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Title: Long-term effect of temperature on bioaccumulation of dietary metals and metallothionein induction in Sparus aurata.

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Keywords: metal; metallothionein; bioaccumulation factor; temperature; Sparus aurata

Corresponding Author: Dr. Amparo Torreblanca, PhD

Corresponding Author's Institution: University of Valencia

First Author: Diana Guinot

Order of Authors: Diana Guinot; Rocio Ureña, Ph.D.; Agustin Pastor, Ph.D.; Inmaculada Varó, Ph.D.; Jose Juan del Ramo, Ph.D.; Amparo Torreblanca, PhD

## Reviewer #1

1- Title, the expression "elevated temperatures" is not a clear expression, perhaps authors could search for another expression.

## The expression "elevated temperatures" has been replaced by "temperature"

Page 7, Lines 145-..., results on metal contents must be adjusted attending at significant figures: 0.378 ± 0.05 must be expressed as 0.38 ± 0.05 11.63 ± 0.74 must be expressed as 11.6 ± 0.7 172.64 ± 8.89 must be expressed as 173 ± 9 40.14 ± 2.78 must be expressed as 40 ± 3 0.052 ± 0.01 must be expressed as 0.05 ± 0.01 0.15 ± 0.09 it is OK 138.94 ± 8.76 must be expressed as 139 ± 9

## Values have been expressed as suggested by reviewer.

3- Table 3. Secondary horizontal lines could be eliminated.

## Secondary horizontal lines have been eliminated.

## Reviewer #2

4- Although well organized, the manuscript should be reviewed from a linguistic point of view, being misspelled many words and not clear some sentences.

Manuscript has been reviewed from a linguistic point of view. Changes made are marked in the revised manuscript.

5- The procedure used for the determination of the metallothionein content in the liver must be described.

## Procedure for metallothionein determination has been described.

6- Possibly, metallothionein content and induction should be investigated also in the muscle. Although we agree with reviewer #2 in the interest of determining metallothionein induction also in muscle, liver is recognized to have a greater importance in metal handling than muscle. Furthermore, there are not muscle samples left from this experiment.

#### 1 Long-term effect of elevated temperature on bioaccumulation of dietary

#### 2 metals and metallothionein induction in Sparus aurata.

- 3 Diana Guinot<sup>1</sup>, Rocío Ureña<sup>1</sup>, Agustín Pastor<sup>2</sup>, Inmaculada Varó<sup>3</sup>, Jose del
- 4 Ramo<sup>1</sup> and Amparo Torreblanca<sup>1</sup>
- 5 <sup>1</sup>Department of Functional Biology, Faculty of Biological Sciences, University of
- 6 Valencia. Dr. Moliner, 50, 46100 Burjassot (Valencia). Spain. Phone +34 6 35443378
- 7 FAX +34 6 3543202 Amparo.Torreblanca@uv.es (Corresponding author)
- 8 <sup>2</sup>Departament of Analytic Chemical, University of Valencia. Dr. Moliner, 50, 46100
- 9 Burjassot (Valencia), Spain.
- <sup>3</sup>Institute of Aquaculture Torre de la Sal (CSIC), 12595 Ribera de Cabanes, Castellón.
- 11 Spain.

#### 12 Abstract

Previous studies have demonstrated that the commercial feed of aquacultured 13 fish contains trace amounts of toxic and essential metals which can accumulate 14 in tissues and finally be ingested by consumers. Recently rising temperatures, 15 associated to the global warming phenomenon, have been reported as a factor 16 17 to be taken into consideration in ecotoxicology, since temperature--dependent alterations on bioavailability, toxicokinetics and biotransformation rates can be 18 expected. Sparus aurata were kept at 22°C, 27°C and 30°C for three months in 19 order to determine the temperature effect on metallothionein induction and 20 21 metal bioaccumulation from a non-experimentally contaminated commercial 22 feed. A significant temperature-dependent accumulation of cadmium (Cd), copper (Cu), mercury (Hg), zinc (Zn), lead (Pb) and iron (Fe) was found in liver, 23 together with that of manganese (Mn), Fe and Zn in muscle. Hg presented the 24

- 25 highest bioaccumulation factor, and essential metal homeostasis was disturbed
- 26 in both tissues at warm temperatures. An enhancement of hepatic
- 27 metallothionein induction was found in fish exposed to the highest temperature.
- 28 **Keywords:** metal; metallothionein; bioaccumulation factor; temperature; *Sparus*
- 29 aurata

#### 30 1. Introduction

31 The aquatic ecosystems are undergoing a warming in their deep and surface waters, a fact which may have significant consequences on the organisms 32 33 inhabiting them (Noyes et al., 2009). Gilthead sea bream Sparus aurata is a 34 cosmopolitan species distributed inthroughout the Mediterranean and in the of-35 the NE Atlantic, and which is of great economic interest since it is one of the mainprincipal species of Mediterranean aquaculture. AsBeing an ectothermic 36 species, it is vulnerable to the effects of temperature variations on its 37 metabolism and physiology. These variations may produce changes in the 38 39 toxicokinetics, bioavailability, biotransformation, homeostasis, absorption rate 40 and elimination of different compounds (Douben, 1989; Köck et al., 1996; Yang and Chen, 1996). Other key physiological mechanisms, such as respiration, 41 feeding rate, growth and reproduction may also be affected (Bowen et al., 2006; 42 Heugens et al., 2001). The thermic induced variations in the toxicokinetics of 43 44 pollutants, together with the increase of exposure to the same as a 45 consequence of climate warming-climate (Carrie et al., 2010) may present a risk for the development and survival of species of commercial interest, affecting the 46 quality of the end product (Noyes et al., 2009). 47 In cultured fish, it has been demonstrated that the feed is the principal source of 48 contamination by metals is the feed. Cadmium (Cd), lead (Pb) and mercury 49 50 (Hg), among others, are potentially toxic and tend to accumulate in the tissues, which in the end are ingested by the consumers (Fernandes et al., 2009; Maule 51 et al., 2007; Cretì et al., 2010). Their accumulation in the organisms depends on 52 the concentration, route of absorption, environmental conditions and other 53

54	intrinsic factors (Bowen et al., 2006; Chowdhury and Blust, 2001; Jezierska and
55	Witeska, 2006; Karakoç and Dinçer, 2003; Lemus and Chung, 1999).

56	Due to the growth in the <u>activity of</u> aquaculture <del>activity</del> in recent decades, it has
57	become of special relevance to learn the influence of the increase in
58	temperature on the possible routes of absorption, accumulation and elimination
59	of metals in these organisms in the context of global warming. The-
60	temperatureTemperature may be a determining factor in the capture,
61	transporting and metabolism of the metals incorporated through the feed, both
62	of the essential metals which may become toxic at high concentrations in the
63	tissues, and of the non-essential metals.
64	Metallothionein (MT) is a low-molecular-weight metal binding protein and is
65	known to play an important role in protection against heavy metal toxicity. In
66	addition to the detoxification of toxic metals such as Cd and Hg, MT is involved
67	in the maintenance of homeostasis of essential trace elements such as zinc
68	(Zn) and copper (Cu) (Coyle et al., 2002; Hamilton and Mehrle, 1986). Its role in
69	the protection against xenobiotics or in the cellular protection against oxidative
70	stress should be underlined (Coyle et al., 2002; Van Cleef-Toedt et al., 2001).
71	Although its synthesis is related to the metal exposure, its levels can be affected
72	by endogenous and exogenous factors such as the reproductive cycle or the
73	temperature (Van Cleef-Toedt et al., 2001). Variations in the water temperature
74	could modify, directly or indirectly, modify the behavior of this protein as regards
75	the bioaccumulation of metals, as well as its participation in toxicokinetic
76	processes (Baykan et al., 2007; Gorbi et al., 2005; Rotchell et al., 2001).
77	Given the lack of information as regards the effect of temperature on the
78	bioaccumulation of metals via dietary sources, and on the synthesis of MT, an

- understanding of these processes in the light of the problem of global warming
  and its repercussions on species of commercial interest such as *Sparus aurata*
- 81 is required. As such, the aims of this work are:
- 1) To determine the effect of temperature on the bioaccumulation of essential
- 83 (Cu, Fe, Mn and Zn) and non-essential (Cd, Hg and Pb) metals experimentally
- via a non-contaminated commercial feed.
- 2) To discover whether the temperature has any influence on the homeostasis
  of the essential metals.
- 3) To elucidate the <u>role of</u> metallothionein-role in the above mentioned
- 88 temperature induced changes.

