



15 **Abstract**

16 In the past decade, there have been increasing concerns over the effects of pharmaceutical  
17 compounds in the aquatic environment, however very little is known about the effects of  
18 antidepressants such as the Selective Serotonin Re-uptake Inhibitors (SSRIs). Many  
19 biological functions within invertebrates are under the control of serotonin, such as  
20 reproduction, metabolism, moulting and behaviour. The effects of serotonin and fluoxetine  
21 have recently been shown to alter the behaviour of the marine amphipod, *Echinogammarus*  
22 *marinus* (Leach, 1815). The purpose of this study was to observe behavioural and  
23 transcriptional modifications in this crustacean exposed to the two most prescribed SSRIs  
24 (fluoxetine and sertraline) and to develop biomarkers of neurological endocrine disruption.  
25 The animals were exposed to both drugs at environmentally relevant concentrations from  
26 0.001 to 1 µg/L during short-term (1 hour and 1 day) and medium-term (8 days) experiments.  
27 The movement of the amphipods was tracked using the behavioural analysis software during  
28 12 min alternating dark/light conditions. The behavioural analysis revealed a significant  
29 effect on velocity which was observed after 1 hour exposure to sertraline at 0.01 µg/L and  
30 after 1 day exposure to fluoxetine as low as 0.001 µg/L. The most predominant effect of  
31 drugs on velocity was recorded after 1 day exposure for the 0.1 and 0.01 µg/L concentrations  
32 of fluoxetine and sertraline, respectively. Subsequently, the expression of several *E. marinus*  
33 neurological genes, potentially involved in the serotonin metabolic pathway or behaviour  
34 regulation, were analysed in animals exposed to various SSRIs concentrations using RT-  
35 qPCR. The expression of a tryptophan hydroxylase (*Ph*), a neurocan core protein (*Neuc*), a  
36 Rhodopsin (*Rhod1*) and an Arrestin (*Arr*) were measured following exposure to fluoxetine or  
37 sertraline for 8 days. The levels of *Neuc*, *Rhod1* and *Arr* were significantly down-regulated to  
38 approximately 0.5, 0.29 and 0.46 fold respectively for the lower concentrations of fluoxetine  
39 suggesting potential changes in the phototransduction pathway. The expression of *Rhod1*  
40 tended to be up-regulated for the lower concentration of sertraline but not significantly. In  
41 summary, fluoxetine and sertraline have a significant impact on the behaviour and  
42 neurophysiology of this amphipod at environmentally relevant concentrations with effects  
43 observed after relatively short periods of time.

44 **Keywords:** Antidepressants, SSRIs, Neuro-endocrine disruptor, Behaviour, Biomarker,  
45 Crustacean

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## 47        **1. Introduction**

48        The issue of anthropogenic contaminants released in the aquatic environment acting as  
49        endocrine disruptors has been well studied, but the research effort has mainly consisted of the  
50        study of estrogenic substances and their effects on vertebrates (Hutchinson 2007; Weltje and  
51        Schulte-Oehlmann 2007). The increasing use (over 60% in just the past decade) of  
52        antidepressants, the improper disposal of unused pharmaceuticals, and their limited  
53        biodegradability has raised concerns about their potential effects in the aquatic environment  
54        (Demeestere et al. 2010; Santos et al. 2010). Antidepressants represent about 4% of the  
55        therapeutic drugs found in the environment, and are present in coastal waters and estuaries  
56        (Santos et al. 2010). Indeed, 30 to 90% of ingested drugs are excreted and released in the  
57        environment in an active form (Kashiyama et al. 2010) which can potentially have an impact  
58        on the organisms that inhabit these areas.

59        Much is still unknown about the ecotoxicological effects of pharmaceutical and personal care  
60        products in aquatic organisms (Crane et al. 2006; Santos et al. 2010). However, recent  
61        concerns regarding the impact of antidepressants, especially selective serotonin re-uptake  
62        inhibitors (SSRIs), on aquatic organisms has been increasing (Johnson et al. 2007; Minagh et  
63        al. 2009; Demeestere et al. 2010; Guler and Ford 2010; Styris have et al. 2011). SSRIs inhibit  
64        the serotonin re-uptake into the pre-synaptic nerve inducing an increased neuro-stimulation of  
65        the post-synaptic nerve (Stahl 1998). These compounds act by modulating or mimicking the  
66        effects of serotonin (Santos et al. 2010). Since its approval by the US Food and Drug  
67        Administration in 1987, fluoxetine has become one of the most widely prescribed  
68        antidepressants, being in the top five psychiatric drugs prescribed in 2011 after citalopram  
69        and sertraline (Grohol 2012). Fluoxetine and sertraline are both SSRIs, primarily prescribed  
70        for depression but also used to treat compulsive behaviour, social anxiety, panic and  
71        personality disorders (AHFS 2013).

72        These drugs have been detected in the surface water and in wastewater effluent respectively  
73        at levels up to 0.54 µg/L and 0.929 µg/L for fluoxetine and up to 0.08 µg/L and 0.087 µg/L  
74        for sertraline (Brooks et al. 2003; Metcalfe et al. 2010; Styris have et al. 2011; Silva et al.  
75        2012). Fluoxetine has also been detected in groundwater at 0.056 µg/L (Silva et al. 2012).  
76        The only record of fluoxetine in seawater is in the Chesapeake Bay (Maryland, Virginia,  
77        USA) at 0.0026 µg/L (Pait et al. 2006). These findings make it clear that animals inhabiting  
78        aquatic ecosystems impacted by sewage effluent can be/are subjected to chronic exposure to  
79        SSRIs. Concentrations of fluoxetine and its metabolite norfluoxetine has been found at

80 extremely high level (10 µg/kg) relative to the environmental background in the tissues of  
81 fish collected near a municipal wastewater treatment plant, suggesting that these compounds  
82 have the capacity to bioaccumulate (Orem and Dolph 2002). The chronic effects of SSRIs on  
83 aquatic life are diverse (Brooks et al. 2003). For example, negative impacts of fluoxetine  
84 have been found on the reproduction and growth of invertebrates, vertebrates as well as algae  
85 (Péry et al. 2008; Lister et al. 2009; Santos et al. 2010). The effects of sertraline on aquatic  
86 organisms have been less studied. According to several studies comparing the effects of  
87 SSRIs on diverse species, sertraline is the most toxic, seemingly more potent on daphnia  
88 species than on fish (Christensen et al. 2007; Paterson and Metcalfe 2008; Minagh et al.  
89 2009).

