

1 **Efficacy of a side-mounted vertically oriented bristle pass for improving upstream**
2 **passage of European eel (*Anguilla anguilla*) and river lamprey (*Lampetra fluviatilis*)**
3 **at an experimental Crump weir.**

4
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15 *barrier.*

16
17 **Abstract**

18 Globally, populations of diadromous anguilliform morphotype fish, such as eel and
19 lamprey, have experienced substantial declines, partly as a result of habitat
20 fragmentation caused by river infrastructure. In the UK, a new configuration of
21 hydraulically unobtrusive bristle pass (side-mounted and vertically oriented) has been
22 developed to help upstream moving European eel (*Anguilla anguilla*) negotiate gauging
23 weirs. The efficacy of vertically oriented bristle passes remains untested, despite their
24 potential as a low-cost low-maintenance solution to improve habitat connectivity at low-

25 head structural barriers worldwide. This study assessed the ability of small (82 – 320
26 mm) and large (322 – 660 mm) European eel and adult (291 – 401 mm) river lamprey
27 (*Lampetra fluviatilis*) to pass upstream over an experimental Crump weir installed in a
28 large open-channel flume with (treatment) and without (control) side-mounted vertically
29 oriented bristle passes under three different hydraulic regimes. Both species were highly
30 motivated to explore their surroundings and move upstream during the trials. Under
31 flooded control conditions, passage efficiency (the total number of times fish passed the
32 structure as a percentage of total attempts) and passage success (the number of fish that
33 passed the structure as a percentage of those that attempted) were high, delay was short,
34 and number of failed attempts before passage was low for both species. When
35 difference in head was at its greatest (230 mm) and velocity and its variation
36 downstream were high (maximum u and σ : 2.43 ms^{-1} and 0.66 ms^{-1} , respectively), the
37 upstream movement of small eel and lamprey was blocked, and passage efficiency and
38 success for large eel low (4.6% and 17.2%, respectively). For large eel that successfully
39 passed, delay was long, and number of failed attempts before upstream passage was
40 high. When bristle passes were installed, passage efficiency for small (91.5%) and large
41 eel (56.7%), and passage success for large eel (76.5%) and lamprey (36.7%) was higher,
42 while delay and the number of attempts before passage was lower for both species.
43 Bristle passes helped European eel and river lamprey pass a small experimental Crump
44 weir, although interspecific variation in efficacy was evident.

1. Introduction

45

46

47 Impacts of infrastructure, such as dams, weirs and barrages, on the physical and
48 chemical processes of rivers are well established (Petts, 1980). Impoundments alter flow
49 and sediment regimes (Nilsson *et al.*, 2005; Xu and Milliman, 2009), channel
50 morphology (Gordon and Meentemeyer, 2006), and nutrient and oxygen availability
51 (Bellanger *et al.*, 2004; Gresh *et al.*, 2000). Ecological impacts include changes in
52 invertebrate communities (Boon, 1988), and for fish the loss of, or reduced access to,
53 critical habitat (Pess *et al.*, 2008), delayed migration (Caudill *et al.*, 2007), population
54 isolation (Morita and Yamamoto, 2002), and reduced productivity and diversity
55 (Agostinho *et al.*, 2008; Matzinger *et al.*, 2007). As a consequence, populations of
56 riverine fish have declined worldwide (Aparicio *et al.*, 2000; Dekker, 2007; Kruk, 2004;
57 Nelson *et al.*, 2002). For diadromous species these declines are often due to impeded
58 migration between essential habitats (Feunteun, 2002; Lucas and Baras, 2001;
59 Ojutkangas *et al.*, 1995; Yoshiyama *et al.*, 1998).

60

61 In an effort to re-establish fluvial connectivity and reverse population declines a range
62 of mitigation strategies have been developed, including the installation of fish passes at
63 structural barriers to migration (Beach, 1984; Clay, 1995; Larinier and Marmulla,
64 2004;). Unfortunately, fish passes, such as those developed for upstream migrating
65 salmonids, often perform poorly for weaker-swimming non-salmonid species (Bunt *et*
66 *al.*, 1999, 2000, 2001; Cooke *et al.*, 2005; Noonan *et al.*, 2012; Slatick and Basham,
67 1985). For example, anguilliform morphotype fish, such as eel (*Anguilla* spp.) and
68 lamprey (e.g. *Lampetra* spp. and *Petromyzon Marinus*), exhibit distinctly different

69 forms of locomotion (Sfakiotakis *et al.*, 1999) and behaviour (Russon and Kemp,
70 2011a), compared to those with a subcarangiform morphology. Although anguilliform
71 morphotypes have good acceleration and are highly manoeuvrable (Muller *et al.*, 2001;
72 Sfakiotakis *et al.*, 1999), they do not leap at barriers and their burst swimming speeds
73 are relatively low (Beamish, 1978; Clough *et al.*, 2004, Russon and Kemp, 2011b;
74 Keefer *et al.*, 2012). Instead, if required, eel and lamprey adopt alternative strategies to
75 ascend obstacles; juvenile eel climb wetted slopes using substrate surface irregularities
76 (Legault, 1988; Tesch, 2003), while lamprey use their oral disk to attach to structures to
77 rest between intermittent bouts of activity (Kemp *et al.*, 2009; Quintella *et al.*, 2004;
78 Russon *et al.*, 2011). In recognition of these adaptations, and in response to
79 environmental legislation (e.g. The Eels [England and Wales] Regulations 2009; CITES;
80 European Habitats Directive [92/43/EEC]; EU Water Framework Directive
81 [2000/60/EC]; Bern convention [COE, 1979]) enacted in an attempt to reverse
82 population declines (Dekker, 2003; Dekker, 2007; ICES, 2012; Kelly and King, 2001;
83 Moriarty and Tesch, 1996; Renaud, 1997), specialist fish passes have been developed
84 and employed for several anguilliform morphotype fishes (Moser *et al.*, 2011; Solomon
85 and Beach, 2004).

86

87 For upstream migrating juvenile eel, specialist fish passes predominantly rely on their
88 ability to climb (Legault, 1988; Tesch, 2003). A variety of substrates have been
89 developed to facilitate climbing (Environment Agency, 2011; Porcher, 2002), including
90 those that incorporate clusters of bristles (usually synthetic), set at regular intervals,
91 protruding from a solid surface (see Environment Agency, 2011). This 'bristled
92 substrate', when used in a traditional configuration (where the base is oriented

93 horizontally, or slightly off horizontal, with water flowing through the bristles), has
94 proved effective at facilitating the upstream passage of a large number (hundreds of
95 thousands per year) (Briand, 2005; Jellyman and Ryan, 1983; Moriaty, 1986) and a
96 broad size range (60-500mm) (Moriarty, 1986, Robinet *et al.*, 2003) of eel worldwide.
97 Further, there is some evidence that lamprey passage can also be enhanced by the
98 judicial use of a bristled substrate (Laine *et al.*, 1998). Bristled substrate is now being
99 used as a cost effective and hydraulically unobtrusive (Environment Agency, 2010)
100 addition to low-head gauging structures, such as Crump weirs (common in the UK), to
101 facilitate the upstream passage of eel (Environment Agency, 2011) and possibly other
102 anguilliform morphotype species. However, to minimise flow interference and negate
103 the need for a separate water source (i.e. as required for 'up and over' installations - see:
104 Environment Agency, 2011), the bristled substrate is oriented vertically and attached
105 with the bristles protruding perpendicularly towards the wing wall of a gauging
106 structure. The efficacy of this configuration of bristle pass is currently untested, despite
107 regional implementation and the recommendation of nationwide deployment in England
108 and Wales (Environment Agency, 2011).

109

110 This study investigated the behaviour of European eel (*Anguilla anguilla*) and European
111 river lamprey (*Lampetra fluviatilis*) as they attempted to pass an unmodified (control),
112 or modified (treatment - with bristle passes installed) Crump weir, under experimental
113 conditions. The experiment was repeated under three hydraulic regimes (low, medium
114 and high velocity) that represent flow conditions similar to those encountered at Crump
115 weirs in the field (see: National River Flow Archive). Passage and delay were quantified
116 and the influence of hydraulic regime and treatment assessed.