#### 89 2. Materials and Methods

#### 90 2.1. Animal collection and maintenance

Adult Sparus aurata were distributed and acclimated in 500 L tanks containing 91 92 seawater (37<del>‰),</del>‰) at a constant temperature of 22°C, with continuous aeration and natural photoperiod, in a closed circuit for two months before starting prior to 93 94 the experiment. Subsequently, the temperature of two of the experimental 95 groups was gradually increased to reach 27°C and 30°C respectively. Control groups of fish were maintained at 22 °C throughout the experiments. The 96 animals were fed with commercial pellets (1.5% of body mass per day) and the 97 survival percentage was 99%. Fish were kept under constant conditions for 98 99 three months. After this period, 6-7 fish were removed at random from each experimental group and heldplaced in water containing 30 mg L<sup>-1</sup> of anesthetic 100 101 clove oil. Lengths and weights of gilthead sea bream were recorded. Fish were 102 sacrificed; livers and a piece of dorsal muscle tissue were dissected and 103 immediately frozen in liquid nitrogen and stored at -80 °C.

## 104 2.2. MT determination by differential pulse polarography

105	Approximately 0.2 g wet weight portions of frozen liver were homogenized using	
106	ultra-turrax in 20 mM Tris-HCI buffer, 1 mM DTT and 0.2 mM PMSF pH 8.6 in	
107	an ice bath. The homogenates were centrifuged at 30 000 g for 45 min at 4°C.	
108	The supernatant was heated at 80°C for 10 min in order to denature high	
109	molecular weight proteins and subsequently centrifuged at 30 000 g for 45 min	
110	at 4°C, The heat-treated supernatant, containing thermally stable MT, was	
111	separated from precipitated proteins. MT was measured using differential pulse	
112	polarography <del>as described by (</del> Ureña et al- <del>(,,</del> 2007).	
113	An aliquot of the heat-treated supernatant was added to the polarographic cell,	Formatted: Space Before: 0 pt
114	containing 20 mL hexammincobalt chloride buffer (the supporting electrolyte),	
115	together with Triton-X (0.025% v/v). The cell was purged for 2 min with purified	
116	$N_2$ prior to analysis. The polarographic response was measured during a	
117	potential scan between -1.38 V and -1.7 V (Model 757 VA Computrace	
118	Analyser, Methrom, Switzerland) in SMDE mode. Quantification of MT was	
119	performed by using the standard addition method with rabbit liver MT I+II	
120	(Sigma). Results are expressed as $\mu g g^{-1}$ wet weight (w.w.) of tissue.	
121	2.3. Metal analysis	
122	Samples of 0.1-0.5 g wet weight of liver and muscle were digested in	
123	concentrated nitric acid 65% (Baker <del>),</del> at room temperature overnight, and were	
124	heated at 80°C for 2 h. In order to determine the trace amounts of metal	
125	contained in the feed, samples of commercial pellets were also digested (n=3).	
126	Within each digestion series, appropriate blanks with ultra-pure water were also	
127	subjected to the same procedure in order to account for <u>take</u> background	

128 contamination levels into acount. After cooling, solutions were transferred to a standard volume with ultra-pure water. Determination of metals (Cd, Cu, Zn, Fe, 129 130 Pb, Hg and Mn) was undertaken using an ICP-Mass (Elan DRC-I, Perkin-Elmer 131 Sciex). Samples of similar weight of certified reference material (DOLT-3 and 132 LUTS-1, National Research Council of Canada, Ottawa), were digested and analyzed during each analytical run. The values of all elements found were in 133 good agreement with the certified values, with the recoveries ranging from 91% 134 to 104%. 135

136 2.4. Data analysis and calculations

Metal content ratio between muscle and liver was calculated in order to detect
changes in the distribution of metals as a consequence of the thermal
experimental conditions, and to showreveal the proportion of accumulated metal
in each tissue.

Bioaccumulation factor (BAF) was calculated for each metal-tissue-temperature
combination in order to detect the effects of temperature on the global
toxicokinetics among metals. <u>ItThis</u> was calculated as a quotient between the
metal concentration in each tissue and the metal content in the commercial
pellets as described by Dabrowska et al. (1996).

146 2.5. Statistical analysis

Statistical analyses were carried out using the software Stata 10 (Stata Corp).
Transformations of the data were performed when the assumption of normality
of residuals werewas not met. One-way ANOVA was used in the analysis of
data to check the influence of temperature inon each variable. The Scheffe test

151	was used as <u>a post hoc test to provedemonstrate</u> the differences among the
152	three temperature groups. Two-way ANOVA, with temperature and metal as
153	fixed factors, was used in the statistical analysis of BAFs, followed by a post
154	hoc Bonferroni test. The Kruskal Wallis test was used when transformed data
155	did not present homoscedasticity. Pearson correlation coefficients (r) were
156	calculated between MT levels and metal content in liver, in order to measure the
157	strength of association between these variables. Results are presented as
158	means $\pm$ SEM and a p value lower than 0.05 was considered as statistically
159	significant.
160	3. Results
161	3.1. Metal content in food and biometry
162	Metal contents of food pellets were $0.37838 \pm 0.05\mu g g^{-1}$ w.w. for Cd, $11.636 \pm 1000$
162	
163	0. <mark>74<u>7</u> µg g<sup>-1</sup> w.w. for Cu, <del>172.64 ± 8.89<u>173 ± 9</u> µg g<sup>-1</sup> w.w. for Fe, 40<mark>.14 ±</mark>-</del></mark>
163	0.74 <u>7</u> $\mu$ g g <sup>-1</sup> w.w. for Cu, 172.64 ± 8.89173 ± 9 $\mu$ g g <sup>-1</sup> w.w. for Fe, 40.14 ± $\frac{2.78 \pm 3}{100}$ $\mu$ g g <sup>-1</sup> w.w. for Mn, 0.052 05 ± 0.01 $\mu$ g g <sup>-1</sup> w.w. for Hg, 0.15 ± 0.09 $\mu$ g g <sup>-1</sup>
163 164 165	$0.74\underline{7} \ \mu g \ g^{-1} \ w.w.$ for Cu, $172.64 \pm 8.89\underline{173} \pm 9 \ \mu g \ g^{-1} \ w.w.$ for Fe, $40.14 \pm 2.78\underline{+3} \ \mu g \ g^{-1} \ w.w.$ for Mn, $0.052\underline{-05} \pm 0.01 \ \mu g \ g^{-1} \ w.w.$ for Hg, $0.15 \pm 0.09 \ \mu g \ g^{-1} \ w.w.$ for Pb and $138.94 \pm 8.76\underline{139} \pm 9 \ \mu g \ g^{-1} \ w.w.$ for Zn.
163 164 165 166	$0.74\underline{7} \ \mu g \ g^{-1} \ w.w.$ for Cu, $172.64 \pm 8.89173 \pm 9 \ \mu g \ g^{-1} \ w.w.$ for Fe, $40.14 \pm 2.78 \pm 3 \ \mu g \ g^{-1} \ w.w.$ for Mn, $0.052 \pm 0.01 \ \mu g \ g^{-1} \ w.w.$ for Hg, $0.15 \pm 0.09 \ \mu g \ g^{-1} \ w.w.$ for Pb and $138.94 \pm 8.76139 \pm 9 \ \mu g \ g^{-1} \ w.w.$ for Zn. The weight and length as well astogether with the condition index calculated for
163 164 165 166 167	$0.74\underline{7} \mu g g^{-1}$ w.w. for Cu, $172.64 \pm 8.89173 \pm 9 \mu g g^{-1}$ w.w. for Fe, $40.14 \pm 2.78 \pm 3 \mu g g^{-1}$ w.w. for Mn, $0.052 \pm 0.01 \mu g g^{-1}$ w.w. for Hg, $0.15 \pm 0.09 \mu g g^{-1}$ w.w. for Pb and $138.94 \pm 8.76139 \pm 9 \mu g g^{-1}$ w.w. for Zn. The weight and length as well astogether with the condition index calculated for gilthead sea bream exposed to three experimental temperatures for three
163       164       165       166       167       168	$0.74\underline{7} \ \mu g \ g^{-1} \ w.w.$ for Cu, $172.64 \pm 8.89173 \pm 9 \ \mu g \ g^{-1} \ w.w.$ for Fe, $40.14 \pm 2.78 \pm 3 \ \mu g \ g^{-1} \ w.w.$ for Mn, $0.052 \pm 0.01 \ \mu g \ g^{-1} \ w.w.$ for Hg, $0.15 \pm 0.09 \ \mu g \ g^{-1} \ w.w.$ for Pb and $138.94 \pm 8.76139 \pm 9 \ \mu g \ g^{-1} \ w.w.$ for Zn. The weight and length as well astogether with the condition index calculated for gilthead sea bream exposed to three experimental temperatures for three months are shown in Table 1. As can be seen, temperature has no effect on the
163       164       165       166       167       168       169	$0.74\underline{7} \mu g g^{-1}$ w.w. for Cu, $172.64 \pm 8.89173 \pm 9 \mu g g^{-1}$ w.w. for Fe, $40.14 \pm 2.78 \pm 3 \mu g g^{-1}$ w.w. for Mn, $0.052 \pm 0.01 \mu g g^{-1}$ w.w. for Hg, $0.15 \pm 0.09 \mu g g^{-1}$ w.w. for Pb and $138.94 \pm 8.76139 \pm 9 \mu g g^{-1}$ w.w. for Zn. The weight and length as well astogether with the condition index calculated for gilthead sea bream exposed to three experimental temperatures for three months are shown in Table 1. As can be seen, temperature has no effect on the condition index andor weight, in. In contrast, fish kept at 30°C (p<0.05) waswere
163       164       165       166       167       168       169       170	$0.74\underline{7} \mu g g^{-1}$ w.w. for Cu, $172.64 \pm 8.89173 \pm 9 \mu g g^{-1}$ w.w. for Fe, $40.14 \pm 2.78 \pm 3 \mu g g^{-1}$ w.w. for Mn, $0.052 \pm 0.01 \mu g g^{-1}$ w.w. for Hg, $0.15 \pm 0.09 \mu g g^{-1}$ w.w. for Pb and $138.94 \pm 8.76139 \pm 9 \mu g g^{-1}$ w.w. for Zn. The weight and length as well astogether with the condition index calculated for gilthead sea bream exposed to three experimental temperatures for three months are shown in Table 1. As can be seen, temperature has no effect on the condition index andor weight, in. In contrast, fish kept at 30°C (p<0.05) waswere found to be significantly shorter than those kept at 22°C and 27°C.

