# **Seasonal variations in population dynamics and biomass of two**  *Unio pictorum mancus* **(Mollusca, Unionidae) populations from two lakes of different trophic state**

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#### *ABSTRACT*

*Samples of* Unio pictorum mancus *collected monthly from spring 2003 to summer 2004 in the meso-oligotrophic Lake Maggiore and the eutrophic Lake Candia were compared to evaluate the influence of lake trophic conditions on mussel populations. Shell form, maximum age (8 years) and percentage of organic matter in the shell and soft tissues were similar in both lakes. However, comparisons between same size classes revealed that mussels from Lake Maggiore were generally older, their shells heavier and their soft tissues lighter than those from Lake Candia. Recruitment occurred in both populations in June-July. The frequency distributions of adult mussels (>30 mm) and their soft tissues and shell biomasses in all size classes (range = 2 mm) were normal, but the curves of the mussels from Lake Candia were flat in comparison to those of mussels from Lake Maggiore. The median body size of Lake Maggiore mussels was 58 mm, while that of the Lake Candia population was 72 mm. During the study period the population density of each lake was fairly constant, whereas the population structure, and consequently the biomass, showed seasonal variations. The greatest differences were found between population density and shell and tissue biomass of the two populations, which were respectively 46.86 ind. m-2; 408 g d.w. m-2 and 38.67 g d.w. m-2 in Lake Maggiore, and 6.93 ind. m-2; 92.29 g* d.w.  $m^2$  and 14.04 g d.w.  $m^2$  in Lake Candia. These values testify to the phenotypical plasticity of Unio pictorum mancus in *relation to environmental characteristics, without excluding a possible genetic influence as a result of geographical isolation and the respective selection mechanisms. The influence exerted by the lake trophic level and by the physical environment on some population characteristics is discussed.* 

*Key words: freshwater bivalve, eutrophication, population density and structure, shell and soft tissues biomasses* 

#### 1. INTRODUCTION

Much more information is available on the population ecology of commercial marine bivalves than on freshwater mussel populations, mainly because the latter are exploited to any great extent only in very restricted areas and for relatively limited periods. As an example, until the middle of the XX century freshwater mussels were exploited principally to produce nacre buttons, for natural pearls and to feed poultry (Tudorancea 1972; Bowen *et al*. 1994). Mussel populations accordingly declined in areas where these were important activities. Nowadays, the decline of many species is the consequence of eutrophication (Patzner & Müller 2001), toxic pollution, running water regulation and the increasing frequency and extent of invasions by allochthonous bivalve species (e.g., *Dreissena polymorpha*). For instance, following invasion by *Dreissena*, more than 70% of North American bivalve species are declining, or are already extinct (Bauer & Wachtler 2001); this has caused changes in the benthic communities, with related effects on the diet and growth of fish (McNickle *et al*. 2006).

Studies on freshwater bivalves have been increasing steadily since the 1960s, when they began to be used as indicators of radioactive and stable pollutants in the environment (Gaglione & Ravera 1964; Bedford *et al*. 1968; Manly & George 1977; Millington & Walker 1983; Storey & Edward 1989). In addition, various environmental warning systems using valve movement of some freshwater mussels have been adopted (England & Heino 1994; Ham & Peterson 1994; Borcherding 2006). During the last decades, in addition to researches on freshwater mussels as pollutant concentrators, studies on populations in the field have yielded a large amount of information; for example, density, biomass, and age distribution were considered in relation to mussel size and production rate. Negus (1966) provided the first quantitative data on growth rate and production of unionids, while Magnin & Stanczykowska (1971) were the first to quantify freshwater mussel production in Canadian lakes.

Mussel soft tissues may be utilized as food by various predators, and after the mussel's death its tissues are degraded in about two weeks (Lomstein *et al*. 2006). In addition, the considerable amount of calcium carbonate hitherto immobilized in the shell can be returned to the environment. Mussels mineralize a great amount of organic matter through their high respiration rate, and take up suspended particles by their filtration activity (Hily 1991), while the indigestible fraction is released in the form of pellets which modify the structure of the



**Fig. 1**. Sampling stations (white spots) on Lake Maggiore (a) and Lake Candia (b).

sediment surface (Wotton 2006). In addition, by filtration mussels reduce the seston concentration, and by excretion release ammonia, a useful nutrient for phytoplankton, into the water. Thus, at high population densities, mussels play an important role in the conversion of energy and matter (Welker & Walz 1998; Vaughn & Hakenkamp 2001).