90 The majority of studies on the impact of antidepressants within invertebrates have focused on  
91 reproduction and growth effects but few data sets are available on their behavioural effects  
92 (Fong 1998; Péry et al. 2008; Gust et al. 2009; Minagh et al. 2009; Campos et al. 2012b).  
93 Behavioural studies provide a link between physiological and ecological impacts, providing a  
94 major endpoint to assess population health and fitness (Craddock and Sklar 2013). Light is  
95 critical to a diverse range of behavioural and physiological processes such as diurnal rhythms,  
96 reproduction and predator avoidance (Henry et al. 2004). Indeed, light exposure regulates  
97 several neuro-modulatory systems; the activation of diverse photo-receptors modulates  
98 neurological components which in turn adjust behaviour. Serotonin, also named 5-  
99 hydroxytryptamine (5-HT), acts as a neurotransmitter or a hormone depending on its location  
100 and is a common modulator of animal behaviour in response to light. It is involved in many  
101 biological endpoints in invertebrates, such as growth, maturation, reproduction, visual  
102 perception and behaviour (Cezilly et al. 2000; Campos et al. 2012a). It has recently been  
103 demonstrated that exogenous serotonin and fluoxetine in amphipods increase phototaxis  
104 activity (Guler and Ford 2010). Acanthocephalan and trematode parasites can also act by  
105 increasing the serotonergic activity leading to an increase in phototaxis activity (Tierney et al.  
106 2004; De Lange et al. 2006; Guler and Ford 2010; Underwood et al. 2010) which can  
107 increase susceptibility to predation (Cezilly et al. 2000; Lagrue et al. 2007; Perrot-Minnot et  
108 al. 2007). Guler and Ford (2010) suggested that altered phototaxis behaviour in amphipods  
109 following SSRI exposure could then conceivably make them more prone to predation.

110 Gammarid amphipods are fundamental to many food chains and have an important role in  
111 ecosystem dynamics (Donner et al. 1994). Therefore, they have often been used in  
112 ecotoxicology studies, being considered as excellent bioindicators to monitor the health of

113 aquatic biotopes and the effects of anthropogenic contaminants (De Lange et al. 2006; Felten  
114 et al. 2008; De Lange et al. 2009; Guler and Ford 2010; Issartel et al. 2010).  
115 *Echinogammarus marinus* (Leach, 1815) is a ubiquitous intertidal marine amphipod which is  
116 widely found throughout the coasts of northwest Europe. The aim of this study was to  
117 develop behavioural biomarkers of SSRI antidepressants exposure and elucidate the  
118 molecular mechanism of action through components of the serotonin pathway.

119

## 120 **2. Materials and methods**

### 121 **2.1. Animals and exposure experiment**

122 *Echinogammarus marinus* were collected on the intertidal zone beneath seaweed and stones  
123 at low tide from Langstone Harbour, Portsmouth, UK (50°47'23.13N 1°02'37.25W). This  
124 area is used for light recreational sailing and is a Special Protection Area (SPA), Site of  
125 Special Scientific Interest (SSSI) and Special Area of Conservation (SAC) due to the use of  
126 expansive mudflats by wading birds. Animals were sorted and adult males with no visual sign  
127 of infection by trematodes (incorrectly reported as acanthocephalans by Guler and Ford,  
128 2010) were isolated. These parasitised individuals were excluded for their known impacts on  
129 host behaviour in response to light and modulation of serotonin in some species. The  
130 amphipods were kept individually in plastic containers filled with 80 mL of mechanical-  
131 filtrated natural seawater (from Langstone Harbour) at 10 °C under a 12 hrs light/12 hrs dark  
132 photoperiod and fed with fucoid seaweed.

133 After a week of acclimation, amphipods were exposed to the antidepressants fluoxetine and  
134 sertraline. In addition to an unexposed control group of 30 animals, groups of 15 animals  
135 were exposed to four nominal concentrations (0.001, 0.01, 0.1 and 1 µg/L) of each  
136 compound. Mortality was recorded and water was renewed every 3 d. Both fluoxetine (CAS  
137 no. 56296-78-7) and sertraline (CAS no. 79559-97-0) were obtained from Sigma-Aldrich®  
138 (St. Louis, MO, USA).

### 139 **2.2. Behavioural analysis**

140 Behavioural assays were performed after 1 hr, 1 d and 8 d of exposure to each condition  
141 using DanionVision™ (Noldus Information Technology, Wageningen, The Netherlands) and  
142 its software EthoVision® XT. Animals were put in 6-wells plates and placed within the  
143 DanioVision hardware for 2 min to allow settling prior to recording. The velocity (mm/s)

144 measurements of amphipods were recorded every 0.1 second (s) during 3 cycles of 2 minutes  
145 (min) dark and 2 min light, thus for a total period of 12 min (Fig. 1).

146 Due to the complexity of the dataset, an average velocity of every 10 s of the raw data were  
147 used to make heat maps for each condition by highlighting in green the 5th percentile, in  
148 black the 50th percentile and in red the 95th percentile (percentile calculated on the entire  
149 pool of data). This enabled a visual representation of periods when the amphipods were very  
150 active (red) or inactive (green). Statistical analyses were conducted using SPSS<sup>®</sup> Statistics  
151 v.20.0.0 software (IBM<sup>®</sup>) on the velocity of each amphipod during the 3 time periods (1 hr, 1  
152 d and 8 d) of behavioural assays. The data was normalised with a cube-root transformation  
153 and tested using a Kolmogorov-Smirnov test. Repeated Measure Analysis of Variance  
154 (ANOVA) with Dunnett multiple comparison tests was used to determine whether significant  
155 differences occurred over the 12 min recording period and between concentrations for both  
156 drugs. This enabled us to determine whether the velocity of the amphipods changed over the  
157 12 min dark/light regime, with SSRI concentrations or an interaction occurred between time  
158 and concentration. Within subject factors (time over the dark-light cycles and interactions  
159 between time and concentration) were tested using the Greenhouse-Geisser adjustments  
160 whereby sphericity of data is not assumed. All statistical analysis used a significance level of  
161  $p < 0.05$ .

### 162 **2.3. DNA/RNA isolation, purification and reverse transcription**

163 After the 8 d behavioural assays, animals were anaesthetised using a mixture of clove oil and  
164 seawater (0.2  $\mu\text{L}/\text{mL}$ ). The head of each amphipod was rapidly dissected, snap frozen in  
165 liquid nitrogen and stored in Tri Reagent<sup>®</sup> (Ambion<sup>®</sup>, Life Technologies, Carlsbad, CA,  
166 USA) at -80 °C before the extraction. DNA and RNA were extracted according to the  
167 manufacturer protocol and used for the infection screening and real-time PCR, respectively.  
168 After a DNase step using DNase I (RNase free) (New England Biolabs, Ipswich, MA,  
169 USA), RNA samples were cleaned on RNA clean and concentrator 5 columns (Zymo  
170 Research, Orange, CA, USA) following the manufacturer instructions. Quantification of total  
171 RNA and genomic DNA was performed with a NanoDrop<sup>®</sup> ND-100 Spectrophotometer  
172 (Nanodrop Technology Inc., Wilmington, DE, USA) and the integrity was checked using  
173 1.5% agarose gel electrophoresis. For each sample, 250 ng of total RNA isolated was used to  
174 obtain cDNA by reverse transcription using the GoScript<sup>™</sup> Reverse Transcription System

175 (Promega, Fitchburg, WI, USA) following the manufacturer protocol and using Oligo(dT)<sub>15</sub>  
176 primers and recombinant RNasin<sup>®</sup> Ribonuclease Inhibitor.

