

Spreading recruitment over time to cope with environmental variability

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Received: 26 March 2010 / Accepted: 17 July 2010
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Abstract Seedling establishment is one of the most vulnerable life cycle stages, and a key component for the population dynamics in short-lived plants. In unpredictable environments, timing of emergence is critical for the success of plant performance, and different adaptive bet-hedging strategies have evolved to reduce the risk of failure in recruitment. In this study we describe the spatio-temporal pattern of seedling emergence (overall rate and timing) and survival in four contrasting Mediterranean habitats for *Plantago coronopus*, a small herb with dimorphic seeds. We then explore the importance of spreading germination within years, as well as the role of the two types of seeds from a broader temporal perspective. Populations strongly differed for all recruitment components analyzed in a given year, but this spatial differentiation diluted when a longer period was considered. Apical (smaller) seeds germinated later and in a significantly lower proportion than basal (larger) seeds. Both late emergents and seedlings from apical seeds had lower survival probability in a rainy year. However, our results suggest that in a population

having the lowest production of apical seeds, late emergents coming from apical seeds may constitute a large fraction of yearly recruitment and that their performance was non-significantly different from that of early emergents over the 4-year study period. This study provides evidence of the importance of two related traits (spreading seedling emergence through time by producing dimorphic seeds) as bet-hedging strategies to cope with environmental unpredictability. This is at least partly accomplished by increasing the potential of recruitment in favourable years, instead of buffering such important process in extremely bad years.

Keywords Annual and perennial populations · Bet-hedging · Mediterranean · *Plantago coronopus* · Seed dimorphism · Seedling survival

Introduction

Seedling establishment is a critical stage in the life cycle of plants. It determines the number of individuals entering the next generation, which has important consequences for the demography and evolutionary ecology of the species (Kitajima and Fenner 2000). The importance of recruitment for species' persistence, however, varies among organisms with different longevities (Silvertown et al. 1993; García et al. 2008). Many short-lived species strongly depend on

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58 the short-term success of recruitment, because in the
59 absence of a persistent seed bank, failure of seedling
60 establishment can result in local extinction.

61 Seedling recruitment is also one of the most
62 vulnerable life cycle stages, and often the stage
63 most intimately linked to environmental conditions
64 (Fenner 1987). It has been demonstrated that biotic
65 and abiotic factors strongly affect plant performance
66 during this stage through variation in soil resources,
67 light, drought, pathogens, herbivory or neighbours
68 (de Jong and Klinkhamer 1988; Ostfeld and Canham
69 1993; Herrera 2002; Gustafsson and Ehrlén 2003).
70 Because environmental factors might also vary
71 during the season, timing of emergence is critical
72 for the success of seedling performance (Verdú and
73 Traveset 2005). The larger size reached by early
74 emerging seedlings may entail benefits in terms of
75 increased survival advantages in competitive situa-
76 tions (Dyer et al. 2000), or in arid and Mediterranean
77 environments where species must survive drought
78 periods (Escudero et al. 1999). On the other hand,
79 interactions with negative environmental factors
80 might also be longer for early emergents too, which
81 would favour late emergence (Weekley et al. 2007).
82 Thus, the optimal time of emergence will depend on
83 the particular set of environmental conditions influ-
84 encing seedling performance at each site.

85 Recruitment may also vary considerably in time,
86 and strong temporal variation in vital rates like
87 recruitment has important consequences for life-
88 history evolution and population dynamics (Real
89 1980; Tuljapurkar 1989; Boyce et al. 2006). Under
90 such situation, natural selection usually favours
91 mechanisms that spread the risk of reproduction
92 through time in what is known bet-hedging strategies
93 (Philippi and Seger 1989). Spreading germination
94 within a year, or among several years (through a
95 permanent seed bank) is an example of adaptive bet-
96 hedging strategies (Cohen 1966; Evans et al. 2007;
97 Venable 2007). Although these strategies may be
98 suboptimal in 'average' years, they protect parental
99 fitness from extremely bad years (Mathias and Kisdi
100 2002).

