



This is a repository copy of *Farming with crops and rocks to address global climate, food and soil security*.

White Rose Research Online URL for this paper:
<http://eprints.whiterose.ac.uk/130923/>

Version: Accepted Version

Article:

Beerling, D.J., Leake, J.R. orcid.org/0000-0001-8364-7616, Long, S.P. et al. (13 more authors) (2018) Farming with crops and rocks to address global climate, food and soil security. *Nature Plants*, 4. pp. 138-147. ISSN 2055-026X

<https://doi.org/10.1038/s41477-018-0108-y>

Reuse

Items deposited in White Rose Research Online are protected by copyright, with all rights reserved unless indicated otherwise. They may be downloaded and/or printed for private study, or other acts as permitted by national copyright laws. The publisher or other rights holders may allow further reproduction and re-use of the full text version. This is indicated by the licence information on the White Rose Research Online record for the item.

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

PERSPECTIVE

Farming with crops and rocks to address global climate, food and soil security

David J. Beerling^{1*}, Jonathan R. Leake¹, Stephen P. Long^{2,3,4}, Julie D. Scholes¹, Jurriaan Ton¹, Paul N. Nelson⁵, Michael Bird⁵, Euripides Kantzas¹, Lyla L. Taylor¹, Binoy Sarkar¹, Mike Kelland¹, Evan DeLucia^{2,3}, Ilsa Kantola², Christoph Müller⁶, Greg Rau⁷ and James Hansen⁸

¹Leverhulme Centre for Climate Change Mitigation, Department of Animal and Plant Sciences, University of Sheffield, Sheffield S10 2TN, UK

²Carl R. Woese Institute for Genomic Biology, University of Illinois at Urbana-Champaign, Urbana, IL 61801, USA

³Department of Plant Biology, University of Illinois at Urbana-Champaign, Urbana, IL 61801, USA

⁴Lancaster Environment Centre, Lancaster University, Lancaster, LA1 4YQ, UK

⁵College of Science and Engineering and Centre for Tropical Environmental and Sustainability Science, James Cook University, Cairns, Queensland, Australia

⁶Potsdam Institute for Climate Impact Research, Telegraphenberg A31, D-14473 Potsdam, Germany

⁷Institute of Marine Sciences, University of California, Santa Cruz, CA 95064 USA

⁸Earth Institute, Columbia University, New York, NY10025, USA

*e-mail: d.j.beerling@sheffield.ac.uk

Limiting future climate change requires urgently decreasing CO₂ emissions through decarbonisation of energy supplies and developing approaches for carbon dioxide removal (CDR) from the atmosphere. Here, we discuss the potential for biogeochemical improvement of croplands by amending soils with crushed fast-reacting silicate rocks as a CDR strategy to address the threats of climate, food and soil security. By improving crop production and protection from pests and diseases, and restoring soil fertility and structure, this strategy could generate financial incentives for widespread adoption in the agricultural sector. Managed croplands worldwide are equipped for frequent rock dust additions to soils, making rapid adoption at scale feasible. However, audited field-scale assessments of CO₂ capture efficacy are urgently required together with detailed environmental monitoring. Recycling substantial quantities of silicate waste materials might help meet the rock requirements in a cost-effective manner. Issues of public perception, trust and acceptance must also be addressed.

Rising concentrations of atmospheric CO₂, and other greenhouse gases emitted by human activities, are already having substantial adverse climate impacts that threaten global food security^{1,2}. These impacts include more intense heat waves and droughts, as well as more extreme and variable rainfall, floods, and storms fuelled by latent energy in water vapour². This situation is unfolding at a time of unprecedented increase in food demand linked to dietary changes and a growing population that may reach ~11 billion by 2100, with agriculture itself a growing contributor to climate change^{2,3}. Crop yields are being further compromised by arable top soil losses that exceed natural rates of formation by an order of magnitude and the depletion of nutrients, such as phosphorus (P) and potassium (K)⁴. Soil nutrient stripping is being addressed with fertilizers, but these are based on finite resources that drive price inflation⁴.

Here, we examine in detail one option to help provide the required increases in yields, while reversing the negative impact of agriculture on sustainability and climate change.

