

University of Warwick institutional repository: <http://go.warwick.ac.uk/wrap>

This paper is made available online in accordance with publisher policies. Please scroll down to view the document itself. Please refer to the repository record for this item and our policy information available from the repository home page for further information.

To see the final version of this paper please visit the publisher's website. Access to the published version may require a subscription.

Author(s): CR Rahn, K Zhang, R Lillywhite, C Ramos, J Doltra, J M de Paz, H Riley, M Fink, C Nendel, K Thorup Kristensen, A, Pedersen, F Piro, A Venezia, C Firth, U Schmutz, F Rayns, K Strohmeyer  
Article Title: EU-Rotate\_N – a Decision Support System – to Predict Environmental and Economic Consequences of the Management of Nitrogen Fertiliser in Crop Rotations

Year of publication: 2010  
Link to published version:  
<http://www.ulmer.de/>  
Publisher statement: None

1 **EUROPEAN JOURNAL OF HORTICULTURE.**

2 **ACCEPTED FOR PUBLICATION 13 JULY 2009**

3  
4 **EU-Rotate\_N - A Decision Support System To Predict Environmental**  
5 **And Economic Consequences Of The Management Of Nitrogen**  
6 **Fertiliser In Crop Rotations**

7 CR Rahn<sup>1</sup>, K Zhang<sup>1</sup>, R Lillywhite<sup>1</sup>, C Ramos<sup>2</sup>, J Doltra<sup>2</sup>, J M de Paz<sup>3</sup>, H Riley<sup>4</sup>, M Fink<sup>5</sup>, C  
8 Nendel<sup>5,6</sup>, K Thorup Kristensen<sup>7</sup>, A. Pedersen<sup>7</sup>, F Piro<sup>8</sup>, A Venezia<sup>8</sup>, C Firth<sup>9</sup>, U Schmutz<sup>9</sup>, F  
9 Rayns<sup>9</sup>, K Strohmeyer<sup>10</sup>.

10 <sup>1</sup>)Warwick HRI, University of Warwick, Wellesbourne, Warwick, CV35 9EF, England.

11 clive.rahn@warwick.ac.uk <sup>2</sup>)Instituto Valenciano de Investigaciones Agrarias (IVIA), Ctra.

12 Moncada - Apdo. Oficial, 46113 Moncada, Spain. <sup>3</sup>) Consejo Superior de Investigaciones

13 Científicas (CSIC). Apdo. Oficial, 46470 Albal/Valencia, Spain. <sup>4</sup>) Bioforsk, Apelsvoll Research

14 Centre, 2849 Kapp, Norway. <sup>5</sup>) Institut für Gemüse- und Zierpflanzenbau (IGZ), Theodor-

15 Echtermeyer-Weg 1, 14979 Großbeeren, Germany. <sup>6</sup>) present address: Leibniz-Centre for

16 Agricultural Landscape Research, Institute for Landscape System Analysis, Eberswalder Straße

17 84, 15374 Müncheberg, Germany. <sup>7</sup>) Department of Horticulture, University of Aarhus, Research

18 Centre Aarslev, Kirstinebjergvej 10, 5792 Aarslev, Denmark. <sup>8</sup>) Consiglio per la ricerca e la

19 sperimentazione in agricoltura - Istituto sperimentale per l'orticoltura (CRA-ISOR). Via dei

20 Cavalleggeri 25, Casella Postale 48, 84098 Pontecagnano, Italy. <sup>9</sup>) Henry Doubleday Research

21 Association (HDRA), Ryton Organic Gardens, Coventry, CV8 3LG, England. <sup>10</sup>) BOLAP GmbH,

22 Obere Langgasse 40, 67346 Speyer, Germany.

23

24

25 **Summary**

26 A model has been developed which assesses the economic and environmental performance of  
27 crop rotations, in both conventional and organic cropping, for over 70 arable and horticultural  
28 crops, and a wide range of growing conditions in Europe. The model, though originally based on  
29 the N\_ABLE model, has been completely rewritten and contains new routines to simulate root  
30 development, the mineralisation and release of nitrogen (N) from soil organic matter and crop  
31 residues, and water dynamics in soil. New routines have been added to estimate the effects of  
32 sub-optimal rates of N and spacing on the marketable outputs and gross margins. The model  
33 provides a mechanism for generating scenarios to represent a range of differing crop and fertiliser  
34 management strategies which can be used to evaluate their effects on yield, gross margin and  
35 losses of nitrogen through leaching. Such testing has revealed that nitrogen management can be  
36 improved and that there is potential to increase gross margins whilst reducing nitrogen losses..

37

38 Key Words: vegetables, organic production, N-fertiliser management, nitrate leaching,  
39 modelling, gross margins.

40

## 41 **Introduction**

42 Large amounts of nitrogen are applied to intensively cultivated land, especially where field  
43 vegetables are grown. DEMYTTENAERE ET AL. (1990) and GOULDING (2000) showed that growing  
44 field vegetable crops can lead to large amounts of potentially leachable nitrate being left in the soil  
45 after harvest. Since the value of the produce is high in comparison to the cost of additional fertiliser,  
46 the temptation to over-fertilise is high, leading to greater risks of nitrate pollution. Increasing  
47 environmental concerns about high nitrate levels in drinking water from such intensive land use now  
48 demands effective systems of fertiliser recommendation.

49 NEETESON and CARTON (2001) reviewed the multiple pathways by which nitrogen applied to field  
50 vegetable crops could pollute the environment. Many EU directives and national regulations are  
51 now in place, which seek to regulate the use of fertilisers. Many of these were identified by an  
52 EU concerted action, the NUMALEC project (DE CLERCQ et al. 2001).

53 In some countries, supermarkets are demanding that the produce they sell has been grown  
54 according to environmentally sound practices and have introduced assurance schemes as a result.  
55 Model based decision support systems can be valuable tools for consultants and farmers to help  
56 meet these increasingly tight standards and regulations.

57 Two existing decision support models: N Expert (FINK and SCHARPF 1993) and WELL\_N (RAHN  
58 et al. 1996) are available to supply fertiliser advice for field vegetable production in Germany and  
59 the UK respectively. WELL\_N is based on routines in the N\_ABLE model (GREENWOOD 2001).  
60 The N\_ABLE model, however, only operates on single season crops and RAHN et al. (1992,  
61 1998) demonstrated that crops can be more effectively fertilised if N fertiliser is managed over  
62 whole crop rotations.

63 A new model, EU-Rotate\_N, was developed, with EU funding, as a tool for assessing the effects  
64 of different fertiliser and rotational practices on losses of nitrogen to the environment and gross  
65 margin returns across Europe. This paper describes the model, its validation using a German  
66 dataset, and demonstrates its use in examining the effects of different agricultural practices under  
67 Norwegian conditions.

68

## 69 **Materials and Methods**

70  
71 The model consists of a number of modules which simulate: plant growth both below and above  
72 ground, nitrogen mineralisation from the soil and crop residues and subsequent N uptake. These  
73 processes are regulated by weather factors such as rainfall, temperature and radiation. Modules  
74 simulate the flow of water and nitrogen in the soil, into the plant and subsequent  
75 evapotranspiration or leaching. The modules operate on a daily basis, utilising data from soil  
76 properties, crop residues, fertiliser and weather data where appropriate (Figure 1). The model can  
77 simulate any number of crops in the rotation with a maximum limit of 30 years.

78

79 *<Figure 1>: The organisation of the main model modules.*

80

### 81 **Description of the soil**

82 In the model, soil is divided into 40 vertical layers of 0.05m thickness. After planting, these  
83 layers are split horizontally into 0.05 m wide cells. The number of cells horizontally depends on  
84 row width. When the crop is harvested or the residues are incorporated the horizontal cells are  
85 merged into one unit until the next crop is planted. Describing the soil in this way allows for  
86 more accurate simulation of root growth of row crops compared to the original N\_ABLE model.  
87 While the crop is growing all the processes described below are simulated at the cell level.

88

89 The basic properties of the soil layers are provided by the user of the model and include the water  
90 content at permanent wilting point, field capacity, and at saturation. These hydraulic properties  
91 control water availability to the plant and allow calculation of drainage. Mineralisation and losses  
92 of nitrogen by denitrification are adjusted for water content. Other inputs include pH, which  
93 allows for simulation of N losses where urea fertilisers are used, and the organic matter content of  
94 the soil, which affects the supply of N from mineralisation. The clay and sand content is used to  
95 calculate denitrification, hydrolysis of urea, and ammonia volatilisation from the top layer.

96

### 97 **The water module**

98 Crop evapotranspiration is calculated using the FAO approach (ALLEN et al. 1998). The main  
99 parameters are those related to the evaporative demand of the atmosphere, summarized by the  
100 reference evapotranspiration ( $ET_0$ ) and a crop coefficient that varies with crop development.

101  
102 The effects of water stress on plant growth are considered and it is assumed that the reduction in  
103 dry matter accumulation due to water deficit is proportional to the transpiration reduction (HANKS  
104 1983; SHANI and DUDLEY 2001).