101 Mediterranean ecosystems show a high ecological
102 variability in space and time: in space, as several
103 different types of habitats can be found in areas of
104 Mediterranean climate; in time, because of high
105 temporal unpredictability, particularly in the amount
106 and seasonal pattern of precipitation, alternating dry

and rainy years. Mediterranean regions are therefore
suitable systems to explore the consequences of
spatio-temporal variations in recruitment patterns,
and on the adaptive strategies of the flora to mitigate
such variation. They have been the focus of several
recruitment studies (see for example Herrera et al.
1994; Gómez-Aparicio et al. 2005; Lázaro et al.
2006), but herbs have received considerably less
attention than woody plants (but see Vilà and Lloret
2000; Garrido et al. 2007).

In this study we use a widespread herb mainly
distributed around the Mediterranean region (*Plan-
tago coronopus* L.), as a model system to explore
seedling recruitment in heterogeneous and unpre-
dictable environments. *P. coronopus* is a small herb
occurring in a wide variety of habitats, and the life
cycle can be either annual or short-lived perennial
(Chater and Cartier 1976). Seedling emergence
occurs from autumn to March, usually in terms of
subsequent cohorts associated with rainfall. *P. cor-
onopus* has dimorphic seeds because fruits may
contain two types of seeds that differ in size, shape,
and ecological behaviour (Dowling 1933). Seed
heteromorphism is considered a bet-hedging strategy
because it spreads offspring in space (through
differential dispersal) and/or time (through differ-
ential germination), covering a wider number of
possible environmental scenarios, and reducing the
risk of parental investment (Imbert 2002).

Here we describe the spatio-temporal patterns
of seedling emergence and survival in contrasting
Mediterranean habitats, and ask the following
questions:

- (1) Is the observed spatial variability of recruitment components consistent among years?
- (2) Is temporal variation in seedling emergence within seasons advantageous?
- (3) Do the two types of seeds differ in recruitment characteristics? If so, in which way can a mixed strategy be beneficial?

Materials and methods

Study species

Plantago coronopus L. subsp. *coronopus* (Plantagin-
aceae), the most widespread subspecies across the

151 species geographic range (Hultén and Fries 1986), is
 152 a small herb with leaves arranged in one to a few
 153 rosettes. Flowering starts in March in annual popu-
 154 lations, and lasts until the plant dies in May–June.
 155 The age of adult plants in these populations can thus
 156 vary from 1 to 6 months. Flowering occurs later in
 157 perennial populations (June–July), where in most
 158 individuals do not flower in their first year.

159 Adult plants produce several inflorescences with
 160 numerous flowers arranged into dense spikes. Fruits
 161 are capsules containing two types of seeds: large
 162 basal (up to four, with a thick mucopolysaccharid
 163 coat), and small apical seeds (zero or one per fruit,
 164 with a much thinner coat). The coat of basal seeds
 165 becomes mucilaginous when in contact with water or
 166 humidity, making them stick to soil surface, and
 167 helping roots to penetrate the soil (Dowling 1933;
 168 Schat 1981). Apical seeds, on the contrary, seem to
 169 be released together with the capsule lid, which may
 170 facilitate dispersal by wind or water (Dowling 1933).

171 Populations and study years

172 Four populations in the southern part of the Iberian
 173 Peninsula located in contrasting environments were
 174 selected for the study: shrubland (s), marsh (m), coastal
 175 dune (c) and mountain meadow (m) (Table 1). In two
 176 of them, individuals behave as annuals (A; thereafter
 177 named as A-s and A-m, attending the habitat), while, in
 178 the other cases, they are perennials (P; thereafter
 179 named as P-c and P-m). The climate of the region is
 180 typically Mediterranean: with little temperature fluc-
 181 tuations, 4 months of summer drought (June–Septem-
 182 ber), rainy winters with high inter-annual variability in
 183 amount, and seasonal distribution of precipitation. The
 184 annual populations occur at sites that experience more
 185 extreme temperatures and lower precipitation than the
 186 perennial ones (Table 1). One of the perennial popu-
 187 lations (P-m) has a particularly wet microclimate
 188 because of its position in the mountain facing the
 189 Strait of Gibraltar, where fogs are very frequent.

