

1 **The effects of different endocrine disruptors defining compound-specific**
2 **alterations of gene expression profiles in the developing testis**

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21

22 **Abstract**

23
24 Environmental contaminants considered endocrine disruptors have been shown to affect testis
25 development and function but the mechanisms of action are not clear. We now have analyzed the
26 effects on the transcriptome in testes of mice exposed to mono-(2-ethylhexyl)-phthalate (9.2; 46.3
27 or 92.7 mg/kg/d), zearalenone (1.3; 3.9 or 6.6 mg/kg/d), lindane (16.6; 32.2 or 64.4 mg/kg/d),
28 bisphenol-A (0.16; 16 or 64 mg/kg/d) or 17 β -estradiol (0.006; 0.012 or 0.048 mg/kg/d). The
29 compounds were orally administered in the drinking water during distinct developmental periods:
30 A) mothers were exposed only during the two weeks before mating; B) the exposure was
31 continued during pregnancy until birth or C) exposure was continued for a further four weeks
32 after birth. Testes were studied at four weeks of age. Mono-(2-ethylhexyl)-phthalate and
33 zearalenone, both produced specific alterations of gene signatures. Interestingly, this was
34 irrespective of the concentration of the toxicant or the developmental period during which
35 exposure occurred.

36
37 **Key Words**
38
39 Endocrine disruptor; testis; 17 β -estradiol; Lindane; Bisphenol-A; Mono-Ethylhexyl Phthalate;
40 Zearalenone; DNA microarrays; gene expression profile.

41
42 **Abbreviations**
43
44 ED, Endocrine Disruptor. TDS, Testicular Dysgenesis Syndrome. MEHP, Mono-EthylHexyl
45 Phthalate. DEHP, Di 2-EthylHexyl Phthalate. BPA, Bisphenol-A. LIN, Lindane. ZEA,
46 Zearalenone. E2, 17- β -estradiol. pn, post-natal.

47

48 **1. Introduction**

49

50 There is significant documentation on the adverse effects of environmental pollutants on
51 reproductive health [1-3]. Environmental toxicants that act as agonists or antagonists of natural
52 hormones, generically considered as endocrine disruptors (EDs), can affect the development of
53 the reproductive system and associated organs [4, 5]. There is some controversy as to the effects
54 and mechanisms by which EDs act [6, 7], although the most accepted hypothesis holds that EDs
55 interfere with steroid hormone action through disruption of steroid biosynthesis, the hormone
56 balance, signaling pathways of downstream consequences. There is currently significant concern
57 regarding the increase in male and female hormone-related disorders detected in epidemiological
58 studies. In mammals, the male reproductive organs have been clearly identified as a target for the
59 deleterious action of many environmental toxicants, and the Testicular Dysgenesis Syndrome
60 (TDS) could be a consequence of developmental exposure to such compounds [8, 9]. TDS groups
61 four clinical and etiologically related traits: hypospadias, cryptorchidism, low sperm counts and
62 testicular tumors [10]. These dysfunctions could originate through changes in the
63 microenvironment that affect different target cells during embryonic differentiation [11].
64 Nevertheless, the molecular mechanisms by which toxicants or potential EDs alter
65 spermatogenesis and testicular function are yet to be fully established. Indeed, their effects cannot
66 be simply explained by direct interactions with hormone pathways [12], and interference with
67 gene expression regulation could occur at diverse levels during development, either as a direct or
68 indirect consequence of exposure to these toxicants. The nature of the compounds, the dose and
69 extent of exposure, as well as the developmental period at which exposure occurs are also factors
70 that should be taken into account when considering the mechanisms causing the adverse effects
71 on testicular development and function.

72 There is evidence that the origin of the adverse effects lies in fetal exposure [13].
73 Potential effects of endocrine disruptors via placental transport during pregnancy and via milk
74 during lactation are well documented. Transplacental absorption and fetal detection of different
75 environmental estrogenic compounds has been reported in experimental animals [14-16] and in
76 humans [17, 18]. Similarly, persistence in breast milk was detected in different populations [19-
77 21]. Nevertheless, the mechanisms underlying the influence of the different xeno-compounds are

78 not well established, especially since daily exposure usually involves contact with a mixture of
79 contaminants.

80 Over 150 different contaminants have been reported in individuals in the US and some,
81 such as phthalates, can be found in nearly 100% of the population [22]. In humans, basic studies
82 associating chemical exposure and testis development disorders are very difficult to carry out
83 [13]. Indeed, most of the biomarkers that could serve as endpoints, such as semen quality or
84 hormone levels, cannot be considered as direct or adequate evaluation of toxicological exposure
85 [23].

86 To date, studies into the environmental induction of changes in gene expression and
87 regulation as part of disease etiology have mainly focused on specific genes or on the relevant
88 genetic pathways related to particular diseases, including reproductive dysfunction [24].
89 However, most emphasis has been placed on altered gene expression mediated by estrogen
90 receptors [25]. Some EDs can also act as anti-androgens [26, 27] producing specific alterations in
91 gene expression [28]. Considering that developmental exposure to EDs interferes with gene
92 expression in the testis, a comparison was performed here to evaluate whether specific EDs, with
93 potential different molecular mechanisms of action, induce specific signatures at a global level of
94 gene expression. In addition, we studied whether these compounds act in similar pathways of
95 estrogen signaling and render similar signatures of gene expression, as well as whether specific
96 patterns of gene modulation can be associated to developmental windows and/or doses of
97 exposure and to cytological/histological changes in the seminiferous epithelium.

98 To address these questions, we used DNA microarrays that represent the global mouse
99 transcriptome (31,769 printed 70-mer DNA probes corresponding to 24,878 expressed or
100 predicted genes) to analyze the effects on gene expression of five environmental pollutants
101 considered as EDs. Bisphenol-A (BPA) is one of the most intensely produced plasticizers
102 worldwide that can leach into food and beverages [29]. BPA has been detected in blood samples
103 and other fluids, and there is still considerable controversy regarding its potential effects [30].
104 Lindane (γ -hexachlorocyclohexane, LIN) is one of the oldest synthetic pesticides and despite
105 being considered a persistent toxicant that adversely affects reproductive functions in animals
106 [10, 31], it is still in use worldwide. Mono 2-ethylhexyl phthalate (MEHP) is the active
107 metabolite of di 2-ethylhexyl phthalate (DEHP) and is widely used as plastic flexibilizer despite

108 its reported estrogenic/antiandrogenic effect [11, 32, 33]. Zearalenone (ZEA) is a toxic substance
109 considered a phytoestrogen that is produced by *Fusarium spp.*, a contaminant of grain, and is
110 thought to cause male germ cell toxicity [34]. Finally, we assessed the effects of 17- β -estradiol
111 (E2) as a natural estrogen.

112 The effects of each of these compounds on the transcriptome were compared after
113 exposure during different developmental periods, and at different concentrations for each period.
114 The objective was to evaluate the level of global gene expression modification in the testes of
115 mice exposed to EDs and the signature of gene expression they provoked considering three
116 factors: compound specificity, the developmental window and the dosage of exposure.
117 Morphological effects on testicular development and spermatogenesis were also evaluated after
118 the various experimental conditions. The parameters studied were: body and testis weight, the
119 numbers of apoptotic cells in the testis, the percentage of tubule sections showing abnormalities
120 such as missing generations of germ cells or abnormal cell associations, and the number of
121 diploid spermatids as a measure of problems occurring at meiotic divisions.

122

123 **2. Materials and methods**

124

125 *2.1. Ethics Statement*

126 All animal care and the procedures for sacrificing the animals were in accordance with the
127 regulations laid down by the CSIC Bioethics Committee and the relevant European Commission
128 (EC) guidelines (directive 86/609/EEC). The present study was approved by the CSIC Bioethics
129 Committee (ID number: CB/CIB-PI071007-2007).

130

131 *2.2. Animal Exposure*

132 CD-1 mice were supplied by our own animal facility, the CIB-CSIC bioterium. Breeding
133 and production of the mice were carried out under specific pathogen-free (SPF), controlled
134 temperature (22 \pm 1 $^{\circ}$ C) and regulated humidity (50-55%) conditions; periods of light/dark 12h and
135 diet available *at libitum*.

136 CD-1 mice were exposed to different doses of EDs *in vivo* following a defined regimen,
137 detailed in Figure 1. In all cases, females were mated with unexposed males. The day when the

138 vaginal plug was detected was recorded as day 0. The age-range of the parental mice used in this
139 study was 2-3 months. At least, three adult females were exposed to each dose and ED during
140 developmental exposure A, B and C. The male offspring were sacrificed at four weeks of age to
141 obtain their testes for the different analysis. At least, three males, offspring from different
142 mothers, including those from exposure C which were also exposed during four weeks after birth
143 were used for RNA purification. For the histological and morphological analysis we proceeded
144 using the same protocol of exposure and number of animals. The number of animals tested for
145 histological analysis is explicitly indicated in the figures S2-S11. The control groups were
146 comprised of the same number of animals and the exposure route for the appropriate vehicle was
147 similar to that for mice that received an ED.

148 To emulate the regular intake of the environmental toxicants studied and to approximate
149 the route of administration to whole body exposure, the compounds were administered orally in
150 the drinking water at different dosages during different periods of development to reach the doses
151 indicated in Table 1. As a result, 45 different experimental conditions were compared in this
152 study. The estimated intakes were calculated on the basis of average of drinking and body weight
153 as recorded in pilot experiments and in agreement with the data in the literature referred to these
154 parameters. *In utero* and neonatal exposure was assumed to occur via placental transport and via
155 milk, respectively. Ethanol was used as the vehicle for E2, BPA and ZEA, and DMSO for MEHP
156 and LIN. Comparative control testes for microarray hybridizations were obtained from animals
157 exposed to the vehicle alone, following the same exposure regimens as for EDs, detailed in
158 Figure 1: ethanol at estimated intake of 0.060 g/kg/day and DMSO at 0.029 g/kg/day. The
159 NOAEL (No-observed-adverse-effect-level) has been established at 2.4 g/kg/day for ethanol
160 (OECD-SIDS, www.jetoc.or.jp/HP_SIDS/pdf/files/64-17-5.pdf) and at 2.5 g/kg/day for DMSO
161 (<http://www.epa.gov/oppt/chemrtk/pubs/summaries/dimthslf/c14721rr.pdf>). All control mice
162 were exposed to doses of vehicle equivalents to those for each ED.

163

164 2.3. *Histological Parameters Studied*

165 At least, 3 different mice were analyzed after every experimental condition. From each
166 animal, body weight and testis weight were registered and relative testis weight were calculated
167 (absolute testis weight proportioned to the body weight). Epithelial abnormalities of the

168 seminiferous tubules, namely vacuolisation, missing germ cell generations or abnormal germ cell
169 associations were observed in PAS-hematoxylin stained sections and the numbers of tubules
170 showing these abnormalities were counted and expressed per 100 tubule sections. In the same
171 slides, the diameters of 25 cross sections of seminiferous tubules were measured per animal, and
172 the average tubule diameter was calculated. To assess problems at the second meiotic division,
173 tubules containing diploid spermatids were counted and the percentage of the tubules containing
174 diploid spermatids was determined. The number of TUNEL positive cells was counted in all
175 testes and calculated as numbers per 100 tubule cross sections. At least 100 tubules per animal
176 were studied (all histological parameters studied are shown in Supplementary figure S12).
177 Numbers were expressed as means \pm SEM and statistical analysis was performed using the
178 Univariate Analysis of Variance.

179
180 *2.4. Apoptosis detection by way of Terminal deoxynucleotidyl transferase-mediated deoxyuridine*
181 *5-triphosphate-biotin nick end labeling (TUNEL)*

182 Five μ m Bouins fixed, paraffin embedded testis sections were incubated for 5 min in
183 10mM citric buffer (pH6.0) at 98°C and slowly cooled to room temperature (RT). Endogenous
184 peroxidase was blocked with 3% H₂O₂ in water for 5 min. Sections were washed three times with
185 PBS before a 60 min incubation in TUNEL mix at 37°C. TUNEL mix consisted of 0.3 U/mL calf
186 thymus terminal deoxynucleotidyl transferase (GE Healthcare Bio-Sciences AB, Uppsala,
187 Sweden), 6.66 mM/mL biotin dUTP (Roche Applied Science, Penzberg, Germany) in terminal
188 transferase buffer (GE Healthcare Bio-Sciences AB, Uppsala, Sweden). The TUNEL reaction
189 was stopped by incubation in 300mM NaCl, 30mM Sodium citrate in water for 15 min at RT.
190 After washing in PBS, sections were blocked with 2% BSA (Sigma-Aldrich, St. Gallen,
191 Switzerland) in PBS at RT for 10 min. Sections were treated for 30 min at 37°C in a moist
192 chamber with a 1:20 dilution of ExtrAvidin peroxidase antibody. After three washes in PBS,
193 detection was performed with DAB+ (Dako, Glostrup, Denmark). Sections were counterstained
194 with Mayer's hematoxylin, dehydrated and mounted with Pertex (Cellpath, Newtown, UK). No
195 statistically significant differences in the evaluated parameters were seen between samples from
196 animals exposed only to the vehicle of the compounds used as controls versus unexposed, i.e.,
197 EtOH or DMSO versus H₂O.