3.2. Temperature effect on metals contentmetal contents and their distribution
betweenamong tissues

Metal content achieved<u>reached</u> in liver and muscle of gilthead sea bream is shown in Figure 1. The ANOVA test shows a significant effect of temperature on Cd, Hg and Pb content (p<0.001), and on concentration of the essential metals Cu (p<0.01), Zn and Fe (p<0.001) in liver, but not on Mn content. Moreover, the results show a significant effect of temperature on Mn, Fe and Zn content (p<0.01) in muscle.

The muscle/liver ratios for each metal and temperature were calculated when 180 181 possible and are shown in Table 2. They indicate the distribution of metals after temperature exposure and the proportion of each one in muscle in relation to 182 183 liver. The highest muscle/liver ratio for fish maintained at 22 and 27 °C was reached by Hg; however, whereas at 30 °C the highest ratio was reached by 184 185 Mn. As can be seen in Table 2, calculated ratios for Hg and Fe decreased significantly when temperature rose. Notably, and unlike for the other metals, 186 Mn ratio rose significantly with increasing temperature. 187

188 Insert Figure 1

189 Insert Table 2

- 3.3. Temperature effect on toxicokinetic related processes: Bioaccumulation
  Factor and MT induction
- 192 The BAF values obtained for each metal-tissue-temperature combination are
- 193 shown in Table 3. The two-way ANOVA test shows significant differences
- among <u>between</u> metals in muscle (p<0.001) at all the <u>temperaturetemperatures</u>
- 195 tested. Hg BAF in this tissue is considerably higher in comparison with that
- calculated for the other measured metals, reaching a value of 0.45 at 30°C. This

197	means that muscle presents Hg content of almost a half of the feed content. In
198	contrast, Mn BAF exhibitexhibits the lowest BAF for all temperature exposures.
199	Statistical analysis of BAFs also showed significant differences among metals in
200	liver (p<0.001). BAF for Hg in liver differed from the BAFs for other metals at
201	27°C and 30°C, and reached the highest BAF registered in this study (3.20 at
202	30°C). It should be noted that Hg BAFBAFs in liver at 27°C and 30°C are.
203	respectively, nearly 3 and 5 fold higher than those determined at 22°C-
204	respectively. A similar pattern to Hg was found for Cd, and in this case BAFs
205	are 3 and 4 fold higher at 27°C and $30^{\circ}C_{1}$ respectively, than those determined
206	at 22ºC. On the other hand, Mn BAFs in liver are the lowest among the studied
207	metals.
208	Insert Table 3
209	MT levels determined in liver for the three experimental temperature groups are

shown in Figure 1. As can be seen, there is a significant thermal effect on hepatic MT content. A<u>An</u> MT significant induction is found in fish exposed to the higher<u>highest</u> temperature (p<0.001). Furthermore, significant positive correlation<u>correlations</u> between MT and Cd<u>in liver</u> (r= 0.52; p<0.05), Hg (r= 0.78; p<0.001), Zn (r= 0.82; p<0.001), Cu (r= 0.82; p<0.001) and Fe (r= 0.78; p<0.001) in liver have been found, but not with Pb and Mn.

216 4. Discussion

#### 217 4.1. Metal content in food and biometry

- Length differed significantly between animals exposed to 22°C and 30°C (Table
- 1), which could be due to an increase in energetic demand related to an
- 220 increase in metabolic rate. A reduced growth and length at high temperature

221 was reported in *Oncorhynchus kisutch,* as a consequence of the energetic

deficit after an elevated metabolism (Bowen et al., 2006).

Metal content of food pellets was in the same range as most of the values reported in the literature for commercial pellets (Ciardullo et al., 2008; Dang and Wang, 2009; Mackee et al., 2008; Minganti et al., 2010). Cd, Pb and Hg feed content are also below the established maximum permitted level in fish feed according to the-EU Directive 2002/32/EC and 2005/87/EC amending Annex I to Directive 2002/32/EC.

229

4.2. Effect of temperature on metal content and its distribution betweenamong
tissues

232 Metal contents in tissues (Figure 1) were in the same range as most of the values reported in the literature for farmed gilthead sea bream (Cretì et al., 233 2010; Minghetti et al., 2008). It should be noted that Cd, Pb and Hg 234 235 concentrations measured in muscle of Sparus aurata in the present work are 236 below the established maximum permitted level in muscle of several species according to the EU Regulation (EC) No 1881/2006. We have found a general 237 238 trend of increasing metal content in gilthead sea bream tissues with increasing temperatures. The positive relationship between temperature and increased 239 240 metal uptake in different aquatic organisms, has been explained by accelerated 241 biochemical and physiological processes, changes in the membrane 242 permeability, and alterations of metal transport systems, and binding to several ligands (Baykan et al., 2007, Chowdhury and Blust, 2001; Jezierska and 243 244 Witeska, 2006; Lemus and Chung, 1999).

