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FOULING EFFECTS ON REJECTION IN THE MEMBRANE FILTRATION OF NATURAL WATERS

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ABSTRACT

Membrane processes in drinking water applications are micro- (MF), ultra- (UF), and nanofiltration (NF). These processes remove turbidity and bacteria (MF), viruses and macromolecules (UF) and small molecules and hardness (NF). Of particular concern in water treatment is the removal of natural organic matter (NOM) which contains potential disinfection by-product precursors. The presence of colloids, multivalent ions and organics in surface waters may cause substantial fouling of membranes.

A study was carried out which looked at the rejection abilities of a range of membranes targeting hematite colloids (40-500nm), NOM and cations, fouling conditions and cost of treatment of these processes with consideration of chemical pretreatment with ferric chloride [1]. In this paper the effect of membrane fouling on rejection is presented.

The study was based on experiments with two MF membranes (GVWP, GVHP, 0.22 µm, Millipore), six UF membranes (1, 3, 5, 10, 30, 100 kDa, regenerated cellulose, Millipore), and four organic nanofiltration (NF) membranes (TFC-SR, TFC-S, TFC-ULP, CA-UF, Fluid Systems, U.S.).

Three different types of organics (IHSS humic acid, IHSS fulvic acid and an Australian concentrated NOM) in a carbonate buffer containing calcium chloride and a background electrolyte were used. Experiments were carried out in perspex (MF, UF) and stainless steel (NF) stirred cells of a volume of 110-185 mL and a membrane area of 15.2 - 21.2*10⁻⁴ m² at transmembrane pressures of 1, 3, and 5 bar for MF, UF, and NF, respectively.

UF removes 10 - 95% of NOM depending on the molecular weight cut-off (MWCO) of the membrane. Pore sizes of < 6 nm are required to remove about 80% of NOM, where a 6 nm pore size corresponds to a MWCO of about 10 kDa. Colloids are fully rejected.

NF removes NOM effectively (70 - 95% as dissolved organic carbon (DOC) and 85 - 98% as UV absorbance). Cation rejection is very membrane dependent and varies for the investigated membrane types between 13 and 96% for calcium and 10 to 87% for sodium.

Fouling was also dependent on pore size and was caused by large colloids (250 nm) or coagulant flocs in MF, small colloids, organic-calcium flocs and aggregates with a dense structure (formed slowly) in UF, and by a calcium-organic precipitate in NF. The fouling influenced the rejection of colloids in MF and that of NOM in UF and NF. If a highly charged layer was deposited on the NF membranes, cation rejection was also influenced. The characterisation of permeate organics revealed that low molecular weight acids passed through the NF membranes and that the rejection of these acids was also dependent on the deposit on the membrane.

The mechanisms which can explain such an increase in rejection are different for the three membrane processes. In MF, pore plugging and cake formation was found responsible for fouling. This reduces the pore size and increases rejection. In UF, internal pore adsorption of calcium-organic flocs reduces the internal pore diameter and subsequently increases rejection. In NF, the key factor appears to be the charge of the deposit. This was investigated with the deposition of a ferric chloride precipitate. If the precipitate was of high positive charge, the rejection of cations increased and that of negatively charged low molecular weight acids decreased compared to more neutral or negative precipitates.

In essence, the rejection characteristics of membranes depend more on the fouling state of the membranes and the nature of the foulants than on the initial membrane characteristics.

KEYWORDS

Calcium, Fouling, Microfiltration, Nanofiltration, Ultrafiltration, Natural Organic Matter, Rejection.

1. INTRODUCTION

The extent of rejection of solutes by membranes is perhaps the most critical parameter in membrane filtration. For a clean (unfouled) membrane, the extent of rejection is influenced largely by the pore size or molecular weight cut-off (MWCO) though, in some instances, electrostatic interactions between the solute and membrane may influence the extent of rejection. For example, unfouled MF does not retain natural organics unless they are associated with particulates due to the large pore size [2]. Research in UF mostly focuses on size effects. Wiesner *et al.* [3] and Côté [4] published DOC removal as a function of molecular weight cut-off (MWCO). Wiesner *et al.* found a near linear decline, while Côté showed a steep decline in rejection between 1 and 10 kDa. The graphs were based on a review of publications and therefore represent the MWCO dependence well. In NF combined charge and size effects are well recognised [5].

