duluth, Minnesota / Washington, D.C.  
456 Vandeghechte, M.B., Lemière, F., Vanhaecke, L., Vanden Berghe, W., Janssen, C.R., 2010a. Direct  
457 and transgenerational impact on *Daphnia magna* of chemicals with a known effect on DNA  
458 methylation. Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology 151, 278-  
459 285.  
460 Vandeghechte, M.B., Vandenbrouck, T., Coninck, D.D., De Coen, W.M., Janssen, C.R., 2010b. Can  
461 metal stress induce transferable changes in gene transcription in *Daphnia magna*? Aquat. Toxicol. 97,  
462 188-195.  
463 Vandenbrouck, T., Soetaert, A., van der Ven, K., Blust, R., De Coen, W., 2009. Nickel and binary metal  
464 mixture responses in *Daphnia magna*: Molecular fingerprints and (sub)organismal effects. Aquat.  
465 Toxicol. 92, 18-29.  
466  
467