245	The significant increase in Hg content in liver at both <del>,</del> 27ºC and 30ºC
246	comparing compared to at 22 °C, indicates strong temperature dependence.
247	AlsoAdditionally, the muscle/liver ratio (Table 2) confirms the preferential affinity
248	of Hg to lipidic tissues such as the liver. Using an experimental trophic chain,
249	Boudou et al. (1979) demonstrated that raising the temperature of the
250	environment has a synergic effect on the quantities of Hg bioaccumulated by a
251	carnivorous fish.
252	Cd content in liver also increased with rising temperature. Liver, as is well
253	known, is an organ involved in storage and detoxification of Cd, and a
254	progressive accumulation of dietary Cd over time has been explained by the
255	continuous transference via the portal system from the digestive tract to the liver
256	(Handy, 1993). In previous temperature effect studies in fish, the metal was
257	supplied as waterborne Cd (Douben, 1989; Köck et al., 1996) and the
258	temperaturedependent increase of metal content was explained by a greater
259	increase in uptake rate than the increase found in the elimination rate (Douben,
260	1989 <del>). Although), although</del> mechanisms involved in Cd accumulation may differ
261	in <u>to a</u> great extent depending of <u>on</u> the Cd exposure source.
262	The highest Pb content in liver was reached by fish kept at 27°C. In some
263	studies Pb uptake rate washas been affirmed to be temperature-dependent in
264	fish (Köck et al., 1996).
265	An enhancement of Mn in muscle of fish maintained at 30°C in relation to those
266	kept at colder temperatures was determined, nevertheless no significant
267	variation in Mn was found in liver. However, when muscle/liver ratios were
268	calculated, a significant effect of the temperature was detected in such a way
269	that the ratio was higher at the warmer temperature, with the muscle acting as

270	aan Mn receptor tissue as temperature increased. However, little information is
271	available concerning the effect of temperature on Mn accumulation in fish
272	muscle. Adam et al. (1997) described two groups of organs with different
273	elimination kinetics of <sup>54</sup> Mn, the first was the penetration or transit group, and
274	the second the receptor or storage organs. According to this, liver was classified
275	as aan Mn distribution and elimination tissue, and muscle as a receptor-one.
276	The high levels of Cu in liver from <u>of</u> fish kept at 30°C in comparison to those
277	kept at lower temperatures are in agreement with findings reported in several
278	studies. For instance, Petenia kraussii juveniles exposed to 22°C had lower Cu
279	basal concentrations than those exposed to 30°C (Lemus and Chung, 1999).
280	Temperature-related increase in Cu content has been explained by other
281	authors as a consequence of a different extent in the rising of accumulation and
282	elimination rates, resulting in a final positive balance for uptake (Glover et al.,
283	2003; Lemus and Chung, 1999).
284	In concordance with our results on Zn accumulation, Karakoç and Dinçer (2003)
285	reported a temperature induced accumulation of Zn in liver, and Van
286	Campenhout et al. (2007) described a promotion of Zn assimilation by
287	temperature in fish. Also, Zn uptake in the apical cell membrane of rainbow trout
288	intestine was also found to be a temperature-dependent process (Glover et al.,
289	2003). In addition to the above findings, and taking into account the
290	enhancement of Zn in muscle and liver of fish maintained at 30°C, and that no
291	significant variation in the muscle/liver ratios was found in our study, we can
292	suggest that a temperatureinduced disturbance of Zn homeostasis in Sparus
293	aurata may be occurring.

294	Levels of Fe in muscle and liver remained quitefairly constant when temperature
295	increased from 22°C to 27°C, as can be expected for a tightly regulated metal
296	under physiological conditions. The sharp increase in Fe content in both
297	tissues, and the different ratio value found at 30°C may be due to Fe
298	homeostasis disturbance, indicating that some pathology is occurring. It should
299	be born in mind that Fe loading leads to free radical damage by the Fenton
300	reaction. Evidence in this direction has been found in fish and rats, suggesting a
301	role for Fe in hepatic injury following hyperthermia. (Bloomer et al., 2008;
302	Bowen et al., 2006; Skibba and Gwartney, 1997).
303	SummarizingIn summary, essential metals in muscle are strongly regulated
304	under physiological conditions, however,although at 30°C, their homeostasis is
305	disturbed as a result of a warmer temperature. The inverse relation
306	amongbetween muscle/liver ratio and temperature is a general trend for all the
307	metals studied, except Mn, which indicates that warmer temperatures facilitate
308	metal regulation processes, in <u>such</u> a way that there is a preferential
309	accumulation in liver thus protecting peripheral tissues.
310	4.3. Temperature effect on toxicokinetic related processes: Bioaccumulation
311	Factor and MT induction
312	BAF, the quotient between the metal concentration in tissue and the metal
313	content in the feed, allowed us to quantify the accumulation of metals from feed
314	in Sparus aurata, and to compare the temperature effect on bioaccumulation
315	among the determined metals.
316	Metabolic rates of ectothermic organisms may be strongly dependent on
317	temperature, and the same can be said for heavy metalmetals complexing with
318	cellular constituents, which may affect tissue distribution and BAF. We can

319	suggest that variations in toxicokinetics caused by temperature exist in some of
320	the studied metalmetals and tissues, although we cannot determine if whether
321	these changes are a consequence of the temperature effect on assimilation
322	efficiency, or on feeding rate.
323	Hg presents the highest BAF values among metals for both tissues, at all
324	temperatures tested (Table 3). Hg accumulation depends on its speciation,
325	mainly because methylated Hg (MeHg) is better absorbed from the
326	gastrointestinal tract than the inorganic form in such a way that in mammals
327	95% is absorbed (Liu et al., 2008 <del>). Furthermore,), and furthermore</del> it undergoes
328	extensive enterohepatic recycling (Liu et al., 2008). Taking ininto account all the
329	above, together with the fact that MeHg is the predominant form of Hg in fish
330	feed (>80%) (Amlund et al., 2007 <del>)),</del> it is easy to understand the BAF values
331	found for Hg.
332	In agreement with hepatic Cd content, BAF values found in liver are higher at
333	warmer temperatures and, as discussed above, this may be due to the function
334	that this organ has in storage and detoxification of Cd (Jezierska and Witeska,
335	2006; Köck et al., 1996; Yang and Chen, 1996).
336	BAF for Pb in liver is lower than the BAF for the other two toxic metals (Cd and
337	Hg), which couldmight be due to Pb preferential deposition into internal soft
338	tissues other than liver as reported in Oncorhynchus mykiss (Alves et al., 2006).
339	Mn shares a large number of transport mechanisms with other metals, like such
340	as transferrin, or the promiscuous divalent metal transporter (DMT) (Bury and
341	Grosell, 2003), and excretion mechanisms such as bile (Papavasiliou et al.,
342	1966), and for this reason, similar Mn behavior to other metals should be
343	expected. However, in this study Mn has a completely different reaction and

presents the lowest uptake from dietary sources, as the BAF values 344 345 suggestshow. This may mean that Mn stores is stored in other tissues than liver, 346 and this phenomenon may be amplified with increasing temperature. 347 It is important to note that BAFBAFs of Cu and Zn in liver are almost the same 348 when determined for the same temperature, which indicates that a similar fraction of metal contained in feed is retained, and also that temperature has a 349 similar effect on the overall metabolism of both elements. 350 Essential metals have transport systems that include several both specific and 351 general mechanisms. They are strongly regulated in organisms under non-352 353 stressful conditions, and consequently environmental factor variations should 354 not greatly affect their concentration. On the other hand, toxic metals have no 355 specific mechanisms and use essential element transport proteins and channels, such as divalent metal transporter (DMT1), diverse ZIP proteins or 356 MT (Bury et al., 2003; Coyle et al., 2002; He et al., 2006; Vesey et al., 2010). It 357 358 is clear that the common mechanisms may be affected by temperature to a different degree or in a different measure or way depending on the metal and 359 360 tissue. For instance, the temperature--dependent increase of membrane fluidity and physical diffusion rate may be of special relevance for Hg accumulation 361 (Foulkes, 2000). It is difficult to determine at which level the mechanisms 362 363 implicated in the toxicokinetics of each metal are influenced by temperature. The effect of temperature on MT level in liver could be a direct thermal 364 response, or may be related to the increase in metal content. MT synthesis is 365 considered one of the best-known biochemical detoxification mechanisms for 366 metal and it is widely demonstrated that its induction may be influenced by 367 metal contamination. It is very common to try to relate seasonal variability to 368

369	temperature, but in this case the reproductive state may also be involved (Gorbi
370	et al., 2005; Köck et al., 1996; Olsson et al., 1996; Rotchell et al., 2001). There
371	are only a few studies specially designed for elucidating the effect of
372	temperature on MT synthesis in fish. Van Cleef-Toedt et al. (2001)
373	demonstrated that non-spawning Fundulus heteroclitus exposed to thermal
374	stress exhibited significantly elevated liver, gill, and intestine MT mRNA
375	expression compared with controls.
376	In this study we have found an unusual correlation between MT and Fe content
377	in liver of Sparus aurata. MT is known to be a protein involved in protection
378	against oxidative stress (Viarengo et al 1999; Andrews, 2000; Gourgou et al.
379	2010), including heat-induced oxidative stress (Ivanina et al., 2009). On the
380	other hand, Fe may play a role in hepatic injury after hyperthermia and is
381	considered as an indicator of liver damage in mammals (Bloomer et al., 2008;
382	Bowen et al., 2006; Skibba and Gwartney, 1997). Fe excess is believed to
383	generate oxidative stress, understood as an increase in the steady state
384	concentration of oxygen radical intermediates (Puntarulo, 2005). Taking all the
385	above together we can suggest that the MT-Fe correlation may be due to the
386	accumulation of Fe in liver produced by hyperthermia, which can cause
387	oxidative stress which, in turn, produces MT induction.