190 This study lasted from autumn 2003 until summer
 191 2007. In the following, years will be named with the
 192 last two digits of the 2 years involved in the growing
 193 season (for example, 03–04 lasts from September
 194 2003 until August 2004). Years were classified as
 195 ‘rainy’ (03–04), ‘dry’ (04–05) or ‘average’ (05–06
 196 and 06–07) according to their total precipitation
 197 relative to the average for the last 30 years at the

Table 1 Geographic location, main environmental characteristics and sampling variables of the four populations of *P. coronopus* studied in the southern part of the Iberian Peninsula

Population	Life history	Location (m asl)	Habitat	Min–Max temperature (°C)	Yearly precipitation (mm)	No. plots monitored (m of the sides)	No. quadrats (10 × 10 cm ²)	No. early/late emergents labelled in 03–04
A-s	Annual	37°15' N 6°13' W (42)	Shrubland	11–25	558	6 (0.4 × 0.4)	24	112/98
A-m	Annual	36°53' N 6°17' W (2)	Marsh	12–24	578	6 (0.4 × 0.4)	24	105/128
P-c	Perennial	36°02' N 5°38' W (1)	Coastal dunes	15–20	845	3 (1 × 4)	130	112/98
P-m	Perennial	36°06' N 5°32' W (730)	Foggy mountain	11–19	1,045	5 (1 × 1)	25	86/102

198	meteorological station closest to the respective pop-	autumn precipitations to record early emergents	245
199	ulation. Quartiles of the distribution were used to	(November), and during the winter (February) and	246
200	classify dry, average and rainy years.	the spring (April), to record late emergents.	247
201	Sampling method	Data analysis	248
202	Monitoring of seedlings started in the autumn 2003. It	Variation in overall emergence, seedling survival and	249
203	was carried out in several small quadrats ($10 \times$	proportion of seedlings emerging late among the four	250
204	10 cm^2) randomly selected within larger plots. The	populations was examined in a good (rainy) year	251
205	number and size of plots, as well as the number of	(03–04), the only period where data were available	252
206	quadrats, differed among populations according to	from all the studied populations. Population was	253
207	seedling density (Table 1). Populations were repeat-	considered a fixed factor, while plots within popula-	254
208	edly visited over four growing seasons (every	tion, and quadrats within plots, as random effects.	255
209	4–6 weeks during the first year of study, and three	When significant differences among populations were	256
210	times per year in the following years), and the	detected, post-hoc comparisons were conducted (with	257
211	number of new seedlings was counted in each	Bonferroni correction) to find out which populations	258
212	quadrat. The total number of emerging seedlings	differed. Analyses were conducted by fitting <u>gener-</u>	259
213	per quadrat in a year will hereafter be named ‘overall	<u>alised</u> linear mixed models (GLMMs) to the data with	260
214	emergence’. ‘Survival’ was estimated as the propor-	the GLIMMIX procedure of SAS 9.2. In each case,	261
215	tion of seedlings reaching the onset of the seasonal	we chose the error distribution and the link function	262
216	reproductive period at each population (March for	that best fitted our data: poisson error and log link	263
217	annuals, and June for perennials), regardless of	function for seedling counts (overall emergence), and	264
218	whether they flowered or not. ‘Timing of emergence’	binomial error and logit link function for binary	265
219	was classified as early if seedlings germinated right	response variables or proportions (survival, and	266
220	after the first autumn precipitations (October–	proportion of late emergence).	267
221	November) and late when the first autumn precipita-	Differences among populations in overall emer-	268
222	tions recorded later. A total of 7,951 seedlings were	gence, seedling survival, and proportion of seedlings	269
223	monitored during the study.	emerging late were explored again including the	270
224	About 100 early- and 100 late-emerging seedlings	4 years of study, although only three populations	271
225	were individually labelled in each population in the	were used (the A-s was vandalized in the second year	272
226	autumn 2003, and followed until they died (Table 1).	of monitoring). To do so, year (nested within	273
227	To avoid extensive labelling in the field and their	population) was included as another random factor	274
228	accidental losses, in the autumns of 2004–2006, we	into the previous GLMMs, as it was considered	275
229	monitored the fate of seedlings located in the	another source of variation affecting recruitment	276
230	quadrats containing only one kind of seedling (early	within populations. On the other hand, the coeffi-	277
231	or late emergents). The A-s population was destroyed	cients of variation (CV) of the three recruitment	278
232	during the second year of study, and was removed	components above mentioned were also calculated	279
233	from the analysis of temporal variability.	for comparison.	280
234	Differences in the timing of emergence between	Individually labelled seedlings in 2003 were used	281
235	the two types of seeds were examined in field sowing	to test for the differences in survival between early	282
236	experiments. In October 2003 and 2004 (rainy and	and late emergents. We included emergence period,	283
237	dry years, respectively) we sowed basal and apical	population and their interaction as fixed factors in	284
238	seeds in one annual (A-s) and one perennial (P-c)	GLMMs. Plots within populations and quadrats	285
239	population. Groups of 100 or 50 seeds (basal or apical	within plots were considered as random factors. In	286
240	seeds, respectively) were released on sterilized soil	order to explore to what extent results were consistent	287
241	within permeable pots randomly placed among plots.	over time, we compared survival of early and late	288
242	A total of 4,000 seeds (800 basals and 200 apicals in	emergents across 4 years in the P-c population, the	289
243	each population and year) were used for this exper-	only one with quadrats containing only early or late	290
244	iment. Pots were checked three times: after the first	seedlings in the 4 years of monitoring (04–05 was the	291

