

1  
2  
3 **Hydro-physical responses of gypseous and non-gypseous soils to livestock**  
4 **grazing in a semi-arid region of NE Spain**  
5  
6

7 **D. Moret-Fernández <sup>a\*</sup>, Y. Pueyo <sup>b</sup>, C. G. Bueno <sup>c</sup>, C. L. Alados <sup>b</sup>**  
8  
9  
10

11 <sup>a</sup> *Departamento de Suelo y Agua, Estación Experimental de Aula Dei, Consejo Superior de*  
12 *Investigaciones Científicas (CSIC), PO Box 13034, 50080 Zaragoza, Spain.*

13 <sup>b</sup> *Instituto Pirenaico de Ecología (CSIC), Av. Montañana 1005, P.O. Box 13.034, 50080*  
14 *Zaragoza, Spain*

15 <sup>c</sup> *Instituto Pirenaico de Ecología (CSIC), Av. Rgnto. Galicia s/n, P.O. Box 64, 22700 Jaca,*  
16 *Spain*

17  
18  
19 \* Corresponding author. Tel.: (+34) 976 716140; Fax: (+34) 976 716145

20 *E-mail address:* [david@eead.csic.es](mailto:david@eead.csic.es)  
21  
22  
23

# Hydro-physical responses of gypseous and non-gypseous soils to livestock grazing in a semi-arid region of NE Spain

## Abstract

Pasture productivity depends on soil hydro-physical properties, which in turn are deeply affected by livestock grazing. However, the comparative response of different soil types, and particularly gypseous soil types, to grazing has hardly been studied before. This paper compares the effect of grazing on the soil hydro-physical properties of silty gypseous (Gy) and non-gypseous (NGy) soils located in a semi-arid region (Middle Ebro Valley, NE Spain). Two different soil managements were selected: ungrazed natural shrubland (N) and grazed shrubland (GR) soils. The gypsum,  $\text{CaCO}_3$  and organic matter content (OM), soil texture, soil bulk density ( $\rho_b$ ), penetration resistance (PR), saturated sorptivity ( $S$ ), hydraulic conductivity ( $K$ ), and the water retention curve (WRC) for undisturbed soil samples from the 1-10 cm depth soil layer were measured. The  $\rho_b$  and PR in NGy soils were significantly higher than those observed in the Gy ones. Soil compaction due to grazing treatment tended to increase  $\rho_b$  and decrease the  $K$  and  $S$  values. While no differences in PR were observed in the Gy soils between grazing treatments, the PR measured in the NGy soils under GR was significantly higher than the corresponding values observed under N. Differences in  $K$  and  $S$  between GR and N treatments were only significant ( $p < 0.05$ ) in NGy soils, where  $K$  and  $S$  values under the N treatment were almost four times greater than the corresponding values measured under GR. Overall, no differences in the WRCs were observed between soil types and grazing treatments. While the WRCs of NGy soils were not significantly affected by the grazing treatment, Gy soils under N treatment present a significantly higher level of soil macropores than under GR treatment. The hydro-physical features of Gy soils tended to be less affected by grazing than those of the NGy soils. These results suggest that livestock grazing, in both Gy and NGy soils, has a negative effect on

1 the physical soil properties, which should be taken into account by land managers of these semi-  
2 arid regions where silty gypseous and non-gypseous areas coexist.

3  
4 **Keywords:** Water retention curve (WRC); Penetration resistance (PR); Hydraulic conductivity;  
5 Soil water infiltration; Bulk density, Ungrazed natural shrubland, Grazed shrubland.

## 6 7 **1. Introduction**

8 In areas with a long tradition of livestock grazing, moderate grazing pressure is beneficial to  
9 preserve rangeland productivity and biodiversity (McNaughton, 1985; Milchunas et al., 1988;  
10 Montalvo et al., 1993). In addition, livestock can act as seed dispersal agents that reduce  
11 isolation between vegetation remnants in fragmented landscapes (Collins et al., 1998; Pueyo et  
12 al., 2008). Nevertheless, the negative effects of overgrazing are well documented, including loss  
13 of plant cover and soil degradation, which can ultimately lead to desertification in arid and  
14 semi-arid regions (Heady and Child, 1994; Hary et al., 1996; UNCCD, 1994). In such regions,  
15 maintaining a high vegetation cover is essential for soil conservation (Puigdefabregas and  
16 Mendizabal, 1998).

17 Reduction in pasture productivity in semi-arid areas is associated with water availability  
18 (McNaughton, 1979), and plant water absorption is mediated by the soil chemical and physical  
19 properties (Gijssman and Thomas, 1996). Degradation of the soil physical properties in  
20 rangelands depends on the soil texture, type of vegetation, and soil moisture at the time of  
21 grazing (Chanasyk and Naeth, 1995), and it is mainly associated with animal trampling (da  
22 Silva et al., 2003; Quiroga et al., 2009; Hoshino et al., 2009; Kumar et al., 2010). Soil  
23 compaction due to livestock trampling reduces the pore volume and the saturated hydraulic  
24 conductivity, and modifies the water-retention characteristics (Zhao et al., 2007; Krummelbein  
25 et al., 2009). However, these general results contrast with other studies, such as those obtained  
26 by Hoshino et al. (2009), who found that the water available to plants in the ungrazed areas of

1 the Mongolian steppe grasslands was not significantly different from that measured in the  
2 grazed areas.