198

199 *2.5. RNA preparation*

200 Total RNA from the testes was extracted using TRIzol (Invitrogen, part of Life
201 Technologies, Carlsbad, CA, USA) according to the manufacturer's instructions and the RNA
202 was then purified with MegaClear (Ambion, brand of Life Technologies, Carlsbad, CA, USA).
203 To minimize the inter-individual variability and considering the broad number of variables (45)
204 in the global analysis, we used pools prepared from RNA samples from at least three individual
205 testes (5 µg total RNAs each) exposed to every experimental condition: compound, dose and
206 developmental stage and its respective controls. RNA quality was checked using the Experion
207 automated electrophoresis system (Bio-Rad Laboratories, Hercules, CA, USA). Samples with an
208 rRNA ratio (28S/18S) higher than 1.4 were used for the experiments. Total RNA from samples
209 exposed to EDs and vehicles (DMSO or ETOH) (controls) (1.5 µg for each pool) was amplified
210 using the Amino Allyl MessageAmp aRNA kit (Ambion, brand of Life Technologies, Carlsbad,
211 CA, USA) and we obtained 15-60 µg of amino-allyl amplified RNA (aRNA). The mean aRNA
212 size was approximately 1500 nucleotides when measured in the Experion analyzer. For each
213 sample, 2.5 µg of aRNA was labeled with one aliquot of Cy3 or Cy5 mono NHS ester (CyDye
214 Post-labeling Reactive Dye, GE Healthcare Bio-Sciences AB, Uppsala, Sweden) and it was
215 purified with the Amino Allyl MessageAmp aRNA kit. Cy3 and Cy5 incorporation was measured
216 using 1 µl of probe in a Nanodrop spectrophotometer (Nanodrop Technologies, brand of Thermo
217 Fisher Scientific, Waltham, MA, USA). For hybridization, 110-150 pmol of the Cy3 and Cy5
218 probes were mixed, dried in a speed-vacuum and resuspended in 9 µl RNase-free water. Labeled
219 aRNA was fragmented by adding 1 µl of 10x fragmentation buffer (Ambion, brand of Life
220 Technologies, Carlsbad, CA, USA) and incubating for 15 min at 70°C. The reaction was
221 terminated by adding 1 µl of the stop solution (Ambion, brand of Life Technologies, Carlsbad
222 CA, USA).

223 In order to minimize potential biological variations between individuals as effect of the
224 compounds, a pool of at least three individual RNA samples was used in each experimental
225 condition. Sample pooling, by mixing mRNAs from several biological-replicate samples, gives a
226 mixture of individual responses and the results have been statistically demonstrated to be valid
227 [35], [36, 37].

228

229 *2.6. Microarray processing*

230 Mouse Oligoset v3 (OPERON) arrays were used from the Genomics facility at the
231 University of Cincinnati (USA). The slides contained 31,769 spotted probes (70 mer
232 oligonucleotides) corresponding to 24,878 expressed or predicted mouse genes. Information on
233 the printing and the oligo set can be found at <http://microarray.uc.edu>. The slides were
234 prehybridized (42°C, 45-60 min) in 6x SSC, 0.5% SDS and 1% BSA, and they were then rinsed
235 10 times with distilled water. Fragmented Cy5 and Cy3 aRNA probes were mixed (110-150 pmol
236 of each label) with 10 µg PolyA (Sigma-Aldrich, St. Gallen, Switzerland) and 5 µg Human Cot-
237 DNA (Invitrogen, part of Life Technologies, Carlsbad, CA, USA), and then dried in a speed-
238 vacuum. Each probe mix was resuspended in 60 µl of hybridization buffer (50% formamide, 6x
239 SSC, 0.5% SDS, 5x Denhardt's solution) and the probes were denatured at 95°C for 5 min before
240 they were applied to the slide using a LifterSlip (Erie Scientific Company, Portsmouth, NH,
241 USA). The slides were incubated at 48°C for 16 h in hybridization chambers (Array-It
242 Corporation, Sunnyvale, CA, USA); in a water bath. After incubation, the slides were washed
243 twice with 0.5x SSC, 0.1% SDS for 5 min each, three times with 0.5x SSC for 5 min, and finally
244 in 0.05x SSC for 5 min, before they were dried by centrifugation (1600 r.p.m., 1 min). Images
245 from the Cy3 and Cy5 channels were equilibrated and captured with an Axon 4000B scanner, and
246 the spots were quantified using GenePix 5.1 software. Dye swapping replicates were performed
247 for each particular hybridization, as described below (Data Analysis section).

248

249 *2.7. Data Analysis*

250 The raw data were processed using “AlmaZen” (Alma Bioinformatica, Tres Cantos, Spain)
251 software. The background was subtracted from each individual hybridization and they were
252 Lowess normalized [38]. For each condition, a technical replicate was performed by dye-
253 swapping (two hybridizations) in order to minimize a possible bias in the labeling process. The
254 mean of log₂ ratio (exposed/control) from the merge of the two dye-swapping hybridizations
255 were calculated and p-values, standard deviations, and z-scores were obtained [39-41]. Raw and
256 processed data were included in the Gene Expression Omnibus database (GEO) [42] (accession
257 number GSE14774) following the Minimal Information About a Microarray Experiment

258 (MIAME) criteria [43].

259

260 2.8. Cluster analysis

261 To analyze transcriptome profiles in different experimental conditions, hierarchical
262 clustering of array data has been used widely [44, 45]. Data were preprocessed previously to
263 perform Hierarchical Clustering. Preprocessing included following criteria, using the “Preprocess
264 DNA array data files” web tool (GEPAS v3.0) [46] [47]: merging values of gene replicates (by
265 averaging), filtering missing values (70 % was the minimum percentage of missing values
266 accepted), imputing missing values (using KNN impute) and filtering flat patterns by the
267 numbers of peaks (a minimum of 3 peaks and a threshold of +0.7, in log scale, were considered).
268 The distance matrix was calculated for both, genes and experimental conditions from the
269 preprocessed datasets. The similarity metric used was a centered Pearson correlation. In addition,
270 both genes and conditions were clustered by the Average Linkage method using Cluster 3.0
271 software [48], and the clustering results generated by Cluster 3.0 were visualized using Java
272 Treeview [49].

273 We used this unsupervised hierarchical clustering analysis to organize the different
274 conditions of exposure according to the similarity or dissimilarity of the gene expression profiles,
275 situating the defined conditions of exposure with similar profiles together as neighboring
276 columns in the clustergram. In addition, we analyzed the potential participation of gene
277 signatures in molecular pathways as described in the results of the Ingenuity Pathway Analysis
278 (<http://www.ingenuity.com>).

279

280 2.9. Confirmation of microarray expression data by TaqMan® Low Density Arrays (TLDA)

281 The expression of 23 genes randomly selected from the microarray analyses (Table S1 in
282 supplemental material) was measured in custom TaqMan® Low Density Arrays (TLDA)
283 (Applied Biosystems, part of Life Technologies, Carlsbad, CA, USA), following the
284 manufacturer’s instructions. Expression was measured in the same RNA samples studied in the
285 microarray experiments for each experimental condition.

286 From the entire data set corresponding to the expression of the 23 selected genes (Table
287 S1) obtained from all analyzed microarrays, only 130 expression values filtered for statistically

288 significant p-value (>0.05) were used for qualifying purposes. These data were compared with
289 the corresponding data obtained by qRT-PCR. Statistical analyses of Pearson correlation and p-
290 value were performed using GraphPad Prism version 5.00 for Windows (GraphPad Software, San
291 Diego, CA, USA; www.graphpad.com).

292

293 **3. Results**

294

295 The first aim was to evaluate the global change in levels of gene expression as effect of
296 exposure to EDs in the different experimental conditions. From the microarray data, we used the
297 average of the log-ratio approach adopted in the MicroArray Quality Control (MAQC) project to
298 compare the toxicogenomic effects of different compounds [50]. To assess inter-individual gene
299 expression level as a baseline in the testis from the CD-1 mice used, we performed preliminary
300 analyses comparing RNA from three mice (S#1 vs S#2, S#1 vs S#3 and S#2 vs S#3) and from
301 pooled testes (S#Pool vs S#Pool) obtained from at least three four-weeks old mice, neither
302 exposed to compounds nor solvents. Fig. 2 shows the level of the global gene expression
303 variation found in this comparative analysis in terms of the average of log-ratio from genes with
304 an average signal higher than 64 in MA plots. Only raw and processed data from dye swapping
305 replicates coming from pooled samples were included in the Gene Expression Omnibus database
306 (GEO, accession number GSE14774), named as “Testis_Control”.

307 After the analysis of the signal distribution in MA plots (data not shown), we only
308 considered spots with an average signal >64 in order to avoid false background corresponding to
309 genes weakly or negligibly expressed in the testis. We found that MEHP induced the strongest
310 degree of global modification of gene expression of all the compounds analyzed (Fig. 2). The
311 maximum effects were detected in those mice that were exposed throughout the entire period of
312 development until they were analyzed (Fig. 2, exposure C). However, the results obtained from
313 mice whose mothers were only exposed during the preovulatory pre-mating period indicated that
314 some of the EDs analyzed, such as ZEA and MEHP, still induced gene expression modifications
315 in the testes of the offspring, suggesting a maternal developmental reprogramming effect (Fig. 2,
316 exposure A). Interestingly, this effect was more pronounced at lower rather than at higher
317 concentrations. This “low dosage” effect was also detected following prolonged exposure to all

318 the EDs analyzed except for E2 (Fig. 2 exposure C). Furthermore, E2 had a weaker effect, at the
319 dosages used, than most of the other ED compounds assessed, except at the maximum exposure.
320 We extended this first global estimation to assess the level of global gene expression changes
321 determining the total number of genes differentially expressed in the different experimental
322 conditions under common filtering criteria. To increase the stringency of the analysis [50], we
323 filtered the data to eliminate the results corresponding to genes that displayed weak expression
324 and large variability. Hence, in this case, we only considered and analyzed the data corresponding
325 to genes that underwent a change $>+2$ - or <-2 -fold, with a p-value <0.01 and an average signal
326 >64 . These results indicated a similar pattern of global gene expression changes as effect of the
327 different exposures as when the average of the signals was the criterion for analysis (Fig. 3).

328 To answer the question whether the profiles of expression, at the level of specific genes,
329 were associated with specific EDs, the window of exposure during development and/or level of
330 exposure, we carried out an analysis of the entire transcriptome data by unsupervised hierarchical
331 clustering from dataset of all 45 different experimental conditions. The relationship between the
332 exposure and the gene expression profile is depicted graphically as a dendrogram (Fig. 4). The
333 filtering and pre-processing of the datasets indicated in the Materials and Methods defined
334 specific gene expression signatures (involving 2,676 genes) for conditions corresponding to
335 MEHP and ZEA exposures irrespective of the developmental stage or dose used (Fig. 4). Thus,
336 for MEHP and ZEA the toxicogenomic effect in testis appears to be mediated by the nature of the
337 compound rather than by the levels or the window of exposure during development. This also
338 indicates that low dose exposure during the periconceptional period could result in deregulation
339 of key genes in cellular functions of testicular cells in adulthood.

340 ZEA administration did not have a clear effect on morphological parameters other than a
341 significant increase in the numbers of apoptotic cells after the middle dose group A, and the
342 highest dose in exposure group B (Fig. S10 in supplemental material). However, MEHP did not
343 cause any consistent morphological effect at any dose or exposure period (Fig. S11 in
344 supplemental material).

345 By contrast, the effect of E2 on global gene expression was not consistent across different
346 samples and did not segregate these samples hierarchically from those exposed to BPA and LIN
347 exposures. These data suggested that developmental exposure to MEHP or ZEA modulated gene

348 expression in a manner distinct to E2 and the other compounds assessed. Hence, the effects of
349 these compounds on testicular development would appear to be mediated by mechanisms other
350 than those affected by E2. In this hierarchical clustering it was possible to identify groups of
351 genes preferentially expressed in specific cell types that might share putative functional
352 properties. Following previous gene classifications in the testis [51], different "signatures" could
353 be detected such as those expressed in round spermatids (*Rtds I-V* in grey bars, Fig. 4), as well as
354 the groups of genes preferentially expressed in spermatid elongation (Fig. 4, EI-III in orange
355 bars: Table S2 in supplemental material).

356 Considering the results for E2, BPA and LIN, we studied whether we could define,
357 through gene expression profiles, a pattern of response for these three EDs that could clarify
358 similarities in their mechanism of action in testis development. For this reason, only data of the
359 three compounds were reclustered. The profile of the dendrogram was more tightly defined for the
360 exposure to lindane up until birth (LB in the dendrogram Fig. 5) or throughout the entire
361 developmental period (LC: Fig. 5). It is noteworthy that there was a close linkage between the
362 lowest level of exposure to LIN (LC+) and the highest level of E2 (EC+++), after prolonged
363 exposure during development. However, the gene expression profiles for BPA and E2 were more
364 interspersed, suggesting that BPA and E2 could act through similar mechanisms that differ from
365 those used by the other compounds analyzed.

366 The exposure to LIN had the strongest cytopathological effects on the seminiferous
367 epithelium. Germ cell apoptosis strongly increased in all dose groups after exposure schedules B
368 and C. In all these groups of B and C exposures the differences were statistically significant and
369 moreover all three dose groups in exposure C showed decreased tubule diameters. In addition,
370 significant changes in the frequency of epithelial abnormalities were seen for the two highest
371 doses in group C. The percentages of tubules with diploid spermatids tended to increase parallel
372 to the height of applied dose in exposure C, but the difference was only significant at the highest
373 dose (Fig. S12 in supplemental material).

374 Generally, exposure to the various compounds did not cause clear effects on body and
375 relative testis weights. However, at the longest exposure period (group C) E2 did cause a
376 decrease of about 30 % in body weight at all dose levels and an about 15% decrease in relative
377 testis weight, was seen after exposure to LIN (Fig. S3-S7 in supplemental material).

378 Exposure to E2 caused some particular effects on spermatogenic cells. Significant
379 increases in numbers of apoptotic germ cells were detected in mice exposed according to
380 experimental conditions B and C and receiving the highest doses of E2. Furthermore, clear
381 decreases in tubule diameter and increased numbers of tubules with epithelial abnormalities were
382 found with the high doses of E2 following exposure C. Remarkably, following exposure C
383 diploid spermatids increased after all doses of E2 (Fig. S8 in supplemental material).

384 The effects of BPA administration were less than after E2 exposure. Only increases in the
385 numbers of apoptotic germ cells were seen and only after exposure C and highest doses of BPA.
386 No significant effects on the other parameters were seen after BPA administration (Fig. S9 in
387 supplemental material).

388 The data from the arrays were qualified by qRT-PCR (TaqMan probes). On the basis of
389 the expression data in all experimental conditions from 23 genes randomly selected from those
390 annotated in the clustering analysis, we considered eligible those data that showed statistical
391 significance ($p < 0.05$) in the microarray. A total of 130 expression values, fitting the mentioned
392 statistical stringency, were compared with the data of expression by qRT-PCR. We found an
393 expected correlation in this type of replication analysis between gene expression from the
394 microarray analyses and the qRT-PCR results [52] ($r = 0.535$ $p < 0.0001$; Fig. S2 in supplemental
395 material).