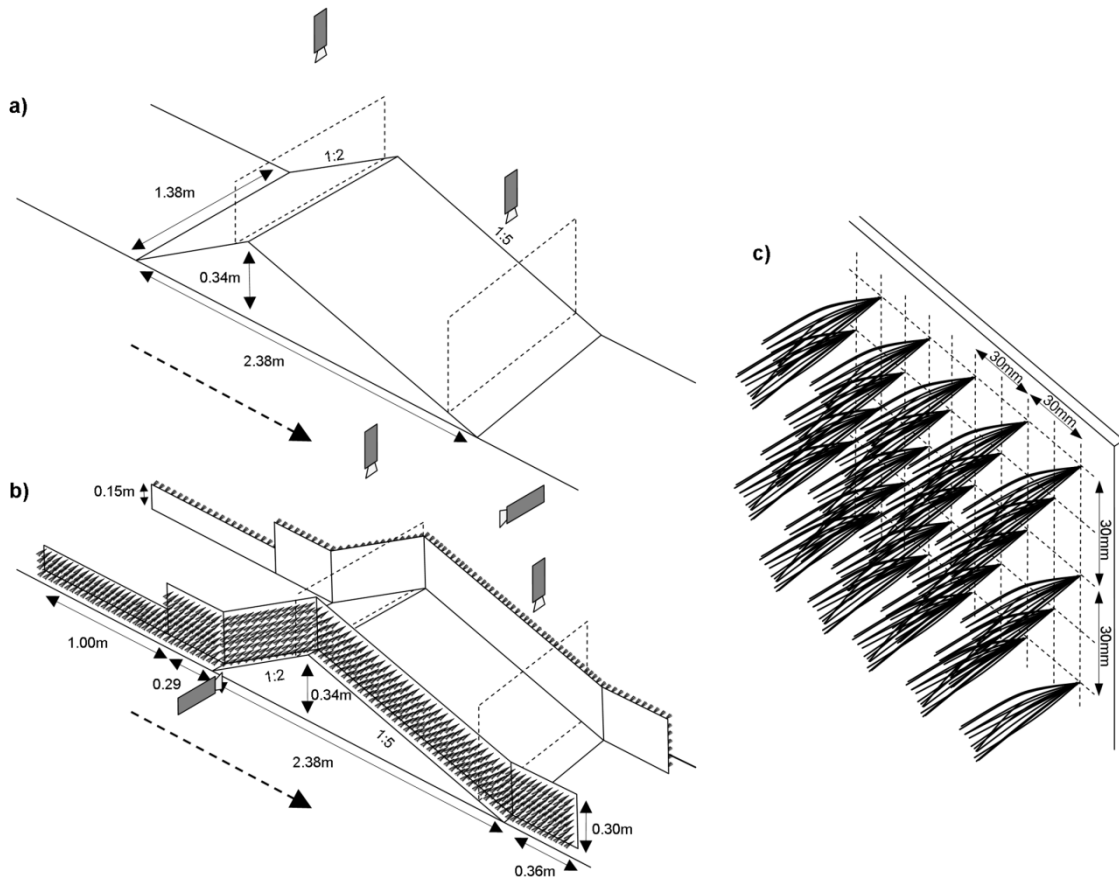
117 **2. Methodology**

118

119 **2.1. Experimental setup**

120

121 A model Crump weir (2.38 m long, 1.38 m wide and 0.34 m high) (Figure 1a) was
122 installed midway along an indoor recirculating flume (21.40 m long, 1.38 m wide, and
123 0.60 m deep) at the International Centre for Ecohydraulics Research (ICER) facility,
124 University of Southampton, UK (50° 57'42.6"N, 1°25'26.9"W). A 14 m long
125 experimental area, sectioned off from the rest of the channel by flow straightening
126 devices (100 mm thick polycarbonate screens with elongated tubular porosity - 7 mm
127 diameter), extended 7 m either side of the weir crest. Under treatment conditions,
128 vertically oriented bristle passes (10 mm thick polypropylene board covered with 30
129 mm spaced orthogonally oriented clusters of *ca.* 24 synthetic fibres [70 mm long x 1.5
130 mm diameter]) were attached with bristles protruding towards the flume wall on each
131 side of the channel (Figure 1b, c). The bristled substrate was installed in accordance
132 with Environment Agency guidelines to maintain a 70 mm cavity (equal to bristle
133 length) between the bristle board and flume wall (see: Environment Agency, 2011).



134
 135 **Figure 1.** The Crump weir under control (a) and treatment (b) setups during which a
 136 bristled substrate (c) was vertically positioned against the channel walls to aid upstream
 137 movement of European eel and river lamprey under various hydraulic conditions. In a
 138 and b dashed lines indicate the position of half-duplex Passive Integrated Transponder
 139 (PIT) antennae coils and the dashed arrows indicates direction of flow.

140
 141 Experiments were conducted under three hydraulic regimes: high (HV), medium (MV)
 142 and low velocity (LV) (Figure 2), created by altering the downstream water level (depth:
 143 220, 330 and 450 mm, respectively) by adjusting an overshoot weir (located at the
 144 downstream end of the channel), under a constant discharge ($0.09 \text{ m}^3 \text{ s}^{-1}$). The HV and
 145 MV regimes were within the modular limits of the experimental weir with upstream
 146 water level (depth: 450 mm) independent of that downstream. The LV regime was

147 outside the modular limits of the weir (flooded conditions - upstream water depth: 455
148 mm). As such, head difference under the HV, MV and LV regime was 230, 120, and 5
149 mm, respectively. Velocities were measured using an Acoustic Doppler Velocimeter
150 (ADV) (Vectrino, Nortek-AS, Norway - frequency 50 Hz, sample volume 0.05 cm³,
151 record length 60 sec), and mean velocity ($V = \sqrt{\bar{u}^2 + \bar{v}^2 + \bar{w}^2}$) and standard deviation
152 ($S.D. = \sqrt{\sigma_u^2 + \sigma_v^2 + \sigma_w^2}$) calculated. Where u , v and w are the instantaneous
153 velocity values corresponding to the x , y and z spatial coordinates, overbar denotes
154 time-average, and σ is the standard deviation of its subscript. $S.D.$ was used as a proxy
155 for the intensity of turbulence. In conditions that precluded using the ADV, i.e. when
156 depth was < 60 mm or air entrainment was high, an electromagnetic flow meter (Model
157 801 Flat, Valeport, UK - frequency 1 Hz, record length 30 sec) was used to measure V
158 and $S.D.$. Spatial maps of the hydraulics associated with the Crump weir were generated
159 in ArcMap v10 (Esri, USA) using a spline interpolation.

160

161 The velocity at the crest of the weir was similar under each regime (*ca.* 0.83 m s⁻¹)
162 (Figure 2). Maximum velocity (2.43, 1.91, and 0.80 m s⁻¹ under the HV, MV, and LV
163 regimes, respectively) was inversely related to head difference (Figure 2) and occurred
164 at the weir crest under the LV and just upstream of the hydraulic jump under the MV
165 and HV regime (Figure 2). The hydraulic jump consisted of a standing wave generated
166 as the super-critical flow along the face of the weir rapidly decelerated on reaching the
167 downstream water level. Despite flooded conditions under the LV regime, a small
168 hydraulic jump occurred *ca.* 100 - 150 mm downstream of the weir crest (Figure 2).
169 Downstream of the hydraulic jump, under all regimes, velocity gradually decreased as
170 the channel deepened (Figure 2).

171

172 Upstream of the weir the intensity of turbulence was low and similar under each regime

173 ($S.D. = ca. 0.05 \text{ m s}^{-1}$). High intensities of turbulence, relative to maximum velocity,

174 were generated at the hydraulic jump ($S.D. = 0.66, 0.27$ and 0.17 ms^{-1} under the HV,

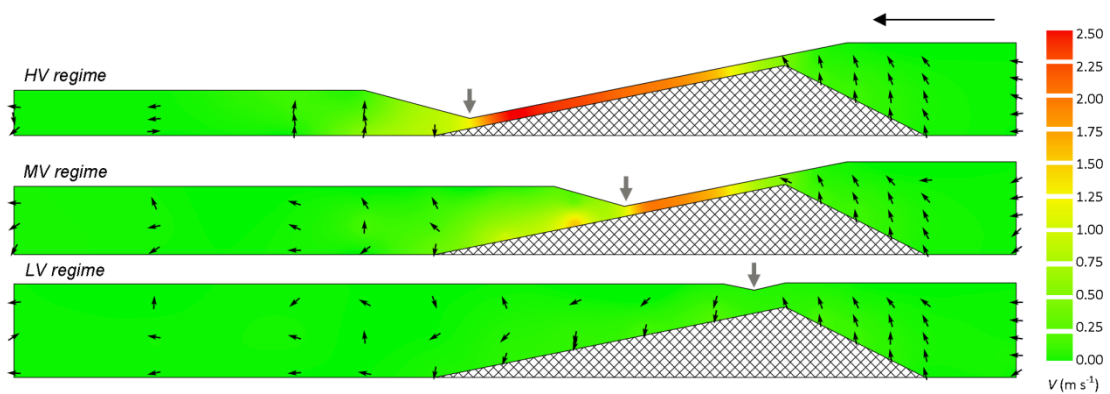
175 MV and LV regime, respectively), and gradually dissipated with distance downstream.

176 At the extent of the hydraulically mapped region (3.74 m downstream of the weir crest),

177 turbulence had almost returned to background levels ($S.D. = 0.10, 0.08$ and 0.05 m s^{-1}

178 under the HV, MV and LV regime, respectively).

179



180

181 **Figure 2.** Velocity (m s^{-1}) profiles for a Crump weir under low (LV), medium (MV) and

182 high (HV) velocity regimes. Small and large black arrows indicate mean and bulk flow

183 direction, respectively. Grey arrows indicate position of a hydraulic jump.

184

185 2.2. Experimental procedure

186

187 Yellow phase European eel were collected by electric fishing from the Rivers Itchen (50°

188 $57' 19.2'' \text{ N}$, $1^\circ 20' 15.8'' \text{ W}$, $N = 208$, Total Length [TL]: $\mu = 397 \text{ mm}$, $\sigma = 108 \text{ mm}$,

189 Range = $149 - 660 \text{ mm}$), Wallington ($50^\circ 51' 45.4'' \text{ N}$, $1^\circ 09' 54.5'' \text{ W}$, $N = 31$, TL: $\mu =$

190 277 mm , $\sigma = 58 \text{ mm}$, Range = $111 - 386 \text{ mm}$) and Meon ($50^\circ 53' 53.2'' \text{ N}$, $1^\circ 11' 14.3''$

191 W, $N = 32$, TL: $\mu = 178$ mm, $\sigma = 72$ mm, Range = 82 – 333 mm) by the Environment
192 Agency between 1 May and 12 July 2011. Actively migrating adult river lamprey were
193 trapped in the River Ouse ($53^{\circ} 53' 26.2''N$, $1^{\circ} 5' 36.8''W$) by a commercial fisherman
194 on 4 December 2012 ($N = 96$, TL: $\mu = 358$ mm, $\sigma = 21$ mm, Range = 291 – 401 mm).
195 Fish were transported to the ICER facility in sealed polyurethane bags (river water and
196 pure oxygen atmosphere - eels) or transportation tanks (aerated river water - lamprey)
197 and held in separate 3000 litre outdoor holding tanks (aerated and filtered, 50% weekly
198 water change) at ambient temperature ($\mu = 16.2$ °C, $\sigma = 1.9$ and $\mu = 7.6$ °C, $\sigma = 3.1$ for
199 eel and lamprey, respectively). All fish were acclimated to holding tank conditions over
200 2 hours via gradual water exchange. Eel >320 mm TL and all lamprey were tagged,
201 under anaesthetic (2-Phenoxy-1-ethanol, 1 ml l⁻¹), with half-duplex Passive Integrated
202 Transponder (PIT) tags (23 mm and 12 mm long, respectively) inserted through a small
203 mid-ventral incision in the posterior quarter of the peritoneal cavity (mortality 0%, tag
204 retention 99.6%). Large eel and lamprey were weighed and measured during the tagging
205 procedure and allowed at least 48 hours to recover from surgery before being used in
206 experiments.