3 Gypsum content has a significant effect on the soil hydro-physical properties. Very high  
4 gypsum content in soils reduces soil aggregation and results in a cemented layer along the soil  
5 profile (Herrero and Boixadera, 2002), increases the irregularities in moisture distribution in the  
6 soil profile (Boyadgiev, 1974), and reduces the soil water infiltration rates (Poch et al., 1998).  
7 Together with the low water retention of gypsum soils (Boukhris and Lossaint, 1975), this  
8 means that the water available for plants is very low (Poch and Verplancke, 1977).

9 Research into the effect of grazing on the hydro-physical properties of gypseous soils is very  
10 scarce (Meyer and García-Moya, 1989). The purpose of the present study was to compare the  
11 grazing effect on the soil hydro-physical properties for silty gypseous and non-gypseous soils.  
12 Measurements of bulk density, penetration resistance, soil hydraulic parameters at saturation,  
13 and the water retention curves sampled on the 1-10 cm depth soil layer of an ungrazed  
14 shrubland soil were compared with the corresponding values measured in grazed areas.

## 16 **2. Material and methods**

### 17 *2.1. Study area and experimental design*

18 The study area is located in the Middle Ebro Valley (NE Spain; 41°30'N, 0°15'W), 250 m above  
19 sea level (a.s.l.) and the largest gypsum outcrop in Europe (Fig. 1). The area is one of Europe's  
20 most arid zones (Herrero and Snyder, 1997). The climate is semi-arid Mediterranean. The mean  
21 rainfall is 353 mm/yr (average of 50 years at 250 m above sea level), and the mean annual  
22 temperature is 14.9°C (M.A.P.A, 1987) (Table 1). The lithology is a gypsum substratum  
23 alternating with carbonate units (marls and limestone) and clays (Quirantes, 1978). The  
24 landscape is characterized by low hills and flat-bottomed valleys with altitudes ranging from  
25 127 to around 800 m a.s.l. Hills are occupied mainly by dwarf-scrubs of *Rosmarinus officinalis*  
26 L., while uncultivated valley bottoms are occupied by *Lygeum spartum* L. steppe and scarce

1 scrub of *Salsola vermiculata* L. and *Artemisia herba-alba* Asso (Braun-Blanquet and Bolós,  
2 1957). Land use in the area is based on a traditional agro-pastoral system involving dry cereal  
3 croplands and extensive sheep production.

4 Two different soil management types (ungrazed natural shrubland, N; and grazed shrubland,  
5 GR), applying to two different types of soils (silty gypseous soils, Gy; and non-gypseous soils,  
6 NGy), were compared (Table 1). Two experimental fields per combination of soil type and  
7 grazing treatment were sampled. All fields were located in a nearly flat area. The grazing  
8 treatment consisted, in all cases, of a moderate grazing intensity ( $< 1 \text{ head ha}^{-1} \text{ year}^{-1}$ ) according  
9 to the traditional use in the area (Pueyo, 2005). Measurements of physical and chemical  
10 properties were performed in May-June 2010.

## 11

### 12 *2.2. Soil texture, chemical properties and organic matter content*

13 In each experimental field, the sampling points, four per treatment, were uniformly distributed  
14 along a 50 m-long straight line. All soil samplings (for soil texture, chemical properties and  
15 organic matter content) per experimental field were taken from the 1-10 cm depth soil layer and  
16 stored in a single bag. The samples were homogenised and sieved to 2 mm-size particles in the  
17 subsequent laboratory analysis. The gypsum content was estimated from the sulphur content,  
18 calculated using a LECO 144DR elemental analyser. The carbonate content was measured with  
19 a Variomax CN elemental analyser (Hanau, Germany). The soil texture was measured using the  
20 laser diffraction technique (COULTER LS230). The organic carbon was determined by an  
21 improved chromic-acid digestion and spectrophotometric procedure (Heanes, 1984), and the  
22 results transformed to organic matter (OM) by multiplying by the factor 1.724.

### 23

### 24 *2.3. Soil hydro-physical properties*

25 All measurements of soil bulk density ( $\rho_b$ ), penetration resistance (PR), saturated sorptivity ( $S$ )  
26 and hydraulic conductivity ( $K$ ), and the water retention curve (WRC) for undisturbed soil

1 samples corresponded to the 1-10 cm depth soil layer. The sampling sites, four per experimental  
2 field, were uniformly distributed along a 50 m-long straight line.

3 The  $\rho_b$  was determined by the core method (Grossman and Reinsch, 2002), with core  
4 dimensions of 50 mm diameter and 50 mm height. The soil samples were dried at 50°C for 48 h  
5 and weighed in order to calculate the bulk density. This sampling was also used to determine the  
6 prior volumetric water content needed to calculate  $S$  and  $K$ . One replication of  $\rho_b$  was carried  
7 out per sampling site.