In consideration of the important ecological role of freshwater bivalves, a research project was designed to evaluate the possible influence of lake trophic state on some important biological (e.g., population density, tissue and shell biomasses, variations in population structure) and chemical (e.g., capacity to accumulate metals in relation to their concentrations in the water) characteristics of mussel populations. To this end, these characteristics were measured in two *Unio pictorum mancus* populations from two Northern Italian lakes: the mesooligotrophic Lake Maggiore and the eutrophic Lake Candia. The present paper deals with the biological characteristics; the chemical characteristics are reported in another paper (Ravera *et al*. 2007).

### 2. MATERIAL AND METHODS

### *2.1. Study sites*

The study was carried out in two Northern Italian lakes (Fig. 1; Tab.1): the deep, meso-oligotrophic Lake Maggiore (latitude N 45°50'19"; longitude E 8°37'17") and the shallow, eutrophic Lake Candia (latitude N 45°19'32"; longitude E 7°54'38"). One sampling station

with the following characteristics was chosen in each lake: similar climate, negligible human impact, moderate amount of toxic pollutants, and abundnce of *Unio pictorum mancus*. Both lakes have already been monitored for several decades (e.g., Manca *et al*. 1992; Giussani & Galanti 1995).

**Tab. 1**. Morphometric parameters of Lake Maggiore (Ambrosetti *et al.* 1992) and Lake Candia (Giussani & Galanti 1992).

	Lake Maggiore	Lake Candia
Mean lake level (m a.s.l.)	194	226
Watershed area $(km^2)$	6599	9.9
Lake area $(km^2)$	212.5	1.5
Mean depth $(m)$	177	3.8
Maximum depth (m)	370	77

A sampling station ("Sabbie d'oro") was set up in the south-eastern part of Lake Maggiore near the village of Brebbia (Province of Varese). The sampling site is a gently sloping bay about 500 m long, with silty-muddy sediments. The submersed macrophyte cover is very limited. On the beach there are some stands of *Phragmites australis* and a few trees (*Salix* sp.). Water transparency is commonly high, except during short periods of phytoplankton blooms. At low water levels, an extensive littoral area dries up, causing periodic mass mortality of mussels. The anthropogenic impact is restricted to a few bathers. Besides a very abundant population of *Unio pictorum mancus*, the bay presents rare specimens



**Fig. 2**. Water temperature (left panel) and level (right panel) in Lake Maggiore (black circles, solid line) and Lake Candia (empty circles, dashed line).

of *Anodonta cygnea*, *Viviparus ater* and some Pulmonates (*Lymnaea ovata* and *Physa acuta*). *Dreissena polymorpha*, an epibiontic on living *Unio*, is fairly frequent.

The sampling station on Lake Candia (Province of Turin) is located in the south-eastern littoral zone of the lake. The bottom of the station has a gentle slope, with sediments composed of a mixture of mud, stones, silt and sand. Most of the lake shore is covered by caricetum and phragmitetum. In addition to *Unio pictorum mancus*, rare specimens of *Anodonta cygnea* and some Pulmonates (*Lymnaea ovata* and *Physa acuta*) are present. The lake is naturally eutrophic, with frequent phytoplankton blooms and episodic hypoxia in the deepest water layers during the summer stratification. The most important water sources are underground springs, precipitation and runoff; in spring-summer the lake receives the surplus water from the Caluso Canal. It receives diffuse nutrient loads from the small drainage area devoted to agriculture.

Seasonal variations in water temperature and level in both lakes are given in figure 2.

#### *2.2. Field and laboratory methods*

We chose the sub-species *Unio pictorum mancus* for our study because of its stronger resistance to eutrophic conditions than other unionid species (e.g., *Unio crassus*). It was identified by its shell morphology and electrophoresis of the soft tissues according to Nagel & Badino (2001). Thirteen samplings were performed in Lake Maggiore from May 2003 to September 2004, with an interruption (November-January) due to the high water level of the lake. The same number of samplings was performed in Lake Candia between June 2003 and August 2004, except for December 2003 and February 2004 when the lake surface was covered with ice.

