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1	Diversity, resource utilization, and phenology of fruiting bodies of
2	litter-decomposing macrofungi in subtropical, temperate, and subalpine forests
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9	
10	AbstractThe diversity, vegetative and reproductive characteristics, and phenology
11	of litter decomposing macrofungi (LDM) were compared between humus forms
12	and climatic regions. Fruiting bodies of LDM were examined for the forest floor of

13 subtropical (ST), cool temperate (CT), and subalpine (SA) forests in Japan. Field

14 surveys during one growing season yielded 35, 32, and 18 species in ST, CT, and

15 SA, respectively. Species richness was generally higher in mull than in moder

- 16 humus and in warmer than in cooler climate. A total of 10 fungal families were
- 17 observed, and species in the Mycenaceae dominated in the LDM assemblages at

18	all study sites. A larger number of species fruited on deeper F layers of the forest
19	floor in SA than in ST, where 74% of species fruited directly on leaf litter. This
20	observation was consistent with the analysis of radiocarbon content in fruiting
21	bodies, implying that LDM tended to utilize older carbon accumulated at deeper
22	layers of the forest floor in cooler climates. Seasonal changes in the fruiting
23	frequency over a growing season exhibited similar two-peak patterns for all the
24	study sites, coinciding with the periods of rainfall and increasing and decreasing
25	air temperatures in early summer and autumn, respectively, but the fruiting
26	period extended longer in warmer than in cooler climate.
27	

28 Keywords Climate · Decomposition · Mycena · Radiocarbon ·
29 Seasonal changes
30
31 Introduction

32

Litter-decomposing macrofungi (LDM) are major components of the diversity of
 soil organisms in terrestrial ecosystems and play central roles in the

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35	decomposition of structural and soluble components in litter that often limit
36	carbon and nutrient cycling in soil (Osono 2007; van der Wal et al. 2013). A suite of
37	LDM with the ability to decompose lignin and other recalcitrant compounds are of
38	particular importance because the colonization of litter materials by these fungi
39	often stimulates the turnover of organic matter and nutrients in soil (Steffen et al.
40	2007; Valášková et al. 2007; Osono et al. 2011a). Fruiting bodies of LDM provide
41	reliable and useful information about their taxonomy, diversity, and reproduction
42	and have been surveyed for their diversity (e.g. Schmit et al. 1999; Mueller et al.
43	2007) and seasonal patterns (Murakami 1989; Yamashita and Hijii 2004) and for
44	the effects on them of vegetation (Hansen and Tyler 1992; Lange 1993; Såstad
45	1995), soil conditions (Tyler 1985; Rastin et al. 1990), and elevational gradient
46	(Gómez-Hernández et al. 2012). Moreover, the observation of vegetative mycelia
47	at the base of fruiting bodies can often yield insights into the substrate utilization
48	and decomposing ability of LDM (Osono et al. 2011a). Currently, however, few
49	studies have investigated the diversity, vegetative and reproductive
50	characteristics, and phenology of LDM simultaneously and compared these
51	between humus forms and climatic regions. It is hypothesized that the pattern of

diversity, substrate utilization, and phenology of fruiting bodies of LDM change
along gradients of soil conditions and climate.

54The purpose of the present study was to investigate fruiting bodies of LDM emerging from the forest floor of subtropical, cool temperate, and subalpine 55forests in Japan. Field sampling of fruiting bodies over growing seasons yielded 56information about the structure, diversity, and species composition of LDM 57assemblages and seasonal patterns of occurrence. Each LDM species was recorded 5859for the soil layer from which its fruiting body emerged to examine the substrate its vegetative mycelia utilized. Radiocarbon (¹⁴C) contents of fruiting bodies were 60 61measured for major LDM species to estimate the age of carbon (i.e. time since 62death of plant litter) utilized by these species. The diameter of the pileus and the 63 length of the stipe were measured for fruiting bodies found in the three forest soils, 64and possible roles of the size variation of fruiting bodies in the seasonal patterns 65of fruiting bodies were discussed.

66

67 Materials and methods

71Samples were collected from three sites in Japan: a subtropical forest (ST), a cool 72temperate forest (CT), and a subalpine forest (SA). ST was located in Okinawa, 73southern Japan. CT was located in Kyoto, Japan. In CT, two study plots were 74established on the upper and lower parts of a northwest-facing slope (approximately 200 m long). SA was located on Mt. Ontake, Gifu, Japan. Details 7576 of the location, climatic conditions, and vegetation are given in Osono (submitted). 77In summary, the three sites differed in mean annual temperature (22°C, 9°C, and 2°C in ST, CT, and SA, respectively), seasonal patterns of change in air 7879temperature, and the duration of the growing season, but they received similar 80 amounts of precipitation annually. The study sites experience a rainy season from May to June in ST and from June to July in CT and SA. Snow covers the forest 81 82 floor of CT from December to April and that of SA from mid-November to early June. Table 1 shows properties of the forest floor of the study sites. The 83 84 accumulation of forest floor material, in terms of the depth and the mass, was in the order: ST, CT (lower) < CT (upper) < SA, whereas the order was generally 85

86	reversed for the leaf fall mass. Consequently, the turnover time of the forest floor
87	was lower in SA and CT (lower) (less than two years) than in CT (upper) (10.4
88	years) and in SA (29.1 years).
89	
90	Study plot and field survey
91	
92	A study plot of 50 \times 10 m (500 m ²) was laid out in each of ST, CT (upper), CT
93	(lower), and SA sites and was divided into 125 grids of 2×2 m. The study area of
94	500 m ² was found to be large enough to describe species richness of macrofungi in
95	CT sites, according to Okabe (1986).
96	Fruiting bodies of LDM were collected in the study plots, seven times at
97	1- to 2-month intervals from March 2007 to January 2008 in ST, nine times at 2-
98	to 4-week intervals from May to November 2001 in CT, and five times at 1-month
99	intervals from June to October 2008 in SA. On each sampling occasion, all fruiting
100	bodies encountered on the surface of the forest floor were recorded, excepting
101	obviously immature or rotting ones. Records were kept of taxa and of grid number
102	and soil horizons (L layer, the border between L and F layers, F layer, or A layer)

from which the fruiting bodies emerged (see Table S1 in Electronic Supplementary Material). Fruiting bodies occurring on logs, twigs, or roots that were fallen or buried were not recorded. Ascomycetes were omitted, but the Xylariaceae on leaf litter were included because of their ligninolytic activity (Osono et al. 2011b). In October and November 2011, fruiting bodies were measured for the diameter of their pileus and length of their stipe at the three sites.