8 The soil penetration resistance (PR) was automatically recorded in the field using a commercial  
9 penetrometer (CP40II Penetrometer). In this case, five replications, close to the infiltration  
10 measurements, were performed per sampling site.

11 The soil hydraulic properties were measured in the field using a tension disc infiltrometer  
12 (Perroux and White, 1988) with a base radius of 50 mm. Infiltration measurements were  
13 performed after removing the crust surface. The base of the disc was covered with a nylon cloth  
14 of 20  $\mu\text{m}$  mesh size, and a thin layer of commercial sand (80–160  $\mu\text{m}$  grain size) was used to  
15 ensure good hydraulic contact between the base of the disc and the soil surface. Only infiltration  
16 measurements at soil saturation conditions, which last up to 15 min, were conducted. Flow  
17 readings were automatically recorded every 5 s from the drop in water level of the water supply  
18 reservoir, using a  $\pm 0.5$  psi pressure transducer that, connected to a datalogger (CR1000,  
19 Campbell Sci.), was installed at the bottom of the water supply reservoir (Casey and Derby,  
20 2002). The method of Vandervaere et al. (2000), which analyses the transient cumulative  
21 infiltration curve, was used to calculate the sorptivity and soil hydraulic conductivity. The final  
22 soil water content, needed to calculate the hydraulic conductivity, was sampled from the upper  
23 centimetres of the soil just after removing the disc infiltrometer from the soil surface. Two  
24 replications were performed per sampling site.

1 The water retention curve (WRC) for undisturbed soil samples was measured in the laboratory  
 2 using a pressure head TDR-cell (Moret-Fernández et al., 2011). The undisturbed soil samples  
 3 were taken with a 50 mm-diameter and 50 mm-high stainless steel core. The volumetric water  
 4 content ( $\theta$ ) was measured by TDR in air-dry soil, which corresponds to a soil suction head of  
 5 about 166 MPa (Munkholm and Kay, 2002), at soil water saturation and at suction heads of 1.7,  
 6 3, 5, 10, 50, 100, 500, and 1500 kPa. Assuming a residual water content equal to zero, the  
 7 WRCs were fitted to the bimodal function proposed by Durner (1992) using the SWRC Fit  
 8 Version 1.2. software (Seki, 2007) (<http://seki.webmasters.gr.jp/swrc/>), according to

$$9 \quad \theta = (\theta_{sat}) \left[ w \left[ \frac{1}{1 + (\alpha_1 \psi)^{m_1}} \right]^{m_1} + (1 - w) \left[ \frac{1}{1 + (\alpha_2 \psi)^{n_2}} \right]^{m_2} \right] \quad (1)$$

$$10 \quad 0 < w < 1$$

$$11 \quad \sum w_i = 1$$

$$12 \quad \alpha_i > 0, m_i > 0, n_i > 1$$

13 where  $n_i$  is the pore-size distribution parameter,  $m_i = 1 - (1/n_i)$ ,  $\alpha_i$  is the scale factor,  $\theta_{sat}$  is the  
 14 saturated volumetric water content, and  $w$  is a weighting factor for the sub-curves. One  
 15 replication of the WRC was performed per sampling site. The same soil cores used to calculate  
 16 the WRC were subsequently dried at 50°C for 48 h and employed to calculate the soil bulk  
 17 density, which allowed additional  $\rho_b$  values to be obtained for each sampling site.

18 To compare the effects of the soil management system on the soil hydro-physical properties, an  
 19 analysis of variance (ANOVA) for a completely randomized design was conducted using SPSS  
 20 (V. 13.0) statistical software. The PR and  $\theta_{sat}$ , and the  $n_i$  variables needed to be normalized  
 21 using the root square and the inverse transformations, respectively. The treatment means were  
 22 compared using Duncan's multiple range test.