#### 388 5. Conclusion

The results obtained in this study show that metal bioaccumulation from nonexperimentalexperimentally contaminated commercial feed and MT induction takes place at warm temperatures in *Sparus aurata*. Furthermore, different patterns of metal distribution among tissues and temperature indicate disturbed

- 393 essential metal homeostasis and higher levels of toxic metal bioaccumulation,
- <sup>394</sup> particularly of Hg and Cd. The interactions among water temperature, metal
- <sup>395</sup> uptake from feed and their distribution between tissues in these species exist,
- but <u>we can still <del>we cannot</del> provide a full mechanistic explanation for itthis</u>.
- 397 Future research on the identification of the temperature-dependent mechanisms
- 398 involved in metal toxicokinetics is required in order to reach a better
- 399 understanding of temperature role in ecotoxicology and its implications in
- 400 aquaculture in a warming world context.

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## Facultat de Ciències Biològiques

- Warming temperatures has an influence on MT in liver and on metal accumulation in liver and muscle in gilthead sea bream.
- The effect of temperature on accumulation differs among metals, being cadmium and mercury accumulation the most sensitive to this variation.
- Essential metal homeostasis was disturbed at the warmest temperature.
- MT was correlated with cadmium, cooper, mercury, zinc and iron content in liver.

*carrer* **del Doctor Moliner, 50** BURJASSOT 46100

 telèfon
 (34)
 96
 354
 3378

 fax
 (34)
 96
 354
 3202

 e-mail
 torrebla@uv.es

## 1 Long-term effect of temperature on bioaccumulation of dietary metals and

- 2 metallothionein induction in Sparus aurata.
- 3 Diana Guinot<sup>1</sup>, Rocío Ureña<sup>1</sup>, Agustín Pastor<sup>2</sup>, Inmaculada Varó<sup>3</sup>, Jose del
- 4 Ramo<sup>1</sup> and Amparo Torreblanca<sup>1</sup>
- <sup>5</sup> <sup>1</sup>Department of Functional Biology, Faculty of Biological Sciences, University of
- 6 Valencia. Dr. Moliner, 50, 46100 Burjassot (Valencia). Spain. Phone +34 6 35443378
- 7 FAX +34 6 3543202 Amparo.Torreblanca@uv.es (Corresponding author)
- <sup>2</sup>Departament of Analytic Chemical, University of Valencia. Dr. Moliner, 50, 46100
- 9 Burjassot (Valencia), Spain.
- <sup>3</sup>Institute of Aquaculture Torre de la Sal (CSIC), 12595 Ribera de Cabanes, Castellón.
- 11 Spain.

### 12 Abstract

13 Previous studies have demonstrated that the commercial feed of aquacultured fish contains trace amounts of toxic and essential metals which can accumulate 14 in tissues and finally be ingested by consumers. Recently rising temperatures, 15 associated to the global warming phenomenon, have been reported as a factor 16 to be taken into consideration in ecotoxicology, since temperature-dependent 17 18 alterations in bioavailability, toxicokinetics and biotransformation rates can be expected. Sparus aurata were kept at 22°C, 27°C and 30°C for three months in 19 order to determine the temperature effect on metallothionein induction and 20 21 metal bioaccumulation from a non-experimentally contaminated commercial feed. A significant temperature-dependent accumulation of cadmium (Cd), 22 copper (Cu), mercury (Hg), zinc (Zn), lead (Pb) and iron (Fe) was found in liver, 23 24 together with that of manganese (Mn), Fe and Zn in muscle. Hg presented the

- highest bioaccumulation factor, and essential metal homeostasis was disturbed
- 26 in both tissues at warm temperatures. An enhancement of hepatic
- 27 metallothionein induction was found in fish exposed to the highest temperature.
- **Keywords:** metal; metallothionein; bioaccumulation factor; temperature; *Sparus*
- 29 aurata

### 30 **1. Introduction**

The aquatic ecosystems are undergoing a warming in their deep and surface 31 waters, a fact which may have significant consequences on the organisms 32 inhabiting them (Noves et al., 2009). Gilthead sea bream Sparus aurata is a 33 34 cosmopolitan species distributed throughout the Mediterranean and the NE Atlantic, and which is of great economic interest since it is one of the principal 35 species of Mediterranean aquaculture. Being an ectothermic species, it is 36 37 vulnerable to the effects of temperature variations on its metabolism and physiology. These variations may produce changes in the toxicokinetics, 38 bioavailability, biotransformation, homeostasis, absorption rate and elimination 39 of different compounds (Douben, 1989; Köck et al., 1996; Yang and Chen, 40 1996). Other key physiological mechanisms such as respiration, feeding rate, 41 growth and reproduction may also be affected (Bowen et al., 2006; Heugens et 42 al., 2001). The thermic induced variations in the toxicokinetics of pollutants, 43 together with the increase in exposure to the same as a consequence of climate 44 45 warming (Carrie et al., 2010) may present a risk for the development and 46 survival of species of commercial interest, affecting the quality of the end product (Noves et al., 2009). 47

In cultured fish, it has been demonstrated that the feed is the principal source of contamination by metals Cadmium (Cd), lead (Pb) and mercury (Hg), among others, are potentially toxic and tend to accumulate in the tissues, which in the end are ingested by the consumers (Fernandes et al., 2009; Maule et al., 2007; Cretì et al., 2010). Their accumulation in the organisms depends on the concentration, route of absorption, environmental conditions and other intrinsic

factors (Bowen et al., 2006; Chowdhury and Blust, 2001; Jezierska and 54 55 Witeska, 2006; Karakoc and Dincer, 2003; Lemus and Chung, 1999). Due to the growth in the activity of aquaculture in recent decades, it has 56 57 become of special relevance to learn the influence of the increase in temperature on the possible routes of absorption, accumulation and elimination 58 of metals in these organisms in the context of global warming. Temperature 59 may be a determining factor in the capture, transporting and metabolism of the 60 metals incorporated through the feed, both of the essential metals which may 61 become toxic at high concentrations in the tissues, and of the non-essential 62 63 metals.

64 Metallothionein (MT) is a low-molecular-weight metal binding protein and is known to play an important role in protection against heavy metal toxicity. In 65 66 addition to the detoxification of toxic metals such as Cd and Hg, MT is involved in the maintenance of homeostasis of essential trace elements such as zinc 67 (Zn) and copper (Cu) (Coyle et al., 2002; Hamilton and Mehrle, 1986). Its role in 68 the protection against xenobiotics or in the cellular protection against oxidative 69 stress should be underlined (Coyle et al., 2002; Van Cleef-Toedt et al., 2001). 70 71 Although its synthesis is related to the metal exposure, its levels can be affected by endogenous and exogenous factors such as the reproductive cycle or the 72 temperature (Van Cleef-Toedt et al., 2001). Variations in the water temperature 73 74 could directly or indirectly modify the behavior of this protein as regards the bioaccumulation of metals, as well as its participation in toxicokinetic processes 75 (Baykan et al., 2007; Gorbi et al., 2005; Rotchell et al., 2001). 76 Given the lack of information as regards the effect of temperature on the 77

bioaccumulation of metals via dietary sources, and on the synthesis of MT, an

- understanding of these processes in the light of the problem of global warming
  and its repercussions on species of commercial interest such as *Sparus aurata*is required. As such, the aims of this work are:
- 1) To determine the effect of temperature on the bioaccumulation of essential
- 83 (Cu, Fe, Mn and Zn) and non-essential (Cd, Hg and Pb) metals experimentally
- via a non-contaminated commercial feed.
- 2) To discover whether temperature has any influence on the homeostasis of
  the essential metals.
- 3) To elucidate the role of metallothionein in the above mentioned temperatureinduced changes.