#### 177 **2.4. Infection screening**

178 The infection of *E. marinus* by parasites capable of inducing an increase in the serotonergic  
179 activity (Guler and Ford 2010) might interfere with the response of this species to SSRIs or  
180 create additional variation within controls. The *E. marinus* population used for this study has  
181 been comprehensively screened and found to contain a single trematode species capable of  
182 neurological modulation of its host (Yasmin Guler, unpublished data). Therefore, an infection  
183 screen using PCR was performed to enable the removal of amphipods infected by this  
184 trematode from the dataset used for the behavioural and transcriptomic analysis. PCR assays  
185 were conducted on genomic DNA using primers designed to amplify the Internal Transcribed  
186 Spacer (ITS) region of the ribosomal RNA gene for this trematode species (Yasmin Guler,  
187 unpublished data). To check the quality of DNA sample, amplification of the glyceraldehyde  
188 3-phosphate dehydrogenase (*Gapdh*) gene was used as a control (Table 1). PCR reactions  
189 were performed in a final volume of 25  $\mu$ L containing 1X GoTaq<sup>®</sup> Flexi Buffer, 2.3 mM of  
190 MgCl<sub>2</sub>, 0.8 mM of each dNTPs, 0.4  $\mu$ M of each primer, 1 U of GoTaq<sup>®</sup> DNA polymerase  
191 (Kit GoTaq<sup>®</sup> Flexi DNA polymerase, Promega) and 30 ng of genomic DNA. The PCR  
192 conditions were: initial denaturation at 95 °C for 2 min, followed by 40 cycles of 95 °C for 45  
193 s, 59 °C for 45 s 72 °C for 2 min and 20 s and a final incubation at 72 °C for 5 min. The PCR  
194 products were then analysed using agarose gel electrophoresis to check the presence of the  
195 amplified trematode ribosomal sequence.

#### 196 **2.5. Primer design and real-time PCR**

197 The transcriptome of *E. marinus* has recently been sequenced (unpublished data). The  
198 generated expressed sequence tags were assembled to create a "transcriptome atlas" of  
199 contiguous sequences (or contigs) and these contigs were annotated by comparison to non-  
200 redundant sequences in the UniProt and FlyBase database (BLASTX, E-value cut-off of 1e<sup>-5</sup>).  
201 The contigs chosen for primer design were selected using the following criteria: (i) selection  
202 of contigs that potentially represent genes involved in behaviour modulation, even if they  
203 were not well annotated (E-value > e<sup>-5</sup>), on the basis that such contigs might represent poorly  
204 conserved genes involved in neurological pathways; or (ii) selection of genes with a  
205 confident annotation (an E-value < e<sup>-5</sup>) potentially involved in serotonin or neurological  
206 pathways; (iii) genes that appeared to show exclusively high expression in the head, on the

207 basis that these are more likely to represent genes with neurological functions. A pooled  
208 sample of cDNA was used to test the suitability of each set of primers. All primers used in  
209 this study, including those used as reference genes *Gapdh* and *Calreticulin* (Table 1) were  
210 designed using Primer-3 software (Koressaar and Remm 2007; Untergrasser et al. 2012) and  
211 synthesised by Eurofins MWG Operon (Ebersberg, Germany).

212 Quantitative real-time PCR (qPCR) analyses were performed using a real-time PCR cyclor  
213 (Eco Illumina<sup>®</sup>, San Diego, CA, USA) on 12 samples per condition (or 3 pools of 4 head  
214 samples to test the primer pairs), using 7.5 µL of LabTAQ<sup>™</sup> Green (LabTech International  
215 Ltd, Uckfield, UK), 1 µL of cDNA, 5.7 µL of ultra-pure water, 0.2 µL each of *Rhod1* forward  
216 and reverse primers and 0.4 µL each for all other primers (all primer volumes taken from a 10  
217 µM stock). The PCR reactions were performed with an initial incubation at 95 °C for 2 min,  
218 followed by 45 cycles of 95 °C for 5 s and 60 °C for 30 s with Rox normalisation. Following  
219 the final cycle, the reactions underwent a 15 s, 95 °C denaturing step followed by a 15 s, 55  
220 °C hybridisation step before PCR product melt curves were determined during a further  
221 temperature increase to 95 °C. Standard curve analysis was used to determine the efficiency  
222 of each primer pairs and melt curve analysis were performed for each gene to confirm the  
223 specificity of the PCR product in each reaction. Ultra-pure water was used in the place of  
224 template in the no template control reactions. Furthermore, minus RT reactions were  
225 performed to control for the potential presence of residual genomic DNA. The control group  
226 of animals (that were not exposed to antidepressants) were used as the reference sample. The  
227 relative expression of each gene was calculated using the  $\Delta\Delta C_t$  method (Livak and  
228 Schmittgen 2001) and normalised with both *Gapdh* and *Calreticulin* as reference genes.  
229 Glyceraldehyde 3-phosphate dehydrogenase (*Gapdh*) is a reference (housekeeping) gene  
230 often used in several species (Barber et al. 2005) and particularly in crustaceans (Underwood  
231 et al. 2010; Leelatanawit et al. 2012). Statistical analyses were conducted on the square-  
232 rooted relative expression of each gene and results are expressed as the mean  $\pm$  standard  
233 deviation (s.d.). The normality was tested using a Kolmogorov-Smirnov test and multiple  
234 comparisons and comparison of two mean values were performed following ANOVA using  
235 the Dunnett's multiple comparison test using SPSS<sup>®</sup> Statistics and at a significance level of p  
236 < 0.05.

237

### 238 **3. Results**



239 The mortality was very low during all exposure experiment with only 2 dead amphipods for  
240 the sertraline exposure at 1 µg/L and 1 ng/L. No trematode-infected amphipods (as detected  
241 by visual inspection and retrospective PCR) were used in the experiments.

### 242 **3.1. Behavioural experiment**

243 The average of each amphipods' velocity during the 3 times 2 min dark/2 min light cycle for  
244 each condition after 8 d exposure are shown in Fig. 1 as an example of the dataset generated.  
245 Generally, when the light was switched on, the amphipods react and the velocity increases  
246 almost instantly for each condition and time of exposure with the velocity gradually abating  
247 after 30 s.

248 Multiple comparison tests (Tukey's Multiple comparison; data not shown) revealed  
249 significant differences ( $p < 0.001$ ) between 30 s time bins occurred overwhelmingly between  
250 the light and dark periods and were more pronounced for the 1<sup>st</sup> thirty seconds into the 2 min  
251 light cycles. This pattern was consistent for all concentrations, drugs and exposure periods.  
252 Interestingly, the 1<sup>st</sup> 30 s bin on the 1<sup>st</sup> of the 3 dark-light periods was also significantly  
253 different ( $p < 0.001$ ) from all other periods within the light apart from after 1 d for both drugs  
254 (Fig. 2A.). For both drugs and at all exposure times, there was a significant effect of the  
255 varying dark-light cycles over the 12 min on the amphipods velocity ( $p < 0.001$ ; Table 2; Fig.  
256 2 and 3).

257 For fluoxetine after 1 hr exposure, there was no significant effect of the different  
258 concentrations ( $p > 0.05$ ; Fig. 2 and Table 2) on velocity (mm/s) but there was a significant  
259 interaction between dark-light cycles and concentration ( $p < 0.05$ ). This interaction appears to  
260 have occurred due to a divergence in velocity between concentrations over the three dark-  
261 light cycle which can be observed on the graph (Fig. 2A.). After 1 d exposure, a significant  
262 difference in the velocity was observed between concentrations ( $p < 0.001$ ). Dunnett's 2-way  
263 multiple comparison tests revealed that significant differences occurred between the controls  
264 and all concentrations (0.001-1 µg/L:  $p < 0.01$ ) with the highest velocities generally observed  
265 in the concentration 0.1 µg/L (about 78% higher than the control) and the lowest increase at  
266 0.001 µg/L (about 43% higher than the control) (Fig. 2). Similarly, a significant interaction  
267 between concentrations and the dark-light cycles was observed ( $p < 0.005$ ). After the 8 d  
268 exposure no significant difference was observed in the velocity of amphipods between  
269 treatments ( $p > 0.05$ ) and the interaction tests failed to reach the significant cut-off ( $p =$   
270 0.098).