23

### 3. Results and discussion

The soils were grouped as gypseous (Gy) (Lomaza and Leciñena) and non-gypseous (NGy) (Planerón and Sariñena), with high and low silty gypsum content, respectively. The taxonomy of the different soils is summarized in Table 2. Soil analysis confirmed that soils of Lomaza and Leciñena had a high content of silty gypsum (Table 2). Soils from the Planerón experimental field, which also had a high gypsum content, were considered non-gypseous since the soil physical properties conferred by the gypsum found in these soils, which crystals were arranged in nodules (up to 2-3 cm diameter), were completely different from the “flour-like” texture observed for the silty gypsum found in Lomaza and Leciñena. The soil texture was sandy loam and loam for the Gy and NGy soils, respectively (Table 2). A qualitative interaction was found in the OM between the soil type and the management. While the N treatment in Gy soils presented 43.3% more OM than the GR treatment, the opposite was found in the NGy soils, where OM under GR was 38.4% higher than under N (Table 2). These results would indicate that, unlike NGy, grazing management in Gy soils has a negative influence on the OM storage. On average, Gy soils present a significantly lower bulk density ( $\rho_b = 1.13 \text{ g cm}^{-3}$ ) than NGy soils ( $1.39 \text{ g cm}^{-3}$ ) (Fig. 2a). These differences can be attributed to the lower particle density of gypsum ( $2.30 \text{ g cm}^{-3}$ ) when compared to the average  $2.62 \text{ g cm}^{-3}$  particle density typically assigned to mineral soils (Hillel, 1998). The  $\rho_b$  measured in both Gy and NGy soils under the GR treatment was significantly higher than the corresponding  $\rho_b$  measured under the N management (Fig. 3a). No significant interaction ( $p = 0.32$ ) was observed in  $\rho_b$  between the soil type and grazing treatment. The lowest  $\rho_b$  value corresponded to the Gy soils under N treatment. As observed by da Silva et al. (2003) and Quiroga et al. (2009), these results indicate that livestock trampling tends to increase the soil bulk density, with consequences for water infiltration, plant root development and seed establishment (Heady and Child, 1994).



1 The volumetric water content in the 1-10 cm soil layer during the PR samplings, calculated  
2 from the soil bulk density measurements, was on average 0.09, 0.07, 0.04, 0.04 m<sup>3</sup> m<sup>-3</sup> for the  
3 Gy-N, Gy-GR, NGy-N and NGy-GR treatments, respectively. The soil penetration resistance  
4 (PR) was significantly higher for NGy soils than the corresponding value measured in Gy soils  
5 (Fig. 2b). A significant interaction was observed between the soil type and grazing treatment ( $p$   
6  $< 0.01$ ). While no statistical differences in PR were observed in the Gy soils between the GR  
7 and N treatments, the PR measured in the NGy soils under GR management was significantly  
8 higher than the corresponding value observed in N (Fig. 3b). The highest PR corresponded to  
9 the NGy soils under the GR treatment. These results indicate that the PR in Gy soils is little  
10 affected by livestock trampling.

11 The average values of saturated hydraulic conductivity ( $K$ ) and sorptivity ( $S$ ) measured in Gy  
12 soils were not statistically different from those calculated for NGy soils (Fig. 2c and d).  
13 However, a significant interaction ( $p < 0.01$ ) was found between the soil type and grazing  
14 treatment in both  $K$  and  $S$  parameters. While no differences in  $K$  and  $S$  were observed between  
15 the N and GR treatments in Gy soils, the grazing treatment was observed to exert a significant  
16 influence in the NGy soils (Fig. 3c and d). This difference can be attributed to the inherent  
17 characteristics of wetted gypsum, which shows a completely different soil structural behaviour  
18 from that observed in NGy soils. During soil wetting, the gypsum that coats the soil macropores  
19 dissolves, and subsequently new gypsum crystals grow, obstructing pre-existing conductive  
20 pores (Poch and Verplancke, 1997). This phenomenon results in a reduction of the water  
21 infiltration rate. This process, which does not take place in many mineral soils, may explain  
22 why Gy soil compaction due to livestock trampling (Fig. 3a and b) does not significantly affect  
23 the infiltration parameters. For the NGy soils, the  $K$  and  $S$  values under the N treatment were  
24 almost four times greater than the corresponding values measured under the GR management.  
25 These results indicate that grazing on NGy soils tends to collapse the soil water conductive

1 macropores of the upper soil layers (Zhao et al., 2007), and consequently reduce the soil water  
2 infiltration.

3 A typical bimodal WRC curve (Eq. 1), which indicates that the studied soils present a clear  
4 double porosity system, was found in all soils and grazing treatments (Fig. 4). Differences  
5 between Gy and NGy were only significant ( $p < 0.05$ ) for the  $\theta_{sat}$  and the weighting factor  $w$ .  
6 The  $w$  was the only WRC parameter that showed significant differences between grazing  
7 treatments and a significant interaction between soil type and grazing treatment. The  $w$   
8 parameter estimated in the Gy soils under N management was significantly higher than the  
9 corresponding value observed under GR (Table 3).

10 All these results suggest that livestock grazing, in both Gy and NGy soils, has a negative effect  
11 on the soil hydro-physical properties. The higher soil compaction due to livestock grazing tends  
12 to slow down the soil water movement through the soil profile. This phenomenon, which leads  
13 to soil water being stored in the upper soil horizons, increases the risk of water loss by  
14 evaporation, and consequently reduces the water available to plant production.

#### 16 **4. Conclusions**

17 This paper compared the effect of livestock grazing on the soil hydro-physical properties for  
18 silty gypseous (Gy) and non-gypseous (NGy) soils. Soil compaction due to grazing treatment  
19 tended to increase  $\rho_b$  and decrease the  $K$  and  $S$  values, resulting in a deterioration of soil hydro-  
20 physical properties. Small differences in the WRCs were observed between soil types and  
21 grazing treatments. In general, the hydro-physical features of Gy soils tend to be less affected by  
22 grazing treatment.