207

208 Treatment replicates were undertaken with multiple small (82 - 320 mm TL) or large
209 (322 - 660 mm TL) eel between the 3 May and 21 July 2011 (temperature: $\mu = 16.2^{\circ}C$,
210 $\sigma = 0.8$) or lamprey (291 - 401 mm TL) between the 24 January and 7 February 2013
211 (temperature: $\mu = 8.8^{\circ}C$, $\sigma = 1.5$) (Table 1). Timings and temperatures were
212 representative of peak migration periods for both species (lamprey: Jang and Lucas,
213 2005; eel: Moriarty, 1986). The duration between capture and experimentation ranged
214 from 2 - 17 and 51 - 65 days for eels and lamprey, respectively. Each replicate lasted 5.5

215 hours and was undertaken at night (23:00 - 04:30) (<0.1 lux) to coincide with peak eel
 216 and lamprey activity (eel: Haro and Kynard, 1997; Laffaille *et al.*, 2007; Tesch, 2003,
 217 lamprey: Kelly and King, 2001; Moser *et al.*, 2002). Fish were acclimated to flume
 218 conditions in a porous container in the channel for 1 hour (22:00 - 23:00) before release
 219 into the experimental area 3 metres upstream of the downstream screen. Small eel were
 220 weighed and measured under anaesthetic (2-Phenoxy-1-ethanol, 1ml l⁻¹) after each
 221 replicate. Each fish was used only once during the study. Due to limited fish availability,
 222 passage experiments with lamprey were conducted only under the LV and HV regime.
 223 Temperature increase during experiments due to the pumps was small for both eel ($\mu =$
 224 0.4°C, $\sigma = 0.5$) and lamprey ($\mu = 0.6^\circ\text{C}$, $\sigma = 0.5$).

225

226 **Table 1.** Conditions encountered by European eel and European river lamprey during
 227 passage over a model Crump weir installed in a recirculating flume under either a high
 228 (HV), medium (MV) or low (LV) velocity regime with (treatment) or without (control)
 229 bristle passes installed during 2011 (eel) and 2013 (lamprey). *N* is the number of fish
 230 used per trial.

Date	Hydraulic regime	Setup	Water depth (mm) ^a		Maximum velocity (m s ⁻¹)	Maximum S.D. of velocity (m s ⁻¹)	Mean water temp (°C)	N	Length range (mm)	PIT tagged
			Upstream	Downstream						
<i>Small European eel</i>										
9 May	HV	Control	450	220	2.43	0.66	16.5	10	195-290	No
10 May	MV	Control	450	330	1.91	0.27	16.8	10	215-317	No
11 May	LV	Control	455	450	0.81	0.17	16.6	10	149-314	No
7 June	LV	Control	455	450	0.81	0.17	15.4	10	220-302	No
8 June	MV	Control	450	330	1.91	0.27	15.8	10	149-290	No
21 June	HV	Treatment	450	220	2.43	0.66	16.0	8	222-297	No
15 July	HV	Control	450	220	2.43	0.66	17.5	10	113-290	No
17 July	HV	Treatment	450	220	2.43	0.66	17.5	12	82-315	No
18 July	MV	Treatment	450	330	1.91	0.27	17.2	10	98-320	No
19 July	LV	Treatment	455	450	0.81	0.17	17.1	10	111-315	No
20 July	MV	Treatment	450	330	1.91	0.27	17.2	10	211-317	No
21 July	LV	Treatment	455	450	0.81	0.17	17.2	10	205-320	No
<i>Large European eel</i>										

3 May	LV	Control	455	450	0.81	0.17	14.7	10	437-660	Yes
4 May	MV	Control	450	330	1.91	0.27	15.0	10	361-582	Yes
8 May	HV	Control	450	220	2.43	0.66	16.2	10	366-575	Yes
12 May	LV	Control	455	450	0.81	0.17	16.3	10	360-585	Yes
16 May	MV	Control	450	330	1.91	0.27	15.3	10	357-630	Yes
17 May	HV	Control	450	220	2.43	0.66	15.9	10	365-540	Yes
18 May	MV	Control	450	330	1.91	0.27	15.8	10	325-481	Yes
19 May	LV	Control	455	450	0.81	0.17	16.3	10	333-501	Yes
9 June	HV	Control	450	220	2.43	0.66	15.8	10	347-549	Yes
13 June	HV	Treatment	450	220	2.43	0.66	15.1	10	405-544	Yes
14 June	HV	Treatment	450	220	2.43	0.66	15.9	10	322-585	Yes
15 June	MV	Treatment	450	330	1.91	0.27	16.6	10	335-543	Yes
16 June	MV	Treatment	450	330	1.91	0.27	16.7	10	373-520	Yes
19 June	HV	Treatment	450	220	2.43	0.66	15.6	10	326-510	Yes
22 June	HV	Treatment	450	220	2.43	0.66	16.2	10	338-537	Yes

River lamprey

24 January	HV	Treatment	450	220	2.43	0.66	5.5	8	329-384	Yes
26 January	HV	Control	450	220	2.43	0.66	6.8	8	320-395	Yes
27 January	LV	Control	455	450	0.81	0.17	7.7	8	320-379	Yes
28 January	LV	Treatment	455	450	0.81	0.17	8.7	8	320-373	Yes
29 January	HV	Treatment	450	220	2.43	0.66	10.2	8	338-401	Yes
30 January	HV	Control	450	220	2.43	0.66	10.6	8	340-388	Yes
31 January	HV	Control	450	220	2.43	0.66	10.6	8	339-395	Yes
1 February	LV	Control	455	450	0.81	0.17	10.2	8	291-388	Yes
2 February	LV	Treatment	455	450	0.81	0.17	9.2	8	322-379	Yes
3 February	HV	Treatment	450	220	2.43	0.66	9.0	8	314-391	Yes
4 February	LV	Treatment	455	450	0.81	0.17	9.3	8	324-371	Yes
6 February	LV	Control	455	450	0.81	0.17	7.6	8	327-388	Yes

231 a: Measured 5 metres upstream or downstream of the weir crest.

232

233 Due to staggered eel availability, source location could not be randomised among
234 treatments. For the purpose of this study it was assumed that there were no differences
235 in behaviour / swimming ability among sources. Mean water temperature did not differ
236 among treatments for any group. Mean TL did not differ among treatments for small
237 and large eel. Despite random allocation, the mean TL of lamprey differed among
238 treatments (one-way ANOVA: $F(3, 8) = 4.578$, $p < 0.05$), being higher under the HV
239 control. Across treatment comparisons were considered acceptable as the difference was
240 deemed small from a biological perspective (8.7 mm).

241

242 2.3. Fish behaviour

243

244 Fish behaviour was monitored using 2-4 low-light digital video cameras (AV-TECH
245 Sony Effio 580TVL CCD) under infrared illumination, enabling visual assessment of
246 movement and differentiation of route selection by individuals. The field of view of the
247 two overhead cameras (control + treatment conditions) spanned the width of the flume
248 at the crest and downstream extent of the weir. The two side cameras (treatment
249 conditions only) monitored fish movement in the bristle passes at the crest of the weir
250 through the glass walls (for camera locations, see Figure 1). Video footage was recorded
251 and reviewed using split-screen multi-channel acquisition and playback software
252 (NUUO ltd., Taiwan). Individual large eel or lamprey were identified during movement
253 over the weir using Half Duplex PIT telemetry (antennae installed at the trailing edge
254 and crest of the weir, Figure 1a, b). Each antenna (3 coils of 2.5 mm² stranded 0.25 mm
255 copper wire) was connected to a PIT detection system incorporating a single reader and
256 two external dynamic tuning units (DEC-HDX-MUX-LOG 134.2 kHz, Wyre Micro
257 Design Ltd., UK), powered using a 110Ah 12v leisure battery, and connected to an
258 external data logger (AntiLog RS232, Anticyclone Systems Ltd., UK). The antenna
259 wiring was attached directly to the face of the weir and had minimal impact on flow due
260 to its low profile. The PIT system was tested by ensuring that tags (either size) held in a
261 clenched fist were consistently detected when passed through each loop at any angle or
262 location.

263

264 For each replicate the video footage and/or PIT data were interrogated and relevant
265 passage events recorded (Table 2). As fish could move freely both up and downstream
266 of the weir throughout the experimental period, multiple upstream passage events per
267 fish were possible during each replicate. Based on the passage events the following

268 metrics were calculated for all fish groups: 1) *number of failed attempts*, 2) *number of*
 269 *upstream passes*, 3) *bristle pass use*, and 4) *passage efficiency* (Table 2). For large eel
 270 and lamprey, to which passage events could be attributed to individual tagged fish,
 271 additional metrics were calculated: 5) *percentage attempts*, 6) *passage success*, 7)
 272 *number of attempts before upstream passage*, and 8) *delay* (Table 2). Tagged fish not
 273 detected at the downstream PIT antenna during the experiments (3 lamprey: 2 LV
 274 treatment, 1 LV control), were considered not to have explored their surroundings or
 275 sampled treatment conditions, and were not included in these metrics. For lamprey,
 276 which have the ability to attach to surfaces using their oral disc (Kemp *et al.*, 2009),
 277 specific attachment metrics were also calculated: 9) *percentage attachment*, 10) *number*
 278 *of attachments*, and 11) *mean duration of attachment* (Table 2).