105  
106 Water infiltration and redistribution in the soil follow a capacitance approach similar to the one in  
107 the N\_ABLE model, but this has been modified using a drainage coefficient that allows the water  
108 transfer between layers above field capacity to be controlled progressively (in more than one day)  
109 and more or less rapidly depending on soil type (RITCHIE 1998). Drainage at any depth is given  
110 as the downward water flow from the cell elements at this depth. The module also accounts for  
111 two-dimensional capillary flow by adopting a soil water normalised diffusion approach (ROSE  
112 1968; RITCHIE 1998). The main parameters that define the hydraulic soil properties, such as the  
113 water content at field capacity and wilting point, are input by the user for the different soil layers.  
114 Values can be estimated from soil texture when not available. (SAXTON et al. 1986)

115  
116 Runoff is calculated using the approach by the U.S. National Resource Conservation Service  
117 (NRCS, formerly the Soil Conservation Service) based on studies of small agricultural  
118 watersheds (< 800 ha) across the United States. (NRCS 2004)

119  
120 ***Mineralisation module***

121 The calculation of N mineralisation from organic matter is based on the routines used in the  
122 DAISY model (HANSEN et al. 1990). Carbon dynamics in the soil are described by three pairs  
123 (slow or rapid decomposition) of conceptual pools (soil organic matter, soil microbial biomass  
124 and added organic matter). Decomposition rate coefficients are temperature and moisture  
125 dependent and reflect the environmental conditions of the simulated site; decay and respiration  
126 rates of soil microbial biomass are additionally influenced by soil clay content. Efficiency  
127 parameters determine the loss of  $CO_2$  during the single turnover processes. N release as  $NH_4^+$  is a  
128 consequence of C lost as  $CO_2$  from the system that maintains fixed C to N ratios in the different

129 pools. Processes of nitrification and denitrification are implemented to complete the turnover  
130 model.

131 Residues of crops simulated with the crop growth module enter the mineralisation routine with a  
132 dynamic C to N ratio determined by crop N content, harvest index and a factor determining the N  
133 content in crop residues relative to the N content in the harvested crop parts, which reflects the  
134 growth conditions of the crop during the season with respect to N supply. A fixed C to N ratio is  
135 assigned to the slow decomposing part of the material whereas the C to N ratio of the fast  
136 decomposable part will then vary depending on total N content in the plant material.  
137 Decomposition rate coefficients of both pools are also fixed (ABRAHAMSEN and HANSEN 2000).  
138 C to N ratios and partitioning coefficients for crop residues are derived from stepwise chemical  
139 digestion experiments (JENSEN et al. 2005). Parameters for the release of N from manure and  
140 slurry were taken from the DAISY model (ABRAHAMSEN and HANSEN 2000).

141 N volatilisation from applied manures and slurries are described using an empirical relation  
142 implemented in the ALFAM model (SØGAARD et al. 2002). A soil pH dependency factor was  
143 introduced by fitting data from HE et al. (1999) to Michaelis-Menten kinetics and subsequently  
144 normalising the relation between pH and volatilisation half-life time to pH 7.0.

145 Hydrolysis of, and gaseous N loss from, applied urea fertiliser is calculated based on routines of  
146 the AMOVOL model (SADEGHI et al. 1988), taking into account the temperature dependent  
147 equilibrium between the ammonium ions in solution and gaseous ammonia, as well as the effect  
148 of soil organic matter, soil temperature and soil water content on the hydrolysis process itself. An  
149 atmospheric resistance parameter finally governs the loss of gaseous ammonia from the top soil.

150

### 151 ***Snow and frost module***

152 The original snow model, developed at the University of Helsinki by VEHVILÄINEN and  
153 LOHVANSUU (1991), was used to calculate water equivalent, but modified by KARVONEN (2003)  
154 to calculate snow depth, which is important for determining soil freezing and thawing. This has  
155 been further modified and calibrated by iterative simulation using a 10-year dataset from  
156 Norway, as described by RILEY and BONESMO (2005). The approach has been validated with  
157 independent data.

158

159 The soil frost module is based on two approaches, one for freezing and one for thawing. The  
160 approach for soil freezing was proposed by OLSEN and HAUGEN (1997) and assumed uniform  
161 thermal properties throughout the profile; values are taken from the SOIL model (JANSSON 1991).  
162 The module requires input of surface temperature as modified by the snow pack. The approach  
163 used for thawing is that of the ECOMAG model (MOTOVILOV et al. 1999). Both freezing and  
164 thawing processes have been validated for Norwegian conditions.

165  
166 The snow and frost calculation routines affect water infiltration and associated processes such as  
167 leaching. In brief, it is assumed that infiltration ceases when soil freezes. If the soil surface is  
168 frozen, it is assumed that precipitation is either stored in the snow pack, if present, or it is lost to  
169 surface runoff. During snowmelt and soil thaw, an amount of melt-water equal to the difference  
170 between field capacity and saturation is stored for later infiltration, whilst the remainder passes to  
171 surface runoff. When complete soil thaw occurs the stored melt-water passes through the profile.

172

### 173 **Root module**

174 The calculations in the root module consist of three main parts: i) first the physical extension of  
175 the root system is calculated, ii) then the total root length of the crop is calculated, and iii) finally  
176 the distribution of the root system with depth and distance from the crop row is calculated. The  
177 root module has been described and tested in PEDERSEN et al. (2009).

178

179 The depth development of the root system ( $r_z$ ) is calculated from the accumulated temperature  
180 sum ( $T_{\text{cumul}}$ ) from crop planting. After a lag period ( $ddg_{\text{lag}}$ ) the rooting depth increases linearly  
181 with accumulated temperature sum from its starting value ( $z_{\text{start}}$ ), using the crop specific rooting  
182 depth develop rate  $Kr_z$ . After a lag period ( $ddg_{\text{lag}}$ ) the rooting depth increases linearly with  
183 accumulated temperature sum from its starting value ( $z_{\text{start}}$ ). The length of the lag period and the  
184 rate of rooting depth development are controlled with crop specific parameter values. This  
185 approach to simulation of crop rooting depth is based on a number of studies showing good linear  
186 relationships between accumulated temperature sum and rooting depth (KRISTENSEN and  
187 THORUP-KRISTENSEN 2004; THORUP-KRISTENSEN 2006; THORUP-KRISTENSEN and VAN DEN  
188 BOOGAARD 1998; KAGE et al. 2000).

189

$$r_z = z_{start} + \left( (T_{cumul} - ddg_{lag}) \cdot Kr_z \right) \quad (1)$$

191

192 Horizontal root extension is calculated in the same way, but for each soil layer the calculation  
193 starts when the roots reach this layer rather than when the crop is planted. In this way horizontal  
194 root growth starts progressively later at larger depths.

195

196 Root biomass is calculated as a fraction of aboveground crop biomass. For all crops this fraction  
197 is reduced with higher crop biomass, but crops are parameterized into three classes with either  
198 high, medium or low fractions of root biomass. The fraction of biomass allocated to the roots  
199 start at 0.65 at very low crop biomass for all root classes, fall to 0.5, 0.3, and 0.2 at 2 Mg dry  
200 matter ha<sup>-1</sup>, and to 0.1, 0.05 and 0.02 when crop biomass exceed 9 Mg ha<sup>-1</sup> for high medium and  
201 low fractions respectively. Total root length is then calculated from the simulated root biomass  
202 and a fixed specific root length which is used for all crops.

203

204 Most vegetable crops are grown as row crops. Simulated root length is distributed spatially into  
205 the 2D array of 0.05 by 0.05 m soil cells used in the model to simulate the effects of the row crop  
206 structure on crop rooting and uptake of water and nitrogen. Root distribution is calculated to a  
207 maximum depth of 2 m, and to a maximum width of half crop row distance. GERWITZ and PAGE  
208 (1974) proposed a logarithmic root length function declining from the topsoil downwards. The  
209 assumption of a logarithmic decline in root density has been used in simulation models (e.g. the  
210 Daisy model, HANSEN et al. 1990), but in these models a rooting depth defined by a very low root  
211 density is assumed, and then the logarithmic function is used to distribute root length in the soil  
212 layers above the rooting depth. This inevitably leads to very low root densities in deeper soil  
213 layers. In our approach the root density at rooting depth is allowed to vary, meaning that we can  
214 simulate higher root densities in deep soil layers. Below rooting depth, root density is simulated  
215 to decline fast to zero, using a simple linear function. The steepness of the logarithmic decline  
216 within the root zone is controlled by one parameter for the vertical distribution (az) and another  
217 parameter for the horizontal distribution (ax). The root length at depth z is calculated as:

$$rootlength_z = L_0 e^{-az} \quad (2)$$

219

220 where  $L_0$  is root density at the soil surface, and  $z$  is the soil depth. Root density decline from  
221 beneath the crop row to the inter row soil is calculated by a similar function.

222 With some crops the plant to plant distance within the row is significant, but the effects of this  
223 cannot be simulated by the 2D approach used here, a 3D approach would be needed. During early  
224 growth this will lead to an overestimation of N availability, as the model will simulate that all N  
225 present close to the crop row will be available to the plants. To avoid this, we use the estimation  
226 of root width to calculate the fraction of the soil between plants within a row which is in contact  
227 with plant roots, and then reduce daily N uptake by this fraction.

228  
229

### 230 **Crop growth and critical N**

231 Crop growth in EU-Rotate\_N uses a total dry matter yield at harvest  $W_{\max}$  t/ha as a target yield.  
232 This approach overcomes difficulties that arise when trying to parameterise the large variety of  
233 different vegetable crops for photosynthesis-driven algorithms, but requires the user to provide the  
234 target. Each day the increment in plant dry matter is calculated from:

235

$$236 \quad \Delta W = \frac{K_2 G_N G_T G_W W}{K_1 + W} \quad (3)$$

237

238 where  $W$  is the cumulative dry weight, and  $K_1 = 1 \text{ t ha}^{-1}$ .  $G_T$  is the effective day degree for the day  
239 divided by the average day degree throughout the entire growing period, where the effective day  
240 degree is the average temperature for the day less a base temperature, with the limitation that if the  
241 average temperature exceeds  $20^\circ\text{C}$  then it is set equal to  $20^\circ\text{C}$ , GREENWOOD (2001).  $G_N$  and  $G_W$  are  
242 the growth coefficients dependent on crop %N and water supply respectively.  $K_2$  is calculated from  
243 the integral of the above equation with  $G_N$   $G_W$  and  $G_T$  set equal to 1. The equation is then

244

$$245 \quad K_2 = \frac{K_1 \ln W_{\max} + W_{\max} - K_1 \ln W_P - W_P}{T_h - T_P} \quad (4)$$

246

247 where  $W_p$  is the dry weight at planting,  $W_{max}$  is the target total dry matter yield ( $t\ ha^{-1}$ ),  $T_h$  is the time  
248 of final harvest and  $T_p$  is the time of drilling or planting in days from Jan 1st.