#### 24 **Acknowledgments**

25 This research was supported by Aragón regional government and La Caixa (Grants: GA-  
26 LC020/2010; GA-LC-010/2008) and the CSIC (Grant: PIE-200930I145). The authors are

1 grateful to A. Bielsa, J. Salvador and R. Gracia for their technical help in several aspects of this  
2 study.

#### 4 **References**

5 Boukhris, M., Lossaint, P., 1975. Aspects écologiques de la nutrition minérale des plantes  
6 gypsicoles de Tunisie. *Rev. Écol. Biol. Sol* 2, 329-348.

7 Boyadgiev, T.G., 1974. *Contribution to the Knowledge of Gypsiferous Soils*. FAO, Rome.

8 Braun-Blanquet, J., Bolòs, O., 1957. Les groupements végétaux du Bassin Moyen de l'Ebre et  
9 leur dynamisme. *Anales de la Estación Experimental de Aula Dei* 5, 1-266.

10 Casey, F.X.M., Derby, N.E., 2002. Improved design for an automated tension infiltrometer. *Soil*  
11 *Sci. Soc. Am. J.* 66, 64–67.

12 Chanasyk, D.S., Naeth, M.A., 1995. Grazing impacts on bulk density and penetrometer  
13 resistance in the foothills fescue grasslands of Alberta, Canada. *Can. J. Soil Sci.* 75, 551–  
14 557.

15 Collins, S. L., Knapp, A. K., Briggs, J. M., Blair, J. M., Steinauer, E. E., 1998. Modulation of  
16 diversity by grazing and mowing in native tallgrass prairie. *Science* 280, 745-747.

17 da Silva, A.P, Imhoff,S., Corsi, M., 2003. Evaluation of soil compaction in an irrigated short-  
18 duration grazing system. *Soil Till. Res.* 70, 83–90.

19 Durner, W., 1994. Hydraulic conductivity estimation for soils with heterogeneous pore  
20 structure. *Water Resour. Res.* 30, 211-223.

21 Gijsman, A.J., Thomas, R.J., 1996. Evaluation of some physical properties of an oxisol after  
22 conversion of native Savanna into legume-based on pure grass pastures. *Trop. Grasslands*  
23 30, 237– 248.

1 Grossman, R.B., Reinsch, T.G., 2002. Bulk density and linear extensibility. In, *Methods of Soil*  
2 *Analysis. Part 4.* (Ed. J.H. Dane and G.C. Topp), SSSA Book Series No. 5. Soil Science  
3 Society of America, Madison WI.

4 Hary, I., H. J. Schwartz, V. H. C. Pielert, and C. Mosler., 1996. Land degradation in African  
5 pastoral systems and the destocking controversy. *Ecol. Mod.* 86, 227-233.

6 Heady, H. F., Child, R. D., 1994. *Rangeland ecology and management.* Westview Press,  
7 Oxford.

8 Heanes, D.L., 1984. Determination of total organic-c in soils by an improved chromic-acid  
9 digestion and spectrophotometric procedure. *Communications in Soil Sci. Plant Analysis*  
10 15, 1191-1213.

11 Herrero, J., Boixadera, J., 2002. Gypsic soils. p. 635–639. In R. Lal (ed.) *Encyclopedia of soil*  
12 *science.* Marcel Dekker, New York.

13 Herrero, J., Snyder, R.L., 1997. Aridity and irrigation in Aragón, Spain. *J. Arid Environ.* 35,  
14 535-547.

15 Hillel, D., 1998. *Environmental soil physics.* Academic Press, London, UK.

16 Hoshino, A., Tamura, K., Fujimaki, H., Asano, M., Ose, K., Higashi, T. 2009. Effects of crop  
17 abandonment and grazing exclusion on available soil water and other soil properties in a  
18 semi-arid Mongolian grassland. *Soil Till. Res.* 105, 228–235.

19 Krummelbein, J., Peth, S., Zhao, Y., Horn, R., 2009. Grazing-induced alterations of soil  
20 hydraulic properties and functions in Inner Mongolia, PR China. *J. Plant Nut. Soil Sci.*  
21 172, 769-776.

22 Kumar, S., Anderson, S.H., Udawatta, R.P., Gantzer, C.J., 2010. CT-measured macropores as  
23 affected by agroforestry and grass buffers for grazed pasture systems. *Agroforest Syst.* 79,  
24 59–65.