### 89 **2. Materials and Methods**

90 2.1. Animal collection and maintenance

Adult Sparus aurata were distributed and acclimated in 500 L tanks containing 91 seawater (37‰) at a constant temperature of 22°C, with continuous aeration 92 and natural photoperiod in a closed circuit for two months prior to the 93 experiment. Subsequently, the temperature of two of the experimental groups 94 95 was gradually increased to reach 27°C and 30°C respectively. Control groups of fish were maintained at 22 °C throughout the experiments. The animals were 96 97 fed with commercial pellets (1.5% of body mass per day) and the survival 98 percentage was 99%. Fish were kept under constant conditions for three months. After this period, 6-7 fish were removed at random from each 99 experimental group and placed in water containing 30 mg L<sup>-1</sup> of anesthetic clove 100 101 oil. Lengths and weights of gilthead sea bream were recorded. Fish were 102 sacrificed; livers and a piece of dorsal muscle tissue were dissected and immediately frozen in liquid nitrogen and stored at -80 °C. 103

### 104 2.2. *MT* determination by differential pulse polarography

105	Approximately 0.2 g wet weight portions of frozen liver were homogenized using
106	ultra-turrax in 20 mM Tris-HCl buffer, 1 mM DTT and 0.2 mM PMSF pH 8.6 in
107	an ice bath. The homogenates were centrifuged at 30 000 $g$ for 45 min at 4°C.
108	The supernatant was heated at 80°C for 10 min in order to denature high
109	molecular weight proteins and subsequently centrifuged at 30 000 g for 45 min
110	at 4°C,The heat-treated supernatant, containing thermally stable MT, was
111	separated from precipitated proteins. MT was measured using differential pulse
112	polarography (Ureña et al., 2007).
113	An aliquot of the heat-treated supernatant was added to the polarographic cell,
114	containing 20 mL hexammincobalt chloride buffer (the supporting electrolyte),
115	together with Triton-X (0.025% v/v). The cell was purged for 2 min with purified
116	$N_2$ prior to analysis. The polarographic response was measured during a
117	potential scan between -1.38 V and -1.7 V (Model 757 VA Computrace
118	Analyser, Methrom, Switzerland) in SMDE mode. Quantification of MT was
119	performed by using the standard addition method with rabbit liver MT I+II
120	(Sigma). Results are expressed as $\mu g g^{-1}$ wet weight of tissue.

121 2.3. Metal analysis

122 Samples of 0.1-0.5 g wet weight of liver and muscle were digested in

123 concentrated nitric acid 65% (Baker) at room temperature overnight, and were

heated at 80°C for 2 h. In order to determine the trace amounts of metal

125 contained in the feed, samples of commercial pellets were also digested (n=3).

126 Within each digestion series, appropriate blanks with ultra-pure water were also

127 subjected to the same procedure in order to take background contamination

levels into acount. After cooling, solutions were transferred to a standard 128 129 volume with ultra-pure water. Determination of metals (Cd, Cu, Zn, Fe, Pb, Hg and Mn) was undertaken using an ICP-Mass (Elan DRC-I, Perkin-Elmer Sciex). 130 131 Samples of similar weight of certified reference material (DOLT-3 and LUTS-1, National Research Council of Canada, Ottawa), were digested and analyzed 132 during each analytical run. The values of all elements found were in good 133 134 agreement with the certified values, with the recoveries ranging from 91% to 104%. 135

136 2.4. Data analysis and calculations

Metal content ratio between muscle and liver was calculated in order to detect
changes in the distribution of metals as a consequence of the thermal
experimental conditions, and to reveal the proportion of accumulated metal in
each tissue.

Bioaccumulation factor (BAF) was calculated for each metal-tissue-temperature
combination in order to detect the effects of temperature on the global
toxicokinetics among metals. This was calculated as a quotient between the
metal concentration in each tissue and the metal content in the commercial
pellets as described by Dabrowska et al. (1996).

146 2.5. Statistical analysis

Statistical analyses were carried out using the software Stata 10 (Stata Corp).
Transformations of the data were performed when the assumption of normality
of residuals was not met. One-way ANOVA was used in the analysis of data to
check the influence of temperature on each variable. The Scheffe test was used

as a post hoc test to demonstrate the differences among the three temperature 151 152 groups. Two-way ANOVA, with temperature and metal as fixed factors, was used in the statistical analysis of BAFs, followed by a post hoc Bonferroni test. 153 154 The Kruskal Wallis test was used when transformed data did not present homoscedasticity. Pearson correlation coefficients (r) were calculated between 155 MT levels and metal content in liver, in order to measure the strength of 156 association between these variables. Results are presented as means ± SEM 157 and a p value lower than 0.05 was considered as statistically significant. 158

159 **3. Results** 

### 160 3.1. Metal content in food and biometry

161 Metal contents of food pellets were  $0.38 \pm 0.05 \mu g g^{-1}$  w.w. for Cd,  $11.6 \pm 0.7 \mu g$ 

162  $g^{-1}$  w.w. for Cu, 173 ± 9 µg  $g^{-1}$  w.w. for Fe, 40.± 3 µg  $g^{-1}$  w.w. for Mn, 0.05±

163  $0.01\mu g g^{-1}$  w.w. for Hg,  $0.15 \pm 0.09 \mu g g^{-1}$  w.w. for Pb and  $139 \pm 9 \mu g g^{-1}$  w.w.

164 for Zn.

The weight and length together with the condition index calculated for gilthead sea bream exposed to three experimental temperatures for three months are shown in Table 1. As can be seen, temperature has no effect on the condition index or weight. In contrast, fish kept at 30°C (p<0.05) were found to be significantly shorter than those kept at 22°C and 27°C.

170 Insert Table 1

171 3.2. Temperature effect on metal contents and their distribution among tissues

Metal content reached in liver and muscle of gilthead sea bream is shown in
Figure 1. The ANOVA test shows a significant effect of temperature on Cd, Hg

and Pb content (p<0.001), and on concentration of the essential metals Cu (p<0.01), Zn and Fe (p<0.001) in liver, but not on Mn content. Moreover, the results show a significant effect of temperature on Mn, Fe and Zn content (p<0.01) in muscle.

The muscle/liver ratios for each metal and temperature were calculated when 178 possible and are shown in Table 2. They indicate the distribution of metals after 179 180 temperature exposure and the proportion of each in muscle in relation to liver. The highest muscle/liver ratio for fish maintained at 22 and 27 °C was reached 181 by Hg, whereas at 30 °C the highest ratio was reached by Mn. As can be seen 182 183 in Table 2, calculated ratios for Hg and Fe decreased significantly when 184 temperature rose. Notably, and unlike for the other metals, Mn ratio rose significantly with increasing temperature. 185

186 Insert Figure 1

187 Insert Table 2

3.3. Temperature effect on toxicokinetic related processes: Bioaccumulation
Factor and MT induction

The BAF values obtained for each metal-tissue-temperature combination are shown in Table 3. The two-way ANOVA test shows significant differences among between metals in muscle (p<0.001) at all the temperatures tested. Hg BAF in this tissue is considerably higher in comparison with that calculated for the other measured metals, reaching a value of 0.45 at 30°C. This means that muscle presents Hg content of almost a half of the feed content. In contrast, Mn BAF exhibits the lowest BAF for all temperature exposures.

197 Statistical analysis of BAFs also showed significant differences among metals in

liver (p<0.001). BAF for Hg in liver differed from the BAFs for other metals at

199 27°C and 30°C, and reached the highest BAF registered in this study (3.20 at

30°C). It should be noted that Hg BAFs in liver at 27°C and 30°C are,

respectively, nearly 3 and 5 fold higher than those determined at 22°C. A similar

202 pattern to Hg was found for Cd, and in this case BAFs are 3 and 4 fold higher at

203 27°C and 30°C, respectively, than those determined at 22°C. Mn BAFs in liver

are the lowest among the studied metals.