249  
250 We use a unified equation to define critical %N (The minimum N content in the plant required for  
251 maximum growth) for different crops, i.e.

$$252 \quad \quad \quad \% N_{crit} = a \cdot (1 + b \cdot e^{-0.26W}) \quad (5)$$

254  
255 where  $\%N_{crit}$  is the critical %N,  $W$  = total dry matter yield ( $t\ ha^{-1}$ ), and  $a$  and  $b$  are crop-specific  
256 coefficients. These coefficients are included for the crops used in the test of the model in Table 1 and  
257 are similar to those described in GREENWOOD (2001).

258  
259 Luxury N consumption is permitted to take place. It is calculated as follows:

$$260 \quad \quad \quad \% N_{max} = R_{lux} \% N_{crit} \quad (6)$$

262  
263 where  $\%N_{max}$  is the maximum possible crop %N, and  $R_{lux}$  ( $>1$ ) is the coefficient for luxury N  
264 consumption (examples shown in Table 1)  
265 ).

266  
267 For each day a growth coefficient  $G_N$  is calculated as:

$$268 \quad \quad \quad G_N = \min\left(\frac{\% N}{\% N_{crit}}, 1.0\right) \quad (7)$$

270  
271 where %N is the actual %N in the dry matter of the whole plant (excluding fibrous roots).

272  
273 Similarly, a growth coefficient  $G_W$  can be activated which regulates growth depending on water  
274 supply which is calculated as:

275

276 
$$G_w = \frac{TR_{act}}{TR} \quad (8)$$

277

278 where  $TR_{act}$  and  $TR$  are the actual and potential transpiration rates.

279 <<Table 1>>

280

281

282

283 ***N uptake***

284 N uptake is calculated as a function of crop N demand on a specific day and the potential root N  
 285 uptake on the same day. The simulated crop N demand is calculated in the crop growth part of the  
 286 model. The potential supply from the soil is calculated as a function of the root length in each soil  
 287 unit and the content of ammonium-N and nitrate-N in each soil unit to control root N uptake  
 288 efficiency. This is calculated separately for ammonium and nitrate N. Equation 9 shows the  
 289 calculation for potential ammonium N uptake.

290

291 
$$N_{potNH_4} = \frac{rootlength \cdot S_N \cdot (NH_4 - S_1)}{S_2 + NH_4} \quad (9)$$

292

293 with  $NH_4$  being the soil ammonium concentration and  $S_N$  a crop specific parameter. Diffusion  
 294 terms are not included in the simulation, since they are assumed to be very small over the  
 295 relevant time spans for the simulations. N in the form of nitrate is highly mobile in the soil, and  
 296 diffusion processes will only limit uptake on the very short term even at low root density. The  
 297 value of  $S_1$  determines the minimum amount of ammonium-N which can be left in the soil (e.g.  
 298 THORUP-KRISTENSEN 2001, 2006). ), and is set to prevent further uptake when less than 5 kg  
 299 ammonium-N is present in the top 30 cm soil layer.  $S_2$  reduces N uptake as these minimal values  
 300 are approached.

301

302 A function is then used to balance actual N uptake according to crop N demand and potential root  
 303 N uptake. At very high or low N supply relative to demand, the uptake will be fully controlled by

304 crop N demand and potential root N supply respectively. When N demand is close to potential N  
305 uptake, the simulated uptake will be below either value.

306

$$307 \quad N_{up} = N_{demand} \cdot \left( 1 - e^{\left( -\frac{N_{pot}}{N_{demand}} \right)} \right) \quad (10)$$

308

309 Often, the calculated actual N uptake will be lower than the potential root N supply. When this is  
310 the case, the actual depletion of soil N will be reduced proportionally from the potential value in  
311 all soil cells. Finally a specific calculation is made of N taken up from below 0.9 m in the soil.  
312 This is made as N leaching loss and other N balance figures are shown mainly for the 0-0.9 m  
313 soil layer in much of the model output, and it is therefore necessary also to have an output  
314 showing how much N is taken up from below this zone.

315

316

### 317 ***Fertility building crops***

318 As it is difficult to specify an appropriate target yield for a fertility building crop an alternative  
319 approach is used. The user specifies Good, Medium or Bad growth to determine crop growth  
320 rates rather than final DW production. The increment in plant dry matter on each day is calculated  
321 from:

322

$$323 \quad \Delta W = \min(G_{type} G_N G_T W, \Delta W_{type}) \quad (11)$$

324

325 where W is the cumulative dry weight,  $G_{type}$  and  $\Delta W_{type}$  is set to one of three possible values (good,  
326 medium, bad), which categorize growing conditions. Growth rate, varies from 2 to 6% per day for  
327 poor and good crops with a maximum dry weight increment of between 20 and 60 kg/ha dry matter  
328 for poor and good crops respectively..  $G_N$  and  $G_T$  are the growth coefficients dependent on the crop  
329 %N and day degree. The calculation of the growth coefficient  $G_N$  is the same as that for a cash crop.  
330 The growth coefficient  $G_T$  is calculated:

331

$$G_T = \begin{cases} 1.0 & \text{if } \textit{day degree} > 10.0 \\ \frac{\textit{day degree} - \textit{base temperature}}{10.0 - \textit{base temperature}} & \textit{base temperature} \leq \textit{day degree} \leq 10.0 \\ 0 & \textit{day degree} < \textit{base temperature} \end{cases} \quad (12)$$

333  
 334 Another crop parameter, litter loss, specifies the percentage of biomass which is returned to the  
 335 upper layer of the soil each day; it is then mineralised as a crop residue. This is particularly  
 336 significant for longer term leys. The user can specify dates at which the crop is mown – on these  
 337 occasions 50% of the biomass is either mulched or removed from the field.

338  
 339 Most fertility building crops are legumes and nitrogen fixation is the main source of nitrogen in  
 340 organic cropping systems. A crop parameter specifies whether the crop is N fixing or not (this also  
 341 applies to cash crops). The growth of N fixing crops is not limited by nitrogen in the soil as any  
 342 deficiency in soil supply is met by fixation of N from the air.

343  
 344 Annual crops are killed after an appropriate period of time for example after the 1<sup>st</sup> of March.  
 345 regardless of the ‘harvest date’ set by the user. Crops are also killed if the temperature drops below a  
 346 specified value, Phacelia is killed when the temperature drops below -5 °C..

347  
 348 Modelling of the growth of undersown crops begins at the harvest of the crop canopy with an  
 349 appropriate dry matter and nitrogen content; the user can choose between Good, Medium and Bad  
 350 performance as an understorey to provide different starting Dry matter yields which are 2000, 1000,  
 351 500 kg/ha for good medium and bad crops respectively.

352  
 353 ***Estimation of marketable yield***

354 Two strategies were adopted to convert total dry matter yield (TDM) into yield of marketable  
 355 produce.

356  
 357 For the first, our own published and un-published field research data were collected, where both  
 358 total dry matter and marketable yields were measured across Europe. The algorithms developed

359 allow direct conversion of total dry matter yield (TDM) into fresh marketable yield (MFY) at any  
360 given N supply and take into account the effects of both sub- and supra-optimal supply of N.

361

$$362 \quad MFY = TDM \cdot R(N_{av}) \quad (13)$$

363

364  $R(N_{av})$  being the ratio of marketable yield to total dry matter yield and  $N_{av}$  the available nitrogen  
365 in soil and plant to 90 cm. The ratio  $R(N_{av})$  is specific for each crop and depends on the  
366 proportion of available N used for each crop. The formula for  $R(N_{av})$  is a linear or polynomial  
367 relationship of available nitrogen ( $N_{av}$ ).

368

$$369 \quad R(N_{av}) = r_0 + r_1 \cdot N_{av} + r_2 \cdot N_{av}^2 + r_3 \cdot N_{av}^3 \quad (14)$$

370

371 The terms  $r_0$ ,  $r_1$ ,  $r_2$ , and  $r_3$  are empirically chosen for each crop. For a simple constant relationship  
372  $r_1$ ,  $r_2$  and  $r_3 = 0$ . For a linear relationship  $r_2$  and  $r_3 = 0$ . Otherwise, the relationship is non-linear. For  
373 some crops, more polynomial terms may be needed because of different behaviour in the sub-  
374 and supra-optimum ranges.

375 In a second approach, the single plant fresh weight is calculated by using the harvest index (HI)  
376 to calculate the dry weight of the harvested parts. Then, with the dry matter content ( $c_{DM}$ ) and the  
377 plant population ( $n$ ), an average single plant fresh weight yield (PFY) is produced:

$$378 \quad PFY = \frac{TDM \cdot HI}{n \cdot c_{DM}} \quad (15)$$

379 A normal distribution of plant fresh weights are assumed with a coefficient of variation (e.g.  
380 20%) and a lower and upper limit of marketable plant fresh weight can be set (e.g., the EU trade  
381 specifications). With this information, an average fresh weight of marketable plants within these  
382 specifications is calculated. Using the plant population again, the marketable yield (MFY) and the  
383 residues left post-harvest are calculated. A more detailed description of this approach can be  
384 found in NENDEL et al. (2008).

385

386 Plants with a single product per plant use the second approach; other crops, such as those with  
387 multiple products or multiple harvests, use the direct conversion approach. After calculation of

388 marketable yield the fraction of N harvested or left in the field as crop residues is then calculated.  
389 The ratio of N in the marketed part of the crop to the whole crop is taken from the Crop shown in  
390 Table 1.