110 Identification was primarily made macroscopically after Imazeki et al. (1988), Imazeki and Hongo (1987, 1989), and Hongo (1994). Small fruiting bodies 111 112of Mycena and Marasmius that were difficult to distinguish and identify at the 113 species level in the field were classified at the genus or section level, which was 114 referred to as species in the present study for the sake of simplicity (but see 115Discussion). Tissues of some fruiting bodies were further analyzed for the DNA 116 sequence of amplicons of rDNA ITS region obtained using primers ITS5 and ITS4 117 (White et al. 1990) and of the 28S rRNA gene D1/D2 region using primers D1 118(Peterson 2000) and NL4 (O'Donnell 1993), according to the method of Hirose and 119 Osono (2006). The sequences determined were compared with the available rDNA sequences in the GenBank database by means of BLAST+ (Camacho et al. 2009)
and assigned taxonomically. The data of molecular analyses will be given in a
future paper.

123 The frequency of occurrence of LDM was calculated as a percentage of 124 incidences based on the number of grids in which the fruiting body was 125 encountered relative to the total number of grids (125) at each study site. Relative 126 frequency of an individual species was calculated as the percentage of its 127 frequency of occurrence with respect to the grand sum of the frequency of 128 occurrence of all species at each study site. Data of fruiting bodies of mycorrhizal 129 fungi were excluded from the following analyses.

130

131 Radiocarbon analysis

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Samples of fruiting bodies were ground in a laboratory mill to make particles that
would pass through a 0.5-mm screen and sent to the Institute of Accelerator
Analysis Ltd, Kanagawa, Japan, for accelerator mass spectrometry
measurements of radiocarbon. The methods are described in Hyodo et al. (2006).

137 Radiocarbon values were reported as Δ^{14} C (‰), which is the part per-thousand 138 deviation from the activity of nineteenth century wood, and corrected for the 139 fractionation using stable carbon (C) isotope ratios of the samples.

140The method to estimate the carbon age of fruiting bodies of LDM followed 141 Hyodo et al. (2006). The carbon age of fruiting bodies of fungi was defined as the 142time elapsed since C in their substrates was fixed from atmospheric CO_2 by primary producers. Δ^{14} C values of samples were compared with those of 143144atmospheric CO₂ recorded at Schauinsland, Germany, for 1976-97 (Levin and Kromer 1997). I estimated the Δ^{14} C values of atmospheric CO₂ after 1997 by 145extrapolation of the exponential function: $\Delta^{14}C(t) = 417 \times \exp(-t/16.0)$, where t is 146147the year after 1974 (Levin and Kromer 1997). This method yielded two estimates of the year of C fixation for the measured Δ^{14} C values of fruiting bodies, one before 148and another after the peak of bomb- Δ^{14} CO₂ in mid-1960s, and hence two carbon 149150ages (Hyodo et al. 2006). I adopted the carbon ages estimated from the year of C fixation after the peak bomb- Δ^{14} C, because these estimated carbon ages were 151compatible with the turnover rates of the forest floor (1.5 to 29.1 years in the 152153study sites, Table 1).

155 Statistical analysis

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157The observed number of LDM species at each study site was denoted as S_{obs}. It 158was possible that S_{obs} be underestimated when the abundance of fruiting bodies of LDM encountered (i.e., the sampling effort) was low at any study site, compared 159160to other sites. To avoid this, I used an individual-based Coleman rarefaction curve 161 (Colwell and Coddington 1994) to depict the cumulative number of species versus 162the observation of fruiting bodies (Fig. S2 in Electronic Supplementary Material). 163In the present study, the number of observation of fruiting bodies was variable 164among the study sites, ranging from 77 at CT (upper) to 620 observations at ST. 165Thus, the study sites were compared for the estimated numbers of LDM species at 77 observations (denoted as Sest). Calculations were performed with R version 166167 3.0.2 for Mac (R Development Core Team 2009) and its vegan package (Oksanen 1682013).

169 Simpson's diversity index (D) and equitability (E) were calculated in the 170 following equations (Osono et al. 2002): $D = 1 / \Sigma P_i^2$, $E = D / S_{obs}$, where P_i was a proportion of the frequency of occurrence of *i*th species to the sum of frequency ofall species.

173	Generalized linear models (GLMs) were used with a Gaussian
174	distribution to compare the size of fruiting bodies of LDM between the study sites.
175	The GLMs were performed with the <i>glm</i> function and with the <i>glht</i> function of the
176	R multcomp package for multiple comparisons with Tukey's test.
177	

- 178 Results
- 179
- 180 Species richness and taxonomic composition
- 181

A total of 35, 32, and 18 species of LDM were observed (S_{obs}) in ST, CT, and SA, respectively; and in CT, 25 and 11 species were observed at lower and upper slopes, respectively (see ESM; summarized in Table 2). The number of singleton species (i.e. species encountered in only one grid) accounted for 17% to 49% of the total number of species, in the order: ST > CT (lower) > CT (upper) > SA (Table 2). Simpson's diversity index was in the order: ST > SA > CT (lower) > CT (upper), and equitability was in the order: CT (upper) > SA > ST > CT (lower) (Table 2).
Rarefaction analysis showed that the estimated number of LDM species (S_{est})
were higher in mull [ST and CT (lower)] than in moder humus [CT (upper) and
SA] (Table 2).

192A total of 10 fungal families were observed: five, seven, five, and five families in ST, CT (lower), CT (upper), and SB, respectively (Table 2). Mycena 193 194species in the Mycenaceae dominated in the LDM assemblages at each study site 195 in terms of the number of species (27% to 50% of the total number of species; Table 196 2) and the relative frequency (Fig. 1). The frequencies of occurrence of two major 197 Mycena species reached more than 90% (i.e. the fruiting bodies of these species 198occurred in more than 90% of the 125 grids) in ST and between 13.6% to 60.8% in 199 CT and SB (see ESM). These major *Mycena* species were followed by species in 200 Marasmiaceae in ST and CT (upper), by species in Agaricaceae in CT (lower), and 201 by species in Hymenogasteraceae in SA (Table 2; Fig. 1).