Action on climate change is essential given global mean temperature, already more than 1°C above the pre-industrial level, will exceed the United Nations Paris Agreement 1.5°C aspirational limit⁵ within 30 years with the recent warming rate of 0.18°C per decade⁶. Further warming is ‘in the pipeline’ because of Earth’s present energy imbalance, thermal inertia in the ocean response, and slow amplifying climate feedbacks, including ice-sheet melt⁶. The continued response of the climate system to increased greenhouse gases, and practical difficulties of transitioning to carbon-free energy, makes even a more lenient 2°C warming target⁵ challenging. Consequently, the effective mitigation policy needed for meeting the UN targets requires rapid phasing out of fossil fuel emissions and deployment of scalable approaches for carbon dioxide removal (CDR) from the atmosphere with so-called negative CO₂ emissions in the second half of the 21st century⁷⁻⁹. Danger of sea level rise with loss of productive coastal marine and agricultural ecosystems, resulting displacement of people inland, and effects of increased climate extremes, add further urgency to the need to offset CO₂ emissions^{2,6}.

The United Nations 21st Conference of the Parties in Paris marked a turning point in the climate change debate with the focus shifting from describing climate change to a commitment to seek innovative, sustainable solutions¹⁰. Enhanced weathering aims to accelerate the natural geological process of carbon sequestration with production of alkaline leachate that reduces ocean acidification. It is achieved by amending the soils of intensively managed croplands with crushed calcium (Ca) and magnesium (Mg)-bearing rocks¹¹⁻¹³. Besides removing CO₂ from the atmosphere, we discuss how this strategy has the potential to also rejuvenate soils, stabilize soil organic matter, improve crop yields, conserve geological fertilizer resources, and benefit the marine environment.

Carbon capture with crops and rocks

Enhanced weathering accelerates CO₂ reactions with minerals contained in globally abundant, Mg- and/or Ca-rich rocks, a process that naturally moderates atmospheric CO₂ and stabilizes climate on geological time scales. In soils, chemical breakdown of carbonate and silicate rocks is accelerated during aqueous reactions with the elevated soil CO₂ environment, releasing base cations (Ca²⁺ and Mg²⁺) and delivering bicarbonate (HCO₃⁻), and to a lesser extent carbonate (CO₃²⁻) anions via runoff to surface waters and eventually the ocean. Enhanced weathering, therefore, uses the oceans to store atmospheric CO₂ as these stable dissolved inorganic alkaline forms (**Figure 1**). Given the oceans worldwide store around 38,000 Pg C, >45 times the mass of C in the current atmosphere, their future storage capacity is not an issue¹⁴. The residence time of dissolved inorganic carbon in the global ocean is around 100,000 – 1,000,000 years, making it essentially a permanent C-storage reservoir on human timescales¹⁵. Silicate weathering on land can also sequester atmospheric CO₂ without involving the oceans, if soil pore water chemistry results in precipitation of secondary carbonate minerals from base cation release¹⁵. In this case, the precipitated carbonate becomes the sink for CO₂ rather than ocean alkalinity. Carbonate weathering on acidic agricultural soils can lead to a net CO₂ flux to air^{16,17}, and carbonate minerals lack silica (Si) and other plant nutrient elements. The process of carbonate weathering on land thus delivers fewer benefits to climate, soils and crops. For these reasons, we focus on enhanced silicate weathering.

By adding alkalinity to the ocean, enhanced weathering enables the oceans to store more carbon, counters the effects of ocean acidification and the ongoing decrease in CaCO₃ saturation state, critical issues for protecting marine biocalcifiers, such as corals and shellfish,

from acidification impacts¹⁸⁻²⁰. Untreated, such impacts are estimated to cost the global economy²¹ as much as \$1 trillion a year by 2100.

Like other potential large-scale CDR strategies^{15,22,23}, enhanced weathering is relatively immature and requires further research, development and demonstration, including across a range of crops, soil types, climates and across spatial scales (**Table 1**). Experimental and small-scale evaluation of CO₂ capture efficacy and permanency remain priority research areas to understand its future relevance and contribution. A catchment-scale one-off application of 3.5 t ha⁻¹ of pelletized calcium silicate powder, wollastonite, to the 11.8 ha watershed of the Hubbard Brook Experimental Forest, New Hampshire, USA, confirmed key anticipated effects^{24,25}. These included rapid (12-24 months) 50% increased delivery of weathered calcium and silica dissolved in stream water, alleviation of ecosystem acidification and decreased release of soil aluminium²⁵. An upper estimate for CO₂ capture by wollastonite dissolution in the streambed during the first year of treatment, made assuming Ca²⁺ release is balanced by (bi)carbonate production, suggests a range of 110-224 g CO₂ m⁻², with a CO₂ capture efficiency of ~60% for the mass of wollastonite applied²⁶. This upper bound, however, is not likely to be representative of CO₂ capture by weathering in the forest soil, which remains to be quantified for this experiment²⁶.