Despite efforts to minimise the formation of deposits on membranes, fouling of the membranes however is likely to occur in many instances. This fouling process may be attributed to a number of mechanisms including pore blocking by solutes that are of similar diameter to the pores, formation of a cake from excluded solutes (i.e. solutes unable to pass through the membrane pores) and precipitation or gelation of inorganic and organic particulates at the membrane surface as a result of the localised high concentrations that occur at the membrane-solution interface. While interactions between solutes and the membranes are poorly understood, it is thought that effects like charge interactions, bridging, and hydrophobic interactions may play an important role in fouling [6].

It is clear that foulants at the membrane surface (or within the membrane pores) will exert some influence on passage of solutes through the membrane, either because the porosity of the cake or blocked membrane is lower than that of the clean membrane or because the charge properties of the fouled membrane are very different to those of the unfouled membrane.

While an effect on extent of solute rejection might be expected as a result of membrane fouling, very few studies have documented the nature or extent of this effect. The purpose of this study is to report

on results of ordered studies into the effect of fouling on rejection by microfiltration (MF), ultrafiltration (UF) and nanofiltration (NF) membranes.

2. BACKGROUND

MF, UF, and NF are pressure driven membrane processes. While the various commercial membranes are classed into process groups there is no clear boundary between these membrane classifications, rather MF, UF, NF and RO represents a continuum in 'pore size'. This is illustrated in Table 1, where sizes of water relevant solutes and particulates are compared with the range of membrane separation processes. The different processes overlap in their removal potential for various compounds.

Parameters used to quantify the efficiency of membrane processes, are flux (J) and solute rejection (R) where the flux is defined as

$$J = \frac{1}{A} \frac{dV}{dt} \quad (1)$$

and rejection as

$$R = 100 \cdot \left(1 - \frac{c_p}{c_b}\right) \quad (2)$$

The Resistance in Series Model in equation (3) describes the flux of a fouled membrane. R_M is the resistance of the clean membrane. The resistances R_{CP} , R_P and R_C denote the additional resistances which result from the exposure of the membrane to a solution containing particles or solute. R_{CP} is the resistance due to concentration polarisation, R_P the internal pore fouling resistance, and R_C the resistance due to external deposition or cake formation.

$$J = \frac{\Delta P}{\eta(R_M + R_{CP} + R_P + R_C)} \quad (3)$$

The nature of the resistances shows that different fouling mechanisms may operate and the distinction of these mechanism is the objective of this paper.

The effect of fouling on flux has been well documented in MF, UF and NF ([2], [7], [8]). However, very little has been done to describe the effect of fouling on rejection.

3. MATERIALS AND METHODS

3.1 Membranes, Filtration Systems and Filtration Protocol

The membranes used were characterised in detail in other publications by Schäfer et al. [9], Aoustin et al. [10], and Schäfer et al. [8] for MF, UF, and NF, respectively. Stirred cell systems made from perspex (MF, UF) and stainless steel (NF) were used.

The filtration systems and protocols have also been described previously ([9], [10], [8]). Operating pressures are summarised in Table 2. 1000 mL, 60-400 mL, and 120 mL of feed solution were filtered in MF, UF and NF, respectively. Key membrane characteristics are summarised in Table 2.

3.2 Natural Organic Matter (NOM) and Background Solution

Three types of organics were chosen: Two of these were purified and fractionated organics, purchased from the International Humic Substances Society (IHSS), namely, Suwannee River Reference humic (HA) and fulvic (FA) acid. A third organic type, an Australian NOM was concentrated from Mooney Mooney Dam (Gosford, NSW) using microfiltration and RO to concentrate all surface water constituents. The concentrate was further freeze-dried. The NOM powder obtained includes all naturally occurring inorganic salts and hydrophilic organics, which are part of the surface water. The background solution used, contained 0.5 mM CaCl_2 as a representative of naturally occurring multivalent cations, 1 mM NaHCO_3 as a natural buffer system, and 20 mM NaCl as a background electrolyte enabling pH adjustment without variation of the ionic strength.