- 1 M.A.P.A. (Ministerio de Agricultura, Pesca y Alimentación), 1987. Caracterización  
2 agroclimática de la provincia de Zaragoza. Madrid.
- 3 McNaughton, S. J., 1979. Grazing as an optimization process: grass-ungulate relationships in  
4 the Serengeti. *The American Naturalist* 113, 691-703.
- 5 McNaughton, S.J., 1985. Ecology of a grazing ecosystem: the Serengeti. *Ecol. Monographs*  
6 55, 259-294.
- 7 Meyer, S.E., García-Moya, E., 1989. Plant community patterns and soil moisture regime in  
8 gypsum grasslands of north central Mexico. *J. Arid. Environ.* 16, 147-155.
- 9 Milchunas, D. G., O. E. Sala, and W. K. Lauenroth., 1988. A generalized model of the effect of  
10 grazing by large herbivores on grasslands community structure. *The American Naturalist*  
11 130, 168-198.
- 12 Montalvo, J., Casado, M. A., Levassor, C., Pineda, F. D., 1993. Species diversity patterns in  
13 Mediterranean grasslands. *J. Vegetation Sci.* 4, 213-222.
- 14 Moret-Fernández, D., Vicente, J., Latorre B., Herrero, J., Castañeda, C., Lópe, M.V., 2011 TDR  
15 pressure cell for monitoring water content retention curves on undisturbed soil samples.  
16 *Hydrol. Process.* (in press)
- 17 Munkolm, L.J., Kay, B.D., 2002. Effect of Water Regime on Aggregate-tensile Strength,  
18 Rupture Energy, and Friability. *Soil Sci. Soc. Am. J.* 66, 702-709.
- 19 Perroux, K.M., White, I., 1988. Designs for disc permeameters. *Soil Sci. Soc. Am. J.* 52, 1205–  
20 1215.
- 21 Poch, R.M., Verplancke, H., 1997. Penetration resistance of gypsiferous horizons. *Eur. J. Soil*  
22 *Sci.* 48, 535-543.
- 23 Poch, R.M., De Coster, W., Stoops, G., 1998. Pore space characteristics as indicator of soil  
24 behaviour in gypsiferous soils. *Geoderma* 87, 87-109.

1 Pueyo, Y., 2005. Evaluación de los factores ambientales y del uso antrópico como  
2 condicionantes de la conservación de la vegetación del sector central de la depresión del  
3 Ebro. Tesis doctoral, Universidad de Zaragoza.

4 Pueyo, Y., Alados, C. L., Barrantes, O., Komac, B., Rietkerk, M., 2008. Differences in gypsum  
5 plant communities associated with habitat fragmentation and livestock grazing. *Ecol.*  
6 *Appl.* 18, 954–964.

7 Puigdefabregas, J., Mendizabal, T., 1998. Perspectives on desertification: western  
8 Mediterranean. *J. Arid. Environ* 39, 209-224.

9 Quirantes, J., 1978. Estudio sedimentológico y estratigráfico del Terciario continental de los  
10 Monegros. Institución Fernando el Católico, Zaragoza.

11 Quiroga, A., Fernández R., Noellemeier, E., 2009. Grazing effect on soil properties in  
12 conventional and no-till systems. *Soil Till. Res.* 105, 164–170.

13 Seki, K., 2007. SWRC fit – a nonlinear fitting program with a water retention curve for soils  
14 having unimodal and bimodal pore structure. *Hydrol. Earth System Sci. Discus.* 4, 407-  
15 437.

16 Soil Survey Staff, 2010. Keys to soil taxonomy - Eleven edition, USDA, NRCS

17 UNCCD, United Nations Convention to Combat Desertification in those countries experiencing  
18 serious drought and/or desertification, particularly in Africa. 1994. United Nations  
19 Environment Agency (EEA). Office for Official Publications of the European  
20 Communities, Luxemburg.

21 Vandervaere, J.P., Vauclin, M., Elrick, D.E., 2000. Transient Flow from Tension Infiltrimeters.  
22 Part 1. The two-parameter Equation. *Soil Sci. Soc. Am. J.* 64, 1263-1272.

23 Zhao, Y., Peth, S., Krümmelbein, J., Horn, R., Wang, Z., Steffens, M., Hoffmann, C., Peng, X.,  
24 2007. Spatial variability of soil properties affected by grazing intensity in Inner Mongolia  
25 grassland. *Ecol. Model.* 205, 241–254.

## Figure captions

**Figure 1.** Location of the experimental fields in the study area: (A), Lomaza; (B), Planerón; (C), Leciñena; (D), Sariñena. The two dots per experimental field represent the grazed and ungrazed treatments, respectively.

**Figure 2.** Average (a) soil bulk density, (b) penetration resistance, (c) saturated sorptivity ( $S$ ), and (d) saturated hydraulic conductivity ( $K$ ), measured on the 1-10 cm depth soil layers for silty gypseous and non-gypseous soil. Columns within the same soil type with the same letter indicate no significant differences ( $p < 0.05$ ) between treatments. Vertical lines indicate the standard deviation within each treatment.

**Figure 3.** Average (a) soil bulk density, (b) penetration resistance, (c) saturated sorptivity ( $S$ ), and (d) saturated hydraulic conductivity ( $K$ ), measured on the 1-10 cm depth soil layers for silty gypseous and non-gypseous soil under grazed and ungrazed treatments. Columns within the same soil type with the same letter indicate no significant differences ( $p < 0.05$ ) between treatments. Vertical lines indicate the standard deviation within each treatment.