391

### 392 **Gross margin calculation**

393 With the marketable yield modelled, the calculation of the crop gross margin (GM) uses the  
394 standard equation:

$$395 \quad GM = MFY \cdot Price - (VC_{ind} + VC_{dep} + VC_{Nfert}) \quad (17)$$

396 where the variable costs dependent ( $VC_{dep}$ ) and independent ( $VC_{ind}$ ) of marketable yield is  
397 provided by the user in the model run files.  $VC_{ind}$  should include, for example, cost per hectare of  
398 seed, transplants, fleece, irrigation, crop protection, and weed control. It should also include the  
399 cost of fertiliser application, but not the fertiliser itself. Variable costs dependent ( $VC_{dep}$ ) on the  
400 marketable yield should be provided per unit (e.g. tonnes) marketed and are then multiplied by  
401 the modelled marketable yield. They consist of packaging and drying, transport, harvest casual  
402 labour and market commission cost. The variable costs ( $VC_{Nfert}$ ) are the costs of inorganic and  
403 organic fertilisers, dependent on the fertiliser amounts and the prices of the fertilisers.

404 The triggered amount of N fertiliser and number of applications are multiplied by the cost of  
405 fertiliser and the cost per application as specified in the input file. Subsidies are not considered in  
406 the gross margin calculation. Rotational gross margin is cumulative gross margin of all crops in  
407 the rotation (including the negative gross margin of cover crops) divided by the number of years  
408 simulated.

409

### 410 **Model use**

411 The model requires input data in plain text format to describe soil properties, the initial soil  
412 mineral N and initial soil water content conditions. It can then be supplemented by blocks of text  
413 for each individual crop. These blocks contain planting and harvesting dates and the management  
414 of crop residues. The fertilisation and irrigation of these crops can be controlled by a range of  
415 fixed and automatic triggers. The automatic triggers can be used to fertilise or irrigate when

416 certain threshold values are met. To run the model five other text format files are required, one  
417 containing meteorological data, and four others containing parameters for mineral and organic  
418 fertilisers, crop growth and crop residues. The model, along with example files, can be  
419 downloaded from

420 [www.warwick.ac.uk/go/eurotaten](http://www.warwick.ac.uk/go/eurotaten)

### 421 ***Testing the model***

422 The model was tested against field data acquired from a range of sites in each country  
423 participating in the EU-Rotate\_N project. Within this short paper it is impossible to reproduce all  
424 the results so an example of the validation on an independent data set in Germany is presented.

425 The Palatinate region in South-West Germany covers the area from the banks of the Rhine in the  
426 East to the rising hills of the Palatinate Forest in the West. The Palatinate is one of the  
427 economically most important and at the same time one of the most diverse field vegetable  
428 production areas in Germany. 19 biannual crop rotations on 14 farms have been monitored from  
429 April 2003 until the end of 2004. The growers followed different production strategies, including  
430 fertilizer regimes of various intensities. Five rotations were grown on organic farms. A wide  
431 range of crops, including all major arable and horticultural crops, was represented. In addition,  
432 simulations were performed for 8 rotations similarly monitored at two research stations in eastern  
433 Germany, 4 on sand and 4 on clay soils. All crops were grown with a single (non-limiting) level  
434 of nitrogen fertilizer, reflecting actual user practice. Details of the crop rotations under  
435 observation are given in 21.

436

437 << Table 2 >>

438

439 During the vegetable growing period, soil was sampled every two weeks. Each time, soil samples  
440 from 15 points on each plot were taken from 0-30 cm, 30-60 cm, and 60-90 cm depth. In 2004,  
441 the frequency of sampling was less whilst non-vegetable crops such as cereals, maize, sugar beets  
442 and fertility building crops were grown. In the soil samples, soil moisture and mineral N content  
443 were determined. Total crop dry matter was determined at harvest of each crop Nitrogen content

444 of these samples was determined in a Vario EL element analyser (elementar Analysengeräte  
445 GmbH, Hanau, Germany).

446 .

447 To simulate the monitored crop rotations the model was initialised by running it on the same crop  
448 rotations twice in advance. This was carried out in order to initialise the starting properties of the  
449 soil organic matter pools before the testing against measured data was carried out. Observed  
450 yields were set as crop target yield parameters. Weather data observed at the Karlsruhe weather  
451 station (DWD 2003) was used. Soil hydraulic parameters were determined from texture  
452 information according to the German Soil Survey Manual (AG BODENKUNDE 1994). Crop  
453 parameters that were used are shown in Table 1.

454

455 Model performance for soil mineral nitrogen and soil moisture was calculated by comparing  
456 measured and predicted values for the three soil layers. For above-ground biomass dry matter and  
457 nitrogen concentrations, measured and predicted values at harvest were compared. The following  
458 model assessment statistics were used: root mean square error and mean absolute error (RMSE  
459 and MAE; WILMOTT and MATSUURA 2005), model bias (MBE; ADDISCOTT and WHITMORE  
460 1987), model efficiency (EF; NASH and SUTCLIFFE 1970) and index of agreement (d; WILMOTT  
461 1981). Two example rotations with different N regimes were selected to demonstrate the  
462 applicability of the model: (i) an organic farm crop rotation on a loamy soil (Rotation 8 in 2),  
463 where the use of organic fertilisers occasionally leads to very high soil mineral N contents, and  
464 (ii) a conventional, extensive crop rotation on sand (Rotation 15 in 2), where all year round  
465 ground cover and minimal fertiliser rates result in low soil mineral N levels.

466

### 467 ***Case studies - Norway***

468 A case study was selected where early vegetable crops were planted within a 6 year rotation with  
469 spring cereals as break crops. The case study was selected in contrasting soil types in the southern  
470 coastal regions of Norway to illustrate the effects of N management on nitrate requirement and N  
471 leaching. The study was based on two choices of N management.

472

473 A survey of grower practice revealed that levels of N fertilizer applied to vegetables often exceed  
474 the rates specified by the Norwegian Institute for Agricultural and Environmental Research. The

475 reasons for this include a desire to safeguard against deficiencies as well as a tendency to  
476 overestimate the expected/target yield level (to which current recommendations are linked).  
477 Growers make little use of mineral N measurements to check for early season N supply, as small  
478 field size and limited time in spring combine to make this method impracticable and costly. A  
479 modelling approach is an effective way of taking into account previous leaching losses and N  
480 mineralization from crop residues. The following two scenarios are compared:

- 481 • ‘Current recommendations’ (set according to yield level, based mainly on FYSTRO et al.  
482 2006)
- 483 • ‘Current grower practice’ (based on survey if available, otherwise estimated)

## 484 **Results**

### 485 ***Testing the model***

486 Testing the model against field data of 27 highly diverse crop rotations yielded an index of  
487 agreement (d) which indicates that 71% of the variations in soil mineral N, 82% of the variations  
488 in crop N concentration and more than 87% of the variations in soil water content can be  
489 explained by the model, see Table 3.

490

491 <<Table 3>>

492

493 For dry matter yield, 95% of the variation was explained by the model. However, this was  
494 expected as maximum target yields were an input to the model. On the basis of the statistical tests  
495 referred to in the materials and methods section, overall bias (MBE) is relatively low. The  
496 performance of the simulations for soil mineral N were variable on individual rotations but the  
497 model was still able to simulate the differences in soil mineral N between the two contrasting  
498 rotations, figure 2. Compared to the observations, the model is able to simulate both production  
499 systems with an average Index of Agreement of 0.65 for Rotation 8 and 0.33 for Rotation 15.  
500 MBE for Rotation 8 was 28.2 kg N ha<sup>-1</sup> (0–30cm), –21 (30–60cm) and –12 (60–90cm) and for  
501 Rotation 15 3 kg N ha<sup>-1</sup> (0–30cm), –3 kg N ha<sup>-1</sup> (30–60cm) and –3 kg N ha<sup>-1</sup> (60–90cm),  
502 respectively.

503

<< Figure 2 >>

504  
505 **Case Study Norway**  
506 To parameterise the soil mineralization routine, the EU-Rotate\_N model was run without any  
507 crops to check that the rates of release of N from soil organic matter were similar to those  
508 measured in the field. Once parameterised, the model was run for 3 cropping rotations in the  
509 southern coastal region of Norway. Table 4 shows the simulation results. Survey results revealed  
510 that growers often applied up to 36% more N than recommended as good practice. With  
511 recommended management practices nitrate concentrations in the drainage water were nearer the  
512 50 mg/litre EU limit for drinking water. The model simulated that on light soils (CS) gross  
513 margin increased by 14%, suggesting that higher grower N rates may be economically justified  
514 but not environmentally as simulated leaching was increased by 19%.

515  
516 Examination of the detailed outputs showed that there was a leaching peak during the cultivation  
517 of the third cauliflower crop and that using currently recommended rates the crop could fail –  
518 hence the reason for the higher application rates. Further investigation showed that if the lower  
519 rate of nitrogen was split into 3 rather than 2 applications and applied to coincide with crop  
520 demand, increases in gross margin could be achieved without applying any additional fertiliser  
521 (Table 5). Leaching losses could also be reduced. The most effective treatment to increase gross  
522 margin was splitting the N into 6 applications as it made it much more available to the growing  
523 crop. A technique such as fertigation might be used to deliver this approach but the capital cost  
524 (not included) might outweigh the benefit.

525  
526 <Table 4>

527 <Table 5>

528  
529 **Discussion**

530 The EU-Rotate\_N model enables the effect of different strategies of fertilisation and crop  
531 management over rotations for both field vegetable and major arable crops to be tested. The  
532 example simulations demonstrated that the model is able to predict the soil mineral N dynamics

533 for two contrasting production systems. The model was able to simulate the higher amounts of  
534 soil mineral N in the rotation with large inputs of organic N compared with the rotation receiving  
535 more optimised inputs of mineral fertiliser N. In the case studies the value of the model to match  
536 demand of crops more closely to supply in order to reduce N losses was demonstrated.

537  
538 Most of the modules are based on existing models which have already been extensively validated  
539 but few studies have validated the operation of the entire model. Currently few datasets covering  
540 rotations are available for such a validation to be carried out but this situation should improve in  
541 the future.