All chemicals were purchased from Ajax Chemicals (Australia). 1M HCl and 1M NaOH were used for pH adjustment, and 1M NaCl for ionic strength adjustment.

3.3 Analytical Methods

Samples were characterised by UV/VIS spectrometry using a Varian Cary 1E UV/VIS. Dissolved organic carbon was analysed using a Skalar 12 carbon analyser. A Perkin Elmer Optima 3000 Inductively Coupled Plasma Atomic Emission Spectrometer (ICP-AES) was used to determine cation content of the samples. Samples were diluted with 5% nitric acid following collection and prior to ICP analysis.

4. RESULTS AND DISCUSSION

4.1 Rejection

Rejection results are summarised in Table 3. These results are from experiments where very little or no fouling occurred. As expected (given their pore size), very little of the natural organics and none of the cations are rejected by the MF or large pore size UF membranes. The extent of rejection increases for decreasing UF pore size and reaches near complete rejection for NF. In UF, charge effects begin to play a role in the rejection of organics [11].

Results of organics rejection as a function of membrane pore size are shown in Figure 1. The lines represent the rejection of UV absorbing organics and DOC in the absence of fouling. At a pore size of about 6 nm (in the UF range) a distinct drop in rejection from 85% (DOC) and 95% (UV) to 10% occurs. The difference between UV and DOC can be explained with the stronger absorbance of larger and more aromatic compounds.

4.2 Membrane Fouling

Fouling also depends on membrane pore size. A number of possible fouling mechanism need to be considered. Fouling may occur in pores by partial pore size reduction caused by foulants adsorbing on the inner pore walls, pore blockage and surface fouling such as cake and gel layer formation. The adsorption of compounds can also influence the hydrophilicity of a material and hence cause flux variation. Results are summarised in Table 4 for a feed containing humics and calcium and discussed in detail below.

In most water treatment applications and processes a combination of mechanisms can be expected. Foulants which are larger than the pore size will only cause surface fouling and for this reason the surface fouling becomes more important as the pore size is reduced.

4.2.1 Microfiltration

Calcium-organic complexes caused considerable flux decline in MF, which confirmed findings by [12]. This is most likely due to surface modification and partial pore plugging. In a previous study, colloids of a size similar to the pore size caused most severe flux decline by pore plugging and larger colloids caused cake formation which was less severe with the small colloids used in this study [9]. Cake effects would be expected for larger particles such as clays and microorganisms.

4.2.2 Ultrafiltration

Due to the wide range of UF membranes used, pore adsorption and cake formation were observed. Calcium-organic complexes adsorbed inside the pores of the 100 kDa membranes. This mechanism was identified by blocking law analysis [10], [13], [14]. When inorganic colloids (75nm) were filtered particle interactions were important. Depending on the aggregation regime aggregates of different structures formed and the structure determined the extent of flux decline [15]. The importance of such aggregation effects was confirmed with mixed systems containing organics and aggregates. Hematite stabilised with organics and thus depositing as individual colloids caused a flux decline of 60% with the 100 kDa membrane, while aggregated colloids caused a flux decline of only 15%. Figure 2 shows the different deposits formed by (A) stable colloids and (B) aggregates. Cakes formed from individual particles (or more compact aggregates) appeared to be more compact than cakes from loosely packed aggregates. These cake characteristics reflect in flux behaviour, while issues of cake compressibility is poorly understood.

4.2.3 Nanofiltration

NF retains calcium and most of the organics. This means the concentration of these compounds is relatively high in the boundary layer and precipitation may cause severe flux decline. Large organic compounds with a smaller diffusion coefficient fouled the membranes more due to a lower solubility and higher concentration in the boundary layer [8]. Internal pore fouling was not observed. Colloids did not cause flux decline in NF, most likely due to the high colloid/pore size ratio, but did influence rejection as discussed below.