279
 280 **Table 2.** Definition of the passage events and metrics obtained for the small eel, large
 281 eel (*LE*), and/or lamprey (*L*) as they passed over an experimental Crump weir, and the
 282 statistical tests used.

Event/metric	Definition	Group	Statistical test for variable:	
			Hydraulic regime	Treatment
Events				
Attempt	<i>Progression upstream, of any part of the body onto the downstream face of the weir upstream of the hydraulic jump.</i>	All	N/A	
Upstream pass over the weir	<i>Passage of whole body upstream beyond the weir crest.</i>	All	N/A	
Upstream pass via a bristle pass	<i>Passage of whole body upstream beyond the weir crest via a bristle pass.</i>	All	N/A	
Attachment	<i>Attachment using oral disk on the downstream face of the weir upstream of the hydraulic jump.</i>	L	N/A	
Metrics				
1. Number of failed attempts	<i>Total number of attempts not resulting in upstream passage normalised by the number of fish per replicate.</i>	ALL	One-way ANOVA ^a	Student t tests
2. Number of upstream passes	<i>Total number of upstream passes normalised by the number of fish per replicate.</i>	ALL	One-way ANOVA ^a	Student t tests
3. Bristle pass use	<i>Quotient of the number of upstream passes via a bristle pass and total number of upstream passes per replicate.</i>	ALL	Not assessed	Not assessed
4. Passage efficiency	<i>Total number of times fish passed the weir as a percentage of total attempts per replicate.</i>	ALL	One-way ANOVA ^a	Student t tests

5. Percentage attempts	Number of fish that attempted as a percentage of the total per treatment.	LE, L	Pearson's Chi-square (χ^2) tests. ^b
6. Passage success	Number of fish that passed the weir as a percentage of those that attempted per treatment.	LE, L	Pearson's Chi-square (χ^2) tests. ^b
7. Number of attempts before upstream passage	Number of attempts before first upstream passage event for each fish.	LE, L	Discrete-time hazard model (Logit function) and the Wald statistic (W). ^c
8. Delay	Time between the first detection at the downstream PIT antennae and first upstream passage for each fish.	LE, L	Kaplan-Meier product-limit estimator and the Log Rank (Mantel-Cox) statistic (χ^2_{ml}). ^c
9. Percentage attachments	Total number of fish that attached as a percentage of the total that attempted per treatment.	L	Pearson's Chi-square (χ^2) tests. ^b
10. Number of attachments	Number of attachments normalised by the number of fish per replicate.	L	Two-way ANOVA
11. Mean duration of attachment	Quotient of total duration and number of attachments per replicate.	L	Two-way ANOVA

283 a: Brown and Forsyth F ratio used in cases that violated homogeneity of variance.

284 b: Fisher's exact tests (FET) used if expected frequencies were < 5.

285 c: Event time analysis (Singer and Willet, 2003).

286

287 Discrete attempts (see Table 2) were delineated by a fish drifting back downstream of
288 the hydraulic jump for > 1 second or by an attachment (see Table 2) on the downstream
289 face of the weir (lamprey only). Any further upstream progression observed on the
290 downstream face of the weir was considered a separate attempt as it involved an
291 observable increase in swimming speed to counter the high velocity flow. All statistical
292 analysis was undertaken in SPSS v20 (IBM, USA). Due to low replicate numbers it was
293 not possible to assess interaction effects. Hence, the influence of hydraulic regime was
294 assessed under control conditions only and the influence of treatment was assessed
295 separately under each hydraulic regime. Percentage data were arcsine square root
296 transformed prior to statistical analysis (*see*: Sokal and Rohlf, 1995). *Delay* and *number*
297 *of attempts before upstream passage* were assessed using time to event analysis (Singer
298 and Willet, 2003) (Table 2). This method provides unbiased estimates by including fish
299 that fail to pass the weir (right-censored individuals) in a probability function
300 (Cumulative Probability of Passage [CPP]) at any given time or number of attempts (*see*:
301 Castro-Santos and Haro, 2003).

302 **3. Results**

303

304 A high percentage of the observed passage events were detected by the PIT system
305 (Large eel: 97.2%, Lamprey: 93.0%) allowing identification of the majority of
306 individuals. Passage events with no directly associated PIT data were assigned to
307 individuals with a high degree of confidence by assessing historic and future detections
308 combined with visual tracking of the fish over time.

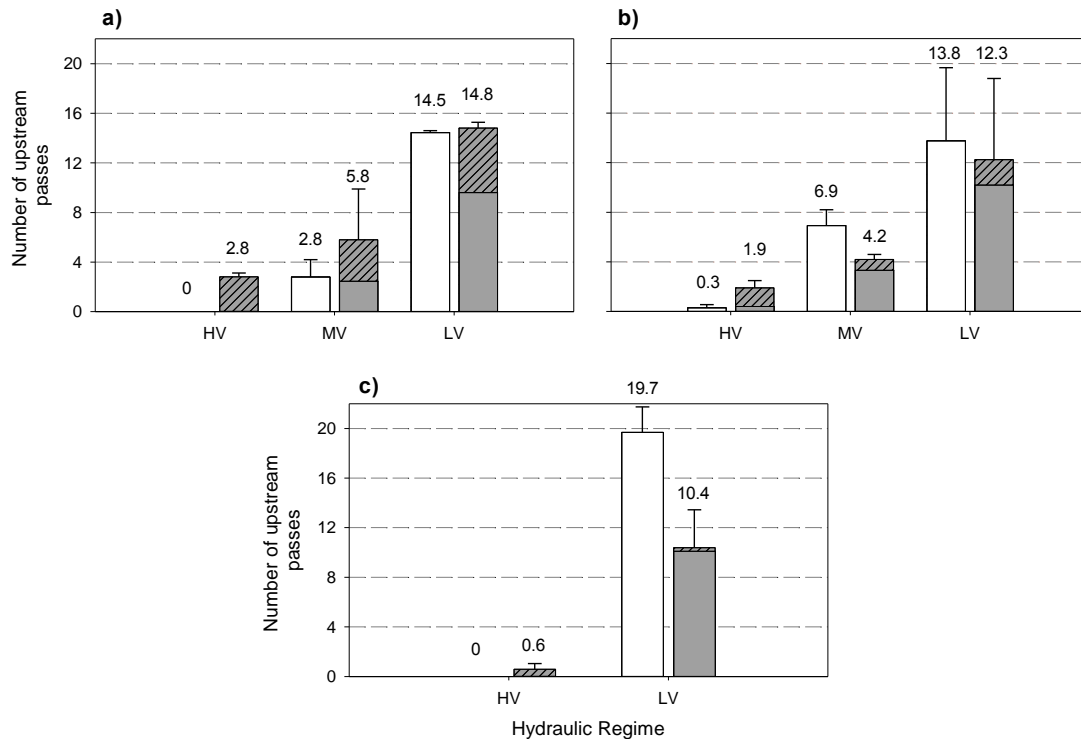
309

310 *Number of failed attempts* was not influenced by hydraulic regime or treatment for any
311 group ($\mu \pm$ S.E.: small eel = 1.87 ± 0.64 , large eel = 3.74 ± 1.10 , and lamprey = $5.24 \pm$
312 1.47).

313

314 *Number of upstream passes* was negatively related to maximum velocity for all groups
315 (small eel: $F(1, 3) = 157.984, p < 0.01$, large eel: $F(1, 6) = 19.020, p < 0.01$, and
316 lamprey $F(1, 4) = 91.240, p < 0.01$), but was not influenced by treatment (Figure 3).

317



318

319 **Figure 3.** Mean number of upstream passes per fish for (a) small eel, (b) large eel, and
 320 (c) lamprey without (control: clear bars) and with (treatment: grey bars) bristle passes
 321 installed under the low (LV), medium (MV), and high (HV) velocity hydraulic regimes.
 322 Hatched sections of the grey bars indicate the proportion of upstream passes that
 323 occurred via the bristle passes. Error bars represent ± 1 S.E..

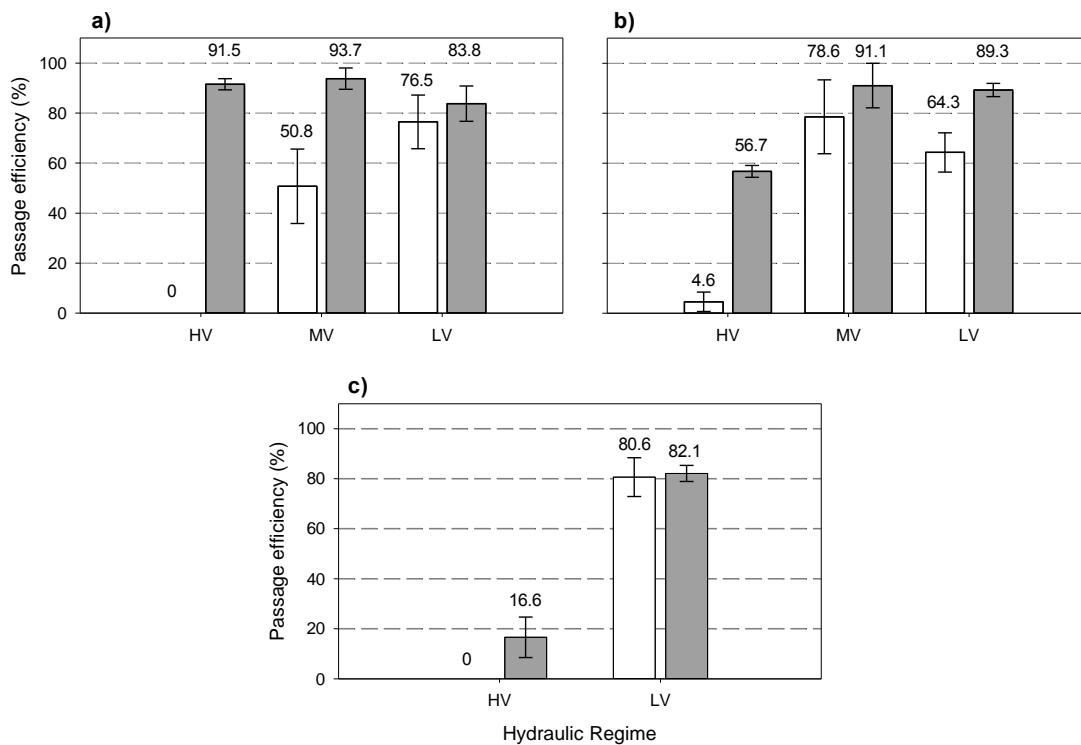
324

325 *Bristle pass use* ($\mu \pm$ S.E.) was highest under the HV, and lowest under the LV regime
 326 for small eel ($100 \pm 0.0\%$; $35.1 \pm 6.0\%$), large eel ($78.3 \pm 6.3\%$; $16.7 \pm 6.1\%$), and
 327 lamprey ($100 \pm 0.0\%$; $2.6 \pm 1.1\%$) (Figure 3).