396 By analyzing the genes that define the clusters using Ingenuity Pathway Analysis
397 software (IPA), the highest scoring gene networks could be identified, those that were relevant to
398 disorders related to the pathologies supposedly induced by EDs. Among the most relevant are
399 diseases of the reproductive system: cancer, developmental and endocrine system disorders
400 (Table 2). The canonical pathways with the highest values identified were the oxidative stress
401 response mediated by *Nrf2*, protein ubiquitination, oxidative phosphorylation and mitochondrial
402 dysfunction (Table 3). These pathways can also be considered as relevant to the mechanisms that
403 potentially cause cell and developmental disorders related to the proposed effects of EDs on
404 mammalian testis.

405 The analysis of the networks of interacting genes was carried out only considering the
406 genes defined in the hierarchical clustering, comparing all EDs and conditions. This means that
407 the genes detected in each network are deregulated in at least three different conditions

408 (considering the criteria of 3 peaks of the clustering), and consequently, this does not mean that
409 the changes in expression affect them equally with all EDs or in all conditions tested. However,
410 these genes could be indicators of the pathways that may be more affected by individual ED or
411 mixtures of EDs. The most representative network described a core of interacting genes that
412 encode proteins involved in the global regulation of translation (Fig. 6 and Table S3 in
413 supplemental material), as represented by the eukaryotic translation initiation factors (EIFS) and
414 the cytoplasmic poly(A) binding protein (PABPC), critical post-transcriptional regulators. In
415 addition members of the CCR4-NOT complex, such as *Cnot1*, *Cnot7* and *Cnot6*, which interacts
416 with PabpC1, were also detected in this most prominently affected network. Through its
417 interaction with this network, the gene encoding the breast cancer anti-estrogen resistance 3
418 protein (*Bcar3*) was also seen to be modified expression by all the ED's analyzed.

419

420 4. Discussion

421

422 In this study we have compared global gene expression in testes of mice exposed to five
423 endocrine disruptors, using three different doses and studying three distinct periods of
424 developmental exposure. The main findings point out that Mono-(2-ethylhexyl)-phthalate and
425 Zearalenone, more than 17β -Estradiol, Lindane or Bisphenol-A, produce specific gene signatures
426 during testis development, irrespective of the concentration of the toxicant or the developmental
427 period during which exposure takes place. These data suggest that prevalent alterations to defined
428 networks can induce the disequilibrium in gene expression programs involved in correct
429 testicular development and spermatogenesis, that can potentially be transmitted to male offspring
430 in an epigenetic-like manner. However, histopathological study only showed moderate
431 morphological abnormalities.

432 In planning the dose range of each compound, we did not consider doses which cause
433 acute toxicity but selected doses above the NOAEL and as the exposures were initiated in the
434 mothers, we used doses that did not induce fetal loss. For some compounds such as E2 or BPA
435 the doses used were below those that were considered to affect the number of births. In pilot
436 experiments, for each compound we evaluated the rates of birth at different doses and
437 developmental periods of exposure. In this way, a dietary concentration of intake of 100

438 $\mu\text{g}/\text{kg}/\text{day}$ of E2 in CD1 mice was found to highly affect the number of successful pregnancies
439 [53]. We confirmed this result and established the maximum dose at $48 \mu\text{g}/\text{kg}/\text{day}$. Similarly,
440 exposure of laboratory animals to BPA was reported to influence fertility [54] [7, 55]. To this
441 respect, considering our scheme of experimental exposure during development, higher doses of
442 compounds as E2 and BPA that potentially can induce severe effects on testis, were not be
443 applied because the pregnancy of exposed mice was also clearly affected. This could explain the
444 relative small effect detected in our study for E2 and BPA with respect to the level of gene
445 deregulation in comparison to the other EDs evaluated.

446 The tolerable daily intake (TDI) for lindane was originally estimated at $12.5 \mu\text{g}/\text{kg}$
447 bw/day [56]. However, lindane is a persistent contaminant and bioaccumulates. Traina et al. [57]
448 using a similar dosage and method of administration as we used, reported long-lasting effects of
449 lindane on mouse spermatogenesis without maternal toxicity or delayed growth and development
450 of pups.

451 With respect to MEHP, the ED with ZEA showing the most conspicuous effect of the
452 compounds studied, a wide spectrum of human exposures to phthalates has been recently
453 reviewed with estimations of intakes of DEHP varying between $7.3 \mu\text{g}$ to $409 \mu\text{g}/\text{kg bw}/\text{day}$ [58].
454 However, in higher exposed groups at risk such as premature neonates in a neonatal intensive-
455 care unit the estimated level can reach up to $6 \text{ mg}/\text{kg bw}/\text{day}$ [59]. Maximal DEHP exposures up
456 to $22 \text{ mg}/\text{kg bw}/\text{day}$ have also been estimated for newborns infants following blood transfusion
457 procedures [60]. These doses are equivalent to $45 \text{ mg}/\text{kg bw}/\text{day}$ and $165 \text{ mg}/\text{kg bw}/\text{day}$
458 respectively in the mouse, which is in the range of the median dose ($46.3 \text{ mg}/\text{kg bw}/\text{day}$) or over
459 maximum dose ($92.7 \text{ mg}/\text{kg bw}/\text{day}$) used in the present study. This indicates that the defined
460 pattern of altered gene expression detected in this study may be representative of the action of
461 phthalates in exposed humans. Relevant expression patterns, with hundreds of deregulated genes,
462 might not be associated to dramatic changes at the morphological or histological level in the
463 testis, as it has been shown in this work. Previous studies reported that MEHP exposures cause an
464 increase in apoptosis of germ cells in adult mice mediated by Sertoli cell injuries at 10 fold [50]
465 or 20 fold the doses used in our study [51].

466 Worldwide, zearalenone is found in a number of cereal crops and derivatives. The
467 concentrations of zearalenone in cereals vary over a wide range between 3 to $8000 \mu\text{g}/\text{kg}$

468 (International Programme on Chemical Safety. WHO food additives series: 44, Geneva 2000
469 <http://www.inchem.org/documents/jecfa/jecmono/v44jec14.htm>). However, high concentrations
470 have been reported. For example: as zearalenone may be transmitted from contaminated grains
471 into beer a very high concentration of zearalenone was found in beers brewed in Africa at levels
472 of 53 mg/L [61]. Assuming an intake of 500 ml of beer per day, this is equivalent to 0.44 mg/kg
473 bw/day. Applying the metabolic factor, this can be calculated to be 3.3 mg/kg bw/day for a
474 mouse which is in the range of the medium dosage used in our study. Similar to MEHP, the
475 specific profile of gene deregulation in testes from mice exposed to ZEA are not directly
476 correlated with morphological alterations neither modifications of cellular composition in the
477 seminiferous epithelium. However, the exposure to E2, BPA and LIN increased the numbers of
478 apoptotic cells after the longest period of exposure but also when the treatment was stopped at
479 birth, indicating long-term effects of these compounds. Epithelial abnormalities and
480 modifications of tubule diameter showed increases at the highest dose of E2 and BPA or LIN
481 after long-term exposure.

482 Studies of the environmental induction of changes in gene regulation as disease etiology
483 have focused on defined genes or genetic pathways related to particular diseases, including
484 reproductive dysfunctions [24]. Estrogen has a crucial role during spermatogenesis [62] and
485 hence, estrogen-like compounds can interfere with the activity of estrogen and its binding to
486 estrogen receptors that mediates the transcription of target genes. However, the mechanisms by
487 which estrogen regulates gene expression are known to be more complex than originally thought
488 [63, 64]. The extensive alterations in gene expression observed in the present work can be
489 explained by a cascade of effects triggered by environmental pollutants with the capacity to
490 interfere with hormonal pathways (acting as EDs) and to lead to gene deregulation.

491 Our results of global gene expression and the hierarchical clustering analysis suggest that
492 some compounds considered as EDs could act via different mechanisms to estrogen. While
493 MEHP [65] and ZEA [66] are considered to have estrogenic activity *in vitro*, they displayed
494 distinct effects on gene expression that could not be considered to mimic those of estrogen,
495 suggesting another mode of action. Indeed, it was recently reported that MEHP alters both
496 steroidogenesis and germ cell number in mice, without involving either the estrogen or androgen
497 receptor [33].

498 In the hierarchical clustering, different signatures included the upregulation or
499 downregulation of the same genes depending on the compound, dose or developmental period of
500 exposure. Hence, it would appear that the effect of EDs cannot solely be attributed the reduced
501 expression of genes specific to spermatids that might produce a lack or decrease in the number of
502 these cell types (round or elongated), as suggested previously [67]. Indeed, alterations of the
503 cytological structure of the seminiferous epithelium that could compromise spermatogenesis
504 were not found in our cytological/histological analysis. Moreover, the results indicate that acute
505 effects of the toxicants may not explain the altered gene expression since low doses and pre-
506 mating exposure of mothers induced the modulations of expression of genes in a clustered
507 manner.

508 In the present study, when the mothers were only exposed to EDs during the pre-mating
509 period (exposure A), the level and type of gene modifications of gene expression observed in the
510 offspring can only be attributed to epigenetic modifications in the oocyte/zygote, or during early
511 embryonic development prior to gonadal sex differentiation if traces of the compound remained
512 in the maternal metabolism. No epigenetic transgenerational effect of ZEA, MEHP or LIN has
513 been reported previously. In rats exposed to vinclozolin, a fungicide considered to be an ED,
514 effects on the male reproductive system were evident until the F4 generation [68] and they were
515 attributed to epigenetic effects. Although, the transgenerational effects of orally administered
516 vinclozolin were recently questioned [69], alterations to the transcriptome have been
517 demonstrated in rat embryonic testis in the F1 to F3 generations when the F0 generation was
518 exposed to vinclozolin from day 8 to 14 of gestation [70]. This exposure coincides with a period
519 when epigenetic reprogramming occurs in the transition of primordial germ cells to sex-
520 differentiated germ cells [71]. The transgenerational effect of some of the EDs that we have
521 analyzed should also be evaluated further at the promoter level on the basis of specific potential
522 epigenetic alterations.

523 By analyzing gene expression signatures in relation to their biological pathways,
524 oxidative stress response mediated by NF-E2-related factor 2 (*Nrf2*) was detected as a
525 particularly relevant pathway that could be altered as response to the different EDs. *Nrf2* is a
526 transcription factor that acts through the antioxidant response element [72]. The expression of
527 mitochondrial and nuclear-encoded subunits of respiratory chain complexes must be closely

528 coordinated, and *Nrf1* and *Nrf2* are the main genes responsible for this coordination. Testicular
529 oxidative stress appears to be a common feature of male infertility and indeed, exposure to
530 toxicants has been correlated with an increase in oxidative stress in the testis [73]. These data
531 suggest that in the testis, EDs modify the expression of genes involved in pathways common to
532 other xenobiotics described in the literature that clearly affect reproduction, even when the
533 compound is administered to the mother prior to fertilization.

534 In the most relevant networks of genes involved in post-transcriptional regulation,
535 regulators such as EIF and poly-A binding protein occupy a central position. Five genes
536 encoding different isoforms of EIF4 (*Eif4A1*, *Eif4A2*, *Eif4E*, *Eif4G1*, and *Eif4G2*) and three
537 encoding EIF3 isoforms (*Eif3B*, *Eif3E*, *Eif3H*), together with *Pabpc*, participate in the core of the
538 43S pre-initiation ribosomal complex, binding mRNAs prior to translation [74]. These nine
539 genes showed different altered expression under different experimental conditions. The EIF4F
540 complex is composed of EIF4E, EIF4G and EIF4A, and it associates with the 5' cap structure of
541 the mRNA as part of the 43S preribosomal complex. This EIF4F complex is largely dependent
542 on the availability of EIF4E, which thereby limits the rate at which translation is initiated [75].
543 *Eif4E* overexpression has long been associated with oncogenesis [76] and there is increasing
544 evidence correlating it with cellular transformation, tumorigenesis and metastatic progression in
545 human cancers [77]. However, experimental downregulation of *Eif4E* induces apoptosis [75].
546 The modification of *Eif* and *Pabpc* expression after exposure to EDs may involve
547 downregulation or upregulation depending on the compound used and the period of
548 developmental exposure. Although, the general tendency was a diminution of their expression,
549 compared to the controls, as seen for *Eif4E*, *Eif3E*, *Eif4A2*, after MEHP exposure we detected an
550 increase in the expression of *Eif4A1*.

551 Post-transcriptional regulation of most mRNAs is mediated by the length of their 3'
552 poly[A] tails. PABP acts by promoting mRNA translation and several deadenylases serve as
553 antagonists, facilitating the degradation of the mRNA [78]. PABPC1 interacts with subunits of
554 the CCR4-NOT complex. Mammalian CCR4-NOT represents a complex of several subunits that
555 acts as a deadenylase for mRNAs [79]. A recent study indicated that CCR4-NOT participates in
556 the regulation of certain endogenous retinoic acid receptors that are essential for normal
557 spermatogenesis [80]. The association and combination of different subunits in this complex

558 modulate different specific cell functions [81]. Three components of this complex in the
559 network: *Cnot1*, *Cnot6* and *Cnot7* were deregulated in different ways. CNOT1 represses the
560 ligand-dependent transcriptional activity of the estrogen receptor (ER) α receptor [82]. Indeed, in
561 all conditions both E2 and MEHP exposure diminishes the amount of *Cnot1* mRNA compared to
562 the controls. CNOT7 acts as a co-regulator of retinoid X receptor beta (Rxb) and significantly,
563 male null mutant mice (*Cnot7*^{-/-}) are sterile due to oligo-astheno-teratozoospermia resulting from
564 Sertoli cell defects [83]. Moreover, the expression of *Cnot7* was downregulated following
565 exposure to all compounds and conditions, except ZEA.

566 The data obtained suggest that global translation of cap-dependent mRNAs, mediated by
567 the initiation complexes, may be deregulated through the disequilibrium in the availability of the
568 components of these complexes. Accordingly, phenotypic alterations may be promoted that
569 range from apoptosis to cellular transformation and cancer, depending on the cell type and
570 developmental stage. In addition, modification of expression of genes encoding proteins that
571 interact with hormone receptors, such as *Cnot*, can have a variety of pathological consequences
572 during testicular development, already proposed to be effects of ED exposure. We speculate that
573 the disequilibrium in different elements of these complexes and associated genes/proteins
574 provoke different pathologies during testes development that can be attributed to different
575 nosological entities.