292 driest year in the decade and the emergence was very
293 reduced compared to 03–04). In this model, year was
294 considered as another random factor.

295 The germination of basal and apical seeds in the
296 hand sowing experiment was examined in terms of
297 overall rates over the study period (g_b and g_a ,
298 respectively) and the proportion of seedlings (origi-
299 nating from each type of seed) that germinated late
300 (g_{bl} and g_{al} , respectively). Because, in these analyses,
301 we are only interested in the intrinsic differences of
302 dimorphic seeds, only the type of seed was consid-
303 ered as a fixed factor in the GLMMs (binomial error
304 and logit link function used), while population and
305 year were included as random factors.

306 Results

307 Overall emergence and survival

308 Our first model tested differences in the amount of
309 seedling emergence and survival among the four
310 contrasted habitats studied in a rainy year (03–04).
311 Populations significantly differed in overall emer-
312 gence (range 5–70 seedlings per quadrat) and
313 seedling survival (Table 2; Fig. 1 left and middle
314 upper row). On average, seedling emergence in
315 annual populations was twice as high as in perennial
316 ones, and also had much higher survival rates.

317 When both spatial and temporal variations were
318 included in the model, the spatial differences found
319 above disappeared (Table 2). The rate of emergence
320 was more variable across years in A-m (CV: 145%)
321 than in the other two perennials (CV for P-c and P-m:
322 74 and 83%, respectively; Fig. 1 left bottom row).
323 Seedling survival was also very variable across years
324 (Fig. 1 middle bottom row), including the possibility
325 of no survival at all in dry years (04–05 in A-m and
326 P-m). The perennial population, P-m, was the most
327 variable one for survival, with a coefficient of
328 variation of 89%, whereas A-m and P-c had values
329 of 77 and 55%, respectively.