4.3 Effect of Fouling on Rejection

Two effects of fouling on rejection were observed; first an increase in rejection due to a reduction of membrane pore size and second a modification of the separation behaviour of the membranes due to the deposit. This modification resulted in increase or decrease in rejection in particular instances. The increase in rejection with fouling is shown as a function of pore size in Figure 1. The cut-off of about 6 nm where rejection drops for unfouled membranes is now increased to >20 nm. This is dominated by effects in the UF range, where internal pore adsorption was predominant as shown by Aoustin et al. [10]. The deposit also modified rejection behaviour and depending on the process different solution compounds were responsible (as shown below).

4.3.1 Microfiltration

In MF, particles of a size very similar to the pore size, blocked pores and subsequently prevented further particles passing through the membrane [9]. This effect is a very direct sieving phenomena.

4.3.2 Ultrafiltration

The influence of the deposit was also observed in UF where the structure of a deposit of inorganic colloids could modify the rejection of an organic. This is shown in Figure 3. The tight cake structure produced from relatively stable colloids retains far more of the organics than the loose structure (produced from aggregates formed under diffusion limited conditions), but both inorganic cakes retain organics effectively as the rejection of the cake-free membrane is <10%. The relation between aggregate structure and effect on membrane flux is described in detail by Waite et al. [15].

4.3.3 Nanofiltration

In NF, the effect of rejection variation by fouling was confirmed by two different types of experiments. Firstly, when the membrane was fouled with a calcium-organic deposit as described by Schäfer et al.

[8], fouling increased the rejection of organics by up to 20%. Secondly, ferric chloride salt was added to remove foulants and inhibit precipitation of a calcium-organic gel layer. In this case a ferric hydroxide precipitate deposited on the membrane. While this prevented flux decline, the rejection behaviour was modified strongly. This is shown in Table 5. The variation of rejection was in this case attributed to the formation of deposits of a different charge. The high ferric chloride dose lead to a positively charged deposit and thus an increase in rejection for the cations and a decrease for the negatively charged organics. The lower rejection for the organics may have also been due to raised concentrations in the unstirred cake layer.

5. CONCLUSIONS

The study of three membrane processes which are widely used in water treatment, MF, UF, and NF, showed that fouling significantly influences membrane rejection behaviour. A reduction in pore size by pore adsorption and pore plugging increased rejection, while deposits on NF membranes can also decrease rejection of some compounds. A pore size to foulant size ratio was important in the determination of the mechanisms involved in rejection and fouling.

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REFERENCES

- [1] A. I. Schäfer, "Natural Organic Matter Removal using Membranes," in *Chemical Engineering*, Sydney, Australia: University of New South Wales, 1999.
- [2] W. Yuan and A. L. Zydney, "Humic acid fouling during microfiltration," *Journal of Membrane Science*, vol. 157, pp. 1-12, 1999.
- [3] M. R. Wiesner, M. M. Clark, J. G. Jacangelo, B. W. Lykins, B. J. Marinas, C. R. O'Melia, B. E. Rittmann, and M. J. Semmens, "Committee report: membrane processes in potable water treatment," *Journal AWWA*, vol. Jan, pp. 59-67, 1992.
- [4] P. Côté, "State-of-the-art techniques in reverse osmosis, nanofiltration and electrodialysis in drinking-water supply," presented at 20th International Water Supply Congress & Exhibition IWSA, Durban, South Africa, 1995.
- [5] J. M. M. Peeters, M. H. V. Mulder, K. Keizer, H. Strathmann, and H. Verweij, "Streaming potential measurements as a characterisation method for nanofiltration membranes," presented at Proc. of Colloid Science in Membrane Engineering, Toulouse, France, 1996.
- [6] J. Baker, T. Stephenson, S. Dard, and P. Côté, "Characterisation of fouling of nanofiltration membranes used to treat surface waters," *Environmental Technology*, vol. 16, pp. 977-985, 1995.
- [7] E. Matthiasson, "The role of macromolecular adsorption in fouling of ultrafiltration membranes," *Journal of Membrane Science*, vol. 16, pp. 23-36, 1983.
- [8] A. I. Schäfer, A. G. Fane, and T. D. Waite, "Nanofiltration of natural organic matter: removal, fouling and the influence of multivalent ions," *Desalination*, vol. 118, pp. 109-122, 1998.
- [9] A. I. Schäfer, U. Schwicker, M. M. Fischer, A. G. Fane, and T. D. Waite, "Microfiltration of Colloids and Natural Organic Matter," *Journal of Membrane Science*, vol. 171, pp. 151-172, 2000.
- [10] E. Aoustin, A. I. Schäfer, A. G. Fane, and T. D. Waite, "Effect of Humic Substance and Colloid Interactions on Ultrafiltration Behaviour," *Separation and Purification Technology*, vol. (in press), 2000.
- [11] A. I. Schäfer, R. Mauch, A. G. Fane, and T. D. Waite, "Charge effects in the fractionation of natural organics using ultrafiltration," *Environmental Science & Technology*, (in preparation).