271 For sertraline, after just 1 hr exposure, significant differences were observed in the velocity  
272 between concentrations ( $p < 0.05$ ; Fig. 3 and Table 2). Velocities were elevated in all  
273 concentrations, apart from the lowest (0.001  $\mu\text{g/L}$ ), relative to the control. However,  
274 Dunnett's multiple comparison tests revealed that significant differences occurred only  
275 between the controls and the 0.01  $\mu\text{g/L}$  concentration (about 73% higher than control  
276 velocity;  $p = 0.002$ ; Fig. 3). A significant difference was also observed after 1 d of exposure  
277 to sertraline with all (apart from 0.001  $\mu\text{g/L}$ ; velocity higher of about 69, 55 and 33%  
278 respectively for 0.01, 0.1 and 1  $\mu\text{g/L}$ ) exposed groups recording higher average velocities  
279 compared to the control ( $p < 0.001$ ; Fig. 3). After 8 d exposure, no significant difference was  
280 observed in the velocity between exposed and control groups, with the associated p-value just  
281 failing to meet the significant criteria ( $p = 0.057$ ; Fig. 3A.). For all sertraline exposure times,  
282 no interaction was observed between concentration and time ( $p > 0.05$ ; Table 2).

### 283 **3.2. The expression levels of neurologically-related genes in amphipods** 284 **exposed to fluoxetine and sertraline**

285 The expression level of reference genes, *Gapdh* and *Calreticulin*, did not change for any  
286 concentrations of both fluoxetine and sertraline: for *Gapdh*:  $df = 8$ ,  $F = 1.380$  and  $p\text{-value} =$   
287  $0.239$ ; for *Calreticulin*:  $df = 8$ ,  $F = 1.648$  and  $p\text{-value} = 0.148$ .

288 RNA pooled from 12 individuals for each exposure group were used to test the suitability of  
289 each primers pair associated to a set of 10 potential neurological biomarker genes (7  
290 annotated with an E-value  $< e^{-5}$  and 3 unannotated). The serotonin receptor 1 (*5HT1*), the N-  
291 acetylserotonin O-methyltransferase-like protein (*Acser*), the inebriated neurotransmitter  
292 (*Ine1*) genes and contig 11430 presented very low expression, making it hard to determine  
293 expression from genomic contamination or the amplification of small amounts of artefact  
294 (results not shown). Those primer sets were then subsequently abandoned. For the two  
295 remaining unannotated genes (contig 9063 & 113810) the Ct value (the cycle number at  
296 which the fluorescent signal ( $\Delta R_n$ ) crossed an arbitrary threshold set within the linear phase  
297 of amplification) for both genes, was less than 22 cycles and no contamination by dimers or  
298 hairpin hybridisation was evident. However, despite this high expression, no variation in their  
299 expression with pooled cDNA was observed between each exposure (results not shown).  
300 Four sets of primers, [Neurocan core protein (*Neuc*), Rhodopsin (*Rhod1*), Arrestin (*Arr*) and  
301 tryptophan hydroxylase (*Ph*)] did present evidence of both high and altered expression using

302 the pooled cDNA and were therefore used to quantify the variation of gene expression among  
303 each condition (drug and concentration after 8 d exposure).

304 The mRNA expression levels of these four genes in the head of *E. marinus* exposed for 8 d to  
305 0, 0.001, 0.01, 0.1 and 1 µg/L of fluoxetine are illustrated in Fig. 4A. Significant differences  
306 were observed between expression of Neurocan core protein ( $F = 6.632$ ,  $df = 4$ ,  $p = 0.007$ ),  
307 Rhodopsin ( $F = 4.367$ ,  $df = 4$ ,  $p = 0.027$ ), and tryptophan hydroxylase ( $F = 3.917$ ,  $df = 4$ ,  $p =$   
308  $0.036$ ) but not for Arrestin ( $F = 1.313$ ,  $df = 4$ ,  $p = 0.330$ ). Where significant differences were  
309 observed, these were predominantly found to be down-regulated in treated samples when  
310 compared to the control group for the lower fluoxetine concentrations (Dunnett's Multiple  
311 Comparison  $p < 0.05$ ; Fig. 4A).

312 The mRNA expression levels of the four genes for 8 d exposure to sertraline are illustrated in  
313 Fig. 4B. Significant differences in expression were observed for Rhodopsin ( $F = 7.868$ ,  $df =$   
314  $4$ ,  $p = 0.004$ ) and Arrestin ( $F = 3.527$ ,  $df = 4$ ,  $p = 0.048$ ) but not for Neurocan core protein ( $F$   
315  $= 2.860$ ,  $df = 4$ ,  $p = 0.081$ ) and tryptophan hydroxylase ( $F = 2.137$ ,  $df = 4$ ,  $p = 0.151$ ).  
316 Multiple comparison tests found no significant differences from the control, although it is  
317 worth noting that the expression of Neurocan core protein just failed to meet the significance  
318 criteria for the lowest concentration (0.001 µg/L,  $p = 0.064$ ) as well as Rhodopsin for 0.1  
319 µg/L ( $p = 0.075$ ).

320

## 321 **4. Discussion**

### 322 **4.1. Effect of light on amphipod behaviour**

323 Amphipods naturally avoid well lighted areas and favour shadowed or dark regions in the  
324 intertidal zone where there is lower risk of predation (Cezilly et al. 2000). In this study, a  
325 significant increase in the velocity was observed in the first 30 s of light periods with a higher  
326 increase for the first of three light periods, at 1 hr and 8 d but not after 1 d of the beginning of  
327 the experiment. Sudden stimulation of the eyes could be interpreted by the amphipod as a  
328 reduction in cover and results in an escape-related behaviour in order to avoid predation. The  
329 decrease in the response to subsequent light periods indicates that the optic nerves may have  
330 been overstimulated and that a time of recovery from the first stimulation is needed. One day  
331 after the start of the experiment, the initial response to the light was reduced across all  
332 exposures indicating that more time may be necessary to recover.