542  
543 One of the new modules simulates the growth of roots for field vegetable and some arable crops  
544 that are grown in wide rows using a two dimensional approach the single dimension approach for  
545 water and N uptake being inadequate (SCHRÖDER et al. 1996; THORUP-KRISTENSEN and VAN DEN  
546 BOOGAARD 1998, 1999). Since the range of plant morphology in field vegetable crops makes  
547 modelling of growth and development of leaf area for photosynthesis too complex  
548 (BARANAUSKIS 2005) EU-Rotate\_N uses the target yield approach used in the N\_ABLE and  
549 WELL\_N models (GREENWOOD 2001). This enables the simulation of dry matter accumulation in  
550 a large variety of field vegetables with different morphologies as well as in multiple harvest crops  
551 such as cucumbers or courgettes. This simplification does lead to a limitation that target yield  
552 has to be estimated before the model can be run, however suitable values for target yields can be  
553 obtained from previous experiments or can be based on growers expert knowledge.

554  
555 The model also simulates recovery of N that has leached below the depth of shallow rooted crops,  
556 by crops with deeper roots allowing the planning of rotations to minimise N losses. The  
557 importance of N supply to successive crops through decomposing crop residues, left in the field  
558 by the preceding one, is often poorly described in dynamic process based -models for agricultural  
559 systems (KERSEBAUM et al. 2007). Automatically triggered fertilisation and irrigation events  
560 allow the calculation of long-term scenarios to assess different strategies for improving the N  
561 efficiency in vegetable crop rotations. Such strategies were demonstrated under drip and furrow  
562 irrigation systems used by Mediterranean producers (DOLTRA et al. 2007), within highly variable

563 input production systems (NENDEL 2009) or within organic low-input production systems  
564 (SCHMUTZ et al. 2006, 2008)

565  
566 Rotation planning is particularly important in organic production systems where the application  
567 of permitted fertilisers and manures must also be optimised. Very simple approaches has been  
568 used for predicting N availability in organic systems (PADEL 2002, CUTTLE 2006), approaches  
569 which avoid many of the difficulties associated with the EU-Rotate\_N approach of handling the  
570 recycling of N as a result of litter loss and mowing residues. However, such simple approaches  
571 are also less able to deal with complex rotations and frequent short term fertility building crops  
572 common in field vegetable production. A more sophisticated approach has been used in the  
573 NDICEA model (KOOPMANS and BOKHORST 2002; VAN DER BURGT et al. 2006), originally  
574 developed for use under Dutch conditions. This model does allow rotations to be built up but it  
575 does not take into account reductions in yield attributable to lack of water or N, neither does it  
576 include any of the economic aspects of EU-Rotate\_N.

577  
578 The ability to calculate gross margins across crop rotations will support farmers in balancing  
579 environmental and economic objectives. This is in contrast to typical practice, where evaluations  
580 of the economic and environmental impact (in terms of N leaching) of farmer's decisions or  
581 political measures range from very simple approaches based on yield and N leaching assessment  
582 with the help of non-feedback functions (HASLER 1998) to quite advanced approaches using  
583 dynamic soil-crop-atmosphere models for specific problems at different scales. The most  
584 frequently employed models in this context are EPIC (HUGHES et al. 1995; TEAGUE et al. 1995;  
585 KELLY et al. 1996; JOHNSON, SOIL-SOILN (VATN et al. 1999, 2002), FASSET (BERNTSEN et al.  
586 2003), CropSyst (FARES 2003; MORARI et al. 2004) and STICS (SCHNEBELEN et al. 2004).  
587 However, these models do not include any economic assessments. An ecological and economical  
588 evaluation of different fertiliser strategies on a regional level using EU-Rotate\_N was presented  
589 by NENDEL (2009).

590

## 591 **Conclusions**

592

593 The case study demonstrated how the EU-Rotate\_N model can be used as a tool to illustrate the  
594 effects of different management strategies on yield and nitrogen losses. It is clear that, following  
595 recommended practice which include assessments of available N in the soil, can reduce the  
596 amounts of applied fertiliser in most cases, thereby reducing N losses, particularly by leaching.  
597 The simulations in southern Norway illustrate that the model will in some situations recommend  
598 higher N rates than those based on National recommendations, e.g. in situations where there is a  
599 risk of significant N leaching loss during crop growth. Helping farmers in general to reduce N  
600 inputs, but also sometimes to increase fertilisation of crops where needed due to soil and weather  
601 conditions, will be a major advantage of using the model for N advice. However, the model could  
602 be further used to refine the management practices to minimise N leaching. These practices  
603 could be tested in the field and demonstrated to farmers.

604  
605 The EU-Rotate\_N decision support system provides a platform for evaluating the impact of  
606 implementing national fertiliser recommendations on crop, environmental and economic outputs  
607 of varied crop rotations, which could subsequently allow the identification of leaky points and  
608 beneficial practices to plug them. Contrasting beneficial practices, which can reduce the  
609 environmental impact with “reasonable” economic costs can be tested against each other.  
610 Fluctuations in input and output prices, subsidies and tax effects can also be analysed, providing  
611 a dynamic feedback that could help both farmers and policymakers in the future.

612  
613 **Acknowledgement**  
614 The authors wish to acknowledge funding from the EU Quality of life Programme Key Action 5  
615 – Sustainable agriculture which led to the development of EU-Rotate\_N . Project number QLK5-  
616 2002-01100

617  
618 **References**  
619 ABRAHAMSEN, P. and S. HANSEN 2000: Daisy: an open soil-crop-atmosphere system model.  
620 Environ. Model. Software **15**, 313-330.  
621 ADDISCOTT, T.M. and A.P. WHITMORE 1987: Computer-simulation of changes in soil mineral  
622 nitrogen and crop nitrogen during autumn, winter and spring. J. Agric. Sci. **109**, 141-157.

- 623 AG BODENKUNDE 1994: Bodenkundliche Kartieranleitung. 4. Auflage. E. Schweizerbartsche  
624 Verlagsbuchhandlung, Hannover.
- 625 ALLEN, R.G., L.S. PEREIRA, D. RAES and M. SMITH 1998: Crop evapotranspiration. Guidelines for  
626 computing crop water requirements. FAO Irrigation and Drainage Paper n° 56, FAO, Roma.
- 627 BARANAUSKIS, K., V. BANDZEVICIUTE, G. SAMUOLIENE, G. SABAJEVIENE, R. ULINSKAITE, J.  
628 SAKALAUŠKAITE and P. DUCHOVSKIS 2005: Dimensions of the determination methodology of  
629 white cabbage (*Brassica oleracea* L. var. *capitata* f. *alba* L.) assimilation area. *Sodininkyste ir*  
630 *Darzininkyste* **24**, 42-47
- 631 BERNTSEN, J., B.M. PETERSEN, B.H. JACOBSEN, J.E. OLESEN and N.J. HUTCHINGS 2003:  
632 Evaluating nitrogen taxation scenarios using the dynamic whole farm simulation model FASSET.  
633 *Agr. Syst.* **76**, 817-839
- 634 CUTTLE, S.P. 2006: Development of the FBC model to estimate the nitrogen  
635 available from fertility-building crops in organic rotations. *Aspect.*  
636 *Appl. Biol.* **79**, 259-226
- 637 DE CLERCQ, P., A.C. GERTSIS, G. HOFMAN, S.C. JARVIS, J.J. NEETESON and F. SINABELL (EDS.)  
638 2001: Nutrient Management Legislation in European Countries. Department of Soil Management  
639 and Soil Care, Faculty of Agricultural and Applied Biological sciences. Wageningen Press The  
640 Netherlands.
- 641 DEMYTTENAERE, P., G. HOFMAN, D. RONSE and M. VAN RUYMBEKE 1990: Excessive soil  
642 mineral-N at harvest of field-grown vegetables: impact on nitrate pollution of ground and surface  
643 water. In: Calvert, R. (ed.) Symposium "Nitrates, Agriculture, Water" Paris - November 1990,  
644 INRA. 239-244.
- 645 DOLTRA, J., C. NENDEL and C. RAMOS 2007: Evaluation of different fertilisation practices with  
646 the EU-ROTATE\_N model. In: BOSCH, A., M.R. TEIRA and J.M. VILLAR (eds.): Towards a  
647 better efficiency in N use. Editorial Milenio (Spain). 426-428.
- 648 DWD 2003: Daily weather data for selected stations in Germany.  
649 [http://www.dwd.de/de/FundE/Klima/KLIS/daten/online/nat/ausgabe\\_tageswerte.htm](http://www.dwd.de/de/FundE/Klima/KLIS/daten/online/nat/ausgabe_tageswerte.htm).
- 650 FARES, E. 2003: Agriculture pollution control policies: a case study of nitrate pollution in the  
651 Apulia region (Southern Italy). Master Thesis. Institut Agronomique Mediterranéen de  
652 Montpellier (64) Paris: Centre International de Hautes Etudes Agronomiques Mediterranéennes  
653 Institute, Washington D.C.
- 654 FINK, M. and H.C. SCHARPF 1993: N-expert – a decision support system for vegetable fertilization  
655 in the field. *Acta Hort.* **339**, 67–74
- 656  
657 FYSTRO, G., B. HOEL, H. HOLE, T. LUNNAN and H. RILEY 2006:  
658 [www.bioforsk.no/dok/senter/ost/ape/gjodslingshandbok/gjodslingshandbok.html&ResearchAreaID=2](http://www.bioforsk.no/dok/senter/ost/ape/gjodslingshandbok/gjodslingshandbok.html&ResearchAreaID=2)