- [12] W. Yuan and A. L. Zydney, "Effects of Solution Environment on Humic Acid Fouling during Microfiltration," *Desalination*, vol. 122, pp. 63-76, 1999.
- [13] J. Hermia, "Constant pressure blocking filtration laws-application to power-law non-newtonian fluids," *Trans IChemE*, vol. 60, pp. 183-187, 1982.
- [14] W. R. Bowen, J. I. Calvo, and A. Hernández, "Steps of membrane blocking in flux decline during protein microfiltration," *Journal of Membrane Science*, vol. 101, pp. 153-165, 1995.
- [15] T. D. Waite, A. I. Schäfer, A. G. Fane, and A. Hener, "Colloidal Fouling of Ultrafiltration Membranes: Impact of Aggregate Structure and Size," *Journal of Colloid and Interface Science*, vol. 212, pp. 264-274, 1999.
- [16] M. Cheryan, *Ultrafiltration Handbook*. USA: Technomic Publishing Company, Inc., 1986.
- [17] M. K. A. Agbekodo, "Élimination par nanofiltration des composés organiques d'une eau de surface prétraité. Caractérisation du carbone organique dissous avant et après nanofiltration," Poitiers, France: Université de Poitiers, 1994.
- [18] E. Worch, "Eine neue Gleichung zur Berechnung von Diffusionskoeffizienten gelöster Stoffe," *Vom Wasser*, vol. 81, pp. 289-297, 1993.

SYMBOLS

- A: Membrane Surface [m²]
 c_F: Feed Concentration [mgL⁻¹]
 c_P: Permeate Concentration [mgL⁻¹]
 J: Flux [Lm⁻²h⁻¹]
 J_{w0}: Pure Water Flux prior to Experiment [Lm⁻²h⁻¹]
 J_w: Pure Water Flux after Experiment [Lm⁻²h⁻¹]
 ΔP: Transmembrane Pressure [bar]
 R: Rejection [%]
 R_C: Membrane Resistance due to external deposition or cake formation [m⁻¹]
 R_{CP}: Membrane Resistance due to concentration Polarisation [m⁻¹]
 R_M: Resistance of clean Membrane [m⁻¹]
 R_P: Membrane Resistance due to internal Pore Fouling[m⁻¹]
 t: Time [h]
 V: Permeate Volume [L]
 η: Viscosity of Solvent (water) [Pa s]

TABLES

Table 1 Overview of membrane treatment processes and solute/particle dimensions ([16], [17]).

	Atomic Force Microscopy		Electron Microscope		Optical Microscope		Visible	
	Dissolved Matter			Colloids		Suspended Matter		
	Ions	Molecules	Macromolecules	Microparticles		Macroparticles		
Molecular Weight [Da]	100	1000	10000	100000				
Size [µm]	0.001	0.01	0.1	1	10	100	1000	
Solute/Particle Dimension								
Membrane Separation Process								

Table 2 Pure water membrane characteristics of the membranes used.