202

203 Soil layer from which macrofungi fruited

205	Soil layer from which LDM fruited differed among the study sites: more number of
206	LDM species that fruited from deeper layers of the forest floor at cooler climate.
207	That is, 74% (26/35) of species in ST fruited on the surface L layer (i.e. emerging
208	directly on leaf litter), whereas 73% to 92% from the border between L and F
209	layers in CT, and 78% from the F layer in SA (Table 2) did so. Those that fruited
210	on L layer were 'component-restricted' sensu Osono (2007) in that individual
211	mycelia were limited in extent by the physical boundaries of the litter component
212	they occupied. Conversely, those that fruited on the L-F border and F layer were
213	'component-non-restricted' in that the entire forest floor, rather than an
214	individual litter component, provided a habitat for the fungi.
215	
216	Radiocarbon content
217	
218	The mean $\Delta^{14}\mathrm{C}$ values of fruiting bodies ranged between 48.2‰ and 139.7‰ (Table

- 219 3), indicative of the fungal uptake of bomb- Δ^{14} C that was primarily derived from
- 220 the litter that these LDM utilized (Hyodo et al. 2006). The carbon age of fruiting
- 221 bodies from ST ranged from 2.8 to 9.0 years. These values suggested that these

222	LDM utilized leaves that died at least 1.8 to 8.0 years previously, because tree
223	leaves in ST were mostly evergreen and had leaf longevity of more than one year.
224	In contrast, the carbon age of fruiting bodies from CT ranged from 3.6 to 11.4
225	years, suggestive of the utilization of deciduous leaves that died as long as 11.4
226	years before. The carbon age of fruiting bodies from SA reached as old as 20.3
227	years, suggestive of the utilization of evergreen leaves (maximum ages of 6 to 11
228	years, Mori and Takeda 2004) that had died at least 10 years before. These results,
229	together with the results of direct observation of fruiting bodies, suggested that
230	LDM tended to utilize older carbon accumulated at deeper layers of the forest
231	floor in cooler climates.
232	
233	Size and phenology of fruiting bodies
234	
235	The mean size of the fruiting bodies, measured as the diameter of pileus and
236	length of stipe, was significantly different among the study sites (ANOVA, p<0.05),
237	in the order: $CT > SA > ST$ (Table 4). This was accounted for by the difference in
238	size of the major <i>Mycena</i> species in these study sites (Table 4).

239	The frequency of occurrence of fruiting bodies was generally higher in ST
240	than in CT or SA (Fig. 2). Seasonal changes in the frequency over a growing
241	season exhibited similar two-peak patterns for all the study sites: a peak during
242	the rainy season in early summer and another in autumn (Fig. 2). That is, the
243	peaks were found in June and in September to January at ST, in June and in
244	September to November at CT, and in July and in September-October at SA. The
245	number of LDM species followed similar seasonal patterns as the frequency of
246	occurrence of fruiting bodies, except that there was a rapid increase in the
247	number of species in June in ST.

The major LDM species differed in the seasonal patterns of their 248249frequency of occurrence over a growing season (Fig. 3). In ST, fruiting bodies of 250some major species, such as Mycena sp.ST1 and Marasmius sp.ST1, occurred 251relatively constantly over the growing season, whereas Mycena sp.ST2, Xylaria 252sp.ST1, and Crinipellis sp.ST1 displayed fruiting peaks in June and/or in 253September to January. In CT and SA, the frequencies of major species increased 254once in autumn (My. polygramma in CT and My. epipterygia, G. atkinsoniana, and Mycena sp.SA1 in SA) or twice (in early summer and in autumn) (My. filopes 255

256 in CT and *My. aurantiidisca* in SA) over the growing season.

257

258 Discussion

259

The numbers of species of LDM observed in CT and SA (Table 2) were within the 260261range reported previously in temperate and boreal forests (Tyler 1985; Hintikka 2621988; Villeneuve et al. 1989; Brunner et al. 1992; Miyamoto et al. 2000; 263Outerbridge 2002; Richard et al. 2004; Gates et al. 2011; O'Hanlon and Harrington 2012), despite the short survey period (one growing season) at each 264study site. The dominance of Mycena in terms of the number of species and the 265266 frequency of occurrence is consistent with these previous reports. López-Quintero 267 et al. (2012) also observed the occurrence of Mycena and Marasmius species in 268Amazon tropical rainforests, but comparative studies on the diversity of LDM in 269tropical and subtropical forests have been relatively scarce, especially in Asian tropical regions (Mueller et al. 2007). 270

The number of observed and estimated species and Simpson's diversity index of LDM were generally higher in mull than in moder humus (Table 2). This

273	was evident in CT, where the richness of LDM was higher at the lower (mull) than
274	the upper (moder) slope. The two most frequent species were common to the two
275	sites, so that infrequent species accounted for the low similarity of species
276	composition (Table 2, Fig. 1). Rastin et al. (1990) also compared LDM between the
277	lower and upper slope of a spruce forest in Germany and reported that the species
278	composition was generally similar between those parts, in contrast to the results
279	of the present study. This discrepancy may be partly due to the slope length [30 m
280	in Rastin et al. (1990) versus 200 m in the present study]. The causal factors for
281	the higher LDM richness in mull of CT remain unclear, but the relatively
282	favorable moisture condition at the lower slope could possibly favor the fruiting
283	and co-occurrence of more LDM species on the forest floor.
284	The LDM diversity was generally higher in warmer than in cooler
285	climates, suggesting a climatic gradient of diversity. Similar climatic gradients of
286	fungi have been found for litter decomposing microfungi (Osono 2011) and foliar
287	endophytic fungi (Arnold and Lutzoni 2007; Ikeda et al. 2014). At least two
288	explanations are possible for the putative higher diversity of fruiting bodies of
289	LDM in warmer locations. First, the warmer condition throughout the year and

290	lack of snow cover period in winter of ST can favor the fruiting (and possibly, the
291	co-existence) of multiple LDM. This is illustrated in the fruiting phenology of
292	major LDM species (Fig. 3): fruiting bodies occurred throughout the year or with
293	multiple peaks in ST, whereas in CT and SA they peaked once or twice over the
294	growing season. Such differences may be partly due to suitability of the climatic
295	conditions for establishment, growth, and fruiting of more LDM species in ST.
296	Secondly, differences in the quality of resources can also affect the
297	diversity of LDM. I found that LDM from a cooler climate produced fruiting bodies
298	that originated from deeper soil layers than those from a warmer climate (Table 2).
299	In accordance with this, LDM from a cooler climate appeared to utilize more aged
300	dead carbon than those from a warmer climate (Table 3). Given that more
301	decomposed materials in deeper layers contain less readily available organic
302	carbon sources, such as non-lignified holocellulose and soluble carbohydrates
303	(Berg 1986; Osono et al. 2003), the utilization of resources at deeper layers in a
304	cooler climate may be unfavorable for the growth and fruiting of LDM. Osono
305	(2011) demonstrated that non-ligninolytic microfungi are major components of
306	fungal assemblages on recently fallen litter in a cooler climate, suggesting that