- 659 GERWITZ, A., and E.R. PAGE 1974: Empirical mathematical-model to describe plant root systems.  
660 J. Appl. Ecol. **11**, 773-781.
- 661 GOULDING, K. 2000: Nitrate leaching from arable and horticultural land. Soil Use Manag. **16**,145-  
662 151.
- 663 GREENWOOD, D.J. 2001: Modelling N-response of Field Vegetable Crops grown under Diverse  
664 Conditions with N\_ABLE: A review. J. Plant Nutr. **24**, 1799-1815
- 665 HANKS, R.J. 1983: Yield and water-use relationships: An overview. In: H.M. Taylor et al. (eds.):  
666 Limitations to efficient water use in crop production. ASA, Madison, WI., 393-411.
- 667 HANSEN, S., H.E. JENSEN, N.E. NIELSEN and H. SVENDSEN 1990: Daisy - A Soil Plant Atmosphere  
668 System Model. NPO Research from the National Agency of Environmental Protection. No. A 10.  
669 272 pp.
- 670 HASLER, B. 1998: Analysis of environmental policy measures aimed at reducing nitrogen  
671 leaching at the farm level. Environ. Pollut. **102**, 749-754.
- 672 HE, Z.L., A.K. ALVA, D.V. CALVERT and D.J. BANKS 1999: Ammonia volatilization from  
673 different fertilizer sources and effects of temperature and soil pH. Soil Sci. **164**,750-758.
- 674 HUGHES, D., W. BUTCHER, A. JARADAT and W. PENARANDA 1995: Economic analysis of the long-  
675 term consequences of farming practices in the barley cropping area of Jordan. Agr. Syst. **47**, 39-  
676 58.
- 677 JANSSON, P.E. 1991: Simulation model for soil water and heat conditions. Report 165, Swedish  
678 University of Agricultural Sciences, Uppsala, 72 pp.
- 679 JENSEN, L.S., T. SALO, F. PALMASON, T.A. BRELAND, T.M. HENRIKSEN, B. STENBERG, A.  
680 PEDERSEN, C. LUNDSTRÖM and M. ESALA 2005: Influence of biochemical quality on C and N  
681 mineralisation from a broad variety of plant materials in soil. Plant Soil **273**, 307-326.
- 682 JOHNSON, S.L., R.M. ADAMS and G.M. PERRY 1991: The on-farm costs of reducing groundwater  
683 pollution. Am. J. Agric. Econ. **73**, 1063-1073..
- 684 KAGE,H., M. KOCHLER and H. STÜTZEL 2000: Root growth of cauliflower (*Brassica oleracea* L.  
685 botrytis) under unstressed conditions: measurement and modelling. Plant Soil **223**, 131-145.  
686
- 687 KARVONEN, T. 2003 (internet publication): Model of snow accumulation and snowmelt.  
688 [www.water.hut.fi/wr/kurssit/Yhd-12.135/kirja/paa\\_e.htm](http://www.water.hut.fi/wr/kurssit/Yhd-12.135/kirja/paa_e.htm)
- 689 KELLY, T.C., Y.C. LU, and J. TEASDALE 1996: Economic-environmental tradeoffs among  
690 alternative crop rotations. Agr. Ecosyst. Environ. **60**, 17-28.
- 691 KERSEBAUM, K.C., J.M. HECKER, W. MIRSCHEL and M. WEGEHENKEL 2007: Modelling water  
692 and nutrient dynamics in soil-crop systems: a comparison of simulation models applied on

- 693 common data sets. In: Kersebaum, K. C., J.M. Hecker, W. Mirschel and M. Wegehenkel (Eds.):  
694 Modelling water and nutrient dynamics in soil crop systems. Springer, Stuttgart, 1-17.
- 695 KOOPMANS, C.J. and J. BOKHORST 2002: Nitrogen mineralisation in organic farming systems: a  
696 test of the NDICEA model. *Agronomie* **22**, 855-862.
- 697 KRISTENSEN, H.L. and K. THORUP-KRISTENSEN 2004: Uptake of <sup>15</sup>N labeled nitrate by root  
698 systems of sweet corn, carrot and white cabbage from 0.2 to 2.5 m depth. *Plant Soil* **265**, 93-100.
- 699 MORARI, F., E. LUGATO and M. BORIN 2004: An integrated non-point source model-GIS system  
700 for selecting criteria of best management practices in the Po Valley, North Italy. *Agr. Ecosyst*  
701 *Environ.* **102**, 247-262
- 702 MOTOVILOV, Y.G., L. GOTTSCHALK, K. ENGELAND and A. BELOKUROV 1999: ECOMAG:  
703 Regional model of hydrological cycle. Application to the NOPEX region. Dept. Geophysics,  
704 University of Oslo, Report no. 1051, 88 pp.
- 705 NASH, J.E. and J.V. SUTCLIFFE 1970: River flow forecasting through conceptual models, Part I -  
706 A discussion of principles. *J. Hydrol.* **10**, 282-290.
- 707 NEETESON, J.J., and O.T. CARTON 2001: The environmental impact of Nitrogen in Field  
708 Vegetable Production. In proceedings of ISHS/ENVEG conference 1999. *Acta Hort.* **563**, 21-28.
- 709 NENDEL, C. 2008: Evaluation of Best Management Practices for N fertilisation in regional field  
710 vegetable production with a small scale simulation model. *Eur. J. Agron.*, in press.
- 711 NENDEL, C., U. SCHMUTZ, A. VENEZIA, F. PIRO and C.R. RAHN 2008: Modelling marketable yield  
712 for economic analysis of N fertilisation strategies for vegetables. *HortScience*, under review.
- 713 NRCS 2004: Estimation of direct runoff from storm rainfall. In: *National Engineering Handbook*,  
714 Part 630 – Hydrology. USDA.
- 715 OLSEN, P.A. and L.E. HAUGEN 1997: Jordas termiske egenskaper. Dept. Soil and Water Science,  
716 Agricultural University of Norway, Rapport nr. 8, 14 pp. (in Norwegian)
- 717 PADEL, S. 2002: Development of software to plan conversion to organic production (OrgPlan). In:  
718 Powell, J. et al (eds.) *UK Organic Research 2002: Proceedings of the COR Conference*,  
719 Aberystwyth, 169-172.
- 720 PEDERSEN, A., ZHANG, K., THORUP-KRISTENSEN, K. and JENSEN, L.S. (2009) Modelling diverse  
721 root density dynamics and deep nitrogen uptake - A simple approach, *Plant Soil* (DOI:  
722 10.1007/s11104-009-0028-8)
- 723 RAHN, C.R., L. VAIDYANATHAN and C. PATERSON 1992: Nitrogen residues from brassica crops.  
724 *Aspect. Appl. Biol.* **30**, 263-270.
- 725 RAHN, C.R., D.J. GREENWOOD and A. DRAYCOTT 1996: Prediction of nitrogen fertiliser  
726 requirement with HRI WELL\_N Computer Model. In: Van Cleemput, O., G. Hofman and A.

- 727 Vermoesen (eds.) Progress in Nitrogen Cycling (Proceedings of the 8th Nitrogen Fixation  
728 Workshop, Ghent, 5-8 September 1994), Kluwer, 255-258.
- 729 RAHN, C.R., C.D. PATERSON and L.V.V. VAIDYANATHAN 1998: The use of measurements of soil  
730 mineral N in understanding the response of crops to fertilizer nitrogen in intensive cropping  
731 rotations. *J. Agr. Sci.* **130**, 345-356.
- 732 REJESUS, R.M. and R.H. HORNBAKER 1999: Economic and environmental evaluation of  
733 alternative pollution-reducing nitrogen management practices in central Illinois. *Agr. Ecosyst.*  
734 *Environ.* **75**, 41-53.
- 735 RILEY, H. and H. BONESMO 2005: Modelling of snow and freeze-thaw cycles in the EU-rotate\_N  
736 decision support system. *Grønn Kunnskap (e)* vol. 9, no. 112, 8 pp.
- 737 RITCHIE, J.T. 1998: Soil water balance and plant water stress. In: G.Y. Tsuji et al (eds.):  
738 Understanding options for agricultural production . Dordrecht. Kluwer Academic Press, 41-54.
- 739 ROSE, D. 1968: Water movement in porous materials III. Evaporation of water from soil. *Brit. J.*  
740 *Appl. Phys. ser 2*(1), 1779-1791.
- 741 SADEGHI, A.M., K.J. MCINNES, D.E. KISSEL, M.L. CABRERA, J.K. KOELIKER and E.T.  
742 KANEMASU 1988: Mechanistic model for predicting ammonia volatilization from urea. In: B.R.  
743 Bock und D.E. Kissel (Hrsg.), Ammonia volatilization from urea fertilizers. National Fertilizer  
744 Development Centre, Tennessee Valley Authority, Muscle Shoals, Alabama, 67- 92.
- 745 SAXTON, K.E., W.J. RAWLS, J.S. ROMBERGER and R.I. PAPENDICK 1986: Estimating generalized  
746 soil-water characteristics from texture. *Soil Sci. Soc. Amer. J.* **50**, 1031-1036.  
747
- 748 SCHRÖDER, J.J., J. GROENWOLD and T. ZAHARIEVA 1996: Soil mineral nitrogen availability to  
749 young maize plants as related to root length density distribution and fertilizer application method.  
750 *Neth. J. Agr. Sci.* **44**, 209-225.
- 751 SHANI U., and L. M. DUDLEY 2001: Field studies of crop response to water and salt stress. *Soil*  
752 *Sci. Soc. Am. J.* **65**, 1522–1528.
- 753 SCHMUTZ, U., F. RAYNS, C. FIRTH, A. ROSENFELD, K. THORUP-KRISTENSEN, K. ZHANG and C.  
754 RAHN 2006: Environmental and economic modelling of organic, stockless, horticultural crop  
755 rotations. In: Andreasen, C.B., L. Elsgaard, S.L. Sondegaard and G. Hansen (eds.): Proceedings  
756 of the European Joint Organic Congress, Odense, Denmark, May 30-31, 2006, 238-241.
- 757 SCHMUTZ, U., F. RAYNS, C. FIRTH, C. NENDEL, R. LILLYWHITE, K. ZHANG and C. RAHN 2008:  
758 National-scale modelling of N leaching in organic and conventional horticultural crop rotations –  
759 policy implications. In: Neuhoff, D., N. Halberg, T. Alföldi, W. Lockeretz, A. Thommen, I.A.  
760 Rasmussen, J. Hermansen, M. Vaarst, L. Lueck, F. Caporali, H.H. Jensen, P. Migliorini, H.  
761 Willer (eds.): Cultivating the Future Based on Science. Volume 1: Organic crop production, 282-  
762 285.