Given that farmers routinely apply granular fertilizers and lime, annual applications of, for example, ground basalt (an abundant, weatherable Ca- and Mg-rich rock) is feasible at large scale with existing farm equipment. Global cropland (arable, forage, fibre, fruit etc.) covers approximately 12 × 10⁸ ha (12 million km²)²⁷ and an additional 1-10 × 10⁸ ha of marginal agricultural land may be available where basalt treatment could rejuvenate degraded soils²⁸. Effectively, nearly 11% of the terrestrial surface is managed for crop production and this may offer an opportunity to deploy a means of carbon sequestration at scale within a decade or two. Rapid deployment of CDR strategies is an essential requirement for significantly offsetting carbon emissions in the latter half of the twenty first century to avoid CO₂ and temperatures peaking and then declining with potentially adverse ecological and economic consequences^{8,23}. A first assessment might be achievable in areas of high-intensity agriculture where basalt, rock-crushing machinery, transportation infrastructure and agricultural spreaders are available, e.g., North America²⁹ or the UK³⁰.

Investigations of potential CO₂ sequestration by enhanced weathering with forested lands¹², and the oceans³¹⁻³⁴, have tended to focus on fast-weathering ultramafic olivine-rich rocks for which commercial mines are already in operation. Olivine comprises well over half the content by weight of ultramafic rocks, and is one of the fastest-weathering silicate minerals at pH < 6, potentially able to capture 0.8-0.9 t CO₂ per t of rock dissolved³⁰. However, a synthesis of published chemical analyses indicates that olivine-rich ultramafic rocks (i.e., peridotite: dunite, harzburgite, iherzolite and wehrlite), contain relatively high concentrations of either chromium (Cr) or nickel (Ni) or both (**Figure 2**). Weathering experiments reveal fast release of bioavailable Ni from olivine, and suppression of plant calcium uptake, because of competition with magnesium³⁵; experimental work with a soil column dosed with olivine suggested accumulation of Ni and Cr in the soil profile³⁶. Widespread application of olivine to agricultural soils, therefore, could introduce harmful metals into the food chain, and the wider environment, as well as causing nutritional imbalances and warrant further research¹⁸.

In contrast to ultramafic olivine-rich rocks, major continental flood basalts have lower concentrations of Ni and/or Cr (**Figure 2**) but significantly higher concentrations of phosphorus, suggesting their greater utility for croplands. Cultivation of crops on rich fertile soils that develop on flood basalts across continents is consistent with fewer environmental risks associated with this rock³⁷. Basalt is widely recognized as producing productive soils

because it weathers rapidly releasing elements essential for plant growth³⁸, including P, K, Ca, Mg and Fe. In terms of comparative weathering rates, olivine dissolution rates at oceanic pH~8 (10^{-10} to 10^{-11} mol olivine-Si $m^{-2} s^{-1}$) are within the range of those for basalt dissolution rates at pH 4 and above expected in soils (10^{-10} to 10^{-12} mol Si $m^{-2} s^{-1}$)³⁹.

Significant potential exists for deployment at scale to remove atmospheric CO₂ with ground basalt. A maximum carbon capture potential of ~0.3t CO₂ t⁻¹ is suggested for basalt, assuming a sufficiently fine particle size for effective dissolution on decadal time scales³⁰. The actual particle size will depend upon the mineralogy of the basalt, climate and biological activity, and requires further investigation and verification, but initial calculations suggest particles of 10-30 μm diameter. On this basis, basalt applications of 10 to 50 t ha⁻¹ yr⁻¹ to 70 \times 10⁶ ha of the annual crops corn/soy in the corn-belt of North America could sequester 0.2–1.1 PgCO₂, up to 13% of the global annual agricultural emissions, in the long run²⁹. Theoretical estimates of CO₂ capture and sequestration schemes involving global croplands and silicate rocks are very uncertain. Provisional estimates^{22,40} suggest that amending two thirds of the most productive cropland soils (9×10^8 ha) with basalt dust at application rates of 10-30 t ha⁻¹ yr⁻¹ could perhaps extract 0.5-4 PgCO₂ yr⁻¹ by 2100 depending on climate, soil and crop type. These numbers still need to account for full life cycle assessment, but suggest enhanced weathering could make a significant contribution to the negative emissions required by deep decarbonisation strategies^{8,9,23} and the ~1 Pg CO₂ eq. yr⁻¹ reduction from agriculture⁴¹ by 2030. Involvement of extensive marginal lands classified as not productive, or cost-effective, for food crops, further increases the potential for offsetting anthropogenic CO₂ emissions, although these lands would tend to be less accessible. Better constraining the appropriate particle size distribution for effective dissolution of basalt grains and, ultimately, the technical potential of the approach requires integrated biogeochemical modelling of the plant-soil-atmosphere system to capture interactions between crops, rocks, soils and fertilizers (inorganic and organic)⁴². Subsequent experimental validation at adequate scale will be critical (**Table 1**).