330 Emergence timing and survival

331 The proportion of seedlings emerging late also varied
332 significantly among localities (Table 2; Fig. 1 right
333 upper row). In a rainy year (03–04), most of the
334 emergence took place after the first autumn rains, so

335 that less than 30% could be considered as late
336 emergents (except in P-c where the proportion of
337 early and late emergents was equal). This pattern was
338 not consistent among years (Fig. 1 right bottom row,
339 Table 2). Overall, the proportion of late emergents
340 ranged between 10 (A-m in 06–07, and P-c in 04–05)
341 and 68% (P-m in 06–07) of the yearly emergence.

342 In the rainy 03–04 year, late emergents had a
343 significantly lower survival than early emergents
344 (Table 2; Fig. 2). When several years were consid-
345 ered in the P-c population, however, this effect of
346 timing disappeared (Table 2).

Seed heteromorphism and emergence timing

347 Apical seeds germinated in a significantly lower
348 proportion than basal seeds, irrespective of popula-
349 tion and year ($g_a = 20%$, $g_b = 57%$; $N = 48$ pots;
350 $F = 46.54$; $P < 0.0001$). Only 2% of seedlings
351 originating from basal seeds appeared late in the
352 season (average of the two populations and years)
353 compared with 9% of apical seeds ($N = 44$;
354 $F = 19.09$; $P = 0.0003$).

355 To estimate the proportion of early and late
356 emergents in the field coming from basal and apical
357 seeds we used the recruitment at the P-c population
358 and the 2 years of hand sowings, which corresponded
359 to a rainy (03–04) and a dry (04–05) year (it could not
360 be estimated in the A-s population because emer-
361 gence was incomplete for the second year after
362 vandalization). The probability of early-emerging
363 seedlings to come from basal and apical seeds was
364 estimated straightforward from the proportion of each
365 kind of seed released by mother plants ($P_b = 0.92$
366 basal seeds and $P_a = 0.08$ for apical ones, Braza
367 et al. in press) and germination rates early in the
368 season (results from hand sowings each year; Fig. 3).
369 The proportion of new seedlings later in the season
370 (late emergents) coming from each type of seed,
371 however, will depend not only on the specific
372 germination rate of each kind of seed later in the
373 season each year (for example $g_{bl03-04}$ would corre-
374 spond to the germination of basal seeds late in the
375 03–04 year), but also on their availability in the soil
376 after early germination ($1 - g_{be03-04}$). According to
377 that, the vast majority of seedlings found in the field
378 after the first autumn rains come from basal seeds
379 (95–98%, depending on the year; Fig. 3). Later in the
380 season, however, the source of new emerged
381

Table 2 Results of the GLMMs

	Spatial variability				Temporal variability					
	<i>n</i>	Fixed factors	<i>F</i>	Random effects	Variance estimate	<i>n</i>	Fixed factors	<i>F</i>	Random effects	Variance estimate
Seedling emergence	203	Population	11.07 ^{***}	Plot	0.5 (0.21)	720	Population	0.82	Year	1.15 (0.62)
				Quadrat	0.39 (0.06)				Plot	0.68 (0.17)
Seedling survival	176	Population	30.83 ^{***}	Plot	0.42 (0.19)	536	Population	0.75	Quadrat	0.53 (0.05)
				Quadrat	0.54 (0.14)				Year	4.93 (2.9)
Late emergence	172	Population	4.37 [*]	Plot	0.14 (0.08)	552	Population	1.46	Plot	0.87 (0.31)
				Quadrat	0.49 (0.13)				Quadrat	0.83 (0.17)
Survival of late emergents	841	Emergence time	40.45 ^{***}	Plot	1.05 (0.49)	266	Emergence time	1.67	Year	1.36 (0.79)
		Population	17.77 ^{***}	Quadrat	0.39 (0.24)				Plot	0.68 (0.29)
		Interaction	0.2						Quadrat	1.56 (0.2)
									Year	1.04 (1.28)
									Plot	1.18 (0.71)
									Quadrat	1.01 (0.33)