#### 4.2. Effect of fluoxetine and sertraline on amphipod behaviour

The first purpose of this investigation was to assess the effect of two SSRIs on the swimming behaviour of the amphipod *E. marinus*. In this study, amphipods were exposed to concentrations from 0.001 to 1 µg/L of fluoxetine and sertraline, these concentrations fall well within those currently being found in the aquatic environment (0.929 µg/L and 0.087 µg/L respectively) (Brooks et al. 2003; Metcalfe et al. 2010; Styrishave et al. 2011; Silva et al. 2012). Interestingly, a significant interaction between the dark-light cycling and concentration was observed for fluoxetine at short-term (1 hr and 1 d). This interaction was due to a divergence in the response to light between the animals exposed to various concentrations of fluoxetine and demonstrates that these antidepressants have an effect on amphipod behaviour. There was a significant increase in the velocity over the 12 min time period at 1 d exposure to 0.1 µg/L of fluoxetine (of about 78%) compared to the control, which is consistent with the concentration used in the experiment to produce maximum phototaxis behaviour of this species exposed to fluoxetine (Guler and Ford 2010). Guler and Ford (2010) highlighted the non-monotonic concentration response curve, noting a peak of phototaxis activity in the animals exposed at 0.1 µg/L of fluoxetine. The lack of significant or reduced effects in higher concentrations of fluoxetine could be due to the inhibition of a finite amount of endogenous serotonin or desensitisation, as also suggested by Guler and Ford (2010). Amphipods exposed to 0.01 µg/L of sertraline showed a significant higher velocity than the control after 1 hr exposure (about 69%), as well as from 0.01 to 1 µg/L after 1 d. Sertraline's mode of action is similar to fluoxetine, both being SSRIs. The effect of sertraline was most prominent for the 0.01 µg/L concentration compared to the higher concentrations for which the velocity was lower. This suggests that as well as for fluoxetine, higher concentrations of sertraline might tend to more quickly reach a maximum level of serotonin re-uptake inhibition or lead to a desensitisation. The larvae of the fish, *P. promelas* has a suppression of predator avoidance after less than a week of exposure to 0.025 µg/L of fluoxetine (Painter et al. 2009), although no alteration of this behaviour was found at higher concentrations. In adults, a decrease of the predator avoidance behaviour has also been demonstrated when exposed at a concentration of 3 µg/L of sertraline for 28 d (Valenti et al. 2012). However, contrary to fluoxetine, the response to higher concentrations (10 and 30 µg/L) of sertraline was the same as for 3 µg/L in *P. promelas*.

The increased light-induced velocity of amphipods exposed to SSRIs is consistent with an increase of the serotonin amount. This study did not test the preference of the amphipods to

366 lit areas [although this was observed by Guler and Ford (2010)], but rather the velocity of  
367 shrimp within light or dark environments. The most consistent results from this experiment  
368 indicated that amphipods are significantly more active both in light and dark phases of the  
369 experiment (with some interactions between light and concentration observed) when exposed  
370 to SSRIs as compared to untreated amphipods. Furthermore the recovery time (time to return  
371 to the basal velocity level) to light stimulation is altered between exposures and control. It is  
372 possible that the increased activity could also be due to the influence of serotonin on other  
373 hormones [e.g. Crustacean Hyperglycemic Hormone, CHH; (Fingerman 1997)] and/or  
374 locomotor activity (McPhee and Wilkens 1989). However, changes in the transcription of  
375 genes relating to phototransduction pathways measured during study add some weight for  
376 linking the behavioural and gene responses. It will be beneficial in future studies to lengthen  
377 the periods of light and dark to differentiate the behaviours further.

378 Studies investigating the effect of SSRIs on aquatic organisms have been mainly performed  
379 using concentrations higher than those found in the environment and used in this study.  
380 Impacts of fluoxetine on the reproduction of *C. dubia* were observed at 56 µg/L with a  
381 decrease of fecundity (Brooks et al. 2003), and around 10 µg/L in *D. magna* (Péry et al.  
382 2008). The acute toxicity of sertraline on animals has been demonstrated with a LC50 of 380  
383 µg/L in fish following 96 hr of exposure (Minagh et al. 2009) and change in the behaviour of  
384 fish was found from 3 µg/L (Valenti et al. 2012). Relatively few studies have been carried out  
385 using environmentally relevant concentration of SSRIs (Painter et al. 2009; Guler and Ford  
386 2010; Fong and Hoy 2012). However, the current study has found significant impacts as low  
387 as 0.001 µg/L that fall well within concentrations considered environmental relevant in the  
388 aquatic environment close to wastewater effluent and inhabited by this species (about 0.0026  
389 µg/L in US estuaries, Paint et al. 2006). Furthermore, the degree of degradability of these  
390 antidepressants in water is generally low and their half-lives is from 2 days to indefinite  
391 (Johnson et al. 2005; Kwon and Armbrust 2006). The benthos is a reservoir for these  
392 compounds as they tend to be absorbed by sediments or sludge (Kwon and Armbrust 2006).  
393 The amount of SSRIs in this compartment should also be investigated in order to better  
394 evaluate the effects of antidepressants on amphipods. In this study, fluctuations in fluoxetine  
395 and sertraline concentrations might have occurred due to the static renewal of water every 2  
396 days and the potential binding to the exposure chamber. Furthermore, insignificant results  
397 from the lower concentration range need to be carefully interpreted in light of the nominal  
398 concentrations used and the potential for chemical breakdown.

399 The presence of antidepressants in the environment can be chronic due to a constant release  
400 from the sewage water (Santos et al. 2010), thus a long-term analysis is essential to truly  
401 understand the effect of prolonged exposure times on aquatic organisms. Our results  
402 indicated that the most enhanced effects of fluoxetine and sertraline were observed following  
403 short-term exposure, after 1 hr (sertraline only) and 1 d of exposure. Although, contrary to  
404 this, Guler and Ford (2010) found a significant and continued preference of lit arenas still  
405 after 3 weeks exposure to fluoxetine at 0.1 µg/L compared to controls. As suggested by our  
406 higher concentrations of SSRIs in this study, a longer term exposure might lead to a  
407 desensitisation effect or a lack of serotonin availability and explain why no significant effect  
408 of both drugs was found after 8 d exposure. In mammals, it has been shown that the  
409 responsiveness to fluoxetine decreases following chronic exposure due to a critical decrease  
410 in the tryptophan levels, the precursor of serotonin (Delgado et al. 1999). Therefore, after  
411 several days of exposure to SSRIs, the haemolymph tryptophan content might be nearly  
412 depleted, reducing the drug effect on amphipods. Another hypothesis could be a negative  
413 feedback loop in the serotonin pathways; amphipods might be compensating for the change  
414 by producing less serotonin to flood the synapse or by increasing the expression of serotonin  
415 re-uptake transporter (Pineyro et al. 1994). It would then be interesting to compare the impact  
416 of these drugs on the serotonin pathway at short-term and long-term in further research.

#### 417 **4.3. Effect of fluoxetine and sertraline on amphipod gene transcription**

418 The second aim of this study was to elucidate the molecular mechanism by which  
419 behavioural changes may be taking place. The absence of variations in *Calreticulin* and  
420 *Gapdh* expression supports their utilisation as reference genes.