A key issue affecting carbon capture efficiency is the energy cost associated with mining, grinding and spreading the ground rock, which could reduce the net carbon drawdown by 10-30%, depending mainly on grain size⁴³. Relatively high energy costs for grinding, as influenced by rock mineralogy and crushing processes, call for innovation in the industrial sector, such as grinding and milling technology powered by renewable energy sources (solar, wind, water), to significantly increase the net CO₂ benefit. The benefit will increase as future energy sources are decarbonized, the grinding process becomes more energy efficient, and by utilizing already ground waste silicate materials previously or currently produced by the mining industry. By driving down costs for grinding in this way, carbon sequestration costs would be correspondingly cheaper.

Current cost estimates are uncertain and vary widely, and better understanding the economics involved is a priority. The most detailed analysis for operational costs drawn-up for using a basic rock, such as basalt, gives values of US\$52-480 tCO₂⁻¹, with comminution and transport the dominant components³⁰. This cost range falls below that estimated for bioenergy with carbon capture and storage (BECCS) of \$504-1296 tCO₂⁻¹ (US\$140-\$360 tC⁻¹)²². Deployment costs may be partially or completely offset by gains in crop productivity, and reduced requirements for lime, fertilizer, pesticide and fungicide applications, discussed later. Co-deployment of enhanced weathering with other strategies such as reforestation and afforestation, and with feedstock crops used in BECCS and biochar, could also reduce costs and significantly enhance the combined carbon sequestration potential of these methods.

Rocks for food and soil security

The amount of rock required to deploy enhanced weathering at scale is straightforward to calculate. We analysed illustrative application rates across 9×10^8 ha of the most productive managed lands based on FAO yield statistics for the dominant (in terms of area) annual crops²⁷, with the assumption that crop production is a reasonable proxy for good weathering conditions; both crops^{12,13} and rock³⁹ weathering require sufficient warmth and water. A wide range of experimental studies also point to annual crops as accelerating basalt weathering^{38,44-47} but this aspect is not considered further here. Calculated in this way, application rates of 10 to 30 t ha⁻¹ yr⁻¹ require 9-27 Pg rock yr⁻¹, although in practice, optimization of application rates will follow crop and soil type. These rock-dust application rates compare with recommended liming rates for UK arable soils⁴⁸ of 0.5-10 t lime ha⁻¹. For context, the global aggregate industry extracts ~50 Pg rock yr⁻¹ for construction, global mining for raw mineral materials⁴⁹ extracts ~17 Pg yr⁻¹ and the global cement industry extracts around 7 Pg yr⁻¹ of raw material (mainly limestone, shale, and/or clay)¹⁵. The mass of rock distributed onto land could be reduced if applications were optimized by, for example, restricting to 90% of the most productive regions to improve cost effectiveness. This is equivalent to 75% of agricultural land for annual crops (6.8×10^8 ha) (**Figure 3a**), and reduces the required rock mass to 7-20 Pg rock yr⁻¹. However, these amounts would change if deployment kept pace with projected expansion of arable cropland, which is subject to population growth, dietary choices and land use practices⁵⁰.

Analysed by national crop production (area \times productivity), these data indicate China, USA and India are the leading countries potentially able to sequester CO₂ in this way, with Russia and European countries, mainly Germany and France, next best placed (**Figure 3b**). Russia's relatively high agricultural productivity on moist Steppe soils, and warm summer temperatures over much of its growing region, may be conducive to CDR with enhanced weathering. These countries are the largest contributors to cumulative global CO₂ emissions from fossil fuels and industry (**Figure 3c**) since the pre-industrial era (1870) (565 ± 55 PgC)⁵¹ that is driving global warming^{51,52}.

Demand for reactive silicate rocks could be partially met if the 7-17 Pg yr⁻¹ freshly produced plant nutrient-containing silicate mining and industrial waste materials are utilized⁵³, more if legacy reserves are exploited. Assuming uncarbonated minerals and compounds remain, recycling these wastes might meet a considerable fraction of the demand given the application rates considered here. Mining of igneous rocks for construction generates an estimated 3 Pg yr⁻¹ of fine grained materials, too small for use as aggregates, which may be suitable for carbon capture with crops via enhanced weathering, with a considerably lower energy penalty for grinding⁵³. Increased construction and building activities in Brazil have promoted exploitation of basaltic reserves, and interest is growing in recycling accumulating fine basalt dust waste (particle size distribution peaking in the fine silt range of 10-20 μ m diameter) as a natural agricultural fertilizer⁵⁴. Mining of rocks for minerals, ores and metals produces a further 2-7 Pg yr⁻¹ of overburden material that may also be suitable for CDR⁵³, depending on host geology, with total waste in the USA alone of ~40 Pg accumulating between 1910 and 1980.