Qualitative samples of mussels were collected monthly, at random, visually and tactilely, by hand and hand-net from the littoral zone of both lakes at a maximum depth of 1 m (Strayer & Smith 2003). To evaluate the seasonal variations in the population size structure, 4048 individuals were taken from Lake Maggiore and 3338 individuals from Lake Candia; the number of mussels in the samples taken from the former lake varied between 221 and 389, from the latter lake between 93 to 368. Each individual was returned to its environment after measurement of shell length, using a calliper, to prevent impoverishment of the population.

To allow us to evaluate whether an easily measurable variable such as shell length was correlated with other mussel characteristics (shell height and width, shell and soft tissue fresh and dry weight), samples were collected seasonally from both lakes (139 individuals from Lake Maggiore and 80 from Lake Candia). The external shell surface of each individual was cleaned by scrubbing off sediment particles and periphyton coating the periostracom with a nylon nailbrush, and by removing the zebra mussels attached to the Lake Maggiore specimens. The cleaned mussels were preserved in plastic bags placed in an ice-box and brought to the laboratory, where they were kept at  $-20$  °C until the variables listed above were measured. Additionally, the dry weight was measured after shell and soft tissue drying for 24 hours at 105 °C, and the content of ashes was determined after ignition for 6 hours at 450 °C. The annual rings on the external shell surface were counted to determine age (Strayer & Smith 2003).

The population density in both lakes was estimated by seasonally sampling all the mussels present on the sediment (30 replicates) using a frame of 0.5 m  $\times$  1.0 m. Another 35 samples were taken with the help of a Petersen grab  $(500 \text{ cm}^2)$ .

Due to the relatively long life-span of the mussels, the order of magnitude of the population density might be expected to be fairly constant over relatively short time periods such as that of this study (16 months). Since this assumption was matched by the similarity of the values obtained by sampling each lake in different seasons, the mean population density calculated for each



**Fig. 3**. Relationships between *Unio pictorum mancus* shell length and height (**A**), length and width (**B**), and height and width (**C**), as well as between shell length and age (**D**). The full circles and solid lines refer to Lake Maggiore mussels ( $n = 139$ ), the open circles and dashed lines to Lake Candia mussels  $(n = 80)$ .

lake was regarded as constant for the entire study period and used in the following calculations. To follow the seasonal variations in the population size structure, the density of individuals in each size class (2 mm intervals) was calculated for each monthly sampling according to formula (1):

$$
x_i = a_i \frac{N}{n_i} \tag{1}
$$

where:  $x_i$  = number of individuals belonging to the  $L_i$ size class collected at time  $t_i$  per square meter;  $a_i$ number of individuals belonging to the  $L<sub>i</sub>$  size class collected at time  $t_i$ ;  $n_i$  = total number of individuals collected at time  $t_i$ ;  $N =$  mean population density (number of individuals  $m<sup>2</sup>$  in each lake).

The seasonal variations of soft tissues and shell biomass (dry weight) of each size class were estimated in the same manner.

A curvilinear regression was fitted by least squares to experimental data on the dry weight of soft tissues and shell and on the number of annual rings in relation to shell length. The best fitting of the data was performed by comparing the AIC value of the exponential model ( $y = a \exp^{bx}$ ) with that of the logarithmic model

 $(y = ax^b)$  where *y* is respectively soft tissue dry weight, shell dry weight or n° of rings, *x* is shell length (mm), *a* is the regression intercept and *b* the regression coefficient. For each lake two equations were obtained from the experimental data on the dry weight of soft tissues and shell, in relation to shell length (Fig. 3). These equations were applied to estimate the dry weight of shell and soft tissues of the mean individuals of each length class. These values, multiplied by the population density of the corresponding size class, were used to calculate the soft tissue and shell biomass for each class.

Mean shell and soft tissue dry weight of mussels from the two populations were compared by considering only the mussels belonging to the size classes common to both lakes (i.e., the classes from 24-25 mm to 80-81 mm).