Process	Membrane	Pure Water Flux [Lm ⁻² h ⁻¹]	Water Permeability [Lm ⁻² h ⁻¹ bar ⁻¹]	Pressure [bar]	Membrane Resistance [m ⁻¹]	MWCO [kDa]	Pore Diameter [nm]
MF	GVWP	7968 ± 288	7968	1	4.51 · 10 ⁷	-	220*
MF	GVHP	7803 ± 308	7803	1	4.60 · 10 ⁷	-	220*
UF	PLHK	1320 ± 40	1320	1	0.03 · 10 ¹⁰	100*	18.20 [§]
UF	PLTK	390 ± 20	390	1	0.09 · 10 ¹⁰	30*	9.62 [§]
UF	PLGC	65 ± 5	21.7	3	1.66 · 10 ¹⁰	10*	5.18 [§]
UF	PLCC	28 ± 3	9.3	3	3.85 · 10 ¹⁰	5*	3.72 [§]
UF	PLBC	22 ± 2	7.3	3	4.90 · 10 ¹⁰	3*	2.84 [§]
UF	PLAC	15 ± 2	5.0	3	7.18 · 10 ¹⁰	1*	1.88 [§]
NF	CA-UF	49.9 ± 4.2	10.0	5	3.6 · 10 ¹⁰	5 [#]	3.72 [§]
NF	TFC-SR	45.8 ± 6.1	9.2	5	3.9 · 10 ¹⁰	< 0.18 [#]	< 0.64 [§]
NF	TFC-S	49.4 ± 5.9	9.9	5	3.6 · 10 ¹⁰	< 0.18 [#]	< 0.64 [§]
NF	TFC-ULP	19.4 ± 2.6	3.9	5	9.3 · 10 ¹⁰	< 0.18 [#]	< 0.64 [§]

* information supplied by Millipore.

determined by Fluid Systems using lactose marker tests. Rejections for MWCO as >90% at 1% glucose in MilliQ. CA-UF determined with 5 kDa dextran at 100 mgL⁻¹. Rejection of 10 kDa dextran was 76% (Takigawa (1999)).

§ calculated after Worch [18].

Table 3 Rejection of DOC, UV absorbance at 254 nm and cations as a function of membrane at pH 7-8, 5-15 mgL⁻¹ organics as DOC, 0.5 mM CaCl₂. Cation rejection is in the absence of organics .

	DOC Rejection [%] / UV ₂₅₄ Rejection [%]						Rejection [%]	
	IHSS HA	IHSS FA	NOM HA	NOM FA	NOM Hyd	NOM Hyd	Ca ²⁺	Na ⁺
MF GVWP	10/4 [#]	7/0 [#]	17/2 [#]	-	-	-	0	0
UF 100 kDa	6/4	8/4 [#]	9/9 [#]	8/5 [#]	3/4 [#]	4/5 [#]	0	0
UF 30 kDa	11/12	12/13	10/8	2/7	0/4	14/26	2.8	0
UF 10 kDa	66/88	56/53	51/55	65/76	58/64	41/50	2.6	0
UF 5 kDa	88/91	82/63	74/71	79/87	77/81	60/66	2.0	0
UF 3 kDa	86/93	84/94	77/73	79/90	78/86	60/73	13.6	0
UF 1 kDa	90/96	87/97	86/84	82/79	81/89	68/80	13.2	0
NF CA-UF	76/83	71/91	57/72	-	-	-	14.4	12.6
NF TFC-SR	74/97	94/96	69/97	96/99 [#]	96/98 [#]	84/87 [#]	67.6	37.4
NF TFC-S	90/99	84/95	95/96	-	-	-	94.4	82.0
NF TFC-ULP	76/99	80/88	95/96	-	-	-	90.4	85.0

values are at 2.5 mM CaCl₂, 5-15 mgL⁻¹ organics as DOC. Rejection of IHSS HA at these conditions is 92.7% (TFC-SR) and 57% (UF 100 kDa), for comparison.

Table 4 Flux, flux decline, permeability, membrane resistance and rejection at 2.5 mM CaCl₂ and IHSS HA at pH 7-8.

Process	Flux					Rejection			
	J _F [Lm ⁻² h ⁻¹]	J _F /J ₀ [-]	L _F [Lm ⁻² h ⁻¹ bar ⁻¹]	R _M [10 ¹⁰ m ⁻¹]	R _F [10 ¹⁰ m ⁻¹]	DOC [%]	UV ₂₅₄ [%]	Ca ²⁺ [%]	Na ⁺ [%]
MF GVWP	1732	0.22	1732	0.0046	0.0162	10	4	0	0
MF GVHP	1981	0.25	1981	0.0045	0.0181	16	6	0	0
UF 100 kDa	76	0.12	76	0.057	0.416	57	70	0	0
UF 10 kDa	32	0.94	11	3.168	0.198	70	87	6	4
NF CA-UF	41	0.83	8	3.600	0.780	74	89	98	20
NF TFC-SR	34	0.75	7	3.920	1.360	93	100	96	24
NF TFC-S	19	0.39	4	3.625	5.815	94	99	55	80
NF TFC-ULP	12	0.61	2	9.255	5.710	87	98	21	91