307	ligninolytic LDM are less competitive for readily available resources in the
308	surface litter in a cooler climate (Osono 2007). The dominance in SA of conifers,
309	whose needles are rich in secondary compounds that inhibit the growth of LDM
310	(Bağci and Diğrak 1996), can also affect the colonization of the L layer by LDM.
311	Two peaks were found for the occurrence of fruiting bodies over a growing
312	season at all three climates, but the fruiting period extended longer at warmer
313	than at cooler climates (Fig. 2). The two peaks coincided with the period of rainfall
314	and with the increasing and decreasing air temperatures in early summer and
315	autumn, respectively, at the three forest sites. Similar one- or two-peak patterns
316	of fruiting of LDM have commonly been found in temperate forests (Okabe 1983;
317	Straatsma et al. 2001; Yamashita and Hijii 2004; Gates et al. 2011).
318	The fruiting phenology of major Mycena species may also be associated
319	with the size of fruiting bodies. For example, <i>Mycena</i> species with smaller fruiting
320	bodies fruited more frequently over a growing season than those with larger ones;
321	<i>M. filopes</i> in CT and <i>M. aurantiidisca</i> in SA with smaller fruiting bodies showed
322	two peaks, whereas <i>M. polygramma</i> in CT and <i>M. epipterygia</i> in SA with larger
323	fruiting bodies showed only one peak in autumn (Fig. 3, Table 4), and even

324	smaller Mycena sp. ST1 and ST2 in ST fruited throughout the growing season.
325	Because the production of larger fruiting bodies should require more resources to
326	be utilized, the size of fruiting bodies can set a limit on the reproduction. This
327	discussion is obviously speculative, however, as few data have been available
328	regarding the population structure and reproductive biology of individual LDM
329	species. More studies are needed to examine the life history strategy of LDM,
330	especially in tropical and subtropical regions.

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References

344	
345	Arnold AE, Lutzoni F (2007) Diversity and host range of foliar endophytes: are
346	tropical leaves biodiversity hotspots? Ecology 88:541-549
347	Bağci E, Diğrak M (1996) Antimicrobial activity of essential oils of some Abies
348	(Fir) species from Turkey. Flavour Fragr J 11:251-256
349	Berg B (1986) Nutrient release from litter and humus in coniferous forest soils - a
350	mini review. Scand J For Res 1:359-369
351	Brunner I, Brunner F, Laursen GA (1992) Characterization and comparison of
352	macrofungal communities in an Alnus tenuifolia and an Alnus crispa
353	forest in Alaska. Can J Bot 70:1247-1258
354	Camacho C, Coulouris G, Avagyan V, Ma N, Papadopoulos J, Bealer K, Madden
355	TL (2009) BLAST+: architecture and applications. BMC Bioinformatics
356	10:421

357 Cowell RK, Coddington JA (1994) Estimating terrestrial biodiversity through

358extrapolation. Phil Trans R Soc B 345:101-118 Fukasawa Y, Katsumata S, Mori AS, Osono T, Takeda H (2014) Accumulation and 359360 decay dynamics of coarse woody debris in a Japanese old-growth 361subalpine coniferous forest. Ecol Res 29:257-269 Gates GM, Mohammed C, Wardlaw T, Davidson NJ, Ratkowsky DA (2011) 362 363 Diversity and phenology of the macrofungal assemblages supported by 364litter in a tall, wet Eucalyptus obliqua forest in southern Tasmania, 365 Australia. Fungal Ecol 4:68-75 Gómez-Hernández M, Williams-Linera G, Guevara R, Lodge DJ (2012) Patterns of 366 367 macromycete community assemblage along an elevation gradient: 368 options for fungal gradient and metacommunity analyses. Biodivers 369 Conserv 21:2247-2268 370Hansen PA, Tyler G (1992) Statistical evaluation of tree species affinity and soil 371 preference of the macrofungal flora in south Swedish beech, oak and 372 hornbeam forest. Crypt Bot 2:355-361 373 Hintikka V (1988) On the macromycete flora in oligotrophic pine forests of different ages in south Finland. Acta Bot Fennica 136:89-94 374

375	Hirose D, Osono T (2006) Development and seasonal variations of <i>Lophodermium</i>
376	populations on <i>Pinus thunbergii</i> needle litter. Mycoscience 47:242-247
377	Hongo T (1994) Fungi. Yama to Keikoku Sha, Tokyo, Japan (in Japanese)
378	Hyodo F, Tayasu I, Wada E (2006) Estimation of the longevity of C in terrestrial
379	detrital food webs using radiocarbon (14C): how old are diets in termites?
380	Functional Ecol 20:385-393
381	Ikeda A, Matsuoka S, Masuya H, Mori AS, Hirose D, Osono T (2014) Comparison
382	of the diversity, composition, and host recurrence of xylariaceous
383	endophytes in subtropical, cool temperate, and subboreal regions in
384	Japan. Popul Ecol in press
385	Imazeki R, Hongo T (1987) Colored Illustration of Mushrroms of Japan. Vol. I.
386	Hoikusha, Tokyo, Japan (in Japanese)
387	Imazeki R, Hongo T (1989) Colored Illustration of Mushrooms of Japan. Vol. II.
388	Hoikusha, Tokyo, Japan (in Japanese)
389	Imazeki R, Otani Y, Hongo T (1988) Fungi of Japan. Yama to Keikoku Sha, Tokyo,
390	Japan (in Japanese)
391	Lange M (1993) Maromycetes under twelve tree species in ten plantations on

392	various soil types in Denmark. Opera Bot 120:1-53
393	Levin I, Kromer B (1997) Twenty years of atmospheric ${ m ^{14}CO_2}$ observations at
394	Schauinsland station, Germany. Radiocarbon 39:205-218
395	López-Quintero CA, Straatsma G, Franco-Molano AE, Boekhout T (2012)
396	Macrofungal diversity in Colombian Amazon forests varies with regions
397	and regimes of disturbance. Biodivers Conserv 21:2221-2243
398	Miyamoto T, Igarashi T, Takahashi K (2000) Lignin-degrading ability of
399	litter-decomposing basidiomycetes from Picea forests of Hokkaido.
400	Mycoscience 41:105-110
401	Mori A, Takeda H (2004) Functional relationships between crown morphology and
402	within-crown characteristics of understory saplings of three codominant
403	conifers in a subalpine forest in central Japan. Tree Physiol 24:661-670
404	Mueller GM, Schmit JP, Leacock PR, Buyck B, Cifuentes J, Desjardin DE, Halling
405	RE, Hjortstam K, Iturriaga T, Larsson KH, Lodge DJ, May TW, Minter D,
406	Rajchenberg M, Redhead SA, Ryvarden L, Trappe JM, Watling R, Wu Q
407	(2007) Global diversity and distribution of macrofungi. Biodivers Conserv
408	16:37-48