#### *2.3. Statistics*

The Kolmogorov-Smirnov two-sample test was used to test the differences between the population size distributions in Lake Maggiore and Lake Candia. To identify the differences due to habitat characteristics, and to minimize the seasonal effect, the test was applied to the pooled data of each lake.

Linear regression analysis was used to evaluate for the two populations: the relationship between shell biometrical variables (height, width and length); the relationship between logaritmic transformed data of length and weight of soft tissues and shell; the relationship between logaritmic transformed data of length and number of rings. Analysis of variance was performed on each regression to test whether the slope was significantly different from zero. Regressions between shell biometrical variables, as well as length-weight and length-age (number of annual rings) regressions, for the two populations were compared using t-tests for differences among slopes and for differences among intercepts. Two regressions were taken as different if either the slope or the intercept (or both) were different with a significance level of 0.05 (two-tailed test).

To evidence a seasonal or a dimensional effect on shell and soft tissue composition, two-way ANOVA was applied to the percentage of water, organic substance and ashes, with season and length as factors. Lake Candia data were not analyzed for seasonal and dimensional effects due to the loss of one sample.

#### 3. RESULTS

#### *3.1. Dry weight, organic matter and ashes*

The mean shell and tissue weight of mussels in the 24-81 length classes from Lake Maggiore were respectively 9.00 g d.w. and 0.86 g d.w., those from Lake Candia 7.60 g d.w. and 1.20 g d.w. The ratio between shell and soft tissue dry weight was 10.46 for Lake Maggiore and 6.33 for Lake Candia. The mean wet weight/dry weight ratio was 9.66 for the soft tissues of Lake Maggiore mussels, and 8.65 for those from Lake Candia; the same ratio related to shell weight is 1.01 in both populations.

The mean percentages of water, organic matter and ashes of shell and tissues of mussels from both lakes are reported in table 2. While season had no influence on these variables, the percentage of organic matter decreased and the percentage of ashes increased with mussel size (2-way ANOVA,  $p \leq 0.0001$ ). The most marked difference between the two mussel populations seems to be the higher percentage of ashes and the lower percentage of organic matter in both shell  $(t_8)$ = 10.67 and 10.69,  $p \le 0.0001$ ) and soft tissues ( $t_{146}$  = 20.06 and 11.31, *p* <0.0001) of Lake Maggiore mussels.

## *3.2. Biometrical variables*

A highly significant linear regression of height and width over length and of height over width was found (*p* <0.0001), (Fig. 3). The mean percentages of height and width in relation to mean length of the shells from both lakes are given in table 3. The percentages of height and width over length were significantly higher (one-way ANOVA,  $p \leq 0.0001$  in the mussels from Lake Maggiore than in those from Lake Candia, reflecting the

fact that, although both populations had a very similar shape of shell, the mussels from Lake Candia were slightly more tapering.

**Tab. 2**. Mean percentages of water, organic matter and ashes in soft tissues and shell of mussels from Lake Maggiore and Lake Candia. The percentage of water was calculated on the wet weight (w.w.), that of organic matter and ashes on the dry weight (d.w.).



The relationship between shell length and mussel age is given in figure 3D. A highly significant regression (generalized non linear regression fit,  $p \leq 0.0001$ ) was found for the mussels of both populations, but a steeper curve fits the data of the Lake Candia mussels compared to the Lake Maggiore mussels. Indeed, although the slopes were similar, the intercepts were significantly different (slope n.s., intercept  $p \le 0.001$ ).

**Tab. 3**. Percentage of height and width related to shell length considered as 100%.

	Lake Maggiore	Lake Candia
height	49.34	45.26
width	33.30	32.10

The best representation of the relationship between length of the shell and shell and tissue dry weight was provided by the logarithmic curve which indicates allometric growth. However, the relationship between length of the shell and its dry weight was significantly different from that between shell length and soft tissue weight (slope n.s, intercept  $p \le 0.001$ ), as emerged for the mussels of both lakes (Fig. 4). Since the growth-rate of the shell surface decreases as the mussel ages, while the amount of calcium carbonate secreted by the mantle seems to remain quite constant for the mussel's lifetime, shell thickness increases with shell size.