576 It is interesting to note that in the network the *Bcar3* gene is connected to *Pabpc1* and its
577 level of expression was also altered. The genes of the *Bcar* family contribute to cell proliferation
578 in estrogen-independent breast cancer and hence, these cells are resistant to anti-estrogen
579 endocrine therapy [84]. We speculate that this gene may also participate in some alterations of
580 the endocrine response in germ cells.

581 Besides emphasizing the comparative results of the effect of different compounds, doses
582 and exposure periods, the data points of more than 3 million quantitative values of expression
583 levels of genes expressed in testis that we have included in GEO (Accession number GSE14774)
584 may be of significant value in further comparative studies to determine potential mechanisms of
585 action and biomarkers of the effects of endocrine disrupters.

586

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593

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809
810

811 **FIGURE LEGENDS:**

812
813 **Figure 1. Schedule of ED exposures and their controls.** Three exposure regimes were
814 employed: mothers were exposed for two weeks before mating only (exposure A) or the exposure
815 was maintained during embryonic and fetal development, ceasing at birth (exposure B) or at the
816 end of the prepubertal stage (exposure C). Three ED concentrations were used for each
817 developmental period of exposure. All male offspring were orchidectomized at four weeks of age
818 to obtain the total RNA. The solvents of each ED (DMSO or ethanol) were used as controls.

819
820 **Figure 2. Assessment of global gene expression changes.** Graph representing the calculated
821 average of the \log_2 (R/G) obtained from the microarray expression data. Basal Gene Expression
822 Average (BGEA) from testis was obtained by hybridization of individual and pooled RNA
823 samples from testes neither exposed to compounds nor solvents (S#1 vs S#2 = individual sample
824 1 versus individual sample 2; S#1 vs S#3 = individual sample 1 versus individual sample 3; S#2
825 vs S#3 = individual sample 2 versus individual sample 3; S#pool =pooled sample versus itself).
826 Only the expression data from genes with average signals higher than 64 were considered. Grey
827 triangles indicate the increasing doses of the compounds.

828
829 **Figure 3. Total number of genes with altered gene expression.** The total number of genes
830 induced and repressed (fold change $>+2$ and <-2 , p-value <0.01 and an average signal >64) in all
831 the conditions analyzed.

832
833 **Figure 4. Hierarchical clustering.** Dendogram of the unsupervised hierarchical clustering
834 performed with filtered and pre-processed (3 peaks and a threshold of 0.7) expression data from
835 each experimental condition (see methods). After processing the datasets, 2,676 genes were
836 considered for clustering. Each column represents a specific exposure regime. The codes are
837 made up of the first letter of the compound (L-LIN; E-E2; Z-ZEA; M-MEHP and B-BPA), a
838 second letter representing the developmental period of exposure (pre-mating = A; embryonic and
839 fetal = B; postnatal = C), followed by the dosage in brackets (see Fig. 1). Each experimental
840 compound is depicted with a specific color to readily visualize the clustered conditions (blue for

841 LIN, magenta for E2, yellow for ZEA, light green for MEHP, brown for BPA). Orange and grey
842 bars denote gene clusters that include the genes preferentially expressed during spermiogenesis or
843 in round spermatids, respectively. The color scale of expression (\log_2 ratio) is shown.

844
845 **Figure 5. Clustering of LIN, BPA and E2 conditions.** Unsupervised hierarchical clustering
846 performed with filtered and pre-processed expression data (3 peaks and a threshold of 0.7) after
847 exposure to LIN, BPA and E2 (all exposures excluding MEHP and ZEA). After processing the
848 datasets, 666 genes were considered for clustering. The color scale of expression (\log_2 ratio) is
849 shown.

850
851 **Figure 6. Identified gene networks.** The top ranked networks identified by Ingenuity Pathway
852 Analysis (IPA) software from the genes selected for hierarchical clustering considering all the
853 experimental regimes of exposure. Shaded genes represent those that were included in the
854 geneset analyzed by IPA.

855

856 **Table 1.** Compound dosages administrated to mice in the drinking water (estimated intake in
857 mg/kg/day).

858

Escalation	Estradiol (E2)	Bisphenol-A (BPA)	Zearalenone (ZEA)	Mono (2- EthylHexyl) Phthalate (MEHP)	Lindane (LIN)
Dose (+)	0.006	0.16	1.3	9.2	16.6
Dose (++)	0.012	16	3.9	46.3	32.2
Dose (+++)	0.048	64	6.6	92.7	64.4

859

860

861

862

863 **Table 2.** Top associated network functions identified by Ingenuity Pathway Analysis
864 considering genes selected by hierarchical clustering of all experimental conditions.

Associated network functions	Score
Cancer, cell cycle, reproductive system disease	43
Cellular functions and maintenance, developmental disorders , genetic disorders	43
Molecular transport, protein trafficking, endocrine system development and function	41
Drug metabolism, small molecule biochemistry, cell cycle	38
RNA post-transcriptional modification, protein synthesis, gene expression	36

865

866

867

868

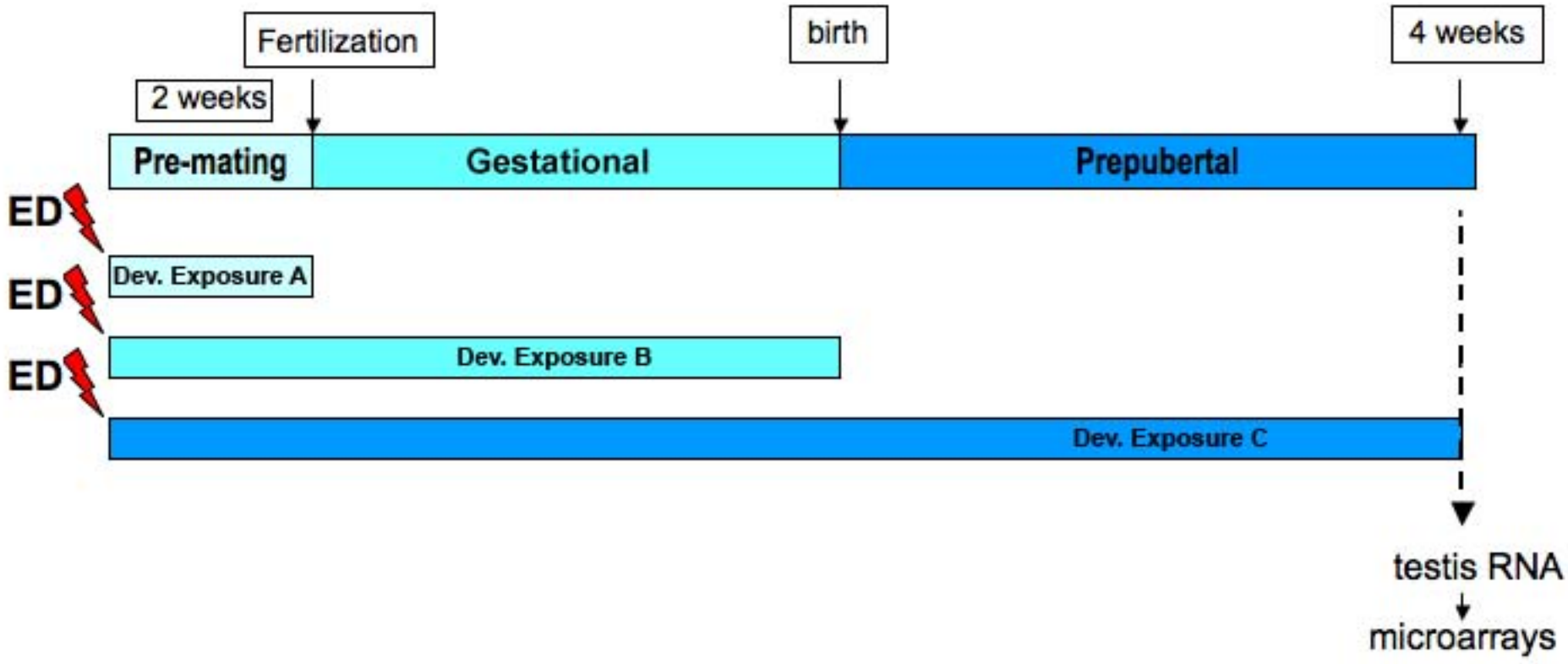
869

870

871 **Table 3.** Top canonical pathways identified by Ingenuity Pathway Analysis considering genes
872 selected by hierarchical clustering of all experimental conditions.

Pathways	p-value
Nrf2-mediated oxidative stress response	0.0000000165
Protein ubiquitination pathway	0.00000011
Oxidative phosphorylation	0.000000464
Mitochondrial dysfunction	0.00000118
Ubiquinone biosynthesis	0.000142

Figure 1



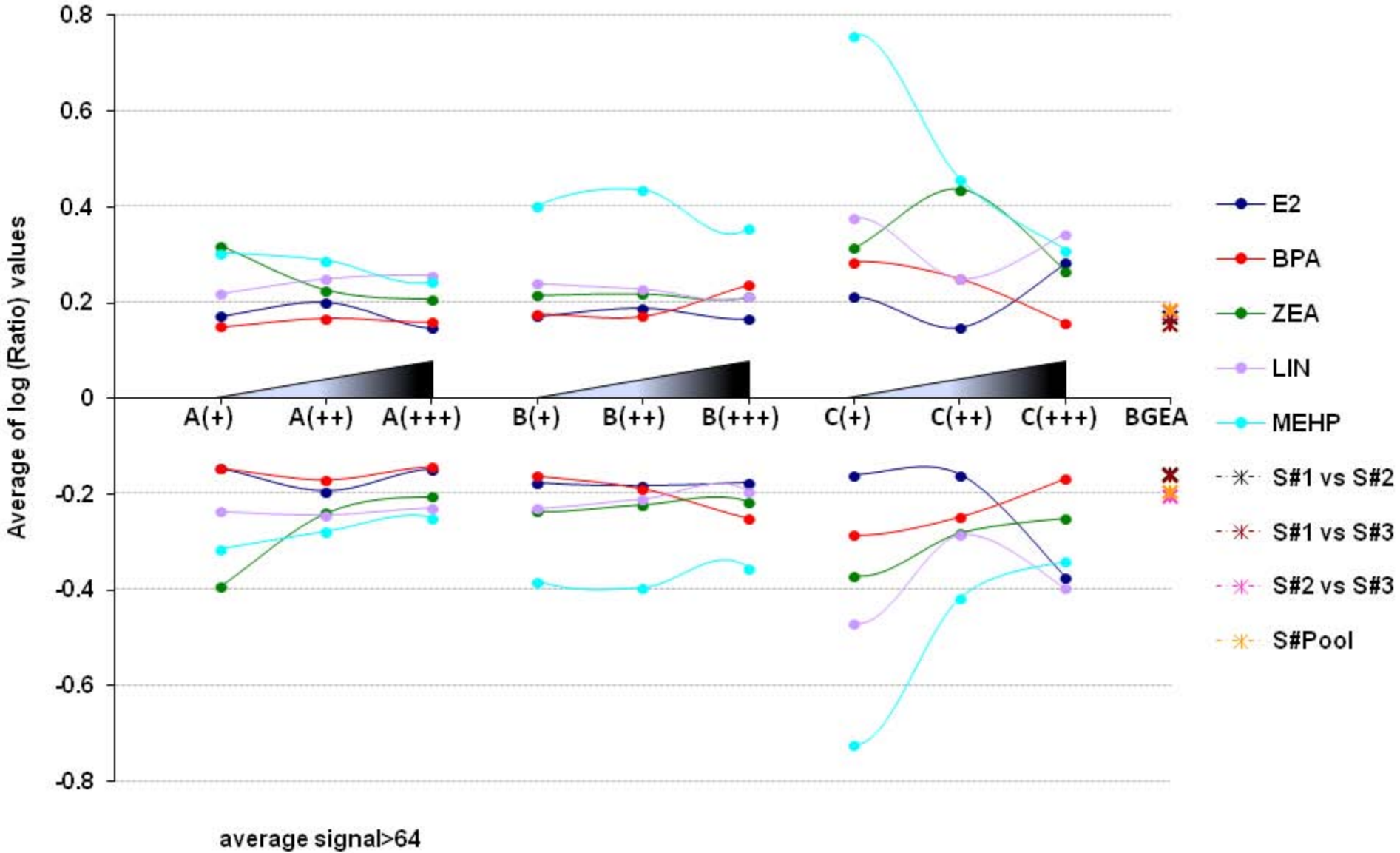


Figure 3.