409	Murakami Y (1989) Spatial changes of species composition and seasonal fruiting
410	of the Agaricales in <i>Castanopsis cuspidata</i> forest. Trans Mycol Soc Japan
411	30:89-103
412	O'Donnell K (1993) <i>Fusarium</i> and its near relatives. In: Reynolds DR, Taylor JW
413	(eds) The fungal holomorph: mitotic, meiotic and pleomorphic speciation
414	in fungal systematics. CAB International, Wallingford, UK, pp 225-233
415	O'Hanlon R, Harrington TJ (2012) Macrofungal diversity and ecology in four Irish
416	forest type. Fungal Ecol 5;499-508
417	Okabe H (1983) Mycosociological research of Agaricales in natural forests (II)
418	Seasonal changes on each stand and life form. Bull Kyoto Univ Forest
419	53:20-32 (in Japanese with English abstract)
420	Okabe H (1986) Ecological study of distribution of fungi within forests. PhD thesis,
421	Kyoto University, Kyoto (in Japanese)
422	Oksanen J (2013) Multivariate analysis of ecological communities in R: vegan
423	tutorial. http://cc.oulu.fi/~jarioksa/opetus/metodi/vegantutor.pdf
424	(accessed 14.5.14)
425	Osono T (2007) Ecology of ligninolytic fungi associated with leaf litter

426 decomposition. Ecol Res 22:955-974 427Osono T (2011) Diversity and functioning of fungi associated with leaf litter 428 decomposition in Asian forests of different climatic regions. Fungal Ecol 429 4:375-385 Osono T. Mycelial biomass in the forest floor and soil of subtropical, temperate, 430 431and subalpine forests. J For Res:submitted 432Osono T, Hobara S, Fujiwara S, Koba K, Kameda K (2002) Abundance, diversity, 433 and species composition of fungal communities in a temperate forest affected by excreta of the Great Cormorant Phalacrocorax carbo. Soil Biol 434 Biochem 34:1537-1547 435436 Osono T, Ono Y, Takeda H (2003) Fungal ingrowth on forest floor and 437 decomposing needle litter of Chamaecyparis obtusa in relation to 438 resource availability and moisture condition. Soil Biol Biochem 439 35:1423-1431 T, Hobara S, Hishinuma T, Azuma JI (2011a) Selective lignin 440 Osono 441 decomposition and nitrogen mineralization in forest litter colonized by *Clitocybe* sp. Eur J Soil Biol 47:114-121 442

443	Osono T, To-Anun C, Hagiwara Y, Hirose D (2011b) Decomposition of wood, petiole
444	and leaf litter by Xylaria species from northern Thailand. Fun Ecol
445	4:210-218
446	Outerbridge RAM (2002) Macrofungus ecology and diversity under different
447	conifer monocultures on southern Vancouver Island. PhD thesis,
448	University of Victoria
449	Peterson SW (2000) Phylogenetic analysis of <i>Penicillium</i> species based on ITS and
450	lsu-rDNA nucleotide sequences. In: Samson RA, Pitt JI (eds) Integration
451	of modern taxonomic methods for <i>Penicillium</i> and Aspergillus
452	classification. Harwood, Amsterdam, the Netherland, pp 163-178
453	Rastin N, Schlechte G, Hüttermann A (1990) Soil macrofungi and some soil
454	biological, biochemical and chemical investigationson the upper and
455	lower slope of a spruce forest. Soil Biol Biochem 22:1039-1047
456	R Development Core Team (2009) R: a language and environment for statistical
457	computing. http://www.r-project.org/ (accessed 14.5.14)
458	Richard F, Moreau PA, Selosse MA, Gardes M (2004) Diversity and fruiting
459	patterns of ectomycorrhizal and saprobic fungi in an old-growth

460	Mediterranean forest dominated by Quercus ilex L. Can J Bot
461	82:1711-1729
462	Såstad SM (1995) Fungi - vegetation relationships in a <i>Pinus sylvestris</i> forest in
463	central Norway. Can J Bot 73:807-816
464	Schmit JP, Murphy JF, Mueller GM (1999) Macrofungal diversity of a temperate
465	oak forest: a test of species richness estimators. Can J Bot 77:1014-1027
466	Steffen KT, Cajthaml T, Šnajdr J, Baldrian P (2007) Differential degradation of
467	oak (Quercus petraea) leaf litter by litter-decomposing basidiomycetes.
468	Res Microbiol 158:447-455
469	Straatsma G, Ayer F, Egli S (2001) Species richness, abundance, and phenology of
470	fungal fruit bodies over 21 years in a Swiss forest plot. Mycol Res
471	105:515-523
472	Takeda H, Kaneko N (1988) Patterns of soil humus accumulation in forests. I.
473	Mull and moder types humus in a broad-leaved forest. Bull Kyoto Univ
474	Forest 60:33-45 (in Japanese with English abstract)
475	Tian X, Takeda H, Ando T (1997) Application of a rapid thin section method for
476	observations on decomposing litter in mor humus form in a subalpine

coniferous forest. Ecol Res 12:289-300

478	Tsukamoto J (1996) Soil macro-invertebrates and litter disappearance in a
479	Japanese mixed deciduous forest and comparison with European
480	deciduous forests and tropical rainforests. Ecol Res 11:35-50
481	Tyler G (1985) Macrofungal flora of Swedish beech forest related to soil organic
482	matter and acidity characteristics. For Ecol Manag 10:13-29
483	Valášková V, Šnajdr J, Bittner B, Cajthaml T, Merhautová V, Hofrichter M,
484	Baldrian P (2007) Production of lignocellulose-degrading enzymes and
485	degradation of leaf litter by saprotrophic basidiomycetes isolated from a
486	Quercus petraea forest. Soil Biol Biochem 39:2651-2660
486 487	<i>Quercus petraea</i> forest. Soil Biol Biochem 39:2651-2660 Van der Wal A, Geydan TD, Kuyper TW, de Boer W (2013) A thready affair:
487	Van der Wal A, Geydan TD, Kuyper TW, de Boer W (2013) A thready affair:
487 488	Van der Wal A, Geydan TD, Kuyper TW, de Boer W (2013) A thready affair: linking fungal diversity and community dynamics to terrestrial
487 488 489	Van der Wal A, Geydan TD, Kuyper TW, de Boer W (2013) A thready affair: linking fungal diversity and community dynamics to terrestrial decomposition processes. FEMS Microbiol Rev 37:477-494
487 488 489 490	 Van der Wal A, Geydan TD, Kuyper TW, de Boer W (2013) A thready affair: linking fungal diversity and community dynamics to terrestrial decomposition processes. FEMS Microbiol Rev 37:477-494 Villeneuve N, Grandtner MM, Fortin JA (1989) Frequency and diversity of