A significant difference was found between the populations in the length-weight relationship of both shell and soft tissues, but while for the tissues the regression slope was similar and only the intercept differed  $(p \le 0.001)$ , for the shell both coefficients were significantly different (slope *p* <0.02, intercept *p*  $< 0.001$ ).

## *3.3. Population density, biomass and their seasonal variations*

The mean population density was  $46.86$  ind. m<sup>-2</sup> in Lake Maggiore and  $6.93$  ind. m<sup>2</sup> in Lake Candia. While the mean population density was fairly constant over time, the abundance of individuals in each size class, and consequently also the relative biomass of shell and



**Fig. 4.** Relationships between shell length and shell dry weight (full circles), and between shell length and soft tissue dry weight (open circles) in from Lake Maggiore ( $n = 139$ ) and Lake Candia ( $n = 80$ ).

soft tissues, varied with the season (Figs 5a and 5b). In the same months, the frequency distribution of the length classes was similar in 2003 and 2004. In Lake Maggiore, the distribution of mussels in the different length classes was generally unimodal and symmetrical, except in July 2003 when abundance of young individuals caused three peaks of frequency: one of adults and two of young mussels, which accounted for about 40% of the total population (Fig. 5a). During the following months the number of these small specimens progressively decreased. Indeed, except for the June-July period, the number of small individuals was never high. In Lake Candia the frequency distribution of mussel length classes was generally multimodal (Fig. 5b) .

On an annual basis, the frequency of both population density and shell and soft tissue biomass shows a normal distribution in both lakes, but the shape is considerably more flattened in Lake Candia than in Lake Maggiore (Fig. 6).

The mussels of Lake Candia are larger than those of Lake Maggiore: the shell length of the former ranged from 25 to 101 mm, the mean and median values being respectively  $70.4 \pm 10.6$  mm and 72 mm, that of the latter from 14 to 81 mm, the mean and median values being respectively  $55.5 \pm 11.3$  mm and 58 mm. The middle sized individuals were numerically dominant in both populations, but there was a significant difference between the size distribution of the two populations (Kolmogoroff-Smirnov test,  $p = 0.001$ ).

In spite of the larger size of Lake Candia mussels, the far higher (about 7 times) population density in Lake Maggiore resulted in a biomass of both shell and soft tissues respectively 4-fold and 3-fold higher than in Lake Candia.

The mean soft tissue and shell biomass of Lake Maggiore mussels were  $38.67$  g d.w. m<sup>-2</sup> and  $408$  g d.w. m<sup>-2</sup>, respectively. The corresponding values of the Lake Candia population were  $14.04$  g d.w. m<sup>-2</sup> and  $92.29$  g d.w. m-2, respectively. The seasonal variations of the soft tissue and shell biomasses in Lake Maggiore and Lake Candia are presented in figure 7.

The coefficient of variation (% CV) calculated for the seasonal variations of shell as well as of tissue biomass of *U. pictorum mancus* from Lake Maggiore (15%) was about twofold higher than that calculated for the Lake Candia population (7%). The pattern of variations over time of the shell biomass was similar to that of the soft tissues in both lakes. In Lake Candia the lowest biomass values were calculated in June-July of both 2003 and 2004. The largest biomass values were found during the cold season, when the large-sized individuals were the dominant component of the population. A similar trend was observed in Lake Maggiore: the lowest biomass values coincided with the occurrence of an abundant number of young mussels in 2003, while a corresponding decrease was not so evident in summer 2004.

# 4. DISCUSSION

Besides confirming the positive relationship between the size of a mussel and the trophic level of its environment (Ravera *et al*. 2003), the results of this study highlighted other important differences between mussel populations which are probably related to the lake trophic level. The population from the meso-oligotrophic Lake Maggiore has higher density, biomass, annual recruitment and shell weight to soft tissue weight ratio than those recorded in the population from the eutrophic Lake Candia.

Eight years is the maximum age attained by mussels from both lakes, but the mussels from Lake Candia have a faster growth and higher mean individual size than those from Lake Maggiore. In both populations the frequency distribution of mussel size is Gaussian, but the median falls at 58 mm for Lake Maggiore and 72 mm for Lake Candia. During the investigation period, the population density was fairly constant, and seasonal variations in size and biomass structure were noticeable only during the annual recruitment period, and especially in the Lake Maggiore population.