205 Insert Table 3

206 MT levels determined in liver for the three experimental temperature groups are

shown in Figure 1. As can be seen, there is a significant thermal effect on

208 hepatic MT content. An MT significant induction is found in fish exposed to the

highest temperature (p<0.001). Furthermore, significant positive correlations

210 between MT and Cd in liver (r= 0.52; p<0.05), Hg (r= 0.78; p<0.001), Zn (r=

211 0.82; p<0.001), Cu (r= 0.82; p<0.001) and Fe (r= 0.78; p<0.001) have been

found, but not with Pb and Mn.

### 213 4. Discussion

4.1. Metal content in food and biometry

Length differed significantly between animals exposed to 22°C and 30°C (Table

1), which could be due to an increase in energetic demand related to an

217 increase in metabolic rate. A reduced growth and length at high temperature

- was reported in *Oncorhynchus kisutch,* as a consequence of the energetic
- deficit after an elevated metabolism (Bowen et al., 2006).

Metal content of food pellets was in the same range as most of the values reported in the literature for commercial pellets (Ciardullo et al., 2008; Dang and Wang, 2009; Mackee et al., 2008; Minganti et al., 2010). Cd, Pb and Hg feed content are also below the established maximum permitted level in fish feed according to EU Directive 2002/32/EC and 2005/87/EC amending Annex I to Directive 2002/32/EC.

226

4.2. Effect of temperature on metal content and its distribution among tissues

228 Metal contents in tissues (Figure 1) were in the same range as most of the values reported in the literature for farmed gilthead sea bream (Cretì et al., 229 230 2010; Minghetti et al., 2008). It should be noted that Cd, Pb and Hg 231 concentrations measured in muscle of Sparus aurata in the present work are below the established maximum permitted level in muscle of several species 232 233 according to EU Regulation (EC) No 1881/2006. We have found a general trend of increasing metal content in gilthead sea bream tissues with increasing 234 235 temperatures. The positive relationship between temperature and increased 236 metal uptake in different aquatic organisms has been explained by accelerated biochemical and physiological processes, changes in the membrane 237 permeability, alterations of metal transport systems, and binding to several 238 239 ligands (Baykan et al., 2007, Chowdhury and Blust, 2001; Jezierska and Witeska, 2006; Lemus and Chung, 1999). 240 The significant increase in Hg content in liver at both 27°C and 30°C compared 241 to at 22 °C indicates strong temperature dependence. Additionally, the 242 muscle/liver ratio (Table 2) confirms the preferential affinity of Hg to lipidic 243

tissues such as the liver. Using an experimental trophic chain, Boudou et al. 244 245 (1979) demonstrated that raising the temperature of the environment has a synergic effect on the quantities of Hg bioaccumulated by a carnivorous fish. 246 247 Cd content in liver also increased with rising temperature. Liver, as is well known, is an organ involved in storage and detoxification of Cd, and a 248 progressive accumulation of dietary Cd over time has been explained by the 249 250 continuous transference via the portal system from the digestive tract to the liver (Handy, 1993). In previous temperature effect studies in fish, the metal was 251 supplied as waterborne Cd (Douben, 1989; Köck et al., 1996) and the 252 253 temperature-dependent increase of metal content was explained by a greater 254 increase in uptake rate than the increase found in the elimination rate (Douben, 255 1989), although mechanisms involved in Cd accumulation may differ to a great 256 extent depending on the Cd exposure source.

The highest Pb content in liver was reached by fish kept at 27°C. In some
studies Pb uptake rate has been affirmed to be temperature-dependent in fish
(Köck et al., 1996).

An enhancement of Mn in muscle of fish maintained at 30°C in relation to those 260 261 kept at colder temperatures was determined, nevertheless no significant variation in Mn was found in liver. However, when muscle/liver ratios were 262 calculated, a significant effect of the temperature was detected in such a way 263 264 that the ratio was higher at the warmer temperature, with the muscle acting as an Mn receptor tissue as temperature increased. However, little information is 265 266 available concerning the effect of temperature on Mn accumulation in fish muscle. Adam et al. (1997) described two groups of organs with different 267 elimination kinetics of <sup>54</sup>Mn, the first was the penetration or transit group, and 268

the second the receptor or storage organs. According to this, liver was classified 269 270 as an Mn distribution and elimination tissue, and muscle as a receptor. The high levels of Cu in liver of fish kept at 30°C in comparison to those kept at 271 272 lower temperatures are in agreement with findings reported in several studies. For instance, *Petenia kraussii* juveniles exposed to 22°C had lower Cu basal 273 concentrations than those exposed to 30°C (Lemus and Chung, 1999). 274 275 Temperature-related increase in Cu content has been explained by other authors as a consequence of a different extent in the rising of accumulation and 276 elimination rates, resulting in a final positive balance for uptake (Glover et al., 277 278 2003; Lemus and Chung, 1999). 279 In concordance with our results on Zn accumulation, Karakoç and Dinçer (2003) reported a temperature induced accumulation of Zn in liver, and Van 280 281 Campenhout et al. (2007) described a promotion of Zn assimilation by temperature in fish. Zn uptake in the apical cell membrane of rainbow trout 282 283 intestine was also found to be a temperature-dependent process (Glover et al., 2003). In addition to the above findings, and taking into account the 284 enhancement of Zn in muscle and liver of fish maintained at 30°C, and that no 285 286 significant variation in the muscle/liver ratios was found in our study, we can suggest that a temperature-induced disturbance of Zn homeostasis in Sparus 287 aurata may be occurring. 288 289 Levels of Fe in muscle and liver remained fairly constant when temperature increased from 22°C to 27°C, as can be expected for a tightly regulated metal 290 under physiological conditions. The sharp increase in Fe content in both 291 tissues, and the different ratio value found at 30°C may be due to Fe 292 homeostasis disturbance, indicating that some pathology is occurring. It should 293

be born in mind that Fe loading leads to free radical damage by the Fenton

reaction. Evidence in this direction has been found in fish and rats, suggesting a

role for Fe in hepatic injury following hyperthermia. (Bloomer et al., 2008;

Bowen et al., 2006; Skibba and Gwartney, 1997).

In summary, essential metals in muscle are strongly regulated under

299 physiological conditions, although at 30°C their homeostasis is disturbed as a

300 result of a warmer temperature. The inverse relation between muscle/liver ratio

and temperature is a general trend for all the metals studied except Mn, which

302 indicates that warmer temperatures facilitate metal regulation processes in such

a way that there is a preferential accumulation in liver thus protecting peripheraltissues.

4.3. Temperature effect on toxicokinetic related processes: Bioaccumulation
Factor and MT induction

BAF, the quotient between the metal concentration in tissue and the metal content in the feed, allowed us to quantify the accumulation of metals from feed in *Sparus aurata*, and to compare the temperature effect on bioaccumulation among the determined metals.

Metabolic rates of ectothermic organisms may be strongly dependent on temperature, and the same can be said for heavy metals complexing with cellular constituents, which may affect tissue distribution and BAF. We can suggest that variations in toxicokinetics caused by temperature exist in some of the studied metals and tissues, although we cannot determine whether these changes are a consequence of the temperature effect on assimilation efficiency, or on feeding rate.

Hg presents the highest BAF values among metals for both tissues at all 318 319 temperatures tested (Table 3). Hg accumulation depends on its speciation, 320 mainly because methylated Hg (MeHg) is better absorbed from the 321 gastrointestinal tract than the inorganic form in such a way that in mammals 95% is absorbed (Liu et al., 2008), and furthermore it undergoes extensive 322 enterohepatic recycling (Liu et al., 2008). Taking into account all the above, 323 324 together with the fact that MeHg is the predominant form of Hg in fish feed (>80%) (Amlund et al., 2007), it is easy to understand the BAF values found for 325 Hg. 326 327 In agreement with hepatic Cd content, BAF values found in liver are higher at

warmer temperatures and, as discussed above, this may be due to the function
that this organ has in storage and detoxification of Cd (Jezierska and Witeska,
2006; Köck et al., 1996; Yang and Chen, 1996).