In addition, waste materials from industrial processes including cement production and steel manufacturing may also be suitable for enhanced weathering⁵³. Cement manufacture contributes ~6% of global CO₂ emissions⁵¹, and cement-based products (mainly concrete) used for construction also contain weatherable calcium-bearing minerals. Huge quantities of construction/demolition waste (1.4-5.9 Pg yr⁻¹), often used for landfill, have potential for enhanced weathering⁵³. Iron and steel manufacturing produces readily weatherable calcium silicate slag waste (0.4-0.5 Pg yr⁻¹), with the accumulation of significant global stockpiles (5.8-

8.3 Pg)^{30,53}. Steel slag contains fertilizer components (CaO, SiO₂, MgO, FeO, MnO and P₂O₅) with alkaline properties for remedying soil acidity. Consequently, these industrial by-products already have a long-history of being used on farms in place of lime, increasing crop production without toxic metal contamination at the application rates used for soil pH adjustment⁵⁵, and may have scope for wider adoption in enhanced weathering. China, a potentially important player in enhanced weathering (**Figure 3b**), is the largest steel producer in the world, but only recycles 22% of its steel slag, with scope for greatly expanding this use⁵⁶.

Residual combustion products from some agricultural sectors produces 0.2-0.4 Pg yr⁻¹ of calcium-bearing ashes, with estimated cumulative reserves of 4-8 Pg since 1980, suitable for enhanced weathering⁵³. Globally, the sugarcane industry produces ~47 Tg ash yr⁻¹, with the Australian sugar industry⁵⁷ alone producing 1 Tg yr⁻¹, enough to apply to 10,000 ha. Mill ash is a base cation, nutrient- and silica-rich by-product of fibrous cane residue combustion that improves cane yields by up to 40% at application rates of 50-60 t ha⁻¹ (dry weight)^{58,59}, with significant enhanced weathering potential.

Use of these mining and industrial wastes might be supplemented with substantial Ca-rich basic igneous silicate-rich rocks available as 38×10^8 ha (38 million km²) of surface exposed continental flood basalts produced episodically by massive volcanic eruptions throughout Earth's history⁶⁰. Major formations are located near to productive agricultural regions where rock might be required with estimated masses⁶⁰ sufficient for the annual requirements for enhanced weathering over many decades. For example, the USA might be served by the Central Atlantic Magmatic Province (eastern US) and the Columbia River basalts (Washington/Oregon), South America by the Paraná-Etendeka Traps and the Caribbean-Colombian Plateau, China by the Emeishan Traps, Russia by the Siberian Traps, the UK by the North Atlantic Igneous Province, western India by the Deccan Traps, and eastern India by the smaller Rajmahal Traps.

Adding crushed silicates to soils, whether residues or purposely mined, will likely have further economic benefit arising from their ability to help replenish eroded soil, and enhance SOC content, both serious global concerns threatening food security^{61,62}. Erosion rates from cropland soils outpace natural rates of formation by a factor of ten (average ~6 t ha⁻¹ yr⁻¹ loss vs. 0.6-0.8 t ha⁻¹ yr⁻¹ formation), limiting agricultural sustainability⁶¹. Erosion rates in US cropland soils, while declining some 50% over the past 30 years, still range from ~3 to ~13 t ha⁻¹ yr⁻¹, depending on agricultural practices⁶¹. In the European Union⁶³, soil erosion rates over 12.7 % of arable land exceed 5 t ha⁻¹ yr⁻¹. Depending on management practices, this situation is likely to worsen with climate change. Increased variations in rainfall patterns and intensity will make soils more susceptible to erosion. If agricultural soil erosion outpaces rates of soil formation, new methods will be needed to sustain and protect soils⁶¹, which have suffered global losses of 133 PgC from the original carbon stocks in the top 2 m over the past two centuries⁶².

Enhanced weathering might help reverse diminishing SOC stocks and retard soil erosion. Cation release from basalt weathering increases the cation exchange capacity of soils and nutrient availability^{64,65} and could improve SOC sequestration by resulting in higher inputs of organic carbon from roots and mycorrhizal fungi, which themselves promote soil aggregate formation and SOC stability⁶⁶. Increased formation of clay minerals from weathering of silicates could further increase SOC retention through a range of organo-mineral interactions, including adsorption reactions and physical protection of organic matter from decomposing organisms, which help build soil while improving quality⁶⁷. Increasing SOC in the rooting zone benefits crop yields in diverse agricultural soils of the tropics and sub-tropics⁶⁸. Operating across timescales from years to several decades, these effects, and others associated with

increasing mineral surface area available to trap soil carbon⁶⁹, could help rebuild soils and retard erosion. It may, therefore, contribute to increasing soil organic matter stocks, the goal of the '4 per 1000 Initiative: Soils for Food Security and Climate', proposed under the Agenda for Action of the 21st session of the United Nations Framework Convention on Climate Change⁷⁰. At present, however, the long-term effects of applying pulverized silicate rocks on the organic carbon content of agricultural soils is not understood and requires further research. Over time, adding crushed rocks to soils will change their porosity, and other factors governing hydrology, with feedbacks on crop performance, trace gas emissions, and the diversity and functioning of soil organisms, that are currently uncertain.