Table 5 Rejection as a function of organic type and FeCl₃ concentration (25 mgL⁻¹ FeCl₃ or 100 mgL⁻¹ FeCl₃, TFC-SR, 5 mgL⁻¹ DOC).

	DOC [%]	UV _{254nm} [%]	Calcium [%]	Sodium [%]
0 FA	94	96	68	36
0 HA	74	98	63	19
0 NOM	69	97	74	40
25 FA	72.4	94.6	44.3	0
25 HA	69.4	93.8	71.2	0
25 NOM	68.8	96.7	63.4	0
100 FA	45.1	99.7	96.1	31.2
100 HA	45.9	98.4	93.2	31.4
100 NOM	47.6	96.8	93.7	21.9

FIGURES

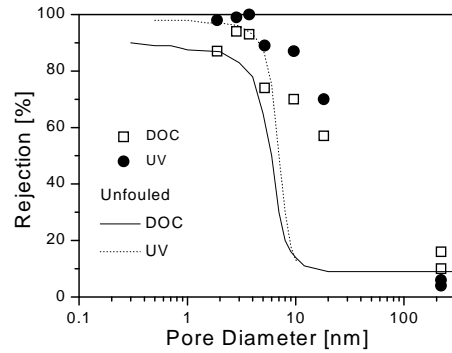


Figure 1 Rejection at fouling conditions as a function of pore diameter. The line graphs are rejection of unfouled membranes. Experimental conditions 2.5 mM CaCl₂ and IHSS HA at pH 7-8. Pore diameter as shown in **Table 1**.

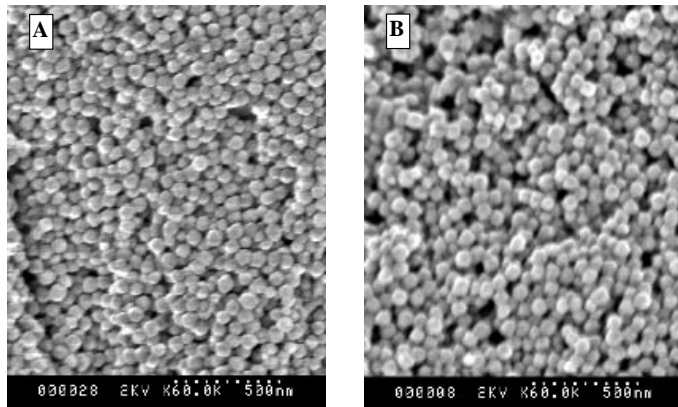


Figure 2 Electronmicrographs of deposits on the 100 kDa membranes of (A) OPS and (B) SPO systems (12.5 mgL⁻¹ as DOC IHSS HA, pH 7-8, 10 mgL⁻¹ hematite, primary colloid size 75 nm, in background solution).

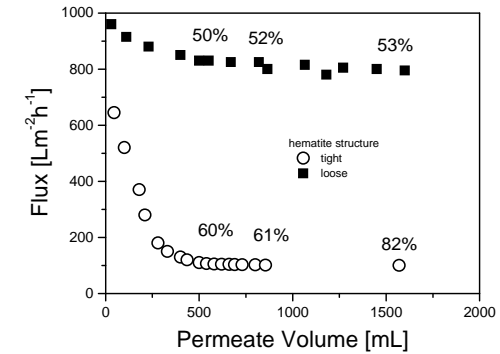


Figure 3 Permeate flux versus permeate volume during UF of humic acid solution with loose or compact structured aggregate coating of the membrane (100 kDa membrane, 5 mgL⁻¹ HA as DOC, pH 7-8, hematite deposition at pH 3 with 40 mM KCl (tight structure) and 80 mM KCl (loose structure)).