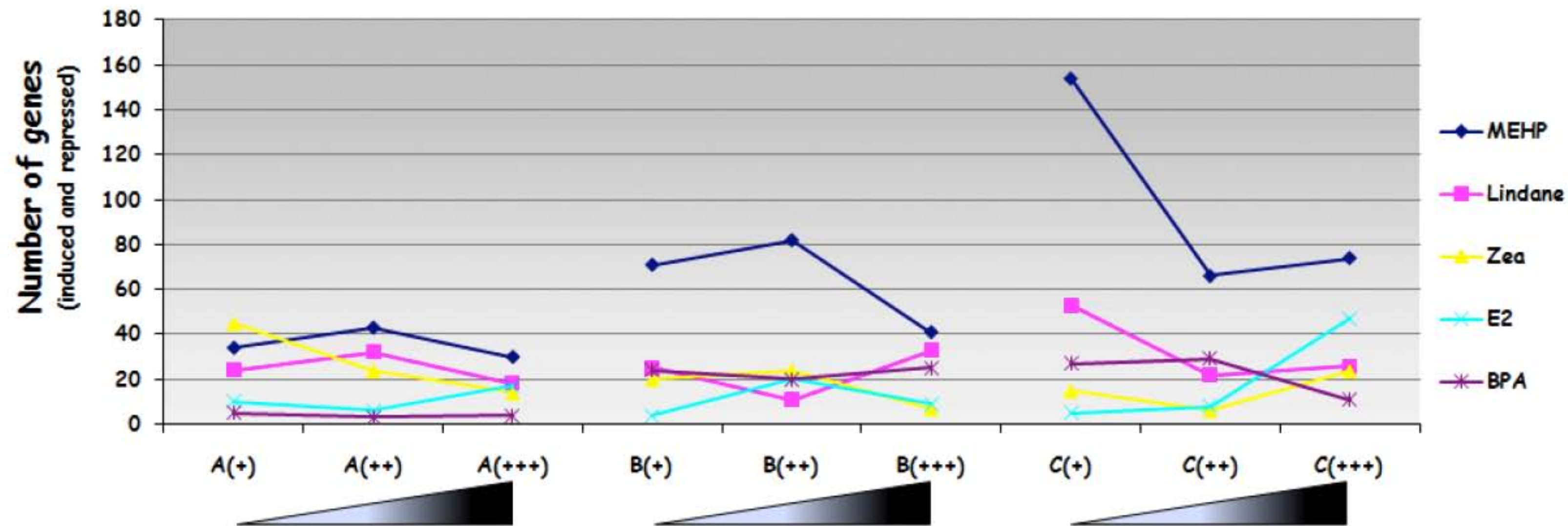
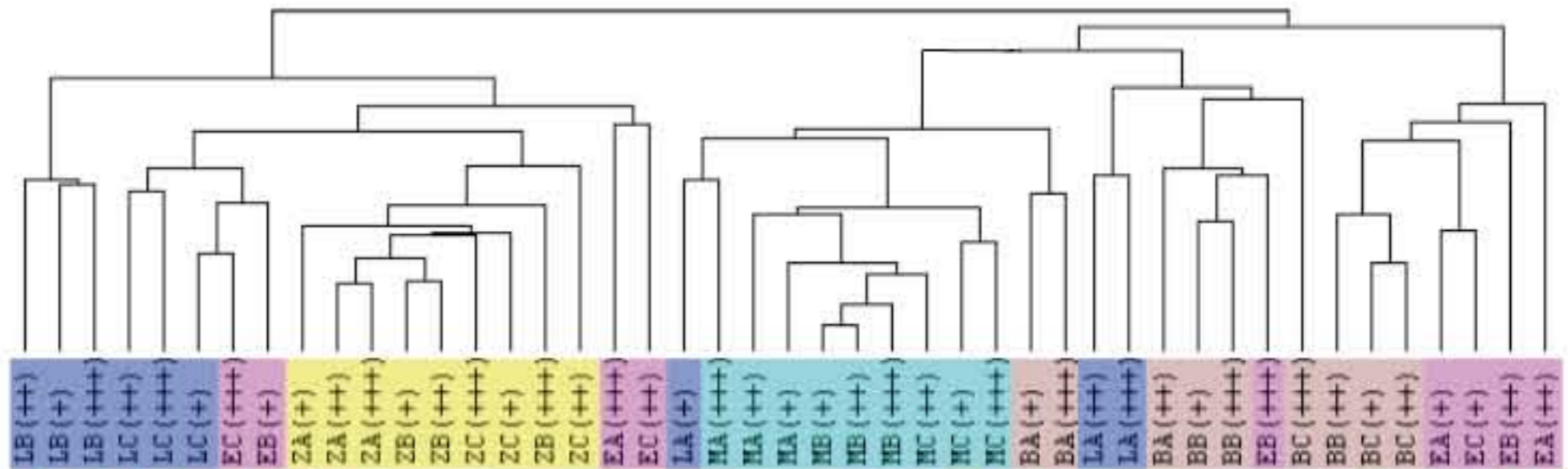


Figure 4

Total Clustering
Peaks 3
Threshold 0.7
2,676 genes



Rtds-I
E-I

II
E-II
III
IV

V
E-III

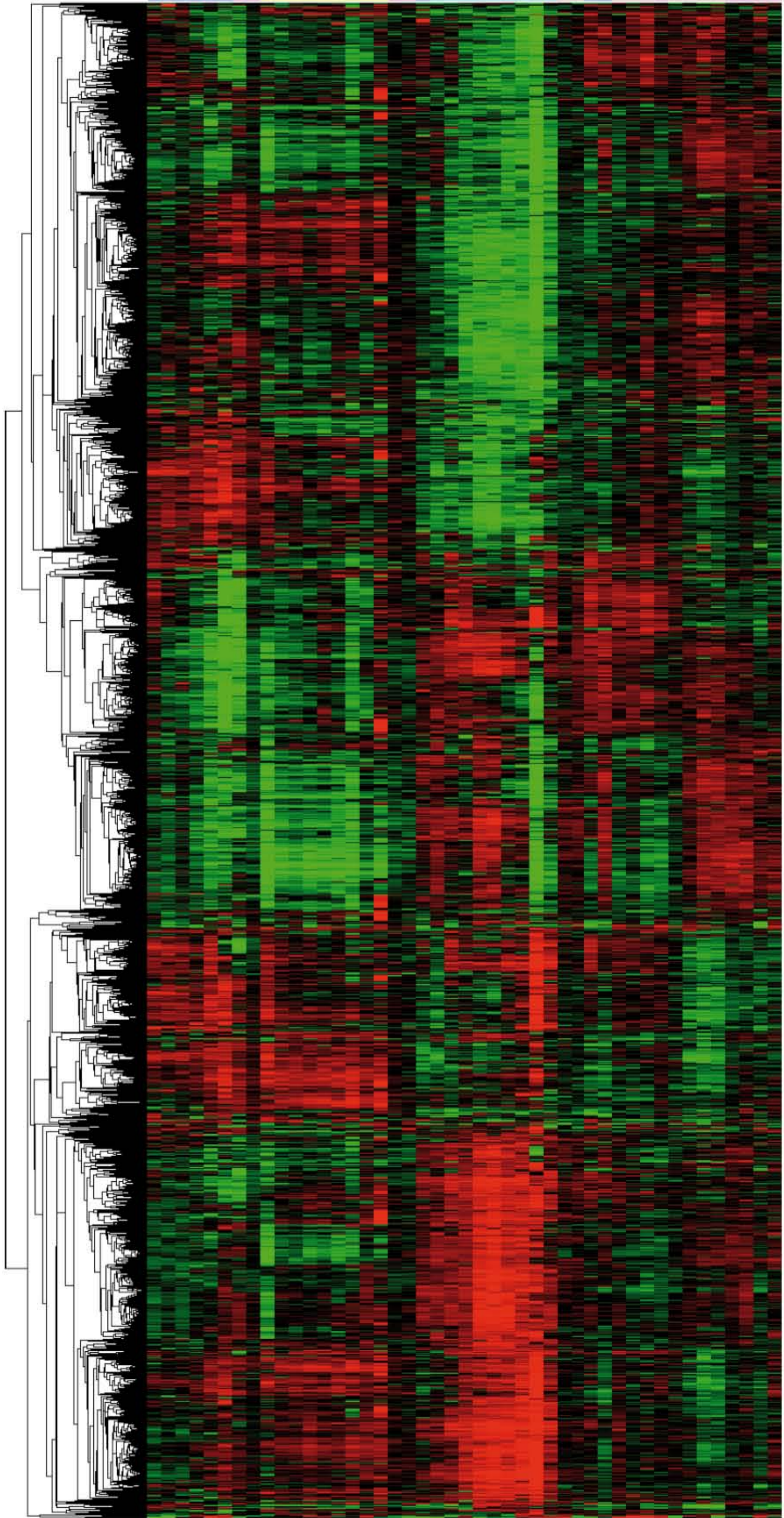
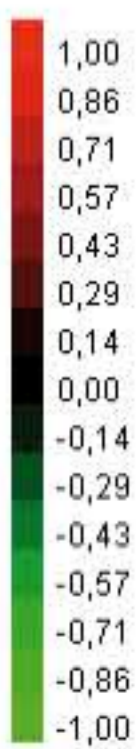


Figure 5

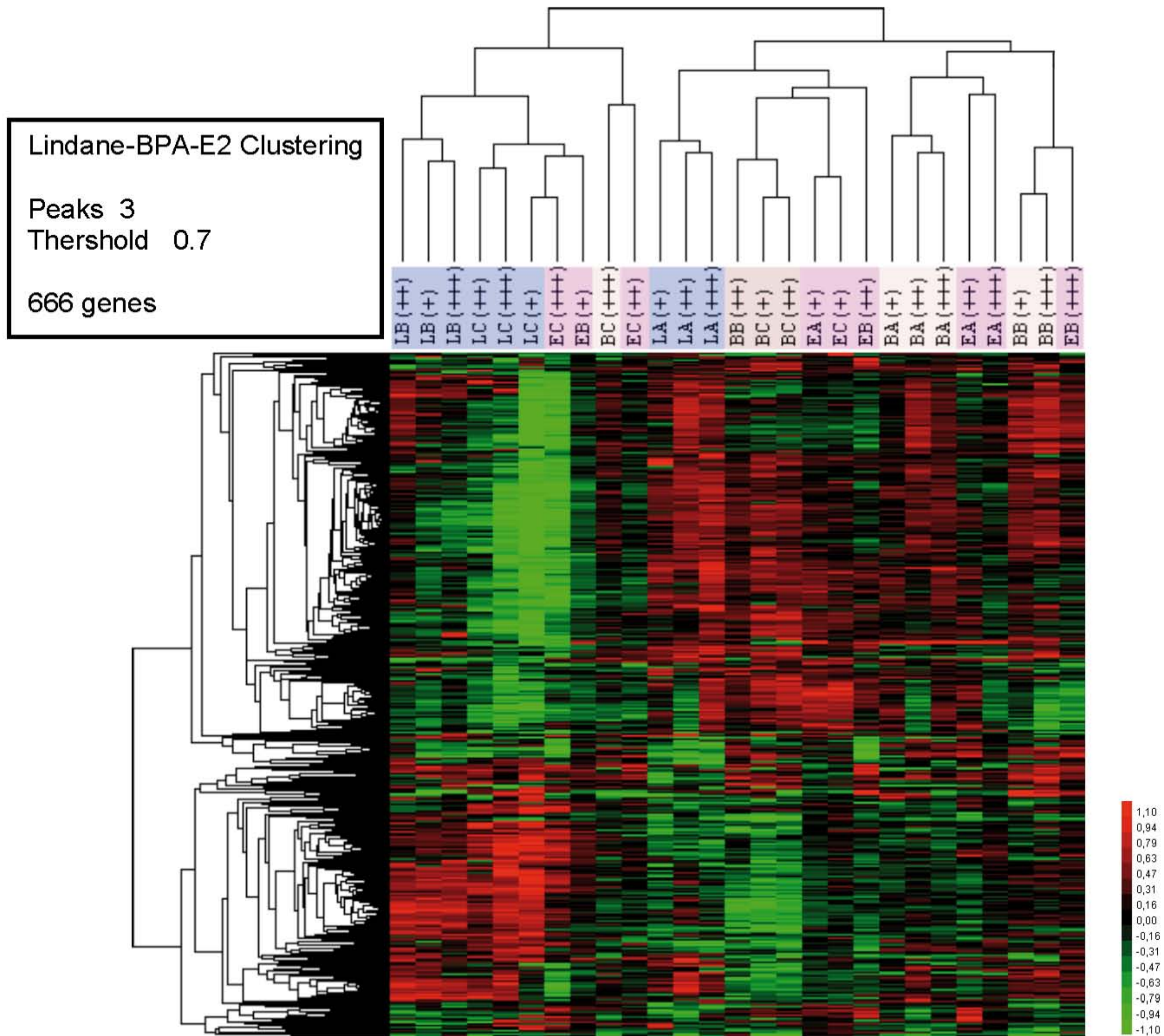
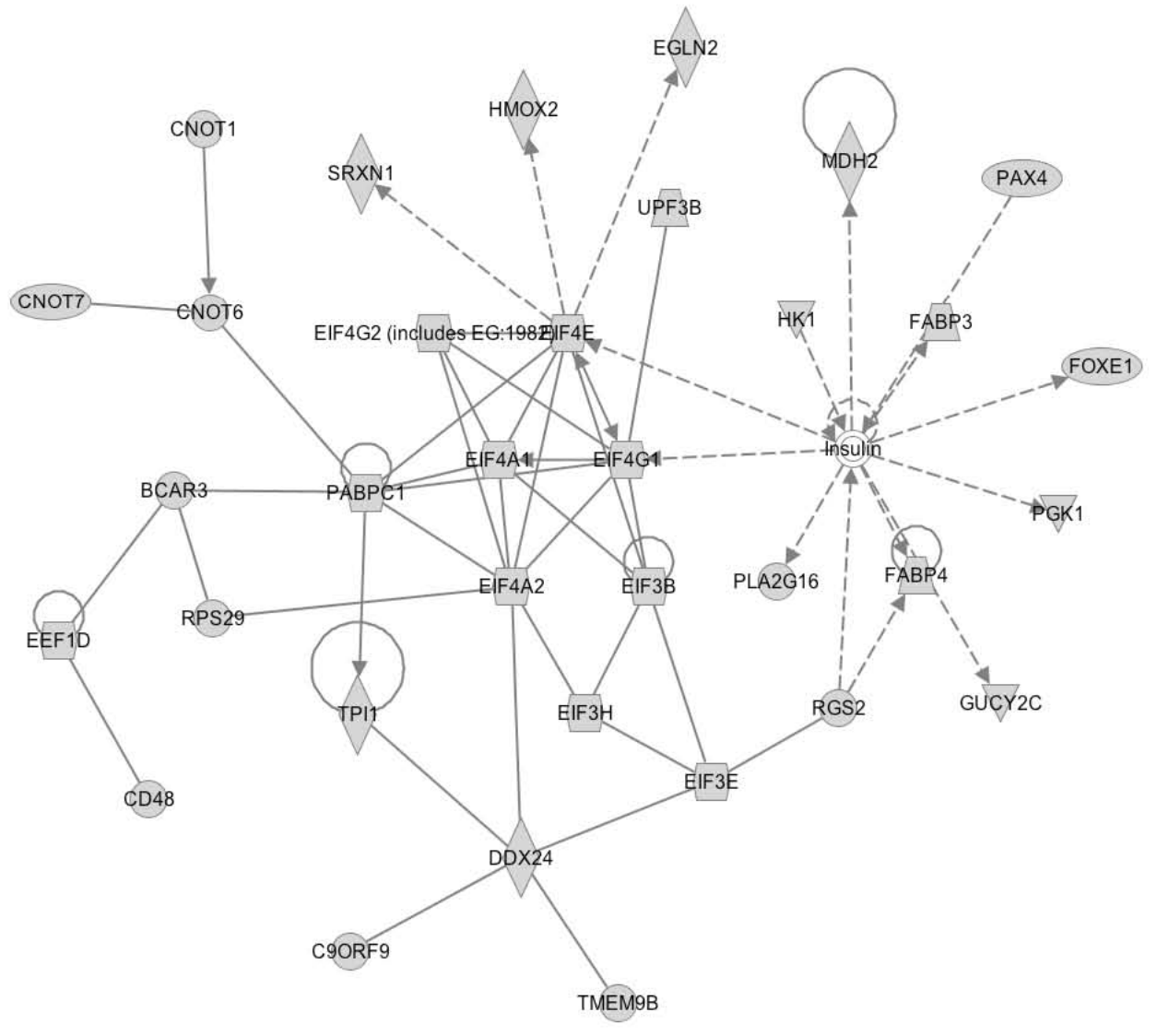
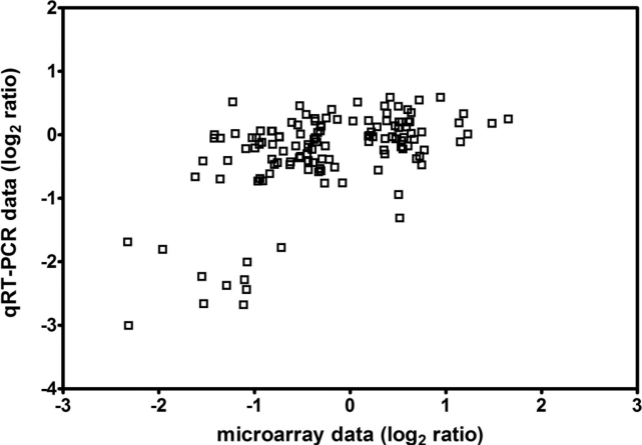