328

329 *Passage efficiency* was negatively related to maximum velocity for small eel ($F(1, 3) =$
 330 43.841 , $p < 0.01$), large eel ($F(1, 5) = 24.961$, $p < 0.01$) and lamprey ($F(1, 4) = 145.462$,
 331 $p < 0.001$) (Figure 4). Under the HV regime, passage efficiency was higher for small

332 (91.5%; $t(1) = -31.658$, $p < 0.05$) and large eel (56.7%; $t(3) = -5.057$, $p < 0.05$) when the
 333 bristle passes were installed (Figure 4). Treatment did not significantly influence
 334 *passage efficiency* for lamprey under the HV regime, or for any group under the MV or
 335 LV regime.
 336



337
 338 **Figure 4.** Mean *passage efficiency* (%) for (a) small eel, (b) large eel, and (c) lamprey
 339 without (control: clear bars) and with (treatment: grey bars) bristle passes installed
 340 under the low (LV), medium (MV), and high (HV) velocity hydraulic regimes. Error
 341 bars represent ± 1 S.E..

342
 343 *Percentage attempts* for large eel was not influenced by hydraulic regime or treatment,
 344 and was consistently high (>85%). For lamprey, *percentage attempts* was not influenced

345 by treatment but was lower under the HV (62.5%) compared to the LV (95.6%) regime
346 ($X^2(1) = 15.034, p < 0.001$).

347

348 For large eel, *passage success* was lower under HV (17.2%) than the MV (92.3%) ($X^2(1)$
349 = 41.85, $p < 0.001$) and LV control (100%) ($X^2(1) = 30.99, p < 0.001$), but not different

350 between the MV and LV control (Figure 5a). For lamprey, *passage success* was lower

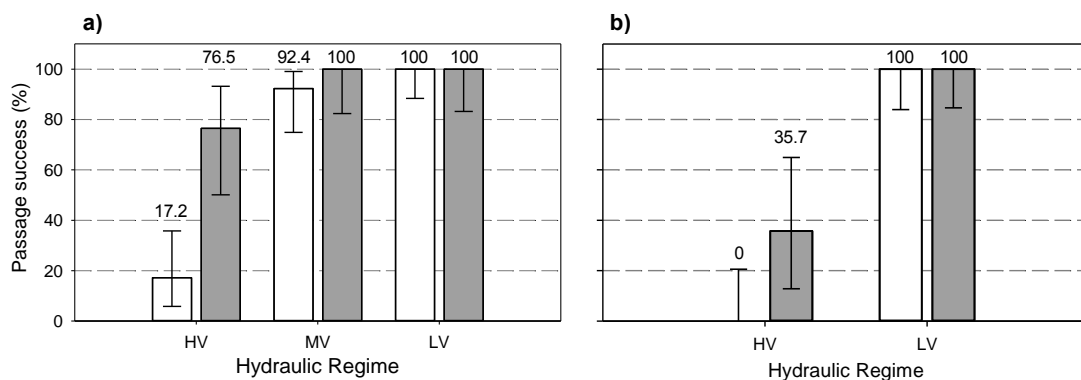
351 under the HV (0%) than LV control (100%) ($X^2(1) = 37, p < 0.001$) (Figure 5b).

352 *Passage success* was higher under the HV treatment than control for both large eel

353 (76.5%; $X^2(1) = 5.785, p < 0.001$) and lamprey (35.7%; $FET: p < 0.05$) (Figure 5).

354 There was no influence of treatment under the MV or LV regime (Figure 5).

355



356

357 **Figure 5.** *Passage success* (%) for (a) large eel and (b) lamprey without (control: clear
358 bars) and with (treatment: grey bars) bristle passes installed under the low (LV),

359 medium (MV), and high (HV) velocity hydraulic regimes. Error bars are 95%

360 confidence intervals calculated using the Clopper-Pearson exact method.

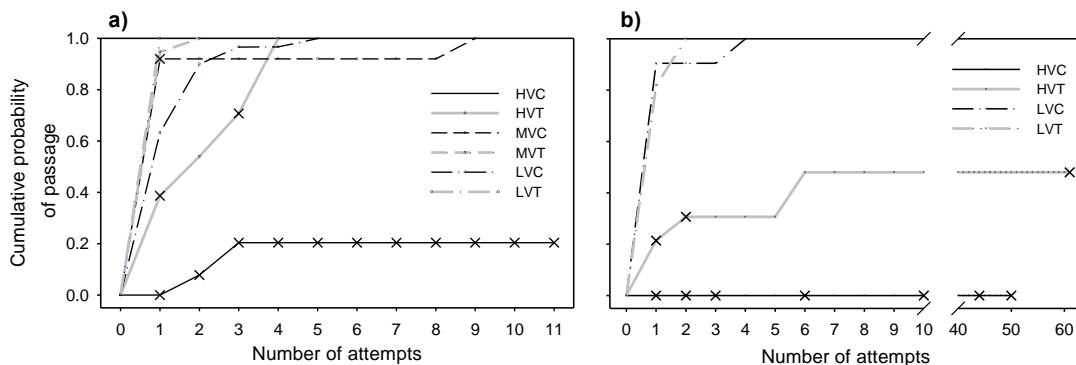
361

362 For large eel, *number of attempts before upstream passage* was higher under HV

363 control (20.5% CPP after 3 attempts) than the MV (>50% CPP after the 1st attempt)

364 ($W_s(1) = 26.729, p < 0.001$) and LV control (>50% CPP after the 1st attempt) ($W_s(1) =$
 365 31.593, $p < 0.001$), but was not different between the LV and MV control (Figure 6a).
 366 For lamprey, *number of attempts before upstream passage* was higher under HV control
 367 (0% CPP despite up to 50 attempts) than the LV control (>50% CPP after the 1st attempt)
 368 ($W_s(1) = 29.176, p < 0.001$) (Figure 6b). *Number of attempts before upstream passage*
 369 was lower under the HV treatment than control for both large eel (>50% CPP after the
 370 2nd attempt; $W_s(1) = 18.275, p < 0.001$) and lamprey (30.6% CPP after the 2nd attempt;
 371 $W_s(1) = 45.702, p < 0.001$) (Figure 6). There was no influence of treatment under the
 372 MV or LV regime (Figure 6).

373



374

375 **Figure 6.** Cumulative Probability of Passage (CPP) upstream with number of attempts
 376 for (a) large eel and (b) lamprey with (treatment: grey lines) and without (control: black
 377 lines) bristle passes installed under the low (LV), medium (MV), and high (HV)
 378 velocity regimes. Crosses represent right censored data.

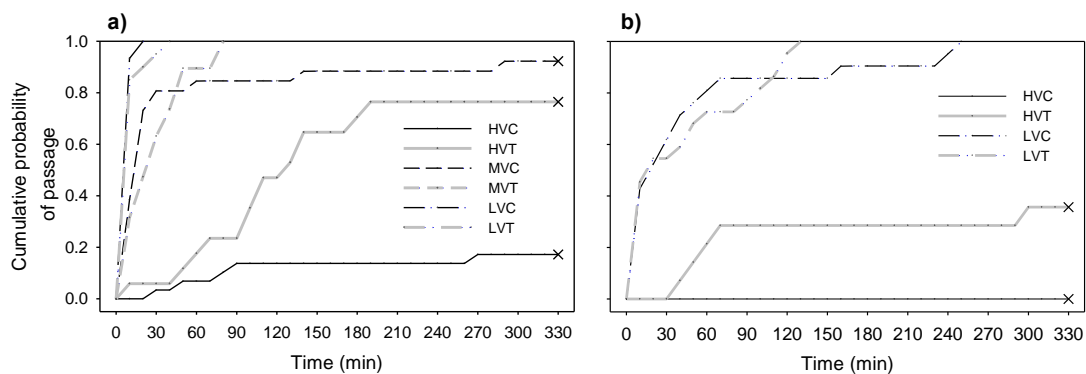
379

380 For large eel, *Delay* was longer under the HV control (17.2% CPP after 330 minutes)
 381 than the MV (50% CPP after 13.3 minutes) ($X^2_{mc}(1) = 44.974, p < 0.001$) and LV
 382 control (50% CPP after 5.36 minutes) ($X^2_{mc}(1) = 69.399, p < 0.001$), and longer under

383 MV control than the LV control ($X^2_{mc}(1) = 22.837, p < 0.001$) (Figure 7a). For lamprey,
 384 *Delay* was longer under HV control (0% CPP after 330 minutes) than the LV control
 385 (50% CPP after 19.28 minutes) ($X^2_{mc}(1) = 38.767, p < 0.001$) (Figure 7b). *Delay* was
 386 shorter under the HV treatment than control for both large eel (50% CPP after 115
 387 minutes: $X^2_{mc}(1) = 16.260, p < 0.001$) and lamprey (35.7% CPP after 330 minutes:
 388 $X^2_{mc}(1) = 6.730, p < 0.01$) (Figure 7). There was no influence of treatment under the
 389 MV or LV regime (Figure 7).