- 763 SCHNEBELEN, N., B. NICOULLAUD, H. BOURENNANE, A. COUTURIER, B. VERBEQUE, C. REVALIER,  
764 A. BRUAND and E. LEDOUX 2004: The STICS model to predict nitrate leaching following  
765 agricultural practices. *Agronomie* **24**, 423-435.
- 766 SØGAARD, H.T., S.G. SOMMER, N.J. HUTCHINGS, J.F.M. HUIJSMANS, D.W. BUSSINK and F.  
767 NICHOLSON 2002: Ammonia volatilization from field-applied animal slurry - the ALFAM model.  
768 *Atmos. Environ.* **36**, 3309-3319.
- 769 TEAGUE, M.L., D.J. BEMARDO and H.P. MAPP 1995: Farm-level economic analysis incorporating  
770 stochastic environmental risk assessment. *Am. J. Agr. Econ.* **77**, 8-19.
- 771 THORUP-KRISTENSEN, K. 2001: Are differences in root growth of nitrogen catch crops important  
772 for their ability to reduce soil nitrate-N content, and how can this be measured? *Plant Soil* **230**,  
773 185-195.
- 774 THORUP-KRISTENSEN, K. and R. VAN DEN BOOGAARD 1998: Temporal and spatial root  
775 development of cauliflower (*Brassica oleracea* L. var. *botrytis* L.). *Plant Soil* **201**, 37-47.
- 776 THORUP-KRISTENSEN, K. and R. VAN DEN BOOGAARD 1999: Vertical and horizontal development  
777 of the root system of carrots following green manure. *Plant Soil* **212**, 145-153.
- 778 THORUP-KRISTENSEN, K. 2006A: Root growth and nitrogen uptake of carrot, early cabbage, onion  
779 and lettuce following a range of green manures. *Soil Use Man.* **22**, 29-38.
- 780 THORUP-KRISTENSEN, K. 2006B: Effect of deep and shallow root systems on the dynamics of soil  
781 inorganic N during three year crop rotations. *Plant Soil* **288**, 233-248.
- 782 VAN DER BURGT, G.J.H.M., G.J.M. OOMEN, A.S.J. HABETS and W.A.H. ROSSING 2006: The  
783 NDICEA model, a tool to improve nitrogen use efficiency in cropping systems. *Nutr. Cycl.*  
784 *Agroecosys.* **74**, 275-294.
- 785 VATN, A., L. BAKKEN, P. BOTTERWEG and E. ROMSTAD 1999: ECECMOD: an interdisciplinary  
786 modelling system for analyzing nutrient and soil losses from agriculture. *Ecol. Econ.* **30**, 189-  
787 205.
- 788 VATN, A., L. BAKKEN, M.A. BLEKEN, O.H. BAADSHAUG, H. FYKSE, L.E. HAUGEN, H.  
789 LUNDEKVAM, J. MØRKEN, E. ROMSTAD, P.K. RØRSTAD, A.O. SKJELVÅG, T. SOGN, N. VAGSTAD  
790 and E. YSTAD 2002: ECECMOD (2.0): An Interdisciplinary Research Tool for Analysing  
791 Policies to Reduce Emissions from Agriculture. Report no. 3/2002, Agricultural University of  
792 Norway.
- 793 VEHVILÄINEN, B. and J. LOHVANSUU 1991: The effects of climate on water discharges and snow  
794 cover in Finland. *Hydrolog. Sci. J.* **36**, 109-121.
- 795 WILLMOTT, C.J. 1981: On the validation of models. *Phys. Geogr.* **2**, 184-194.
- 796 WILLMOTT, C.J. and K. MATSUURA 2005: Advantages of the mean absolute error (MAE) over the  
797 root mean square error (RMSE) in assessing average model performance. *Clim. Res.* **30**, 79-82.

798  
799

800 **Tables**

801

802 **Table 1** Main crop parameter values used for testing the operation of the EU-Rotate\_N model over

803 rotations shown in Table 2

804

805

806 **Table 2** Crop rotations monitored for model testing. FEA = Farmer's environmental awareness, SOM =

807 Soil organic matter content.

808

809 **Table 3** Statistical evaluation of model performance assessed over 27 sites in Germany: root mean squared

810 error (RSME), mean absolute error (MAE), mean bias error (MBE), modeling efficiency (EF), and index

811 of agreement (d).

812

813 **Table 4.** Key annual N-flows (kg/ha) and gross margins (Euro/ha per year) simulated for various

814 early vegetable crops grown in 6-year rotations with spring cereals in coastal regions of Southern

815 Norway, with currently recommended N fertilizer rates (A), and assumed grower N fertilizer

816 rates (B). (All data are means of all six years in the rotation, calculated for the period 2000-2005)

817 <sup>1</sup> Proportion of the rotation time expressed as a % when leached nitrogen was greater than

818 0.1 kg/ha/day.

819 <sup>2</sup> Drainage same for both A+B case studies

820

821 **Table 5.** The simulated effect of different fertilizer management strategies on environmental and

822 economic outputs of Cauliflower crops grown on Sand soils in southern Norway

823

824 **Figures**

825

826 **Figure 1:** The organisation of the main model modules.

827

828 **Figure 2**

829 Soil mineral nitrogen dynamics in 0-30 cm (A, D), 30-60 cm (B, E) and 60-90 cm (C, F) in two

830 different crop rotations (A, B, C: Rotation 8: onion – spinach – spinach – maize on a light sandy

831 loam soil in South-Western Germany; D, E, F:Rotation 15: carrot – winter wheat – lucerne on a

832 sandy soil in Eastern Germany). Symbols: observed data; solid lines: model simulation.

833

834 s

**Table 1**

CROP	a	b	Rl <sub>UX</sub>	Base	ddg <sub>lag</sub>	Kr <sub>z</sub>	a <sub>z</sub>	HI	N_ratio
Dutch_White_Cabbage	3.45	0.6	1	7	100	0.0014	1.5	0.65	0.9
Cabbage_Summer	2.6	1.1	1	7	100	0.001	2	0.75	0.9
Cabbage_Wint/Spring	2.6	1.1	1	7	100	0.001	1.5	0.54	1.2
Calabres	3.45	0.6	1	7	100	0.001	2	0.28	0.6
Carrot	1	1.26	1.5	7	250	0.0007	3	0.83	2
Cauliflower	3.45	0.6	1	7	100	0.001	2	0.45	0.9
Leek	2	4	1.4	7	350	0.0003	8	0.68	1.2
Lettuce_Butterhead	1.35	1.35	1	7	100	0.001	3	0.8	0.8
Lettuce_Crisp	2.6	1.1	1	7	100	0.001	2	0.8	0.8
Maize_grain	0.6	9	1	7	100	0.0014	3	0.8	0.8
Onion	1.35	2.42	1	7	250	0.0003	8	0.75	2
Peas	1.35	3	1	7	100	0.001	3	0.25	0.6
Potato_Early	1.35	3	1.5	7	100	0.0007	3	0.8	2
Potato_Late	1.35	3	1.5	7	100	0.0007	3	0.95	1.9
Radish	1.35	1.87	1.2	7	100	0.001	3	0.5	1.4
Spinach	1.35	3	1	7	100	0.001	3	0.71	0.8
Sugar_Beet	1.11	1.38	1.65	7	250	0.001	2	0.7	2.8
Turnip	1.35	3	2	7	100	0.001	2	0.47	1.5
Wheat	1.35	3	1.2	4	100	0.001	3	0.51	0.3
Lamb's_Lettuce	1.35	3	1.2	4	250	0.0014	3	0.95	1
Kohlrabi	1.35	3	1.4	3	100	0.0014	3	0.7	1.5
Celery	1.35	3	1.3	6	250	0.0004	3	0.7	1
Celeriac	1.35	3	1.2	6	250	0.0004	3	0.71	2
Small_Radish_Spring	1.35	3	1.2	2	100	0.001	3	0.84	0.8
Small_Radish_Summer	1.35	3	1.2	2	100	0.001	3	0.85	0.8
Parsley	1.35	3	1.4	4	250	0.001	3	0.75	0.6
Radicchio	1.6	3	1.3	7	100	0.0012	3	0.4	1
Spring_onion	1.35	2.42	1	7	200	0.0003	8	0.9	1
Barley	1.35	3	1.2	4	100	0.001	3	0.51	0.3
Rye_and_Triticale	1.35	2	1.2	4	100	0.001	3	0.5	0.3
Maize_(corn_cob_mix)	0.6	9	1	7	100	0.0014	3	0.7	0.8
Maize_(silage)	0.6	9	1	7	100	0.0014	3	0.93	0.8

a and b are crop specific parameters for equation 5 - %N,  $R_{lux}$  = coefficient for luxury consumption, Base=Base temperature Degree °C ,  $ddg_{lag}$  = lag period before root growth begins Degree °C days.  $r_z$   $a_z$  = Form parameter for root development in vertical and horizontal directions m. HI= Harvest Index (dry matter basis). N\_Ratio = %N in residue DM / %N in harvested DM