Figure 6





Variable X: microarray data (log₂ ratio)

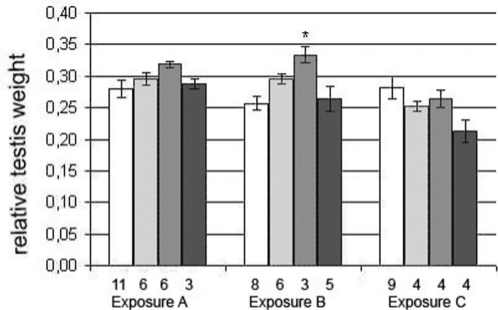
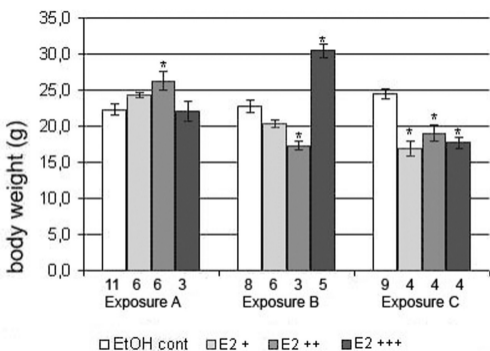
Variable Y: qRT-PCR (log₂ ratio)

Number of XY Pairs= 130

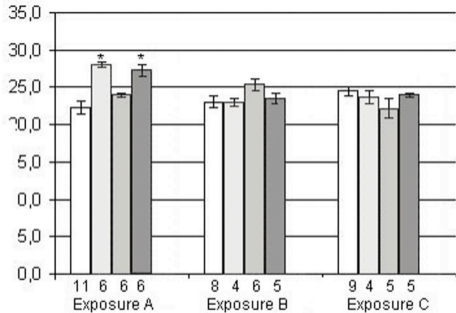
Pearson r = 0.5346

95% confidence interval for r= 0.3991 to 0.6472

P-value (two-tailed): p<0.0001

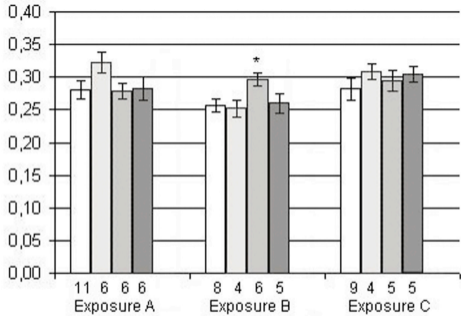


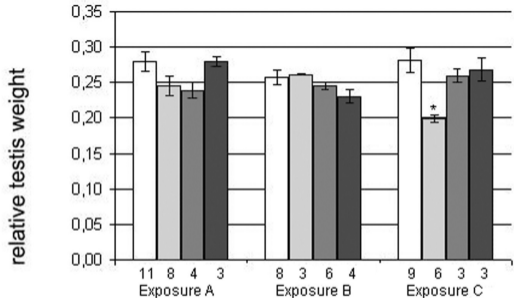
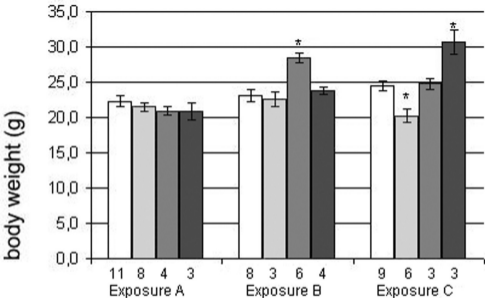
body weight (g)



□ EtOH cont □ BPA + □ BPA ++ □ BPA +++

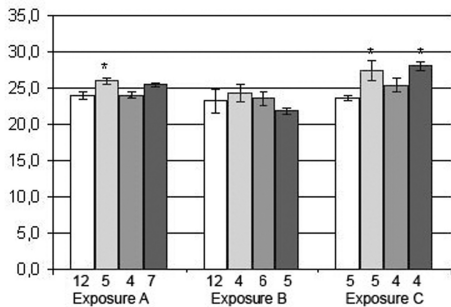
relative testis weight





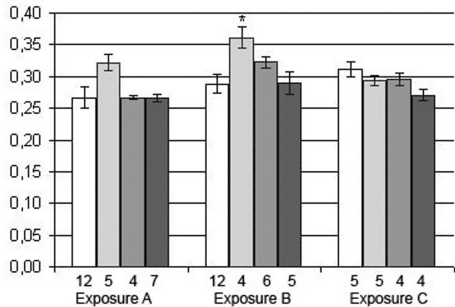
□ EtOH cont □ ZEA + ■ ZEA ++ ■ ZEA +++

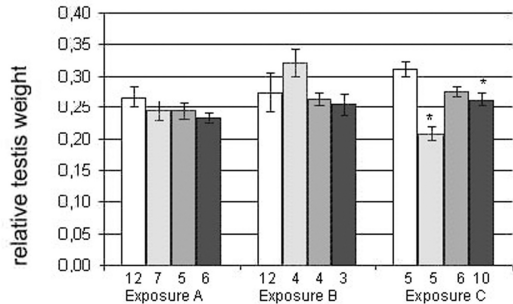
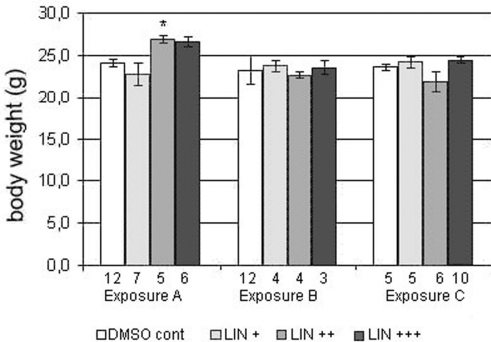
body weight (g)

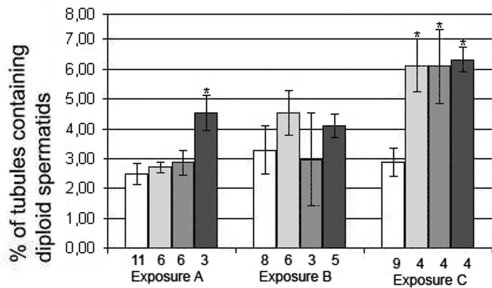
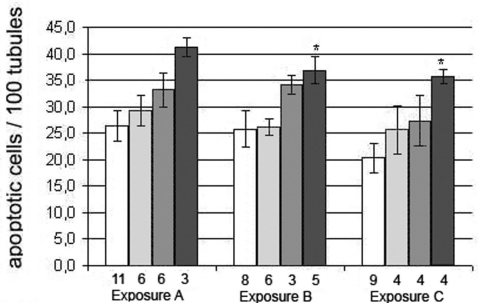
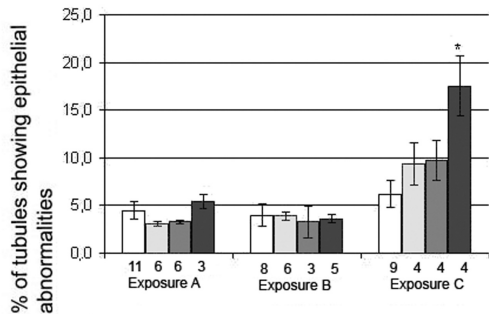
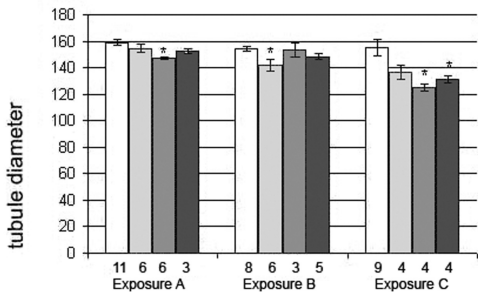


□ DMSO cont □ MEHP + ■ MEHP ++ ■ MEHP +++

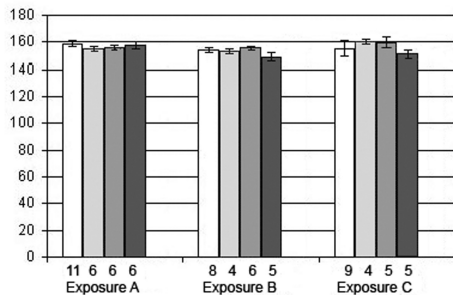
relative testis weight



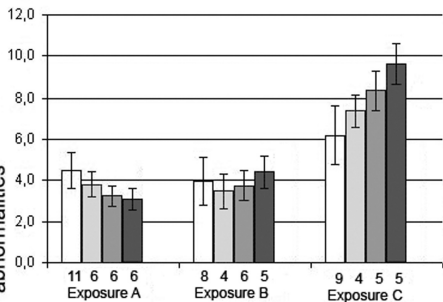




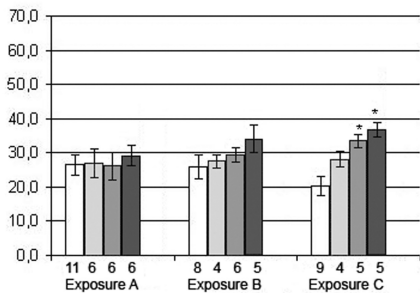
tubule diameter



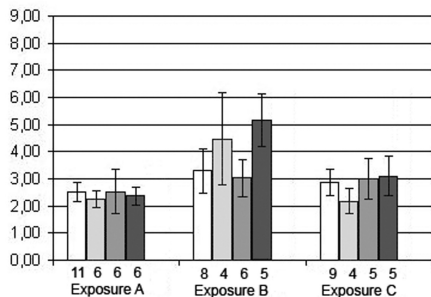
% of tubules showing epithelial abnormalities



apoptotic cells / 100 tubules

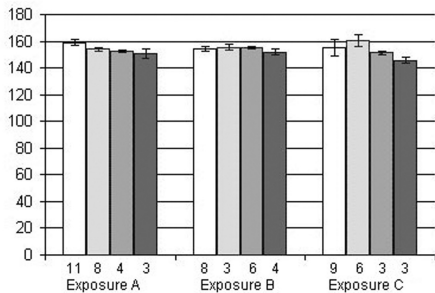


% of tubules containing diploid spermatids

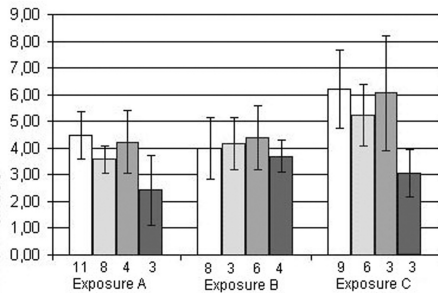


□ EtOH cont □ BPA + ■ BPA ++ ■ BPA +++

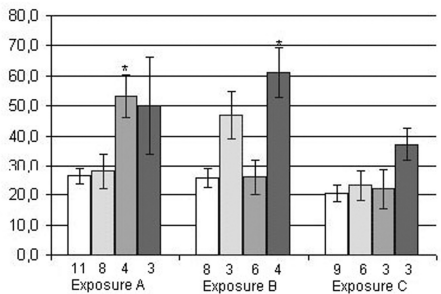
tubule diameter



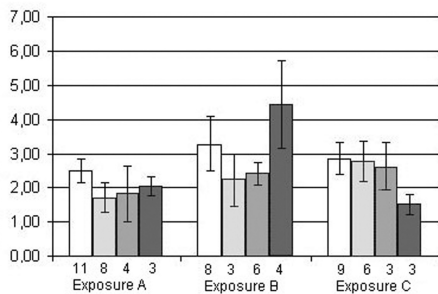
% of tubules showing epithelial abnormalities



apoptotic cells / 100 tubules

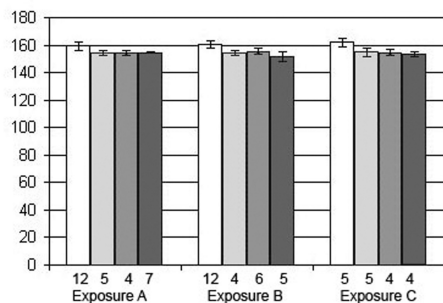


% of tubules containing diploid spermatids

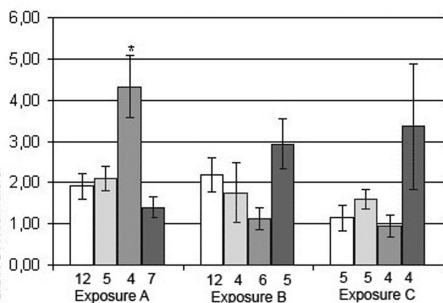


□ EtOH cont □ ZEA + ■ ZEA ++ ■ ZEA +++

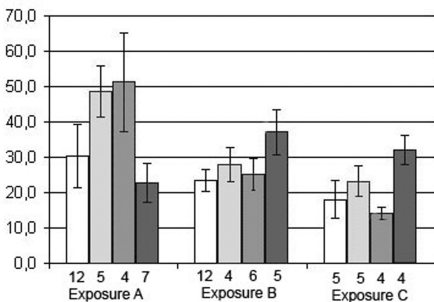
tubule diameter



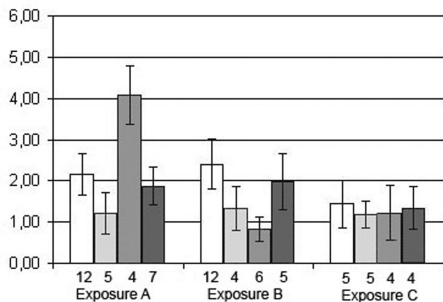
% of tubules showing epithelia abnormalities