390

391



392

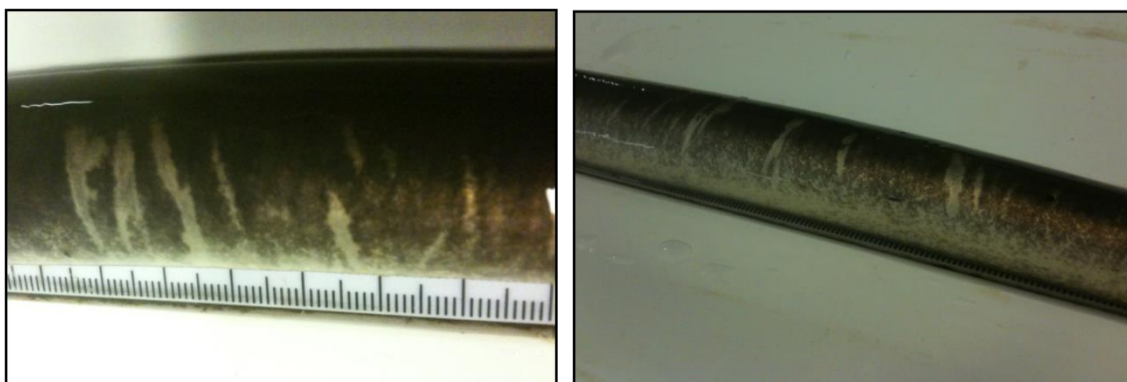
393 **Figure 7.** Cumulative Probability of Passage (CPP) upstream against time for (a) large
 394 eel and (b) lamprey with (treatment: grey lines) and without (control: black lines) bristle
 395 passes installed under the low (LV), medium (MV), and high (HV) velocity hydraulic
 396 regimes. Crosses represent right censored data.

397

398 Neither *percentage attachments* (34.2%) nor *number of attachments* ($\mu \pm \text{S.E.}: 16.0 \pm$
 399 6.8) were influenced by hydraulic regime or treatment. *Mean attachment duration* was
 400 influenced by hydraulic regime ($F(1, 8) = 7.807, p < 0.05$), being longer under the HV
 401 (150.7 \pm 27.0s) than LV regime (46.5 \pm 19.6s), but not by treatment.

402

403 Lamprey were not as proficient at navigating the bristled substrate as eel, often
404 struggling to make progress through the passes. Lamprey were observed to have striated
405 marks along the length of their body after exiting the bristle passes (Figure 8). These
406 were temporary and disappeared within 24 hours. Eel showed no obvious physical
407 external effects of bristle pass use.



408

409 **Figure 8.** Two examples of striated marks on the flanks of lamprey caused by bristle

410

pass use. Scale is in mm.

411 4. Discussion

412

413 This study experimentally assessed the efficacy of a side-mounted vertically oriented
414 bristle pass for improving upstream passage of European eel and river lamprey at a low-
415 head gauging weir. Eel and lamprey were highly motivated to explore their
416 surroundings and move upstream. Bristle passes improved their ability to do so when
417 high flow velocities and turbulence restricted passage. Interspecific variation in efficacy
418 was apparent with the passes being more effective for eel than for lamprey.

419

420 Barriers can block or impede the movement of fish between essential rearing and
421 spawning habitat (Lucas and Baras, 2001). Excessive energetic costs during migration
422 can compromise the physiological and behavioural processes necessary for sexual
423 maturation and successful reproduction (Mesa *et al.*, 2003). Delayed migration can
424 increase predation risk (Peake *et al.*, 1997; Rieman *et al.*, 1991), physiological stress,
425 and susceptibility to disease (Loge *et al.*, 2005). For adult lamprey, as for most
426 anadromous species, additional energetic costs during upstream movement to spawning
427 grounds cannot be compensated as feeding ceases during migration (Lucas and Baras,
428 2001). In this study, bristle passes mitigated to some extent these negative effects by
429 providing higher passage success and efficiency, shorter delay, and fewer failed
430 attempts for both eel and lamprey as they passed the model crump weir.

431

432 A key concern in the design of the experiment was to allow fish sufficient time to pass
433 the obstruction. As such, a single 5.5 hour long trial was undertaken per night. This, in
434 combination with the limited duration of the experimental period, resulted in a low

435 number of replicates. As such, the statistics presented could be considered conservative
436 with a high chance of a type II error (i.e. only large effects being detected as significant).
437 Although not statistically significant the measured mean and variance values indicate
438 that bristle passes may also be affecting the number of upstream passes per night and
439 having further beneficial influences on passage efficiency outside of those identified
440 through the inferential statistics. For example, in addition to the bristle passes
441 significantly improving passage efficiency for small and large eel under the HV regime,
442 the data indicate they may have also improved passage efficiency for lamprey, and for
443 small and large eel under the medium and low velocity regime. Further experimental
444 data would have to be collected to validate these trends.

445

446 This study provides: 1) evidence that bristle passes improve the upstream passage of
447 both eel and lamprey under experimental conditions and 2) a mechanistic understanding
448 of how they function which will help improve future pass design. However, the majority
449 of barriers where bristle passes are likely to be installed are larger than the model weir
450 used in this experiment (e.g. increased head difference and distance for traversal).

451 Larger scale flume trials would provide useful information of the effects of increased
452 barrier size but the facilities to undertake such experiments are rare. In addition, flume
453 trials cannot adequately account for the numerous confounding variables that occur in
454 situ. The next step in validating the effectiveness of side-mounted vertically oriented
455 bristle passes is to undertake robust field studies at larger barriers.

456

457 In good years, juvenile European eel are recruited into the lower catchment of
458 freshwater systems in large numbers (Moriarty, 1990). As there is a causal relationship

459 between body length and absolute swimming performance (Beamish, 1978; Clough *et*
460 *al.*, 2004) small juvenile eel are particularly susceptible to velocity barriers. In this
461 study, bristle passes facilitated the upstream passage of eel as small as *ca.* 100mm.
462 Enhanced dispersal of this life-stage is particularly important as it is likely that density-
463 dependent mortality (see: Vøllestad and Jonsson, 1988) would limit system productivity
464 unless early upstream colonisation is achieved.

465

466 In comparison to small eel, a higher percentage of large eel passed over the weir directly,
467 rather than via a bristle pass under each hydraulic regime. Possibly because bristle
468 spacing was less appropriate for larger eel (restricted manoeuvring space) or their higher
469 absolute swimming capability enabled them to more easily ascend the weir. Similarly, a
470 lower percentage of both large and small eel passed the weir via the bristle passes under
471 the low compared to high velocity treatment. Probably due to it being easier for all sizes
472 to ascend the weir directly under these conditions. Few lamprey passed through the
473 bristle passes under any treatment. Those that did exhibited cutaneous abrasions, which
474 can increase a fish's susceptibility to bacterial infection (Bader *et al.*, 2006). For this
475 species, further research to investigate how design alterations, such as increasing bristle
476 spacing, may improve passage success and reduce abrasion is warranted. The
477 implication of such design modifications on eel passage should be considered in parallel.

478

479 Poor attraction efficiency is known to limit the overall effectiveness of fish passes (Bunt
480 *et al.*, 2001; Moser *et al.*, 2002). In this study, limited downstream area, long trial
481 duration, and the highly active nature of both species resulted in a very high chance of
482 individuals encountering the entrance of a bristle pass. In addition, both eel and lamprey

483 tended to move upstream along the flume walls further increasing their chances of
484 encountering a pass entrance. Actively migrating juvenile eels tend to migrate on mass
485 in the shallow low velocity regions along the banks of estuaries and rivers (Tesch, 2003),
486 and passes located along channel boundaries generally catch more individuals than
487 those in the centre (Piper *et al.*, 2012). As such, the configuration of bristle pass tested
488 in this study (attached directly to the wing wall of a gauging structure) probably
489 represents the optimal location to maximise attraction efficiency. However, it is
490 acknowledged that at complex sites the low flow through this type of pass may limit
491 attraction. In such cases extra attraction flow should be provided (see: Piper *et al.*, 2012).

492

493 Unlike eel, lamprey lack paired fins and struggle to maintain stability in turbulent
494 conditions (see: Liao, 2007). A lower percentage of lamprey attempted to pass the weir
495 under the high compared with low velocity regime, possibly because turbulent
496 conditions associated with the hydraulic jump inhibited upstream movement. Lamprey
497 also frequently attached to the face of the weir and attempted to pass using a burst-
498 attach-rest mode of locomotion thought to enhance performance (Kemp *et al.*, 2011;
499 Quintella *et al.*, 2004). Previous studies indicate that lamprey vary their attachment
500 behaviour in response to hydraulic conditions (Kemp *et al.*, 2011), an observation
501 supported by the results of this study in which mean duration of attachment was longer
502 under the high velocity regime, presumably to facilitate recovery.

503

504 In this study, when high velocity and turbulence restricted passage, bristle passes
505 increased the passage success of large eel and lamprey to 76.5 and 36.5%, respectively.