494	of fungal ribosomal RNA genes for phylogenetics. In: Innis MA, Gelfand
495	DH, Sninsky JJ, White TJ (eds) PCR Protocols: a Guide to Methods and
496	Applications. Academic Press, New York, USA, pp 315-322
497	Xu X, Tokashiki Y, Enoki T, Hirata E (1998a) Characteristics of nutrient
498	accumulation in forest floor under evergreen broadleaved forests in
499	Okinawa Island. Sci Bull Fac Agr Univ Ryukyus 45:185-193
500	Xu X, Enoki T, Tokashiki Y, Hirata E (1998b) Litterfall and the nutrient returns in
501	evergreen broadleaved forests in Northern Okinawa Island. Sci Bull Fac
502	Agr Univ Ryukyus 45:195-208
503	Yamashita S, Hijii N (2004) Relationships between seasonal appearance and
504	longevity of fruitbodies of Agaricales and meteorological factors in a
505	Japanese red pine forest. J For Res 9:165-171

Site	ST	СТ	SA
Humus type	Mull Mull (Moder ^b
		Moder (upper) ^a	
Depth of L layer (cm) ^c	1.1 ± 0.1	1.2 ± 0.1	2.7 ± 0.3
Depth of FH layer (cm) ^c	1.0 ± 0.2	4.0 ± 0.4	13.4 ± 1.2
Forest floor mass (Mg/ha)	12.0^{d}	7.7 (lower)	104.6^{f}
		33.3 (upper) ^e	
Leaf fall mass (Mg/ha/yr)	7.95^{d}	4.10 (lower)	3.59^{f}
		3.20 (upper) e	
Turnover time (yr) ^g	1.5	1.5 1.9 (lower)	
		10.4 (upper)	

Table 1. Forest floor and field survey of fruiting bodies in the study sites.

^aTakeda and Kaneko (1988). ^bTian et al. (1997). ^cValues are means ± standard errors (n=20). Measurement was carried out in the three study sites in October 2012. Values of CT were from the lower slope. ^dXu et al. (1998a, 1998b). ^eTsukamoto (1996). ^fFukasawa et al. (2014). ^gTurnover time = forest floor mass / annual leaf fall mass.

Table 2. Assemblage structure and family composition of macrofungi and the soil layer from which fruiting bodies occurred. Numbers of macrofungal species are indicated for fungal families and the litter layers. Numbers in parentheses indicate the proportion relative to the observed number of species. Values of S_{est} indicate means \pm standard deviations. See text for S_{est} .

	ST	CT (lower)	CT (upper)	SA	
Diversity					
Observed number of	35	25	11	18	
species (S _{obs})					
Singleton species	17 (49)	11 (44)	4 (36)	3 (17)	
Simpson's D	7.68	4.47	3.40	4.86	
Equitability	0.22	0.18	0.31	0.27	
Estimated number	14.2 ± 2.1	17.2 ± 2.0	11.0 ± 0.0	11.9 ± 1.7	
of species (S _{est})					
Family composition					
Mycenaceae	13 (37)	10 (40)	3 (27)	9 (50)	
Marasmiaceae	12 (34)	3 (12)	3 (27)	2 (11)	
Agaricaceae	3 (9)	6 (24)	2 (18)	1 (6)	
Tricholomataceae	5 (14)	2 (8)	0 (0)	3 (17)	
Strophariaceae	0 (0)	1 (4)	2 (18)	0 (0)	
Psathyrellaceae	0 (0)	2 (8)	0 (0)	0 (0)	
Pluteaceae	0 (0)	1 (4)	0 (0)	0 (0)	
Hygrophoraceae	0 (0)	0 (0)	1 (9)	0 (0)	
Hymenogasteraceae	0 (0)	0 (0)	0 (0)	1 (6)	
Xylariaceae	1 (3)	0 (0)	0 (0)	0 (0)	
Unidentified	1 (3)	0 (0)	0 (0)	2 (11)	
Soil layer					
L layer	26 (74)	0 (0)	0 (0)	0 (0)	
L-F border	9 (26)	23 (92)	8 (73)	3 (17)	
F layer	0 (0)	0 (0)	1 (9)	14 (78)	
Alayer	0 (0)	2 (8)	2 (18)	1 (6)	

Osono Table 3

	Species	Soil layer	Collection date	Laboratory code	$\delta^{13}C$	$\Delta^{14}\mathrm{C}$	Carbon age (yr)
ST	<i>Mycena</i> sp. ST1	L layer	Apr 11	IAAA-111556	-27.3 ± 0.5	48.2 ± 2.8	2.8
	<i>Gymnopus</i> sp. ST1	L layer	Oct 11	IAAA-111557	-30.4 ± 0.5	$52.0~\pm~3.0$	4.5
	<i>Marasmius</i> sp. ST4	L-F border	Oct 11	IAAA-111558	-22.9 ± 0.4	$68.6~\pm~2.7$	9.0
\mathbf{CT}	Mycena polygramma	L-F border	Oct 01	IAAA-81685	-25.5 ± 0.3	132.6 ± 3.7	9.5
	Mycena amygdalina	L-F border	Nov 11	IAAA-111554	-20.2 ± 0.4	$79.6~\pm~2.8$	11.4
	Gymnopus peronatus	L-F border	Nov 01	IAAA-111555	-24.3 ± 0.4	$91.0~\pm~2.7$	3.6
SA	Mycena aurantiidisca	L-F border	Oct 08	IAAA-81687	-24.9 ± 0.4	$96.0~\pm~3.8$	11.3
	Mycena epipterygia	L-F border	Oct 11	IAAA-111552	-23.0 ± 0.4	139.7 ± 2.8	20.3
	Galerina atkinsoniana	L-F border	Oct 11	IAAA-111553	$-28.9~\pm~0.4$	56.7 ± 2.7	5.9

Table 3. Radiocarbon content in fruiting bodies of macrofungi. Samples from CT were from the lower slope.

Table 4. Size of fruiting bodies of macrofungi. DP, diameter of pileus; LS, length of stipe. Values are means ± standard errors in cm. Numbers in parentheses indicate the number of samples. The same letters indicate that the values are not significantly different at 5% level by Tukey's HSD test. Data from CT are from the lower slope.