BAF for Pb in liver is lower than the BAF for the other two toxic metals (Cd and

Hg), which might be due to Pb preferential deposition into internal soft tissues

other than liver as reported in *Oncorhynchus mykiss* (Alves et al., 2006).

334 Mn shares a large number of transport mechanisms with other metals such as

transferrin, or the promiscuous divalent metal transporter (DMT) (Bury and

Grosell, 2003), and excretion mechanisms such as bile (Papavasiliou et al.,

1966), and for this reason similar Mn behavior to other metals should be

expected. However, in this study Mn has a completely different reaction and

presents the lowest uptake from dietary sources, as the BAF values show. This

may mean that Mn is stored in other tissues than liver, and this phenomenon

341 may be amplified with increasing temperature.

332

342 It is important to note that BAFs of Cu and Zn in liver are almost the same when 343 determined for the same temperature, which indicates that a similar fraction of 344 metal contained in feed is retained, and also that temperature has a similar 345 effect on the overall metabolism of both elements.

Essential metals have transport systems that include both specific and general 346 mechanisms. They are strongly regulated in organisms under non-stressful 347 348 conditions, and consequently environmental factor variations should not greatly affect their concentration. On the other hand, toxic metals have no specific 349 mechanisms and use essential element transport proteins and channels, such 350 351 as divalent metal transporter (DMT1), diverse ZIP proteins or MT (Bury et al., 352 2003; Coyle et al., 2002; He et al., 2006; Vesey et al., 2010). It is clear that the common mechanisms may be affected by temperature to a different degree or 353 354 in a different way depending on the metal and tissue. For instance, the temperature-dependent increase of membrane fluidity and physical diffusion 355 356 rate may be of special relevance for Hg accumulation (Foulkes, 2000). It is difficult to determine at which level the mechanisms implicated in the 357 toxicokinetics of each metal are influenced by temperature. 358 359 The effect of temperature on MT level in liver could be a direct thermal response, or may be related to the increase in metal content. MT synthesis is 360 considered one of the best-known biochemical detoxification mechanisms for 361 362 metal and it is widely demonstrated that its induction may be influenced by metal contamination. It is very common to try to relate seasonal variability to 363 364 temperature, but in this case the reproductive state may also be involved (Gorbi et al., 2005; Köck et al., 1996; Olsson et al., 1996; Rotchell et al., 2001). There 365 are only a few studies specially designed for elucidating the effect of 366

temperature on MT synthesis in fish. Van Cleef-Toedt et al. (2001)

368 demonstrated that non-spawning *Fundulus heteroclitus* exposed to thermal

369 stress exhibited significantly elevated liver, gill, and intestine MT mRNA

370 expression compared with controls.

In this study we have found an unusual correlation between MT and Fe content 371 372 in liver of Sparus aurata. MT is known to be a protein involved in protection 373 against oxidative stress (Viarengo et al 1999; Andrews, 2000; Gourgou et al. 2010), including heat-induced oxidative stress (Ivanina et al., 2009). On the 374 other hand, Fe may play a role in hepatic injury after hyperthermia and is 375 376 considered as an indicator of liver damage in mammals (Bloomer et al., 2008; 377 Bowen et al., 2006; Skibba and Gwartney, 1997). Fe excess is believed to generate oxidative stress, understood as an increase in the steady state 378 379 concentration of oxygen radical intermediates (Puntarulo, 2005). Taking all the 380 above together we can suggest that the MT-Fe correlation may be due to the accumulation of Fe in liver produced by hyperthermia, which can cause 381 oxidative stress which, in turn, produces MT induction . 382

### 383 5. Conclusion

The results obtained in this study show that metal bioaccumulation from nonexperimentally contaminated commercial feed and MT induction takes place at warm temperatures in *Sparus aurata*. Furthermore, different patterns of metal distribution among tissues and temperature indicate disturbed essential metal homeostasis and higher levels of toxic metal bioaccumulation, particularly of Hg and Cd. The interactions among water temperature, metal uptake from feed and their distribution between tissues in these species exist, but we can still not

- 391 provide a full mechanistic explanation for this. Future research on the
- identification of the temperature-dependent mechanisms involved in metal
- toxicokinetics is required in order to reach a better understanding of
- temperature role in ecotoxicology and its implications in aquaculture in a
- 395 warming world context.

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# **Figure captions**

**Figure 1.** MT levels in liver and metal content in muscle and liver of *Sparus aurata* exposed to three different experimental temperatures. Data represent mean  $\pm$  SEM (n≥6). Means that do not share the same letter differ significantly (p<0.05).



**Table 1.** Biometry of *Sparus aurata* kept under temperature exposure for three months. Data represent mean  $\pm$  SEM (n≥6). Means that do not share the same letter differ significantly (p<0.05).

	22°C	27°C	30°C
Weigth (g)	290 ± 76 <sup>a</sup>	259 ± 56 <sup>a</sup>	$218 \pm 64^{a}$
Lenght (cm)	234 ± 21 <sup>a</sup>	221 ± 22 <sup>ab</sup>	199 ± 19 <sup>b</sup>
CI (g cm <sup>-3</sup> *10 <sup>-5</sup> )	$2.2 \pm 0.3^{a}$	$2.4 \pm 0.5^{a}$	$2.7 \pm 0.4^{a}$

**Table 2.** Muscle/liver metal ratios calculated of *Sparus aurata* exposed long-term to three different experimental temperatures. Data represent mean  $\pm$  SEM (n≥6). Means that do not share the same letter differ significantly (p<0.05).

	22°C	27°C	30°C	
Fe	$0.09 \pm 0.02^{a}$	0.06 ± 0.01 <sup>a</sup>	$0.07 \pm 0.05$ <sup>b</sup>	
Mn	$0.07 \pm 0.04^{a}$	$0.11 \pm 0.02^{ab}$	$0.25 \pm 0.02^{b}$	
Hg	$0.43 \pm 0.23^{a}$	$0.16 \pm 0.06$ <sup>b</sup>	$0.15 \pm 0.09^{b}$	
Zn	0.14 ± 0.08 <sup>a</sup>	$0.08 \pm 0.04$ <sup>a</sup>	$0.10 \pm 0.06$ <sup>a</sup>	

**Table 3** Metal bioaccumulation factor (BAF) in liver and muscle of *Sparus aurata* kept under temperature exposure for three months. Data represent mean  $\pm$  SEM (n≥6). Means that do not share the same letter at the same temperature differ significantly (p<0.05).

Tissue	Т	Cd	Cu	Fe	Mn	Hg	Pb	Zn
Liver	22°C	$0.35 \pm 0.33^{ab}$	$0.19 \pm 0.07^{ab}$	$0.16 \pm 0.06^{ab}$	$0.03 \pm 0.01^{a}$	$0.66 \pm 0.33^{b}$	$0.14 \pm 0.11^{ab}$	$0.20 \pm 0.05^{ab}$
	27°C	1.01 ± 0.23 <sup>a</sup>	0.31 ± 0.01 <sup>b</sup>	$0.16 \pm 0.01$ <sup>b</sup>	$0.03 \pm 0.01$ <sup>b</sup>	1.73 ± 0.57 <sup>°</sup>	$0.43 \pm 0.12^{b}$	$0.30 \pm 0.08$ <sup>b</sup>
	30°C	$1.40 \pm 0.46^{a}$	0.61 ± 0.24 <sup>b</sup>	$0.48 \pm 0.11^{bc}$	$0.03 \pm 0.01$ <sup>c</sup>	$3.20 \pm 0.79^{d}$	$0.27 \pm 0.11^{bc}$	$0.57 \pm 0.07^{bc}$
Muscle	22°C			$0.017 \pm 0.008^{a}$	$0.002 \pm 0.001$ <sup>a</sup>	0.27 ± 0.13 <sup>b</sup>		$0.03 \pm 0.01^{a}$
	27°C			$0.008 \pm 0.003^{a}$	$0.003 \pm 0.001$ <sup>a</sup>	$0.27 \pm 0.06$ <sup>b</sup>		$0.02 \pm 0.01^{a}$
	30°C			$0.04 \pm 0.03^{a}$	$0.006 \pm 0.004^{a}$	$0.45 \pm 0.27$ <sup>b</sup>		$0.06 \pm 0.03^{a}$