421 Rhodopsin (*Rhod1*) is involved in behaviour regulation and is a light receptor and signal for  
422 phototransduction in vertebrates and invertebrates (Orem and Dolph 2002). In invertebrates,  
423 phototransduction cascade is mediated by rhodopsin, a light receptor which is transformed  
424 into metarhodopsin by photo-isomerisation (Orem and Dolph 2002). The metarhodopsin  
425 activates a Gαq-type of G-protein, hydrolysing guanosine triphosphate (GTP) to guanosine  
426 diphosphate (GDP), which then activates a phospholipase C (PLC). Finally, the PLC opens a  
427 transient receptor potential (TRP) channels which induce a depolarisation of the cells. *Rhod1*  
428 was significantly down-regulated in amphipods exposed to low concentrations of fluoxetine  
429 (0.001 and 0.01 µg/L) and slightly up-regulated for those exposed to 0.001 µg/L of sertraline.  
430 One explanation for the opposite gene expression patterns observed in *E. marinus* when

431 exposed to these two antidepressants may be the differences in their mode of action.  
432 Therefore, one might speculate that the mis-regulation of *Rhod1* could then modulate the  
433 transduction of light stimulation and alter the behaviour of amphipods to light. However,  
434 further studies will be necessary to better understand the role of rhodopsin in modulating  
435 amphipod behaviour. The protein encoded by the arrestin (*Arr*) gene is also involved in the  
436 phototransduction. In fact, this gene contributes to the arrest of the phototransduction cascade  
437 (Kashiyama et al. 2010) by binding the active metarhodopsin and inhibits it by uncoupling  
438 rhodopsin from the  $G\alpha$ -subunit protein (Orem and Dolph 2002). An example of their role in  
439 crustacean is that arrestin and rhodopsin promote light-induced hatching in *Triops granarius*  
440 (Kashiyama et al. 2010). In our study, *Arr* is down-regulated only in animals exposed to  
441 0.001 and 0.01  $\mu\text{g/L}$  of fluoxetine, which could be potentially linked to the down-regulation  
442 of *Rhod1* and components of the phototransduction pathway if followed by a protein down-  
443 regulation.

444 The neurocan core protein (*Neuc*) is a protein involved in cell adhesion and migration and is a  
445 factor in bipolar disorder, manic-depressive disorder and schizophrenia (Cichon et al. 2011;  
446 Mühleisen et al. 2012). In our study, *Neuc* mRNA expression significantly decreases for the  
447 two lower concentrations of fluoxetine. Assuming a similar function of *Neuc* in amphipods  
448 and mammals (Livak and Schmittgen 2001), a decrease in the expression of this gene if  
449 followed by a decrease in amount of its protein might lead to behavioural changes. It might  
450 then induce an increase of energy (Livak and Schmittgen 2001) which might tend to reduce  
451 the predator avoidance behaviour. The role of this gene should be investigated in further  
452 studies to define its function in amphipods. The enzyme tryptophan hydroxylase (*Ph*)  
453 catalyses serotonin biosynthesis in the serotonergic nerves (Hasegawa and Nakamura 2010).  
454 However, no significant variation in the expression of this gene has been found between each  
455 condition, which suggests that this gene is not involved in the serotonin regulation inducing  
456 the behavioural change observed when exposed to antidepressants.

457 The expression variations of these four genes were relatively low in *E. marinus* and it is  
458 unclear what impacts their down-regulation may have on amphipod behaviour. There is  
459 paucity of research regarding the molecular processes that underlie serotonin pathways and  
460 behavioural regulation in aquatic invertebrates. Further studies are essential in order to better  
461 understand the role of these genes in crustaceans and their relationship to the behaviour  
462 modification observed following antidepressant exposure. However, this study clearly  
463 demonstrates that exposure to SSRIs can be associated with alteration in the expression of

464 genes with plausible links to amphipod behaviour and serotonergic activity. Recently, it has  
465 been demonstrated in the crustacean *Gammarus pulex* that the histaminergic system is  
466 involved in the reaction to light in association with the serotonergic system (Perrot-Minnot et  
467 al. 2013). In the same study, looking at the influence of several 5-HT receptor antagonists and  
468 agonists, it has been suggested that the serotonin receptor 5-HTR2 subtype might be involved  
469 in the behaviour regulation of *G. pulex*. Furthermore, in *D. magna*, a transcriptomic analyses  
470 using a custom microarray showed that more than 1200 genes have a mRNA expression  
471 change when exposed to fluoxetine (Campos et al. 2013). Serotonin metabolism, neuronal  
472 development processes, carbohydrate and lipid metabolism functions were found to be  
473 differentially expressed when annotated by comparison to the functionally annotated  
474 *Drosophila* genome.

475

#### 476 **4.4. Summary**

477 This study has provided evidence that a crustacean's behaviour and gene expression could be  
478 abnormally altered in waters receiving antidepressants at concentrations as low as 0.001  
479 µg/L. The use of behavioural analysis has been demonstrated as good biomarker of the  
480 exposure of amphipods to antidepressants. The transcriptome of *E. marinus* is a rich resource  
481 for neurological genes that are potentially involved in behavioural regulation and serotonin  
482 related pathways. Therefore, future studies will be able to test an expanding number of  
483 amphipod genes for transcriptional change following exposure to antidepressants. This study  
484 has also provided further evidence for the non-monotonic concentration responses of some  
485 antidepressants, which should be taken into account when designing and evaluating toxicity  
486 tests. Whether other biological systems, for example: reproduction, moulting, metabolism and  
487 the immune system are impacted following low SSRIs exposure remains an important  
488 unanswered question. The effect of other SSRIs and their metabolites (Brooks et al. 2003;  
489 Stanley et al. 2007; Paterson and Metcalfe 2008; Metcalfe et al. 2010) on amphipods should  
490 also be evaluated along with other types of antidepressant such as the serotonin-  
491 norepinephrine re-uptake inhibitors (SNRIs) and the serotonin antagonist and re-uptake  
492 inhibitors (SARIs). The use of other types of antidepressants increases every year, with an  
493 increase of about 60 % for the SNRI duloxetine the last two years (HSCIC and Prescribing  
494 and Primary Care Services 2013). Considering that the mode of action for these other types of  
495 antidepressants is different from the SSRIs, it is important to also determine their potential



496 impact on aquatic organisms. How multiple antidepressants, with multiple modes of action,  
497 will act in mixtures is another challenge faced by ecotoxicologists. For example, it has been  
498 demonstrated that mixtures of antidepressants have additive effects in aquatic organisms  
499 (Christensen et al. 2007; Styris have et al. 2011) and leads to a decrease in the predation  
500 avoidance behaviour in the larvae of the fish *P. promelas* (Styris have et al. 2011). The  
501 organismal and ecological implications of these findings are difficult to deduce but coupled  
502 with previous studies suggest that SSRIs present in the aquatic environment could  
503 conceivably lead to population level effects through impacts on predation, feeding and  
504 reproductive associated behaviour.

505

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513

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717 **Table and Figures captions**

718 **Table 1.** Primer sequences used in this study and target genes associate. The primers couple  
 719 for serotonin receptor 1 have been design on alignment of several invertebrates' sequences of  
 720 this gene and in very conserved area. *Italic*: the reference genes used to normalised the gene  
 721 expression; † Four set of primers found relevant for quantification; \* Target gene unknown,  
 722 no annotation: E-value > e<sup>-5</sup>.