	Total		Major species 1		Major species 2		Major species 3	
	DP	LS	DP	LS	DP	LS	DP	LS
ST	3.7 ± 1.3 b	$13.9\pm2.3~\mathrm{c}$	1.6 ± 0.3	11.3 ± 1.1	4.6 ± 0.7	12.0 ± 0.8	22.5	46.0
	(16)		Mycena sp. ST1 and ST2 (10)		<i>Gymnopus</i> sp. ST1 (4)		<i>Marasmius</i> sp. ST4 (1)	
\mathbf{CT}	12.0 ± 1.2 a	63.8 ± 5.7 a	3.8	30.0	14.0	116.5	15.8 ± 2.3	73.5 ± 7.0
	(23)		Mycena filopes (2)		Mycena polygramma (2)		Mycena crocata (6)	
SA	$5.8\pm0.5~\mathrm{b}$	$35.1\pm2.1~\mathrm{b}$	4.1 ± 0.3	30.9 ± 2.3	6.9 ± 0.5	46.9 ± 5.6	4.6 ± 0.7	25.5 ± 3.3
	(38)		Mycena aurantiidisca (15)		Mycena epipterygia (6)		Galerina atkinsoniana (8)	

1 Figure legends

 $\mathbf{2}$

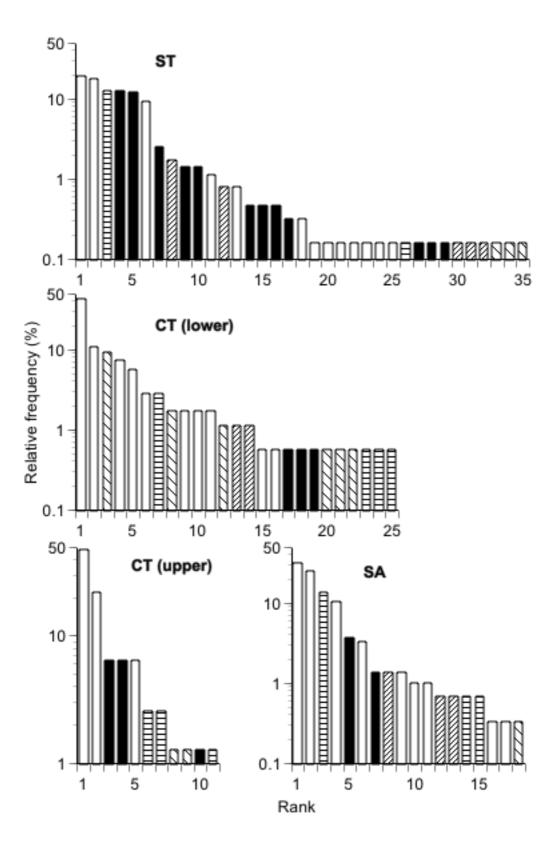
3 Fig. 1. Rank-relative frequency distribution of macrofungal assemblages in subtropical (ST), cool temperate (CT), and subalpine forests (SA). Open, 4 $\mathbf{5}$ Mycenaceae; filled, Marasmiaceae; coarse oblique mesh, Agaricaceae; fine oblique 6 Tricholomataceae; horizontal (Hygrophoraceae, mesh, mesh, others 7Hymenogasteraceae, Pluteaceae, Psathyrellaceae, Strophariaceae, Xylariaceae, 8 and unidentified). The survey in CT was performed at lower and upper parts of a 9 forest slope.

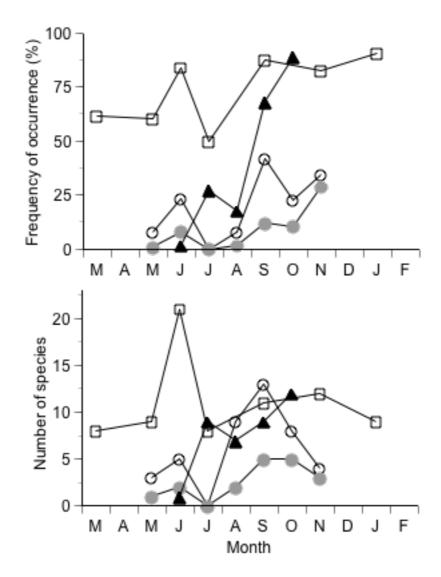
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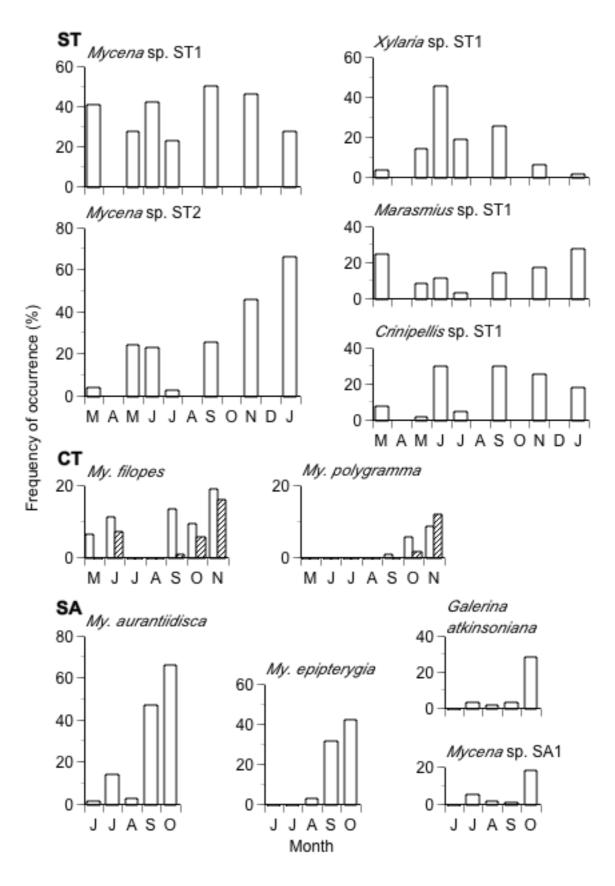
Fig. 2. Seasonal changes in the frequency of occurrence (upper) and the number of
species (lower) of fruiting bodies of macrofungi. □, subtropical forest (ST); ○,
lower part of a slope in cool temperate forest, [CT (lower)]; ●, upper part of a slope
in cool temperate forest [CT (upper)]; ▲, subalpine forest (SA).

15

Fig. 3. Seasonal changes in the frequency of occurrence of major macrofungal
species in subtropical forest (ST), cool temperate forest (CT), and subalpine forest
(SA). For CT, blank and shaded bars indicate lower and upper slopes, respectively.