506 For catadromous European eel, such levels may be adequate to maintain a stable

507 population due to the extended duration of their diffusive upstream migration (i.e. a
508 high probability of being able to pass during a high-flow event). For anadromous river
509 lamprey, which are energetically and temporally constrained during their upstream
510 migration, such levels will likely limit system productivity. It is recommended that new
511 fish passage technologies for both species continue to be investigated. However, for a
512 small barrier the configuration of bristle pass tested would seem to represent a viable
513 low-maintenance and low-cost option to improve habitat connectivity for European eel.
514 For river lamprey, while the wing-wall bristle media shows potential for assisting
515 passage, further studies over a wider range of obstacle heights and bristle spacing are
516 needed to determine whether this approach has merit.

517

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519

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526 **References**

527

528 Agostinho, A.A., Pelicice, F.M., Gomes, L.C. **2008**. Dams and the fish fauna of the
529 Neotropical region: Impacts and management related to diversity and fisheries. *Brazil. J.*
530 *Biol.* **68** (4s), 1119-1132.

531

532 Aparicio, E., Vagras, M.J., Olmo, J.M., de Sostoa, A. **2000**. Decline of native
533 freshwater fishes in a Mediterranean watershed on the Iberian Peninsula: A quantitative
534 assessment. *Environ. Biol. Fish.* **59**, 1-19.

535

536 Bader, J.A., Moore, S.A., Nusbaum, K.E. **2006**. The effect of cutaneous injury on a
537 reproducible immersion challenge model for *Flavobacterium columnare* infection in
538 channel catfish (*Ictalurus punctatus*). *Aquaculture* **253**, 1-9.

539

540 Beach, M.H. **1984**. Fish pass design - criteria for the design and approval of fish passes
541 and other structures to facilitate the passage of migratory fish in rivers. Fisheries
542 Research Technical Report, MAFF Directorate of Fisheries Research, **78**, pp. 46.
543 <http://www.cefas.defra.gov.uk/publications/techrep/tech78.pdf> (Accessed 17/03/2015).

544

545 Beamish, F.W.H. **1978**. Swimming Capacity, in: Hoar, W.S., Randall, D.J. (Eds.), Fish
546 Physiology Vol. 7 Locomotion. Academic Press Inc., New York, pp. 101-187.

547

548 Bellanger, A., Huon, S., Steinmann, P., Chabaux, F., Velasquez, F., Vallès, V., Arn, K.,
549 Clauer, N., Mariotti, A. **2004**. Oxidic–anoxic conditions in the water column of a tropical
550 freshwater reservoir (Peña-Larga dam, NW Venezuela). *Appl. Geochem.* **19**, 1295-1314.
551

552 Boon, P.J. **1988**. The impact of river regulation on invertebrate communities in the UK.
553 *Regul. River.* **2**, 389-409.
554

555 Briand, C., Fatin, D., Fontenelle, G., Feunteun, E. **2005**. Effect of re-opening of a
556 migratory pathway for eel (*Anguilla anguilla*, L.) at a watershed scale. *B. Fr. Peche.*
557 *Piscic.* **378-379**, 67-86.
558

559 Bunt, C. M., Katopodis, C., McKinley, R.S. **1999**. Attraction and passage efficiency of
560 white suckers and smallmouth bass by two Denil fishways. *N. Am. J. Fish. Manage.* **19**,
561 793-803.
562

563 Bunt, C.M., Cooke, S.J., McKinley, R.S. **2000**. Assessment of the Dunnville fishway
564 for passage of walleyes from Lake Erie to the Grand River, Ontario. *J. Great Lakes R.*
565 **26**, 482-488.
566

567 Bunt, C. M., van Poorten, B.T., Wong, L. **2001**. Denil fishway utilization patterns and
568 passage of several warm water species relative to seasonal, thermal and hydraulic
569 dynamics. *Ecol. Freshw. Fish* **10**, 212-219.
570

571 Castro-Santos, T., Haro, A. **2003**. Quantifying migratory delay: a new application of
572 survival analysis methods. *Can. J. Fish. Aquat. Sci.* **60**, 989-996.
573

574 Caudill, C.C., Daigle, W.R., Keefer, M.L., Boggs, C.T., Jepson, M.A., Burke, B.J.,
575 Zabel, R.W., Bjornn, T.C., Peery, C.A. **2007**. Slow dam passage in adult Columbia
576 River salmonids associated with unsuccessful migration: delayed negative effects of
577 passage obstacles or condition-dependent mortality? *Can. J. Fish. Aquat. Sci.* **64**, 979-
578 995.
579

580 Clay, C.H. **1995**. Design of fishways and other fish facilities. Second edition. CRC
581 Press, Inc., Boca Raton, Florida. pp. 249.
582

583

584 Clough, S.C., Lee-Elliott, I.H., Turnpenny, A.W.H., Holden, S.D.J., Hinks, C. **2004**.
585 Swimming speeds in fish: Phase 2. R&D Technical Report No. W2-026/TR3,
586 Environment Agency, Bristol, pp. 82.
587 [https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/290591/s](https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/290591/scho0404bipv-e-e.pdf)
588 [cho0404bipv-e-e.pdf](https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/290591/scho0404bipv-e-e.pdf) (accessed 19/03/2015)
589

590 Cooke, S.J., Bunt, C.M., Hamilton, S.J., Jennings, C.A., Pearson, M.P., Cooperman,
591 M.S., Markle, D.F. **2005**. Threats, conservation strategies, and prognosis for suckers
592 (Catostomidae) in North America: insights from regional case studies of a diverse
593 family of non-game fishes. *Biol. Conserv.* **121**, 317-331.
594

595 Dekker, W. **2003**. Status of the European eel stock and fisheries, in: Aida K.,
596 Tsukamoto K., Yamauchi K. (Eds.), *Eel Biology*. Springer-Verlag, Tokyo, pp. 237-254.
597

598 Dekker, W., Pawson, M., Wickström, H. **2007**. Is there more to eels than slime? An
599 introduction to papers presented at the ICES Theme Session in September 2006. *ICES J.*
600 *Mar. Sci.* **64** (7), 1366-1367.
601

602 Environment Agency. **2010**. Field evaluation of the effect of elver passes on measuring
603 structures. Draft Report, J2578/1.
604

605 Environment Agency. **2011**. Elver and eel passes: A guide to the design and
606 implementation of passage solutions at weirs, tidal gates and sluices. *GEO0211BTMV-*
607 *E-E*.
608

609 Feunteun, E. **2002**. Management and restoration of European eel population (*Anguilla*
610 *anguilla*): An impossible task. *Ecol. Eng.* **18**, 575-591.
611

612 Foulds, W.L., Lucas, M.C. **2013**. Extreme inefficiency of two conventional, technical
613 fishways used by the European river lamprey (*Lampetra fluviatilis*). *Ecol. Eng.* **58**, 423-
614 433.
615

616 Gordon, E., Meentemeyer, R.K. **2006**. Effects of dam operation and land use on stream
617 channel morphology and riparian vegetation. *Geomorphology* **82**, 412-429.
618

619 Gresh, T., Lichatowich, J., Schoonmaker, P. **2000**. An Estimation of Historic and
620 Current Levels of Salmon Production in the Northeast Pacific Ecosystem: Evidence of a
621 Nutrient Deficit in the Freshwater Systems of the Pacific Northwest. *Fisheries* **25** (1),
622 15-21.

623

624 Haro, A., Kynard, B. **1997**. Video evaluation of passage efficiency of American shad
625 and sea lamprey in a modified Ice harbor fishway. *N. Am. J. Fish. Manage.* **17** (4), 981-
626 987.

627

628 ICES. **2012**. Report of the Joint EIFAAC/ICES Working Group on Eels (WGEEL), 3–9
629 September 2012, Copenhagen, Denmark. ICES CM 2012/ACOM:18. pp. 824.

630

631 Jang, M.-H., Lucas, M.C. **2005**. Reproductive ecology of the river lamprey. *J. Fish Biol.*
632 **66**, 499-512.

633

634 Jellyman, D.J., Ryan, C.M. **1983**. Seasonal migration of elvers (*Anguilla* spp.) into Lake
635 Pounui, New Zealand, 1974-1978. *New Zeal. J. Mar. Fresh. Res.* **17**, 1-15.

636

637 Keefer, M.L., Clabough, T.C., Jepson, M.A., Johnson, E.L., Boggs, C.T., Caudill, C.C.
638 **2012**. Adult Pacific lamprey passage: Data synthesis and fishway improvement
639 prioritization tools. Technical report 2012-8. *Prepared for: Department of Fish and*
640 *Wildlife Sciences College of Natural Resources, University of Idaho.*
641 <http://www.uidaho.edu/~media/Files/orgs/CNR/FERL/Technical%20Reports/2012/201>
642 2-

643 8%20Report%20Keefer%20et%20al%20UI%20Lamprey%20data%20synthesis%20FI
644 NAL.ashx (Accessed 29/05/2015).
645
646 Kelly, F.L., King, J.J. **2001**. A review of the ecology and distribution of three lamprey
647 species, *Lampetra fluviatilis* (L.), *Lampetra planeri* (Bloch) and *Petromyzon marinus*
648 (L.): a context for conservation and biodiversity considerations in Ireland. *Biol. Environ.*
649 **101 (B)**, 165-185.
650
651 Kemp, P.S., Tsuzaki, T., Moser, M.L. **2009**. Linking behaviour and performance:
652 intermittent locomotion in a climbing fish. *J. Zool.* **277**, 171-178.
653
654 Kemp, P.S., Russon, I.J., Vowles, A.S., Lucas, M. **2011**. The influence of discharge and
655 temperature on the ability of upstream migrant river lamprey (*Lampetra fluviatilis*) to
656 pass experimental overshot and undershot weirs. *River Res. Appl.* **27**, 488-498.
657
658 Kruk, A. **2004**. Decline in migratory fish in the Warta River, Poland. *Ecohydrol.*
659 *Hydrobiol.* **4 (2)**, 147-155.
660
661 Laffaille, P., Caraguel, J-M., Legault, A. **2007**. Temporal patterns in the upstream
662 migration of European glass eels (*Anguilla anguilla*) at the Couesnon estuarine dam.
663 *Estuar. Coast. Shelf S.* **73**, 81-90.
664
665 Laine, A., Kamula, R., Hooli, J. **1998**. Fish and lamprey passage in a combined Denil
666 and vertical slot fishway. *Fisheries Manag. Ecol.* **5**, 31-44.