Table 2

Nr.	Strategy	FEA	Soil type	SOM %	Total N kg N ha <sup>-1</sup>	Irrig. mm	Crop rotation	
							1 <sup>st</sup> year	2 <sup>nd</sup> year
1	intensive	low	silty loam	1.4	570	210	Spring onion – Lamb’s lettuce	Winter wheat
2	intensive	low	silty loam	1.3	1182	474	2×Small radish – Spring onion	Small radish – Winter rye
3	intensive	high	light sandy loam	1.7	240	840	Kohlrabi – Radish	Spinach – Celery
					280	840	Kohlrabi – Radish	Spinach – Celeriac
4	intensive	high	light loamy sand	1.0	590	755	Phacelia – Lettuce – Phacelia	Cauliflower – Phacelia
					590	755	Phacelia – Lettuce – Phacelia	Romanesco – Phacelia
5	agriculture	high	silty loam	1.4	450	620	Cauliflower – Cauliflower	Sugar beet
6	intensive	low	light clayey loam	1.5	470	685	Broccoli – Lamb’s lettuce	Onion
					470	685	Broccoli – Lamb’s lettuce	Cauliflower
7	organic	high	light sandy loam	1.5	65	145	Potato – Weeds – Winter rye	Lettuce
					65	205	Potato – Weeds – Winter rye	Kohlrabi
8	organic	high	light sandy loam	1.5	250	320	Onion – Spinach – Spinach	Maize
9	organic	moderate	silty loam	1.8	120	165	Pea (ind.) – Lamb’s lettuce	Parsley
					141	165	Pea (ind.) – Lamb’s lettuce	Carrot
10	intensive	high	silty loam	1.5	216	350	3×Parsley	Potato – Spinach
11	intensive	low	light clayey loam	2.3	520	725	2×Broccoli	Potato
12	extensive	high	light sandy loam	1.5	260	150	Onion – Mustard	Potato
13	extensive	high	silty loam	1.4	330	195	Lettuce – Sudan grass	Potato – Sudan grass
14	extensive	very high	light clayey loam	1.5	220	195	Turnip	Radicchio – Ryegrass
15	experiment		sand	1.2	200	503	Carrot – Winter wheat	Lucerne
16	experiment		sand	1.2	380	644	Leek – Winter wheat	Lucerne
17	experiment		sand	1.2	190	334	Summer rye	Carrot
18	experiment		sand	1.2	340	427	Summer rye	Leek
19	experiment		sandy clayey loam	2.2	154	251	Carrot	Summer wheat
20	experiment		sandy clayey loam	2.2	305	175	Broccoli	Summer wheat
21	experiment		sandy clayey loam	2.2	141	100	Summer wheat	Carrot
22	experiment		sandy clayey loam	2.2	287	80	Summer wheat	Broccoli



**Table 3**

		Soil mineral N <i>kg N ha<sup>-1</sup></i>	Soil water <i>kg kg<sup>-1</sup></i>	Dry matter yield <i>t ha<sup>-1</sup></i>	N concentration <i>%</i>
n	<i>no unit</i>	2383	771	89	85
RMSE	<i>unit</i>	62.72	0.07	2.02	1.07
MAE	<i>unit</i>	42.38	0.05	0.97	0.81
MBE	<i>unit</i>	-9.87	0.00	-0.75	-0.16
EF	<i>no unit</i>	-0.14	0.51	0.79	0.47
d	<i>no unit</i>	0.71	0.87	0.95	0.82

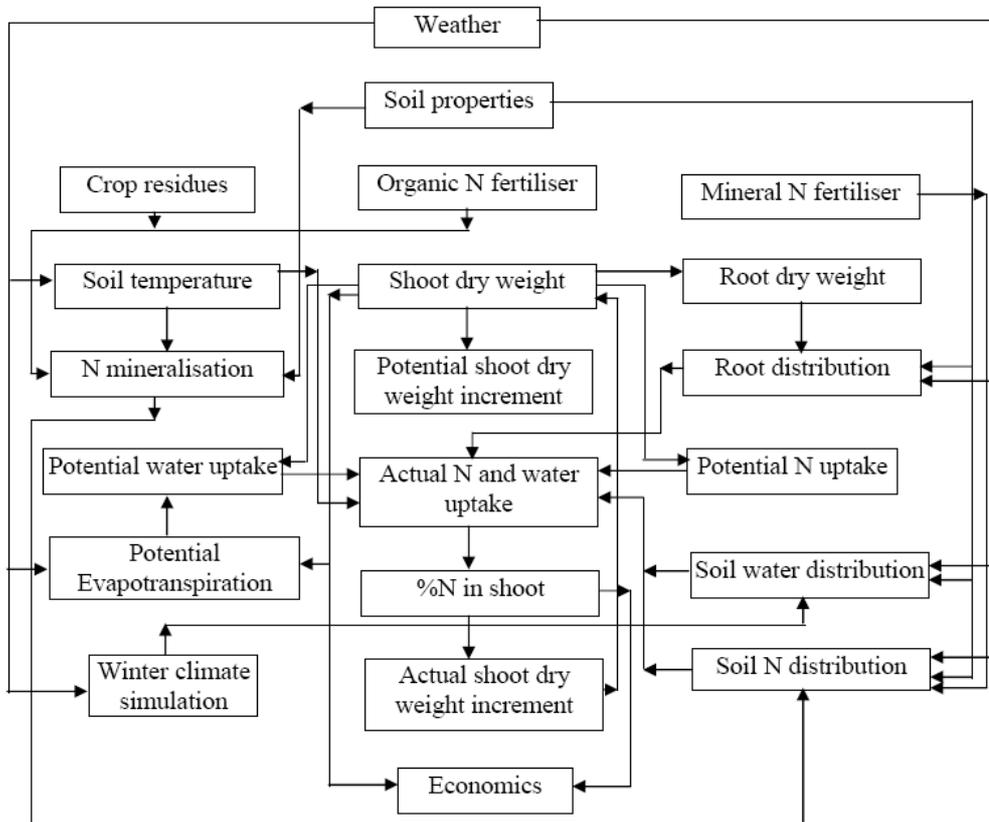
**Table 4**

Soils	Sandy. loam	Sandy. loam	Sand
Rotation name	<b>AS</b>	<b>BS</b>	<b>CS</b>
	early potato	sum. cabbage	sum. cabbage
	early carrot	early. potato	early. potato
	spring wheat	spring wheat	spring wheat
	summer. onion	early cauliflower	early cauliflower
	early carrot	early potato	early potato
	spring barley	spring barley	spring barley
<b>A – Currently Recommended Rates</b>			
Total N fertilizer	110	142	144
Net N mineralisation	59	71	65
Total N uptake	109	133	118
Marketable N offtake	82	99	90
Leaching below 90cm	77	99	109
<sup>1</sup> % Time leaching >0.1	14.1	15.4	16.1
<sup>2</sup> Drainage below 90 cm (mm)	762	749	799
Average Nitrate concentration mg/l	45	58	60
Gr. margin (Euro./ha)	3850	2200	1717
<b>B - Assumed Grower N Rates</b>			
Total N fertilizer	150	183	183
Total N uptake	125	153	143
Marketable N offtake	94	114	107
Leaching below 90cm	105	123	130
<sup>1</sup> % Time leaching >0.1	14.4	15.8	16.2
Average Nitrate <sup>1</sup> concentration mg/l	61	73	72
Gr. margin (Euro/ha)	3933	2350	1967

**Table 5**

<b>Practice</b>	<b>Amount and timing fertilizer kg/ha N</b>	<b>Leaching on 13/5/03 kg/ha N</b>	<b>Gross Margin Cauliflower crop Euro</b>
<b>Recommended Rate (2 splits)</b>	10/4 @ 196 20/5 @ 43	33	- 548
<b>Grower Practice</b>	10/4 @ 237 20/5 @ 52.5	26	555
<b>Modified Recommended Practice (3 splits)</b>	10/4, 5/5, 30/5 @ 80 kg/ha N	12	1360
<b>Regular Feeding 6 applications</b>	<b>10/4,30/4,10/5.20/5,30/5, 9/6 @ 40 kg/ha N</b>	14	2170

Figure 1



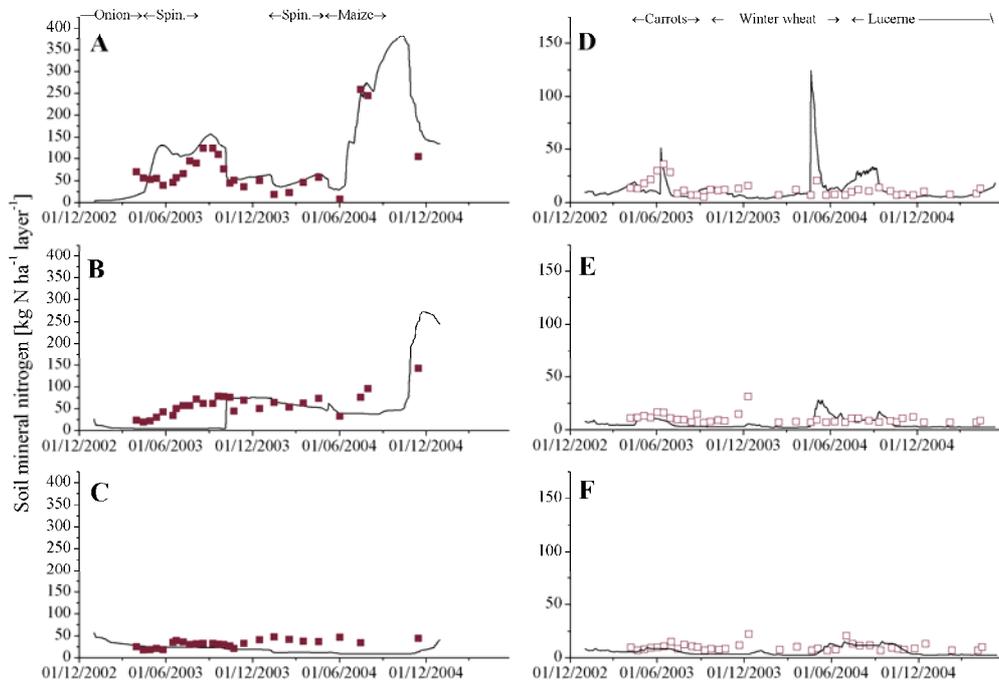


FIGURE 2