Electronic Supplementary Material

Patterns in diversity, resource utilization, and phenology of fruiting bodies of litter-decomposing macrofungi in subtropical, temperate, and subalpine forests

Takashi Osono

Table S1: Frequency of occurrence of fruiting bodies of macrofungi and the soil layer from which the fruiting bodies occurred. Ag, Agaricaceae; Hg, Hygrophoraceae; Hm, Hymenogasteraceae; Mr, Marasmiaceae; My, Mycenaceae; Tr, Tricholomataceae; Pl, Pluteaceae; Ps, Psathyrellaceae; St, Strophariaceae; Xy, Xylariaceae; and Un, unidentified. L/F, L-F border.

Taxa	Family	Soil layer	Frequency (%)
Subtropical forest			
Mycena section Basipedes 'sp. ST1'	My	L	96.8
Mycena section Roridae 'sp. ST2'	My	L	90.4
<i>Xylaria</i> spp. ST1	Xy	L	64.8
Marasmius spp. ST1	Mr	L	63.2
Crinipellis sp. ST1	Mr	L	60.8
Mycena sp. ST3	My	L	46.4
Gymnopus sp. ST1	Mr	L	12.8
Tricholomataceae sp. ST1	Tr	А	8.8
Gymnopus sp. ST2	Mr	L	7.2
Marasmiellus sp. ST1	Mr	L	7.2
Mycena sp. ST4	My	L	5.6
Tricholomataceae sp. ST2	Tr	А	4.0
Mycena sp. ST5	My	L	4.0
Crinipellis sp. ST2	Mr	L	2.4
Gymnopus sp. ST3	Mr	L	2.4
cf. Calyptella sp. ST1	Mr	L	2.4

Marasmius sp. ST2	Mr	L	1.6	
<i>Mycena</i> sp. ST6	My	L	1.6	
Agaricaceae sp. ST1	Ag	А	0.8	
<i>Agaricaceae</i> sp. ST2	Ag	А	0.8	
<i>Gymnopus</i> sp. ST4	Mr	L	0.8	
<i>Mycena</i> sp. ST7	My	L	0.8	
Leucocorpinus sp. ST1	Ag	А	0.8	
Marasmiellus sp. ST2	Mr	А	0.8	
Marasmiellus sp. ST3	Mr	L	0.8	
<i>Mycena</i> sp. ST8	My	L	0.8	
<i>Mycena</i> sp. ST9	My	L	0.8	
Mycena sp. ST10	My	L	0.8	
Unidentified ST1	Un	L	0.8	
Mycena sp. ST11	My	А	0.8	
<i>Mycena</i> sp. ST12	My	L	0.8	
Tricholomataceae sp. ST3	Tr	А	0.8	
Tricholomataceae sp. ST4	Tr	А	0.8	
Tricholomataceae sp. ST5	Tr	L	0.8	
Xeromphalina sp. ST1	My	L	0.8	
Cool temperate forest			Upper	Lower
Mycena amygdalina	My	L/F	29.6	60
Mycena polygramma	My	L/F	13.6	15.2
Gymnopus peronatus	Mr	L/F	4.0	0.8
Gymnopus sp. CT1	Mr	L/F	4.0	0.0
<i>Mycena</i> sp. CT2	My	L/F	4.0	0.0
Stropharia aeruginosa	St	L/F	1.6	0.8
Hygrocybe cantharellus	Hy	L/F	1.6	0.0
Agaricaceae sp. CT1	Ag	L/F	0.8	0.0
Lepiota fusciceps	Ag	F	0.8	0.0
Marasmius sp. CT2	Mr	А	0.8	0.0
Naematoloma sublateritium	St	А	0.8	0.0
Lycoperdon perlatum	Ag	А	0.0	12.8

Mycena pura	My	L/F	0.0	10.4
Mycena sp. CT1	Му	L/F	0.0	8.0
Mycena amicta	My	L/F	0.0	4.0
Psathyrella candolleana	Ps	L/F	0.0	4.0
Agariaceae sp.CT2	Ag	L/F	0.0	2.4
Mycena crocata	My	L/F	0.0	2.4
Mycena luteopallens	My	L/F	0.0	2.4
<i>Mycena</i> sp. CT3	My	L/F	0.0	2.4
Agaricus praeclaresquamosus	Ag	L/F	0.0	1.6
Clitocybe sp. CT1	Mr	L/F	0.0	1.6
Pseudoclitocybe cyathiformis	Tr	L/F	0.0	1.6
Lepiota cf. pseudogranulosa	Ag	L/F	0.0	0.8
Lepiota cygnea	Ag	L/F	0.0	0.8
Lepiota sp. CT1	Ag	А	0.0	0.8
Marasmius pulcheriipes	Mr	L/F	0.0	0.8
Marasmius sp. CT1	Mr	L/F	0.0	0.8
Mycena cf. osmundicola	My	L/F	0.0	0.8
<i>Mycena</i> sp. CT4	My	L/F	0.0	0.8
Psathyrella piluliformis	Ps	L/F	0.0	0.8
Volvariella speciosa var. gloiocephala	Pl	L/F	0.0	0.8
Subalpine forest				
Mycena aurantiidisca	My	F	75.2	
Mycena epipterygia	My	F	60.8	
Galerina atkinsoniana	Hm	F	32.8	
Mycena cf. filopes	My	L/F	24.8	
Clitocybe sp. SA1	Mr	F	8.8	
Mycena sp. SA2	My	L/F	8.0	
Marasmius androsaceus	Mr	L/F	3.2	
Tricholomataceae sp. SA1	Tr	F	3.2	
Mycena cf. stipata	My	F	3.2	
Mycena sp. SA3	My	F	2.4	
Mycena cf. pura	My	F	2.4	

Collybia cookei	Tr	F	1.6
Tricholomataceae sp. SA2	Tr	F	1.6
Unidentified SA1	Un	F	1.6
Unidentified SA2	Un	F	1.6
Mycena sp. SA5	My	F	0.8
Mycena sp. SA4	My	F	0.8
Lycoperdon perlatum	Ag	А	0.8

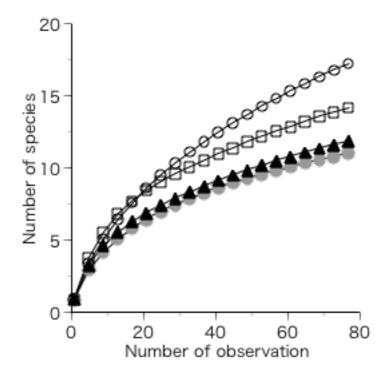


Fig. S2. Rarefaction curves for litter-decomposing macrofungal (LDM) assemblages. □, subtropical forest (ST); ○, lower part of a slope in cool temperate forest, [CT (lower)];
, upper part of a slope in cool temperate forest [CT (upper)]; ▲, subalpine forest (SA).