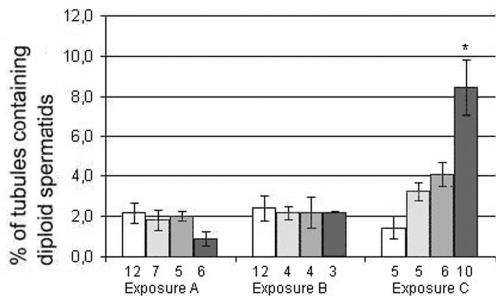
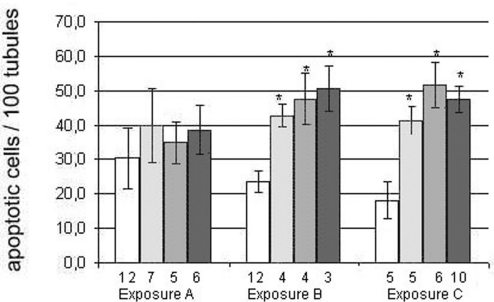
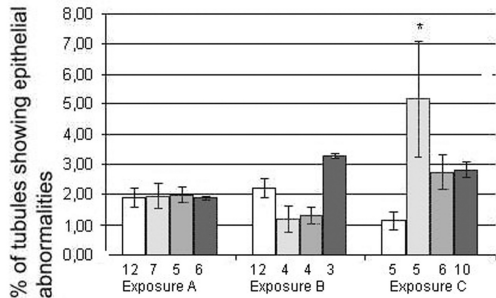
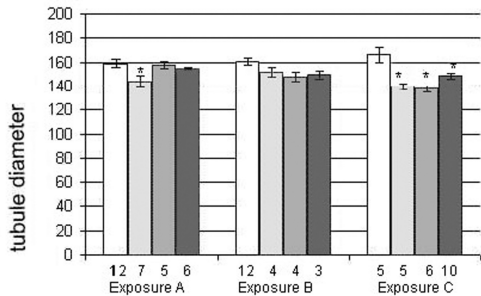
apoptotic cells / 100 tubules



% of tubules containing diploid spermatids



□ DMSO cont ◻ MEHP+ ◼ MEHP++ ◼ MEHP+++



□ DMSO cont □ LIN + □ LIN ++ ■ LIN +++

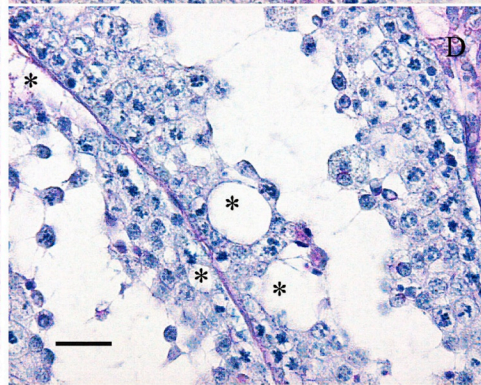
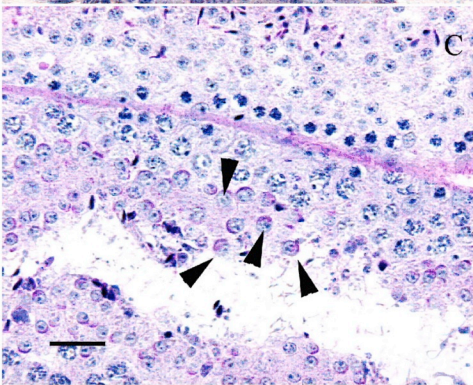
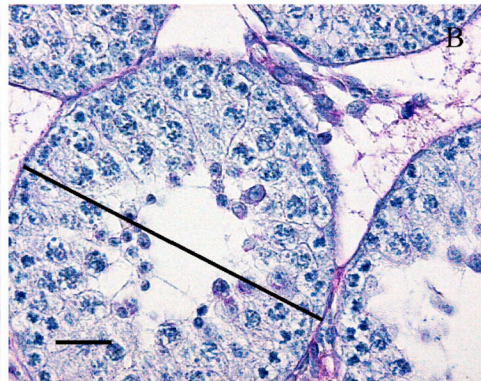
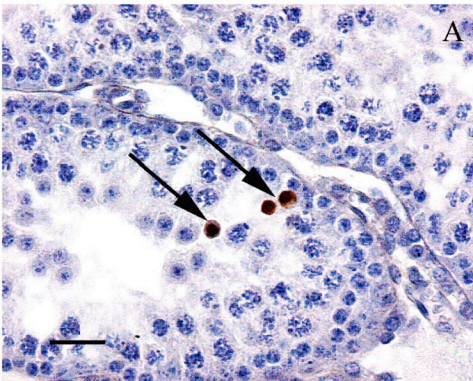


Table S1. Genes selected for quantitative real-time PCR validation purposes and their functions.

Gene Name	Gene Symbol	Function	Process
chromatin modifying protein 4B	Chmp4b	Transfer/carrier protein, Other transfer/carrier protein	Intracellular protein traffic
spermatogenesis associated 18	Spata18	Molecular function unclassified	Biological process unclassified
ect2 oncogene	Ect2	Select regulatory molecule, G-protein modulator	Oncogenesis
eukaryotic translation initiation factor 5A2	Eif5a2	Nucleic acid binding, Translation factor, Translation initiation factor	Protein metabolism and modification, Protein biosynthesis
prominin 1	Prom1	Membrane traffic protein	Intracellular protein traffic, Other intracellular protein traffic
Yip1 domain family, member 5	Yipf5	Molecular function unclassified	Biological process unclassified
DnaJ (Hsp40) homolog, subfamily B, member 8	Dnajb8	Chaperone, Other chaperones	Protein metabolism and modification, Protein folding
spermatid associated	Spert	Molecular function unclassified	Biological process unclassified
RAB5A, member RAS oncogene family	Rab5a	Select regulatory molecule, G-protein, Small GTPase	Signal transduction, Intracellular protein traffic, General vesicle transport, Exocytosis, Regulated exocytosis, Endocytosis, Receptor mediated endocytosis
NADH dehydrogenase (ubiquinone) 1 alpha subcomplex, 13, testis-specific serine kinase 6	Ndufa13,Tssk6	Molecular function unclassified, Kinase, Protein kinase, Non-receptor serine/threonine protein kinase	Biological process unclassified, Protein metabolism and modification, Protein modification, Protein phosphorylation
A kinase (PRKA) anchor protein 3	Akap3	Select regulatory molecule, Kinase modulator	Protein targeting and localization, Protein targeting
ubiquitin specific peptidase 2	Usp2	Protease, Cysteine protease	Protein metabolism and modification, Proteolysis
small nuclear RNA activating complex, polypeptide 2	Snapc2	Transcription factor, Basal transcription factor	Nucleoside, nucleotide and nucleic acid metabolism, mRNA transcription, mRNA transcription regulation
ring finger protein (C3HC4 type) 19	Rnf19	Molecular function unclassified	Protein metabolism and modification, Protein modification, Cell structure and motility, Cell structure
retinoblastoma-like 2	Rbl2	Transcription factor, Other transcription factor, Nucleic acid binding	Nucleoside, nucleotide and nucleic acid metabolism, mRNA transcription, mRNA transcription regulation, Oncogenesis, Tumor suppressor, Cell cycle, Cell cycle control
cell adhesion molecule with homology to L1CAM	Chl1	Cell adhesion molecule, CAM family adhesion molecule	Signal transduction, Cell communication, Cell adhesion-mediated signaling, Developmental processes, Ectoderm development, Neurogenesis, Cell adhesion
defensin beta 19	Defb19	Molecular function unclassified	Biological process unclassified
ring finger protein 10	Rnf10	Molecular function unclassified	Biological process unclassified
glycoprotein, synaptic 2	Gpsn2	Oxidoreductase, Dehydrogenase	Biological process unclassified
zinc finger, FYVE domain containing 9	Zfyve9		
testis derived transcript	Tes	Nucleic acid binding	Electron transport, Other metabolism
phospholipid scramblase 2	Plscr2		
cullin 3	Cul3	Miscellaneous function, Other miscellaneous function protein	Protein metabolism and modification, Proteolysis, Oncogenesis, Cell cycle, Cell cycle control, Apoptosis, Induction of apoptosis

Table S2. Genes preferentially expressed in spermiogenesis (EI-EIII) and round spermatids (Rtds I-V) that appear co-clustered in Figure 3 (EI-III, orange bars, and Rtds I-V, grey bars).

E-I		
Gene	Ref seq DNA	Description
Tax1bp1	NM_025816.1	Tax1 (human T-cell leukemia virus type I) binding protein 1
Col4a3bp	NM_023420.1	Procollagen, type IV, alpha 3 (Goodpasture antigen) binding protein
Clpx	NM_011802.1	caseinolytic peptidase X (E.coli)
lqwd1	AK161893	IQ motif and WD repeats 1
1700007B14Rik	AK077102	Adult male testis cDNA, IQ calmodulin-binding motif containing protei
Anubl1	AK089032	AN1, ubiquitin-like, homolog (Xenopus laevis)
Psm6	NM_011968.1	proteasome (prosome, macropain) subunit, alpha type 6
Mjd	NM_029705.1	Machado-Joseph disease (ataxin 3) homolog (human)
Vashl	NM_144879.1	Vasohibin-like protein
Otud4	AK122429	OTU domain containing 4
Cpeb4	NM_026252.2	cytoplasmic polyadenylation element binding protein 4
Cast	NM_009817.1	Calpastatin, mRNA (cDNA clone MGC:12116 IMAGE:3710078)
Trim36	NM_178872.3	Tripartite motif-containing 36, mRNA
Dgkh	XM_484397	similar to diacylglycerol kinase, eta isoform 1
Tnp1	NM_009407.1	transition protein 1
Hils1	NM_018792.1	histone H1-like protein in spermatids 1
1700010M22Rik	NM_025490.1	SIMILARITY: Contains 1 SCP2 domain
Psme4	NM_134013.2	Proteasome (prosome, macropain) activator subunit 4, mRNA
E-II		
Gene	Ref seq DNA	Description
Prkcd	NM_011103.1	Protein kinase C, delta (Prkcd)
Sh3bp5	NM_011894.1	SH3-domain binding protein 5 (BTK-associated)
Nt5c1b	NM_027588.2	5'-nucleotidase, cytosolic IB
Cdrt4	NM_025496.1	CMT1A duplicated region transcript 4
lsg20l2	NM_177663.2	interferon stimulated exonuclease gene 20-like 2
Q9DC38	AK004585	Adult male lung cDNA hypothetical protein
4931440F15Rik	NM_176829.1	hypothetical protein LOC216622
Sstk	NM_032004.1	Serine/threonine protein kinase SSTK (Sstk)
Tssk2	NM_009436.1	Testis-specific serine kinase 2 (Tssk2)
1700003N22Rik	NM_145538.1	similar to Ubiquitin-conjugating enzyme variant Kua
2810453I06Rik	NM_026050.1	hypothetical protein LOC67238
Aif1	NM_019467.2	Allograft inflammatory factor 1
Nup35	NM_027091.1	nucleoporin 35
D2Ert750e	NM_026412.1	TRAF4 associated factor 1 homolog
1700012A03Rik	NM_029587.1	hypothetical protein LOC76382
Dnajb8	NM_019964.1	DnaJ (Hsp40) homolog, subfamily B, member 8 (Dnajb8)
Csnk1g2	NM_134002.1	casein kinase 1, gamma 2
Spata18	NM_178387.2	Spermatogenesis associated 18 homolog (rat)
Fscn3	NM_019569.2	Fascin homolog 3, actin-bundling protein, testicular (S. purpuratus)
Otub2	NM_026580.1	OTU domain, ubiquitin aldehyde binding 2
Dp111	NM_139292.1	Deleted in polyposis 1-like 1 (Dp111), mRNA
Gsg1	NM_010352.1	Germ cell-specific gene 1
4931407G18Rik	NM_027631.1	testis development protein NYD-SP26
Rnf138	NM_019706.2	Ring finger protein 138, transcript variant 2
lqcf3	NM_026645	IQ motif containing F3
5730596K20Rik	NM_146091.2	hypothetical protein LOC109168
1700029H14Rik	NM_025601.1	hypothetical protein LOC66501
Ppm1a	NM_008910.2	protein phosphatase 1A, magnesium dependent, alpha isoform
Psmf1	NM_144889.2	proteasome (prosome, macropain) inhibitor subunit 1