Spatial variability was tested by comparing seedling emergence, survival, and late emergents across a rainy year (03–04), in a total of *n* quadrats randomly placed in four populations, and survival of *n* late emergents. Temporal variability was tested over 4 years of study in three populations (A-m, P-c and P-m) for seedling emergence and survival, and one population (P-c) for survival of late emergents (*n* indicates the number of quadrats monitored). For fixed factors, *F*, statistic and the significance levels are shown (* 0.05 < *P* < 0.01; *** *P* < 0.0001). For random effects, the covariance parameter estimate and standard error are shown in *parenthesis*

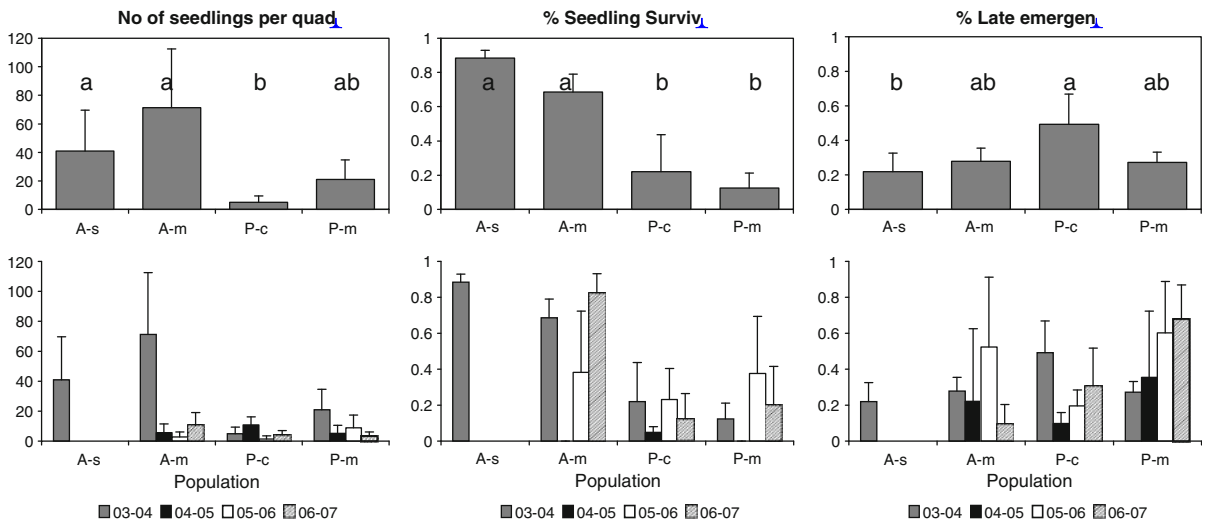


Fig. 1 Components of recruitment in the four populations of *P. coronopus* studied (A-s annual shrubland, A-m annual marsh, P-c perennial coast, P-m perennial mountain) in the rainy year (2003–2004; upper row of histograms; different letters indicate

post-hoc significant differences among populations after Bonferroni correction), and across 4 years (bottom row). Lines on bars correspond to standard deviations of the means

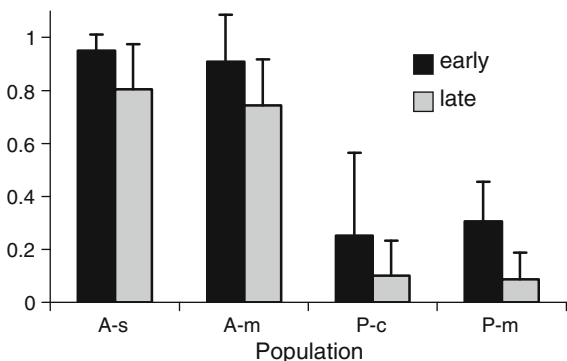


Fig. 2 Survival of early- and late-emerged seedlings in the four populations of *P. coronopus* studied (A-s annual shrubland, A-m annual marsh, P-c perennial coast, P-m perennial mountain) in a rainy year (2003–2004)

nearby localities dramatically differ in the density of emerged seedlings, timing of emergence, as well as in the probability of survival in a given year. The results also show that among-population differences are not consistent among years. This highlights the importance of long-lasting fieldwork studies to make generalizations about recruitment patterns and their variability in space and time, and to understand local adaptation in traits related to germination.