Primer Name	Nucleotide sequences (from 5' to 3')	Target Gene	Uniprot or GenBank ID	Ref. Species	E-value
5HT1-F	CAA CGC AGA GTA CGG GGT TGG T	Serotonin receptor 1			
5HT1-R	GCA AAA CGG CGA AAT CGA ACG GG				
Acser-F	AAA CCC ACA AAC GAC GAC CA	N-acetylserotonin O-methyltransferase-like protein	O95671	<i>Homo sapiens</i>	7E-25
Acser-R	AAG GTT ACT CTC TGC CAC GC				
Arr-F	CTC CTT CGA CTC CAG GCT TG	Arrestin <sup>†</sup>	P32122	<i>Locusta migratoria</i>	5.00E-50
Arr-R	GGC TAA CCT GGG CAT CAA CA				
<i>Calret-F</i>	<i>AGA TCG GAG GCA TTG TTT TG</i>	<i>Calreticuline</i>	Q7Z1E6	<i>Bombyx mori</i>	1.00E-155
<i>Calret-R</i>	<i>AAC ACG TGG GCC GAG TAT AG</i>				
<i>Gapdh-F</i>	<i>ATA GTG TCC AAC GCC TCC TG</i>	<i>GAPDH</i>	P56649	<i>Panulirus versicolor</i>	1E-164
<i>Gapdh-R</i>	<i>CCA GTG GAG GAT GGA ATG AT</i>				
Ine1-F	CGT GGA GGA GCC GTT GCC TG	Inebriated neurotransmitter	NM057664.5	<i>Culex quinquefasciatus</i>	4.00E-05
Ine1-R	CCT GTG CGG CAT CCC TCT GC				
Neuc-F	CCC TAC CCT GTT TGC TCC AG	Neurocan core protein <sup>†</sup>	P55066	<i>Mus musculus</i>	7.00E-19
Neuc-R	CCA TTT TGG TAG TTC GCG GC				
Ph-F	GGT CAA GAC CTG GAG CGC GG	Tryptophan hydroxylase <sup>†</sup>	AY099427.1	<i>Aedes aegypti</i>	6.00E-142
Ph-R	GGT GCT GTG GAA CAC GCG GA				
Rhod1-F	CCC GCC AAC ATG CTG CCT GA	Rhodopsin <sup>†</sup>	DQ85259	<i>Neomysis americana</i>	4.00E-74
Rhod1-R	CGG GTG ACC GCA GGC TCT TG				
9063-F	TCA TCGACG AAC TTG GAG CC				*
9063-R	TCA TTG GCC TCT AGA AGC GC				
11381-F	TTC CGA ACT AAC GCC TGC TC				*
11381-R	CCA ACA GTG CAG CAA CAT CG				
11430-F	GTG AGG AGG AGG TGT GGG TA				*
11430-R	GGT ACA GGC GAG ACA ACA GG				

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726 **Table 2.** Results of statistical analyses of velocity tracking during the 12 min of 2 min dark/2  
 727 min light periods in *Echinogammarus marinus* exposed to each concentrations of fluoxetine  
 728 and sertraline for each time of exposure.

Compound	Exposure Period	Concentration			Time (Light-Dark Cycles)			Interaction: time* concentration		
		F	df	p	F	df	p	F	df	p
Fluoxetine	1 Hour	0.585	4	0.675	27.335	12.35	<b>&lt;0.001</b>	1.482	49.412	<b>0.018</b>
	1 Day	7.199	4	<b>&lt;0.001</b>	14.148	23	<b>&lt;0.001</b>	1.694	53.017	<b>0.002</b>
	8 Days	1.087	4	0.368	13.787	23	<b>&lt;0.001</b>	1.311	39.437	0.098
Sertraline	1 Hour	3.719	4	<b>0.008</b>	14.878	23	<b>&lt;0.001</b>	1.061	53.725	0.358
	1 Day	7.966	4	<b>&lt;0.001</b>	14.341	23	<b>&lt;0.001</b>	1.307	43.656	0.407
	8 Days	2.373	4	<0.057	15.451	23	<b>&lt;0.001</b>	1.321	46.337	0.076

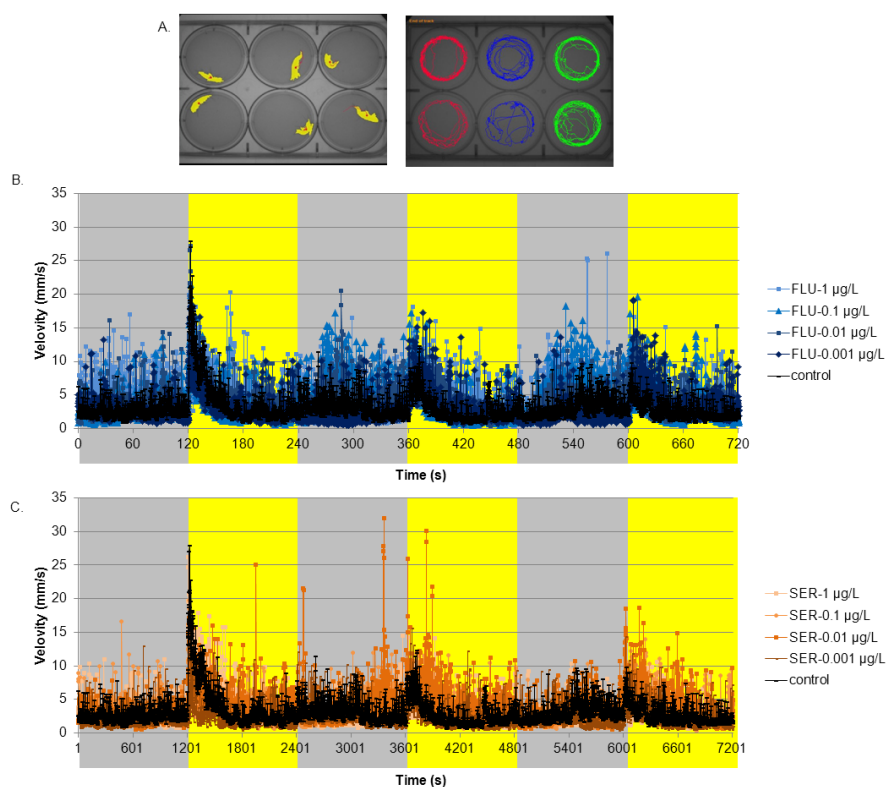
729 F: ratio of the between and within group variance estimates; df: degrees of freedom; p: p-  
 730 value, in bold when significant.