667

668 Larinier, M., Marmulla, G. **2004**. Fish Passes: Types, Principles and Geographical
669 Distribution – An Overview, in: Welcomme, R., Petr, T. (Eds.), Proceedings of the
670 Second International Symposium on the Management of Large Rivers for Fisheries
671 Volume II. FAO Regional Office for Asia and the Pacific, Bangkok, Thailand. RAP
672 Publication 2004/17, pp. 1-14.

673

674 Legault, A. **1988**. Le franchissement des barrages par l'escalade de l'anguille; étude en
675 Sèvre Niortais. *B. Fr. Peche. Piscic.* **308**, 1-10.

676

677 Liao, J.C. **2007**. A review of fish swimming mechanics and behavior in altered flows.
678 *Philos. T. R. Soc. B.* **362**, 1973-1993.

679

680 Loge, F.J., Arkoosh, M.R., Ginn, T.R., Johnson, L.L., Collier, T.K. **2005**. Impact of
681 environmental stressors on the dynamics of disease transmission. *Environ. Sci. Technol.*
682 **39** (18), 7329-7336.

683

684 Lucas, M.C., Baras, E. **2001**. Migration of Freshwater Fishes. Blackwell Science,
685 Oxford. pp. 352.

686

687 Matzinger, A., Pieters, R., Ashley, K.I., Lawrence, G.A., Wüest, A. **2007**. Effects of
688 impoundment on nutrient availability and productivity in lakes. *Limnol. Oceanogr.* **52**
689 (6), 2629-2640.

690

691 Mesa, M.G., Bayer, J.M., Seelye, J.G. **2003**. Swimming performance and physiological
692 responses to exhaustive exercise in radio-tagged and untagged Pacific lampreys. *T. Am.*
693 *Fish. Soc.* **132**, 483-492.

694

695 Moriarty, C. **1986**. Riverine migration of young eels *Anguilla anguilla* (L.). *Fish. Res.*
696 **4**, 43-58.

697

698 Moriarty, C. **1990**. European catches of elver of 1928-1988. *Int. Rev. Hydrobiol.* **75** (6),
699 701-706.

700

701 Moriarty, C., Tesch, F.-W. **1996**. Possible increase in catch of Atlantic elver *Anguilla*
702 *anguilla* in 1993 and 1994. *Ecol. Freshw. Fish* **5**, 213-215.

703

704 Morita, K., Yamamoto, S. **2002**. Effects of habitat fragmentation by damming on the
705 persistence of stream-dwelling charr populations. *Conserv. Biol.* **5**, 1318-1323.

706

707 Moser M.L., Ocker, P.A., Stuehrenberg, L.C., Bjornn, T.C. **2002**. Passage efficiency of
708 adult Pacific lampreys at hydropower dams on the lower Columbia River, USA. *T. Am.*
709 *Fish. Soc.* **131**, 956-965.

710

711 Moser, M.L., Keefer, M.L., Pennington, H.T., Ogden, D.A., Simonson, J.E. **2011**.
712 Development of Pacific lamprey fishways at a hydropower dam. *Fisheries Manag. Ecol.*
713 **18**, 190-200.

714

715 Muller, U.K., Smit, J., Stamhuis, E.J., Videler, J.J. **2001**. How the body contributes to
716 wake in undulatory fish swimming: Flow fields of swimming eel (*Anguilla anguilla*). *J.*
717 *Exp. Biol.* **204**, 2751-2762.

718

719 National River Flow Archive. www.ceh.ac.uk/data/nrfa (Accessed 29/05/2015).

720

721 Nelson, M.L., McMahonald, T.E., Thurow, R.F. **2002**. Decline of the migratory form in
722 bull charr, *Salvelinus confluentus*, and implications for conservation. *Environ. Biol.*
723 *Fish.* **64**, 321-332.

724

725 Nilsson, C., Reidy, C.A., Dynesius, M., Revenga, C. **2005**. Fragmentation and flow
726 regulation of the world's large river systems. *Science* **308**, 405-408.

727

728 Noonan, M.J., Grant, J.W.A., Jackson, C.D. **2012**. A quantitative assessment of fish
729 passage efficiency. *Fish Fish.* **13**, 450-464.

730

731 Ojutkangas, E., Aronen, K., Laukkanen, E. **1995**. Distribution and abundance of River
732 lamprey (*Lampetra fluviatilis*) ammocoetes in the regulated River Perhonjoki. *Regul.*
733 *River.* **10**, 239-245.

734

735 O'Leary, D. **1971**. A low-head elver trap developed for use in Irish rivers. EIFAC
736 Technical Paper 14, pp. 129-133.

737

738 Peake, S., McKinley, R.S., Scruton, D.A. **1997**. Swimming performance of various
739 freshwater Newfoundland salmonids relative to habitat selection and fishway design. *J.*
740 *Fish Biol.* **51**, 710-723.

741

742 Pess, G.R., McHenry, M.L., Beechie, T.J., Davies, J. **2008**. Biological impacts of the
743 Elwha River dams and potential salmonid responses to dam removal. *Northwest Sci.* **82**
744 (s1), 72-90.

745

746 Petts, G.E. **1980**. Long-term consequences of upstream impoundment. *Environ. Conserv.*
747 **7** (4), 325-332

748

749 Piper, A.T., Wright, R.M., Kemp, P.S. **2012**. The influence of attraction flow on
750 upstream passage of European eel (*Anguilla anguilla*) at intertidal barriers. *Ecol. Eng.*
751 **44**, 329-336.

752

753 Porcher, J.P., Travade, F. **2002**. Fishways: Biological basis, limits and legal
754 considerations. *B. Fr. Peche. Piscic.* **364** (s), 9-20.

755

756 Quintella, B.R., Andrade, N.O., Koed, A., Almeida, P.R. **2004**. Behavioural patterns of
757 sea lampreys' spawning migration through difficult passage areas, studied by
758 electromyogram telemetry. *J. Fish Biol.* **65**, 961-972.

759

760 Renaud, C.B. **1997**. Conservation status of northern hemisphere lampreys
761 (Petromyzontidae). *J. Appl. Ichthyol.* **13**, 143-148.

762

763 Rieman, B.E., Beamesderfer, R.C., Vigg, S., Poe, T.P. **1991**. Estimated Loss of Juvenile
764 Salmonids to Predation by Northern Squawfish, Walleyes, and Smallmouth Bass in
765 John Day Reservoir, Columbia River. *T. Am. Fish. Soc.* **120** (4), 448-458.

766

767 Robinet, T., Guyet, S., Marquet, G., Mounaix, B., Olivier, J-M., Tsukamoto, K.,
768 Valade, P., Feunteun, E. **2003**. Elver invasion, population structure and growth of
769 marbled eels *Anguilla marmorata* in a tropical river on Réunion Island in the Indian
770 Ocean. *Environ. Biol. Fish.* **68**, 339-348.

771

772 Russon, I.J., Kemp, P.S., Lucas, M.C. **2011**. Gauging weirs impede the upstream
773 migration of adult river lamprey *Lampetra fluviatilis*. *Fisheries Manag. Ecol.* **18**, 201-
774 210.

775

776 Russon, I.J. and Kemp, P.S. **2011a**. Advancing the provision of multi-species fish
777 passage: Behaviour of adult European eel (*Anguilla anguilla*) and brown trout (*Salmo*
778 *trutta*) in response to accelerating flow. *Ecol. Eng.* **37**, 2018-2024.

779

780 Russon, I.J. and Kemp, P.S. **2011b**. Experimental quantification of the swimming
781 performance and behaviour of spawning run river lamprey *Lampetra fluviatilis* and
782 European eel *Anguilla anguilla*. *J. Fish Biol.* **78**, 1965-1975.

783

784 Sfakiotakis, M., Lane, D.M., Davies, B.C. **1999**. Review of fish swimming modes for
785 aquatic locomotion. *IEEE J. Oceanic Eng.* **24** (2), 237-252.

786

787 Singer, J.D., Willet, J.B. **2003**. Applied longitudinal data analysis: Modelling change
788 and event occurrence. Oxford University Press, New York. pp. 672.

789

790 Slatick, E., Basham, L.R. **1985**. The effect of Denil fishway length on passage of some
791 nonsalmonid fishes. *Mar. Fish. Rev.* **47** (1), 83-85.

792

793 Solomon, D. J., Beach, M. H. **2004**. Fish pass design for eel and elver (*Anguilla*
794 *anguilla*). R&D Technical Report W2-070/TR1. Environment Agency, Bristol, pp. 92.

795

796 Tesch, F.-W. **2003**. The eel. Fifth edition, Blackwell publishing, Oxford. pp. 408.

797

798 Vøllestad, L.A., Jonsson, B. **1988**. A 13-year study of the population dynamics and
799 growth of the European eel *Anguilla anguilla* in a Norwegian river: Evidence for
800 density-dependent mortality, and development of a model for predicting yield. *J. Anim.*
801 *Ecol.* **57**, 983-997.

802

803 Xu, K., Milliman, J.D. **2009**. Seasonal variations of sediment discharge from the
804 Yangtze River before and after impoundment of the Three Gorges Dam.
805 *Geomorphology* **104**, 276-283.

806

807 Yoshiyama, R.M., Fisher, F.W., Moyle, P.B. **1998**. Historical abundance and decline of
808 Chinook salmon in the Central Valley region of California. *N. Am. J. Fish. Manage.* **18**,
809 487-521.