Enhanced weathering strategies not only capture carbon but could also help restore soils and resupply impoverished reserves of trace elements important for human nutrition⁷¹ and crop production⁷². Seven out of the top ten crops ranked according to global production data (sugarcane, rice, wheat, barley, sugar beet, soybean, and tomatoes) are classified as Si accumulators (> 1%)⁶⁵ and intensive cultivation and repeated removal of harvested products from the field is seriously depleting plant available Si in soils^{73,74}. In the US, for example, crop harvesting removes 19 million tonnes of Si annually⁷⁵. Annual depletion of soil Si by continuous intensive farming, coupled with low solubility of soil Si, has led to calls for the development of viable Si-fertilization practices in the near future to increase plant available pools and maintain crop yields⁷⁵⁻⁷⁷. Dissolution of crushed silicates, or Si-containing mining and industrial wastes, releases Si, replenishing the plant available form. The fate and transformation of enhanced weathering derived Si in the soil-plant continuum, and its long-term biogeochemical cycling⁷⁸, warrant future research in the context of mitigating Si-related yield constraints on agricultural crop production.

Crop production and protection

Amending soils with ground Ca/Mg-rich silicate rocks can improve crop yields and has a long history of being practiced on a small scale, especially in highly weathered tropical soils in Africa, Brazil^{79,80}, Malaysia^{81,82} and Mauritius⁸³, as well as rejuvenating lateritic soils and promoting tree establishment in Europe^{84,85}. Consequently, enhanced weathering of crushed silicates has a number of proven and expected benefits for temperate and tropical croplands that could improve its prospects for large-scale deployment^{21,29}. Sugarcane trials with crushed basalt applications in excess of 20 t ha⁻¹ in combination with standard NPK fertilizer treatments increased yields by up to 30% over five successive crops on the highly weathered soils of Mauritius compared with plots receiving fertilizer and no basalt addition⁸³. Sugarcane, grown extensively on acidic, nutrient-poor highly weathered soils, generates approximately \$43 billion a year to Brazil's economy and \$1.5 billion a year in export earnings for Australia, suggesting such effects could offer significant economic incentives for the industry to adopt the practice more widely.

Few field and experimental studies have explicitly investigated basalt treatments on temperate croplands to test directly the effects on yields and soil properties but numerous field and greenhouse studies have documented the benefits of applying silicates and modified silicate wastes to crop production across the USA. This practice extends back to 1871 when the first patent for using Si-rich slag as a fertilizer was granted⁷⁵. Consequently, decades of research has established processed Ca-silicate slag as an effective liming material and Si-fertilizer, without yet recognizing its CO₂ capture potential. Studies include field trials in Florida and Louisiana, where silicate slag applications increased sugarcane, maize and rice production and elsewhere in New Jersey where silicate slag increased yields of a wide range of crops including

winter wheat, oats, cabbage and corn, with residual benefits continuing up to 3-4 years after the last application⁷⁵.

By generating alkalinity as they weather, silicate rocks reduce soil acidification caused by overuse of ammonium and elemental sulphur fertilizers, urea, growth of nitrogen-fixing legumes and repeated crop harvesting. Acidification of agricultural soils is a worldwide problem and reversing it improves nutrient uptake, root growth and crop yields. Neutralizing acidic soils also reduces metal toxicity (e.g., aluminium and manganese) and increases P availability, especially in highly weathered acidic tropical soils, where metal oxides strongly bind remaining P reserves⁶⁴. Plant-induced weathering of basalt supplies trace amounts of P in the form of calcium phosphate, the primary source of P in most ecosystems and fertilizers, and adds plant-essential trace nutrients. For example, most of the nutrient-mined tropical soils in developing countries⁴ are deficient in K, and crushed silicate rocks applied as slow-release K fertilisers can help sustain profitable crop production while achieving the primary goal of carbon sequestration⁸⁶.