Fig. 5a. Monthly variations of population density and biomass in each length class of the population from Lake Maggiore. Shell lengths (mm) are reported on the abscisses, density (ind m<sup>-2</sup>), shell and soft tissue biomass



Fig. 5b. Monthly variations of population density and biomass in each length class of the population from Lake Candia. Shell lengths (mm) are reported on the abscisses, density (ind m<sup>-2</sup>), shell and soft tissue biomass (



**Fig. 6**. Population density, shell and soft tissue biomass in each length class of the *Unio pictorum mancus* populations (pooled over time) from Lake Maggiore and Lake Candia, respectively.



**Fig. 7**. Monthly variations of shell and soft tissue biomasses in *Unio pictorum mancus* from Lake Maggiore (left panel) and Lake Candia (right panel).

From an ecological standpoint, i.e. the role played by the mussels in their environment, population density and biomass are the most important variables. Population density in Lake Maggiore  $(46.86 \text{ ind. m}^{-2})$ , which was far greater than in Lake Candia  $(6.93 \text{ ind. m}^{-2})$ , is very high compared to values reported in literature, which range from 0.74 to 13.04 (Økland 1963; Tudorancea & Florescu 1968; Negus 1966; Ravera & Sprocati 1997; Bauer & Wächtler 2001). One of the hypotheses which could explain the lower population density estimated for Lake Candia than for Lake Maggiore could be a less effective relationship with the host fish of the glochidia. This hypothesis seems to be supported by the lower number of young individuals admitted each year to the Lake Candia population compared to that of Lake Maggiore, but other causes (such as reduced fecundity and/or reduced survival of adult or larval stages) could be involved. The high recruitment of young mussels in Lake Maggiore seems to compensate, at least partly, for the high mortality due to the periodic lowering of the lake water level. This stress does not alter the population structure, but abolishes the fraction of population settled in the area most at risk of drying out.

The far greater density of mussels in Lake Maggiore (about 7-fold that of Lake Candia) meant that their tissue and shell biomass also widely exceeded that of Lake Candia, with values respectively about 3-fold and 4-fold higher. This suggests that mussels not only exert a greater influence on their habitat in Lake Maggiore, but also that the habitat conditions they find in the lake are more congenial. This hypothesis also seems to be supported by the differences in size structure between the two populations. In fact, while in Lake Maggiore the distribution of mussels in the different length classes was generally unimodal and symmetrical, in Lake Candia the frequency distribution of mussel length classes was generally multimodal, suggesting that there may be marked variations in recruitment and/or mortality rates from year to year. One of the probable causes of this is

the wide variability of physico-chemical conditions which typically characterizes eutrophic water bodies.

However, if on the one hand a larger population density can be sustained in the meso-oligotrophic lake, on the other hand better nutritional conditions seem to favour individual growth in the eutrophic lake. Indeed, large-sized individuals are more abundant in the Lake Candia population than in the Lake Maggiore population, and both growth-rate and soft tissue weight per length unit are higher in Lake Candia than in Lake Maggiore mussels, as testified by the observed differences in both length-soft tissue weight and length-age relationship. In fact, the steeper slope of the lengthweight data from Lake Candia indicated a higher weight gain per length unit, while the steeper slope of the length-rings data from Lake Maggiore evidenced a slower length increment per time unit. The presence of epibiontic zebra mussel (*Dreissena polymorpha*, Pallas) on the posterior part of the shell, near the inhalant and exhalant siphons, of almost all middle and large sized *Unio* could be a further reason for a reduced food supply. Settlement in this position enables *Dreissena* to use the food particles provided by the filtering flux of the host (Hörmann & Mayer 2006).

The Lake Maggiore mussels at the same shell length are not only older, but have a heavier shell than Lake Candia individuals. This difference probably reflects adaptation to the habitat, rather than the higher calcium concentration in Lake Maggiore waters (26.39 mg  $L^{-1}$ ) than in Lake Candia (15.15 mg L-1) (Ravera *et al*. 2007, this issue). In fact, the calcium concentration in Lake Candia, although lower than in Lake Maggiore, was not so low as to become a limiting factor for mollusc growth. It is conceivable that the heavier shell of Lake Maggiore mussels is a reaction to continuous water movements and the more compact sandy sediments of the lake.