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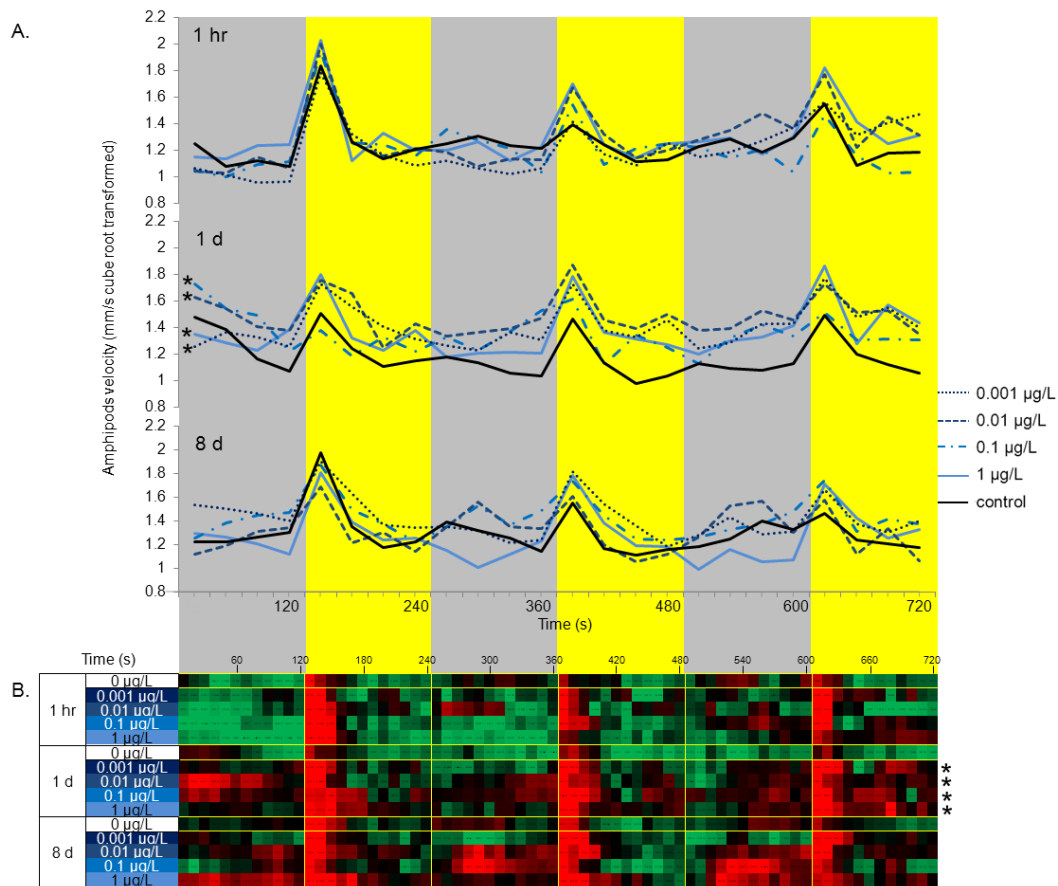
733 **Fig. 1.** Mean velocity (mm/s) of 15 *Echinogammarus marinus* per treatment exposed to  
734 fluoxetine and sertraline for 8 d recorded with DanioVision. 6-wells plates were used to track  
735 the velocity of 6 amphipods at a time every 0.1 s over a 12 min period of alternate 2 min  
736 dark/2 min light periods (A). Lines indicate mean values of replicates specimens. Black:  
737 control, gradation of blue: fluoxetine (FLU) concentrations (B), gradation of orange:  
738 sertraline (SER) concentrations (C).



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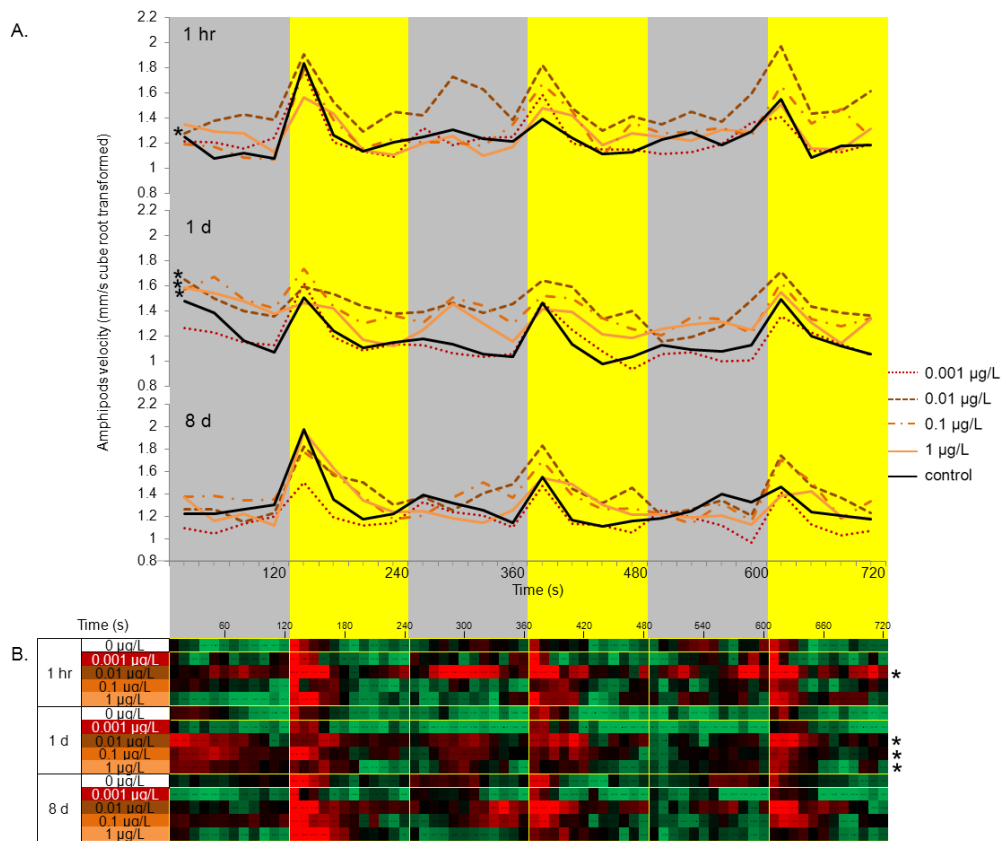
741 **Fig. 2.** Estimated marginal means (A) and heat map (B) of the velocity (mm/s) average every  
742 10 s during the 12 min of 2 min dark/2 min light periods for each fluoxetine concentrations  
743 and time exposure. Heat map: green: the 5th percentile, black: the 50th percentile and red: the  
744 95th percentile. Hr: hour, d: day(s). Asterisks indicate significant differences to the control ( $p$   
745  $< 0.05$ ).



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748 **Fig. 3.** Estimated marginal means (A) and heat map (B) of the velocity (mm/s) average every  
 749 10 s during the 12 min of 2 min dark/2 min light periods for each sertraline concentrations  
 750 and time exposure. Heat map: green: the 5th percentile, black: the 50th percentile and red: the  
 751 95th percentile. Hr: hour, d: day(s). Asterisks indicate significant differences to the control (p  
 752 < 0.05).

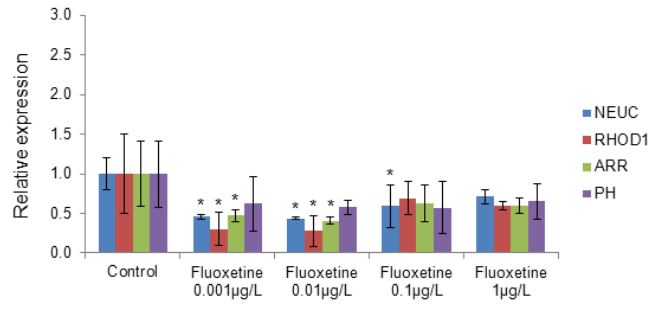


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755 **Fig. 4.** Relative expression of Neurocan core protein (*Neuc*), Rhodopsin (*Rhod1*), Arrestin  
 756 (*Arr*) and tryptophan hydroxylase (*Ph*) mRNA in the head of *Echinogammarus marinus*  
 757 exposed to four fluoxetine (A) and sertraline (B) concentrations for 8 d. The expression was  
 758 normalised according to the expression of *Gapdh* and *Calreticulin*. n = 3 pools of 4  
 759 amphipods. Data are expressed as the mean ± s.d. Asterisks indicate significant differences to  
 760 the control (p < 0.05).

### A. Fluoxetine



### B. Sertraline