Although not regarded as an essential element for plant growth, Si benefits productivity by enhancing the resilience of plants against abiotic stresses include drought, salinity and heat^{72,87}, all of which are expected to worsen with future climate change and sea-level rise². Simultaneous increases in plant available Si in soils amended with silicates reduces the uptake of heavy metals (e.g., cadmium arsenic and lead) in the edible parts of agricultural crops⁸⁸⁻⁹². Increased silica uptake from the soil is a competitive inhibitor of arsenic uptake in rice, for example, which is a widespread human health issue in southeast Asia⁹¹. Cadmium uptake in wheat is also reduced, and this is an important issue where prolonged application of fertilisers, especially single super phosphate, has generated toxicity in agricultural soils worldwide⁹².

Benefits for crop protection against biotic threats from silicate weathering arise from production of soluble silicic acid that is readily taken up by plants, thereby improving stem strength and increasing resistance to pests and diseases in major temperate (e.g., soybean, wheat)²⁹ and tropical (e.g., sugarcane, maize, rice, oil palm)²¹ crops. Greenhouse and field trials have shown Si augments host plant resistance to disease and actively suppresses diseases by influencing incubation period, latent period, lesion number and lesion size⁷⁵. Staple cereal crops, such as rice, maize and barley, are major silica accumulators, with silicic acid transporters responsible for uptake into the root cortex and transfer to the xylem^{93,94}. Silicic acid uptake acts by priming the defence pathways, for example, jasmonic acid (JA)-dependent plant immunity, and strengthens cell walls in leaves and roots⁹⁵. This multi-mechanistic mode of action offers durable and broad-spectrum protection against a wide range of insect herbivores and pathogens.

Accordingly, Si-induced resistance offers tangible opportunities to protect temperate crops and tropical cereals against emerging and enduring pests, an increasing number of which are becoming resistant to pesticides. For example, the recent large-scale invasion of the fall armyworm (*Spodoptera frugiperda*) in Africa reduced maize production. However, Si-treated maize may restrict the spread of this invasive pest by significantly decreasing fecundity⁹⁶. Si-induced resistance against phloem-feeding Hemiptera pests may also reduce the spread of major viral diseases that are transmitted by these insects, such as maize streak virus, the most damaging viral disease for this crop in Africa⁹⁷. Cell wall strengthening and JA dependent defence pathways are involved in resistance against the parasitic weed *Striga*^{98,99}, which causes devastating losses of yields of rain-fed rice, maize, sorghum and millet in sub-Saharan Africa, costing the African economy over seven billion US\$ annually¹⁰⁰.

Genetic assessment of crop attributes, e.g., capacity to recruit and associate with mycorrhizal fungi, could accelerate development of new, faster weathering, crop varieties.

Selection for new cereal varieties with increased performance (e.g., uptake and accumulation of silica) in response to silicate rock/agro-mineral fertilization could be achieved through conventional breeding and/or using gene editing techniques to modify elite varieties (e.g., CRISPR-Cas9). Engineering crop varieties that are effectively able to exploit soil amended with crushed silicate rocks would potentially deliver significant benefits by improving nutrient supply to fertilize production and increasing protection against pests and diseases as well as promoting weathering to raise pH, cation exchange capacity and increase soil organic carbon capture. However, such potential benefits require demonstration in replicated field trials (**Table 1**).

A further co-benefit may arise from agricultural application of crushed silicate rocks to soils suppressing emissions of the powerful and long-lived greenhouse gas N₂O and avert CO₂ emissions by liming. Liming with CaCO₃ can release CO₂ when it is applied to acidic soils (pH <6) typical of agricultural lands^{16,17,101}; in the USA, liming contributes 2% of agricultural greenhouse gas emissions¹⁶. In contrast, silicate weathering consistently consumes CO₂ to produce bicarbonate and carbonate ions. By increasing soil pH as they weather, silicates may also reduce emissions of N₂O, as found with liming¹⁰². Preliminary tests with a replicated field experiment support this suggestion with the soil N₂O flux from heavily fertilized maize plots decreasing by ~50% with the application 10 kg m⁻² of pulverized basalt with no effect on soil respiration¹⁰³. Basalt-treated arable fields may, thus, lower the current substantial global soil-atmosphere flux from croplands¹⁰⁴ of 4-5 Tg N₂O-N yr⁻¹ as a by-product of weathering.

In summary, potential ancillary benefits of CO₂ capture with rocks and agriculture include fertilization of yields and reduced use and cost of fertilizers, including those with finite geological reserves (rock phosphate)⁴, neutralizing soil acidification, suppressing/averting soil greenhouse gas (N₂O and CO₂) emissions, restoration of micro-nutrients important for human nutrition, and replacement of soils lost by erosion (**Figures 1 and 4**). Additionally, increased crop protection from insect herbivores and pathogens, and avoidance of toxic metal uptake, resulting from release and uptake of silica, could decrease pesticide use and cost and improve yields, further safeguarding food security (**Figure 4**).