**Figure 4.** Average measured and modelled water retention curves estimated for the 1-10 cm depth soil layers for (a) silty gypseous and (b) non-gypseous soils under grazed and ungrazed treatments. White and grey points denote the average values of the water content-suction head pair of points measured in the grazed and ungrazed treatments. Vertical lines indicate the standard deviation within each suction head.



Figure 1\_Moret-Fernndez et al\_AGWAT3661R2

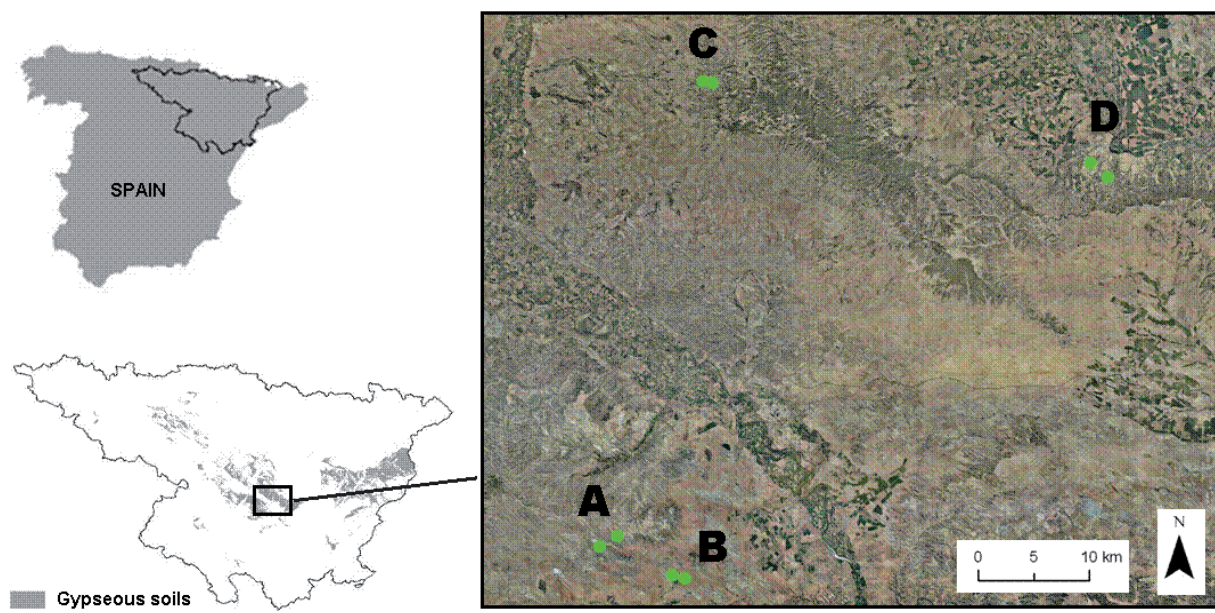


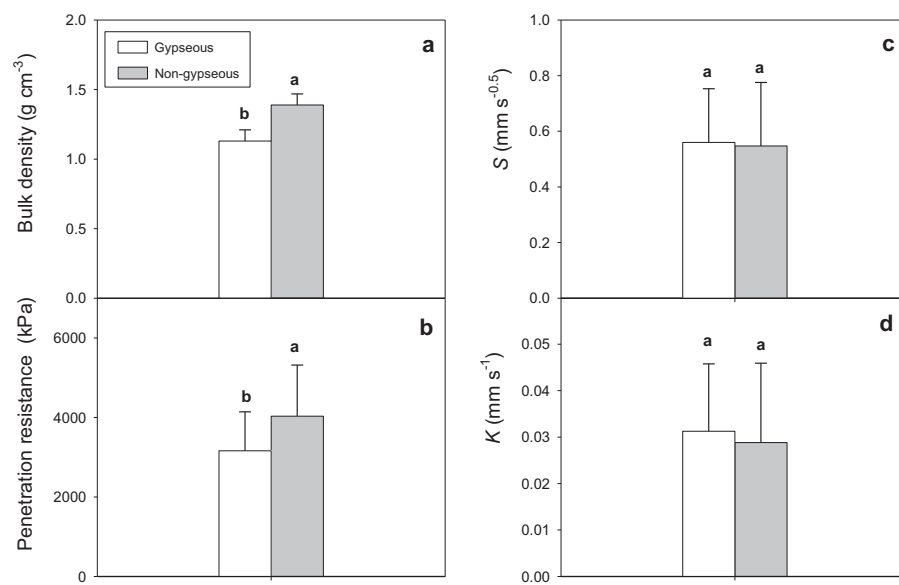
Figure 1.