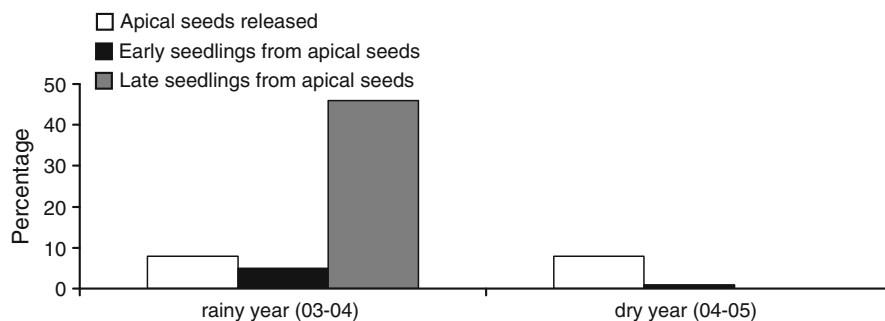
Some spatial differences seem to be associated with the life form. According to the classic life-history theory, annual life forms would be favoured in environments where the probability for a seed to become a flowering plant within one season is greater than the probability for an adult to survive to another flowering season, whereas perennials would be favoured when the contrary occurs (Charnov and Schaffer 1973; Roff 1992). Our results agree with these predictions. At least in favourable years, annuals had higher recruitment success (in terms of seedling emergence and survival) than perennials. Moreover, perennial populations have to survive for one or two harsh summers to become adults. As drought is one of the main causes of seedling mortality in Mediterranean ecosystems (Manzaneda et al. 2005; Garrido et al. 2007; Giménez-Benavides et al. 2007; Rodríguez-Pérez and Traveset 2007), seedling mortality may occur not only during the

seedlings can vary dramatically depending on the precipitation (Fig. 3). In the dry year, no late emergent would come from apical seeds, while in the rainy year, the proportion of late emergents would be almost balanced between both types of seeds (46% of the apical ones).

388 Discussion

389 Recruitment is a crucial process for short-lived
390 plants, and our study on *P. coronopus* shows that

Fig. 3 Proportion of apical seeds produced in the inflorescences of P-c population, and early and late emergents estimated to come from them in two contrasted years (see text for further details)



419 same period as in annuals, but also (mainly) during
 420 the summer, increasing survival differences between
 421 annuals and perennials. In contrast to the association
 422 between recruitment components and life history
 423 reported here, pre-dispersal reproductive components
 424 of this species do not seem to be associated to life
 425 history (fruit set, seeds per fruit, seed weight,
 426 proportion of apical seeds or total seed production
 427 per plant; Braza et al., in press). Thus, pre- and post-
 428 dispersal variables are not necessarily coupled and
 429 driven by the same factors in this widespread plant.
 430 Nevertheless, this study shows that the recruitment
 431 components change from year to year to the extent
 432 that temporal fluctuation may dilute any fingerprint of
 433 the life history.

434 The population of the annual life form fluctuated
 435 more in time than the two perennials for seedling
 436 emergence, while it was one perennial population
 437 which fluctuated most for seedling survival. This is
 438 not a surprising pattern, since seedling emergence
 439 and survival, as well as the advantages of an early or
 440 late emergence, have been found to be highly variable
 441 within species in space and time (Battaglia 1996;
 442 Ibáñez and Schupp 2001; Gómez-Aparicio et al.
 443 2005). Strategies to cope with temporal unpredict-
 444 ability are expected to be of high importance because
 445 strong temporal variation in vital rates like recruit-
 446 ment has important consequences for life-history
 447 evolution and population dynamics (Real 1980;
 448 Tuljapurkar 1989; Boyce et al. 2006). Under such
 449 conditions, short-lived plants like *P. coronopus* are
 450 expected to evolve risk-spreading strategies because
 451 reproduction occurs only once or twice in their
 452 lifetime (Rees 1994).