Skd3	NM_009191.2	Suppressor of K+ transport defect 3 (Skd3), mRNA
Spata19	NM_029299.1	spermatogenesis associated 19
R3hcc1	BC059860	R3H domain and coiled-coil containing 1
Pdzk1	NM_021517.1	PDZ domain containing 1
Nurit	NM_026457.1	Testis-specific leucine zipper protein nurit mRNA,
Adam1a	NM_172126.2	Fertilin alpha precursor (ADAM 1)
4930431B09Rik	BC046309	PREDICTED
Tisp78	NM_144827.3	Spermatogenesis associated 20
Mfap3l	NM_027756.2	Microfibrillar-associated protein 3-like
Chn2	NM_023543.1	chimerin (chimaerin) 2
Tmco5	NM_026104.1	transmembrane and coiled-coil domains 5
5031400M07Rik	NM_020586.1	hypothetical protein LOC80517
Fts	NM_010241.1	Fused toes, mRNA
1700009N14Rik	XM_131323	hypothetical protein LOC75471
Ppp2r2b	NM_027531.1	protein phosphatase 2 (formerly 2A), regulatory subunit B (PR 52)
Akap4	NM_009651.1	A kinase (PRKA) anchor protein 4 (Akap4), mRNA
4921517D22Rik	NM_183290.1	hypothetical protein LOC70900
Tmem56	NM_178936.2	Transmembrane protein 56 (Tmem56)
Mgat4a	NM_173870.1	mannoside acetylglucosaminyltransferase 4, isoenzyme A
1700029J11Rik	AK006503	Adult male testis cDNA, product:hypothetical protein
Ube2v2	NM_023585.2	Ubiquitin-conjugating enzyme E2 variant 2 (Ube2v2)
4933417A18Rik	NM_025750.1	hypothetical protein LOC66761
Odf2	NM_013615.1	outer dense fiber of sperm tails 2
Orc3l	NM_001007589.1	origin recognition complex, subunit 3-like (S. cerevisiae)
4921530L21Rik	NM_025733.1	hypothetical protein LOC66732
4921530D09Rik	AK014983	Adult male testis cDNA, hypothetical protein
1700011H14Rik	NM_025956.2	hypothetical protein LOC67082
Ccdc50	NM_001025615.1	Coiled-coil domain-containing protein 50 (Protein Ymer)
4921528O07Rik	AK019544	Adult male testis cDNA, hypothetical ARM repeat containing protein
Plcz1	NM_054066.2	Phospholipase C, zeta 1 (Plcz1), mRNA
Hip2	NM_016786.2	Huntingtin interacting protein 2
Spag4l	NM_029599.1	Sperm associated antigen 4-like (Spag4l), mRNA
Tes	NM_011570.2	Testin, testis derived transcript (Tes), transcript variant 1
Gpd2	NM_010274.2	glycerol phosphate dehydrogenase 2, mitochondrial
Lnp	NM_027133.1	limb and neural patterns
Zp3r	NM_009581.1	Zona pellucida 3 receptor (Zp3r), mRNA
Sec24b	NM_207209.1	SEC24 related gene family, member B (S. cerevisiae)
Pdia3	NM_007952.1	protein disulfide isomerase associated 3

E-III

Gene	Ref seq DNA	Description
Fabp3	NM_010174.1	fatty acid binding protein 3, muscle and heart
1700023L04Rik	AK076883	hypothetical protein LOC76419
Thrap3	NM_025856.1	thyroid hormone receptor associated protein 3
Dnajb13	NM_153527.2	DnaJ (Hsp40) related, subfamily B, member 13
4930579J09Rik	NM_133689.1	IIIG9 protein, a novel mRNA expressed along brain ventricles
Stard10	NM_019990.1	START domain containing 10 (Stard10), mRNA
Osbp2	NM_152818.2	Oxysterol binding protein 2 (Osbp2), mRNA
Arpm2	NM_028513.2	Actin related protein M2 (Arpm2), mRNA
4930579G22Rik	NM_026916.1	hypothetical protein LOC69034
1700056E22Rik	AK006811	hypothetical protein LOC73363
1700054O13Rik	NM_026096.1	Huntingtin interacting protein M
Nr1d1	NM_145434.1	Nuclear receptor subfamily 1, group D, member 1 (Nr1d1)
BC051628	NM_199312.1	hypothetical protein LOC332713

Rtids-I

Gene	Ref seq DNA	Description
Ddx20	NM_017397.2	DEAD (Asp-Glu-Ala-Asp) box polypeptide 20 (Ddx20), mRNA
Col4a3bp	NM_023420.1	Procollagen, type IV, alpha 3 (Goodpasture antigen) binding protein
4921521K07Rik	NM_027599.1	hypothetical protein LOC70902

Ngly1	NM_021504.2	N-glycanase 1 (Ngly1), mRNA
1700007B14Rik	AK077102	IQ Calmodulin binding motif containing protein
Anubl1	AK089032	AN1, ubiquitin-like, homolog (<i>X. laevis</i>)
PsmA6	NM_011968.1	proteasome (prosome, macropain) subunit, alpha type 6
Tsnaxip1	NM_024445.2	translin-associated factor X (Tsnax) interacting protein 1
Mjd	NM_029705.1	Machado-Joseph disease (ataxin 3) homolog (human)
4631422O05Rik	BC062166	4631422O05 gene
B130052G07Rik	NM_144879.1	Vasohibin-like protein
Otud4	AK122429	OTU domain containing 4/HIV-1 induced protein HIN-1
Cpeb4	NM_026252.2	cytoplasmic polyadenylation element binding protein 4
Trim36	NM_178872.3	Tripartite motif-containing 36, mRNA
Dgkh	XM_484397	<i>M. musculus</i> diacylglycerol kinase, eta, transcript variant 1
4932412H11Rik	NM_172879.1	hypothetical protein LOC242838
Tnp1	NM_009407.1	transition protein 1
Hils1	NM_018792.1	histone H1-like protein in spermatids 1
1700010M22Rik	NM_025490.1	SIMILARITY: Contains 1 SCP2 domain

Rtds-II

Gene	Ref seq DNA	Description
Prkcd	NM_011103.1	Protein kinase C, delta (Prkcd), mRNA
Nt5c1b	NM_027588.2	5'-nucleotidase, cytosolic IB
Cdrt4	NM_025496.1	CMT1A duplicated region transcript 4
Isg20l2	NM_177663.2	interferon stimulated exonuclease gene 20-like 2
Q9DC38	AK004585	Adult male lung cDNA hypothetical protein
Sstk	NM_032004.1	Serine/threonine protein kinase SSTK (Sstk)
Tssk2	NM_009436.1	Testis-specific serine kinase 2 (Tssk2), mRNA
2810453I06Rik	NM_026050.1	hypothetical protein LOC67238
Aif1	NM_019467.2	Allograft inflammatory factor 1, mRNA
1700012A03Rik	NM_029587.1	hypothetical protein LOC76382
Dnajb8	NM_019964.1	DnaJ (Hsp40) homolog, subfamily B, member 8 (Dnajb8)
Csnk1g2	NM_134002.1	casein kinase 1, gamma 2
Spata18	NM_178387.2	Spermatogenesis associated 18 homolog (rat)
Fscn3	NM_019569.2	Fascin homolog 3, actin-bundling protein, testicular (<i>S. purpuratus</i>)
Otub2	NM_026580.1	OTU domain, ubiquitin aldehyde binding 2
Gsg1	NM_010352.1	Germ cell-specific gene 1, mRNA (798)
4931407G18Rik	NM_027631.1	testis development protein NYD-SP26
Rnf138	NM_019706.2	Ring finger protein 138, transcript variant 2, mRNA
1700023E05Rik	NM_027970.1	hypothetical protein LOC71868
Iqcf3	NM_026645	IQ motif containing F3, mRNA
5730596K20Rik	NM_146091.2	hypothetical protein LOC109168
1700029H14Rik	NM_025601.1	hypothetical protein LOC66501
Psmf1	NM_144889.2	proteasome (prosome, macropain) inhibitor subunit 1
Skd3	NM_009191.2	Suppressor of K ⁺ transport defect 3 (Skd3), mRNA
Spata19	NM_029299.1	spermatogenesis associated 19
Agpat3	NM_053014.2	1-acylglycerol-3-phosphate O-acyltransferase 3 (Agpat3)
R3hcc1	BC059860	R3H domain and coiled-coil containing 1
Pdzk1	NM_021517.1	PDZ domain containing 1
Rnf38	NM_175201.2	Ring finger protein 38, mRNA
Nurit	NM_026457.1	Testis-specific leucine zipper protein nurit, alternatively spliced
Adam1a	NM_172126.2	Fertilin alpha precursor (ADAM 1)
4930431B09Rik	BC046309	PREDICTED
Tisp78	NM_144827.3	Spermatogenesis associated 20, mRNA
Wdr51b	NM_027740.3	WD repeat domain 51B
Mfap3l	NM_027756.2	Microfibrillar-associated protein 3-like, mRNA:5025281)
4930435E12Rik	AK077164	PREDICTED: similar to hypothetical protein FLJ32859
	NM_001013783.1	hypothetical protein LOC432552
Chn2	NM_023543.1	chimerin (chimaerin) 2
Tmco5	NM_026104.1	transmembrane and coiled-coil domains 5
5031400M07Rik	NM_020586.1	hypothetical protein LOC80517
1700009N14Rik	AK005799	hypothetical protein LOC75471

Akap4	NM_009651.1	A kinase (PRKA) anchor protein 4 (Akap4), mRNA
4921517D22Rik	NM_183290.1	hypothetical protein LOC70900

Rtds-III

Gene	Ref seq DNA	Description
4921530L21Rik	NM_025733.1	hypothetical protein LOC66732
4921530D09Rik	AK014983	adult male testis cDNA, hypothetical protein
1700011H14Rik	NM_025956.2	hypothetical protein LOC67082
Ccdc67	NM_181816.1	coiled-coil domain containing 67
4921528O07Rik	AK019544	Adult male testis cDNA, hypothetical ARM repeat structure
Plcz1	NM_054066.2	Phospholipase C, zeta 1 (Plcz1), mRNA
Chl1	NM_007697.1	Cell adhesion molecule with homology to L1CAM

Rtds-IV

Gene	Ref seq DNA	Description
Tesp1	NM_009355.1	Testicular serine protease 1 (Tesp1), mRNA
Spag4l	NM_029599.1	Sperm associated antigen 4-like (Spag4l), mRNA
Tekt2	NM_011902.1	Tektin 2, mRNA (cDNA clone MGC:5692 IMAGE:3582971)
1700025E21Rik	NM_029373.1	hypothetical protein LOC75647
Tes	NM_011570.2	Testin, testis derived transcript (Tes), transcript variant 1
Gpd2	NM_010274.2	glycerol phosphate dehydrogenase 2, mitochondrial
Rnf139	NM_175226.2	Ring finger protein 139, mRNA
Lnp	NM_027133.1	limb and neural patterns
Zp3r	NM_009581.1	Zona pellucida 3 receptor (Zp3r), mRNA

Rtds-V

Gene	Ref seq DNA	Description
1700023L04Rik	AK076883	Adult male testis cDNA, hypothetical protein
1700110M21Rik	AK007164	Adult male testis cDNA, hypothetical protein
Thrap3	NM_025856.1	thyroid hormone receptor associated protein 3
Dnajb13	NM_153527.2	DnaJ (Hsp40) related, subfamily B, member 13
4930579J09Rik	NM_133689.1	IIIG9 protein, a novel mRNA expressed along brain ventricles
Arpm2	NM_028513.2	Actin related protein M2 (Arpm2), mRNA
4930579G22Rik	NM_026916.1	hypothetical protein LOC69034
1700054O13Rik	NM_026096.1	Huntingtin interacting protein M
BC051628	NM_199312.1	hypothetical protein LOC332713

Table S3. Genes included in the most representative network from the hierarchical clustering analysis considering all experimental conditions of exposure

Symbol	Entrez Gene Name	Location	Family
BCAR3	breast cancer anti-estrogen resistance 3	Cytoplasm	other
C9ORF9	chromosome 9 open reading frame 9	Unknown	other
CD48	CD48 molecule	Plasma Membrane	other
CNOT1	CCR4-NOT transcription complex, subunit 1	Unknown	other
CNOT6	CCR4-NOT transcription complex, subunit 6	Nucleus	other
CNOT7	CCR4-NOT transcription complex, subunit 7	Nucleus	transcription regulator
DDX24	DEAD (Asp-Glu-Ala-Asp) box polypeptide 24	Nucleus	enzyme
EEF1D	eukaryotic translation elongation factor 1 delta (guanine nucleotide exchange protein)	Cytoplasm	translation regulator
EGLN2	egl nine homolog 2 (<i>C. elegans</i>)	Cytoplasm	enzyme
EIF3B	eukaryotic translation initiation factor 3, subunit B	Cytoplasm	translation regulator
EIF3E	eukaryotic translation initiation factor 3, subunit E	Cytoplasm	translation regulator
EIF3H	eukaryotic translation initiation factor 3, subunit H	Cytoplasm	translation regulator
EIF4A1	eukaryotic translation initiation factor 4A, isoform 1	Cytoplasm	translation regulator
EIF4A2	eukaryotic translation initiation factor 4A, isoform 2	Cytoplasm	translation regulator
EIF4E	eukaryotic translation initiation factor 4E	Cytoplasm	translation regulator
EIF4G1	eukaryotic translation initiation factor 4 gamma, 1	Cytoplasm	translation regulator
EIF4G2 (includes EG:1982)	eukaryotic translation initiation factor 4 gamma, 2	Cytoplasm	translation regulator
FABP3	fatty acid binding protein 3, muscle and heart (mammary-derived growth inhibitor)	Cytoplasm	transporter
FABP4	fatty acid binding protein 4, adipocyte	Cytoplasm	transporter
FOXE1	forkhead box E1 (thyroid transcription factor 2)	Nucleus	transcription regulator
GUCY2C	guanylate cyclase 2C (heat stable enterotoxin receptor)	Plasma Membrane	kinase
HK1	hexokinase 1	Cytoplasm	kinase
HMOX2	heme oxygenase (decycling) 2	Cytoplasm	enzyme
Insulin		Unknown	group
MDH2	malate dehydrogenase 2, NAD (mitochondrial)	Cytoplasm	enzyme
PABPC1	poly(A) binding protein, cytoplasmic 1	Cytoplasm	translation regulator
PAX4	paired box 4	Nucleus	transcription regulator
PGK1	phosphoglycerate kinase 1	Cytoplasm	kinase
PLA2G16	phospholipase A2, group XVI	Nucleus	other
RGS2	regulator of G-protein signaling 2, 24kDa	Nucleus	other
RPS29	ribosomal protein S29	Cytoplasm	other
SRXN1	sulfiredoxin 1 homolog (<i>S. cerevisiae</i>)	Cytoplasm	enzyme
TMEM9B	TMEM9 domain family, member B	Plasma Membrane	other
TPI1	triosephosphate isomerase 1	Cytoplasm	enzyme
UPF3B	UPF3 regulator of nonsense transcripts homolog B (yeast)	Nucleus	transporter