Bigger mussels are usually older than smaller ones, but there is wide variability in this relationship, so that mussels of the same length may have different ages

(e.g., Botnariuc & Tudorancea 1967). Determining age by counting the annual rings is not easy, because of supernumerary rings which arise due to stress from various causes other than the cold season, the corrosion of the umbon area, and the more recent rings being concentrated in a small space on the edge of the shell (Økland 1963; Magnin & Stanczykowska 1971). In spite of these difficulties, and the existence of more reliable methods (e.g., Wellmann 1938; Nelson *et al*. 1966; Garcia & Plante 1992; Garcia 1999; Mutvei & Westermark 2001), ring counting is the most commonly used method because it does not require instruments or sample preparation, and is therefore useful when a high number of mussels has to be analysed. This method yielded a highly significant correlation between number of rings and shell length, although the variability was rather wide.

Instead of using mussel age, which is an essential variable for studying population demography, we preferred to relate the mussel variables (e.g., mussel distribution in size classes, percent composition of shell and soft tissues) to the dry weight of soft tissues and shell and to shell length. These variables are strictly related to biomass and mussel metabolism and, consequently, to their influence on the ecosystem. In addition, according to Bauer (1991), shell length gives an indication of soft tissue mass, which is clearly related to reproductive traits.

The percentage of small (young) individuals is typically low in Unionidae samples (e.g., Negus 1966; Nardi 1972; Bauer 1983; Bauer *et al*. 1991; Hochwald 1997). As an example, in the monthly samples from Lake Candia the proportion of individuals smaller than 31 mm was always lower than 2%. In the samples from Lake Maggiore the proportion of individuals of this size was generally lower than 10% except for the sample of July 2003, when this proportion increased to 40%. The subsequent decrease in the proportion of small specimens was probably due both to their passage into the successive size classes and to mortality during the subsequent months until the next admission. The smallest individuals collected were 14 mm long in Lake Maggiore and 24 mm long in Lake Candia. To evaluate whether the low number of small individuals collected could be due partly to the sampling method (visual and tactile collection), sediment samples were collected with a Petersen grab in the Lake Maggiore station and the sediment collected sieved. The percentage of small individuals was very similar to that obtained by hand collection. It was suggested that in some *Unio* sp. populations the low percentage of young individuals is due to their higher sensitivity to environmental stress compared to adult mussels (e.g., Bauer 1983). In environments not stressed by toxic pollutants, such as our sampling stations, the low percentage of small mussels could be explained by the following hypothesis. From the length-age relationship observed in both populations, the permanence time of mussels in each size class increases exponentially with mussel size. It follows that the flux of individuals from the smallest to the successive size class should be relatively rapid, and the number of individuals in the smallest class comparatively small. This hypothesis seems to be supported by the rapid decrease in the number of small individuals following each annual admission, which is probably only partially accounted for by mortality.

The differences between the population of the mesooligotrophic Lake Maggiore and that of the eutrophic Lake Candia clearly demonstrate the plasticity of *Unio pictorum* in reacting to environmental characteristics. While these differences may be the result of a phenotypical adaptation to the environment, genetic differences between the two populations acquired in the course of evolution cannot be excluded.

Although it is quite clear that eutrophication affects mussels (e.g., Patzner & Müller 2001), it is very difficult to establish a relationship between well-defined variables of the eutrophic environment and the characteristics of a euryoecious species such as *Unio pictorum*. The nutrient enrichment of a water body influences all the physical, chemical and biological characteristics of the ecosystem, some of which may be favourable to the mussel population (e.g., abundance of food, increase in the population density of fish which host the glochidia), while some others may be very dangerous (e.g., oxygen depletion, toxic concentrations of nitrites and ammonia). Our data suggest that while on the one hand eutrophic conditions seem to enhance individual growth rate, on the other hand they appear to limit population increase through a reduction of recruitment and/or of survival rate of the younger individuals.

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