453 There is a large amount of literature predicting or
 454 documenting the advantages of seed dimorphism as
 455 one of such strategies in disturbed or variable
 456 environments. Different kinds of seeds may offer a

457 wider ecological window in terms of seed bank,
 458 delayed germination, higher tolerance to stressful
 459 conditions, or higher dispersability (see, for example,
 460 Imbert 2002; Redondo-Gómez et al. 2008). Seed
 461 dimorphism also allows populations to fit better local
 462 environments by adjusting different proportions of
 463 dimorphic seeds (Venable et al. 1987; Cheptou et al.
 464 2008), and may have a profound effect on the life
 465 cycle and population regeneration in the successive
 466 years (Mandak and Pysek 2005).

467 Our results suggest two traits representing bet-
 468 hedging strategies in *P. coronopus*, which seem to be
 469 related: the distribution of seedling emergence over
 470 the season, and the production of two kinds of seeds.
 471 Both late emergence and the production of apical
 472 seeds could be interpreted as suboptimal traits. Late-
 473 emerged seedlings showed disadvantage in terms of
 474 survival compared to early emergents for all the four
 475 populations in a rainy year. Emerging late, however,
 476 could not be considered disadvantageous under a
 477 long-term perspective, given that the differential
 478 success of early- and late-emerged seedlings in a
 479 given year disappeared in the coastal population.
 480 Thus, spreading germination within a year in that
 481 case, like spreading it among several years (through a
 482 permanent seed bank), can be considered an example
 483 of adaptive bet-hedging strategies (Cohen 1966;
 484 Evans et al. 2007; Venable 2007).

485 According to our model that combines both kinds
 486 of seeds and their respective germination rates and
 487 emergence timing, apical seeds could constitute an
 488 important source of the late emergents even in
 489 populations where apical seeds are produced in a
 490 very reduced proportion compared to others (P-c;
 491 Braza et al., in press). Germination is a fast process
 492 in this species, given that seeds do not need a dormancy
 493 period, and a couple of days of humidity are enough
 494 to activate root emergence in basal seeds. Thus,

apical seeds could play an important role in local recruitment through delayed emergence, besides the theoretical advantages of long distance dispersal. Bet-hedging strategies, despite not being optimal in most conditions, are thought to be advantageous in some circumstances, protecting parental fitness from extremely bad years (Mathias and Kisdi 2002). In our particular case of study, however, the role of apical seeds (and therefore of late emergents) was higher in the good year than in the bad year in the perennial population where it was investigated, contributing significantly to that year's cohort.

The high unpredictability of seasonal precipitation, and the irreversible effects of droughts in the Mediterranean climate, constitutes severe ecological scenarios for seedling settlement, having dramatic consequences for short-lived herbs. This study provided evidence of the importance of bet-hedging strategies like spreading seedling emergence through time by producing dimorphic seeds, to cope with such limitations. In *P. coronopus*, an organism totally dependent on precipitation to reproduce or to simply survive, this is at least partly accomplished by increasing the potential of recruitment in favourable years instead of buffering such important process in extremely bad years.

Acknowledgments We are indebted to Pedro Jordano and F. Rodríguez for their statistical assistance. A. Simon, P. Colomé, and D. Arquellada helped us during field work. F. Braza, B. Ehlers, J. Ehrlén, and J. Vilellas considerably improved an early version of the manuscript. This study was funded by the Spanish Ministry of Science, under projects BOS2002-01162 and CGL2006-08507 to MBG.

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