1



Figure 2\_Moret-Fernndez et al\_AGWAT3661R2

1



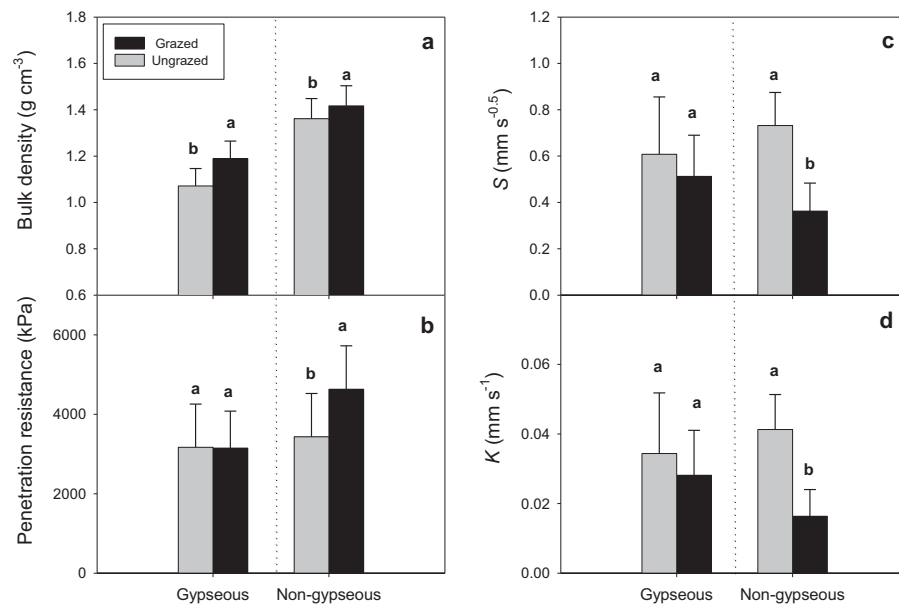
2

3

4 **Figure 2.**

Figure 3\_Moret-Fernndez et al\_AGWAT3661R2

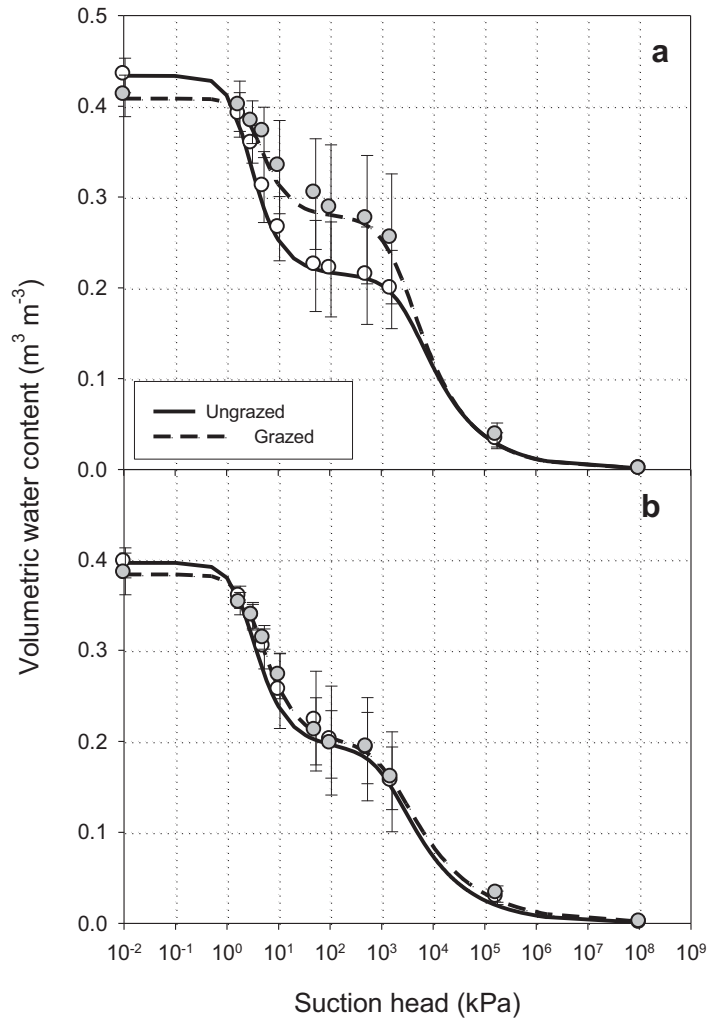
1



2

3 Figure 3.

1



2

3

4 **Figure 4**

5

Table 1\_Moret-Fernndez et al\_AGWAT3661R2

1 **Table 1.** Experimental fields, soil type, grazing treatments, annual cumulative rainfall, and maximal, minimal and  
2 average annual temperature for each experimental field.

Experimental field	Soil type	Grazing treatment	Annual Rainfall (mm) <sup>1</sup>	Average annual temperature (°C) <sup>1</sup>		
				Maximal	Minimal	Average
Lomaza	Gypseous	Grazed / Ungrazed	341.8	20.5	9.3	14.9
Planerón	Non-gypseous	Grazed / Ungrazed	319.3	21.1	9.7	15.4
Leciñena	Gypseous	Grazed / Ungrazed	465.6	19.0	8.3	13.6
Sariñena	Non-gypseous	Grazed / Ungrazed	376.2	20.4	9.1	14.7

3 <sup>1</sup>Atlas Climático de Aragón; <http://anciles.aragon.es/AtlasClimatico>

4

5

6

7