Environmental impacts

Development of widespread mining, grinding and spreading operations will likely have negative environmental and ecological impacts, especially if linked to tropical deforestation near areas of high biodiversity value that would require careful management²¹. However, the severity of the threat to biodiversity and local ecology would depend on the extent to which silicate waste materials are utilized thereby reducing the need for mining operations. Judicious selection of source materials, such as basalt instead of faster-weathering but Ni- and Cr-enriched ultramafic rock types, for example, minimizes dangers of toxic metal contamination (**Figure 2**). Avoiding inhalation of dust particles during mining, grinding and spreading will be important because these can cause silicosis. Additionally, particles washing into rivers, and ultimately the oceans, might cause increased turbidity, sedimentation and pH changes, with unknown impacts for marine biodiversity and function²¹.

In addition to downstream alkalinity addition (discussed earlier), enhanced silicate weathering can be expected to increase dissolved silica fluxes to rivers and oceans. This may partially help mitigate effects of N and P in runoff from agricultural regions. Increased Si:N and/or Si:P ratios in runoff reaching coastal waters from soils amended with silicates might favour the growth of diatoms over problematic non-siliceous algae that produce toxins, red tides (dinoflagellate blooms), foam (Phaeocystis blooms) and scum (cyanobacterial blooms)^{105,106}. Such a changed nutrient balance could also beneficially preserve or increase

downstream food web and fisheries production because diatoms are the preferred diet of pelagic and benthic grazers, mostly copepods and bivalves^{105,106}, and increase marine biological CO₂ drawdown and storage^{12,18}, with economic benefits in particular regions. For example, the Great Barrier Reef is adjacent to the main sugarcane growing regions in Australia, where adding crushed basalt to soils may not only enhance sugarcane production, but also improve runoff and ground water chemistry while countering ocean acidity via alkalinity addition. However, the hypothesized benefits and impacts of land-based enhanced weathering on aquatic food webs have yet to be proven and require further research.

Outlook

Effective climate change mitigation requires an expanding portfolio of actions for extracting and sequestering CO₂ alongside urgent reduction of CO₂ emissions^{2,6-9,107}, as highlighted by the United Nations Environment Programme¹⁰⁸. In our analysis, nations that contributed most to the problem have the potential to be big players in mitigation by addressing the substantial engineering challenge of developing an operational enhanced weathering industry (**Figure 3**). The challenge may be suited to international cooperation between nations, including provision of assets needed for implementation in developing countries. However, like the extensive deployment of any CDR approach, enhanced weathering not only has to be evaluated and proven in field-scale trials, with CO₂ sequestration potential better understood, but also has to be socially and environmentally acceptable. This requires extensive, detailed, risk assessment, public participation and transparency¹⁰⁹⁻¹¹⁰.

Adapting agricultural practices to manage soils, alongside reforestation efforts, for atmospheric carbon removal could help slow the rate of climate change, if combined with near-term emission reductions^{2,6,107,108}. Continued high emissions, on the other hand, may force society to consider more expensive industrial-scale carbon clean-up operations to stabilize climate⁶. Methods of CO₂ extraction, like BECCS and direct air capture (DAC) of CO₂, require large-scale infrastructure development and investment with substantial energy and resource demands and potential land-use conflicts threatening global food security^{6,8,23}. Spurring investment and bringing down costs of CDR options (e.g., BECCS and DAC), requires some form of market linked to the price of carbon. Investment incentives for enhanced weathering are potentially broader and include increased yields, improved soils, reduced agrochemical costs, improved runoff water quality in environmentally sensitive areas, and potential benefits to marine life.

We conclude that substituting a weatherable silicate rock, such as basalt, or silicate waste, for limestone, and increasing application rates over those used in conventional liming operations, may offer a pragmatic, rapidly deployable global carbon cycle intervention strategy. More broadly, if proven effective, and undertaken carefully to minimise undesirable impacts, enhanced weathering may have untapped potential for addressing the United Nations Sustainable Development Goals (SDGs) adopted by 193 countries in 2015¹¹¹. For example, we highlight how sequestering CO₂ constitutes action on climate change (SDG 13), restoring soils and promoting sustainable agriculture contributes to zero hunger (SDG 2), helping protect the oceans from acidification conserves global resources in life below water (SDG 14), reducing agrochemical usage and recycling wastes helps with sustainable consumption and production (SDG 12), and improving agricultural production and restoring degraded soils contributes to land sparing (SDG 15) (**Figure 4**). However, there is an urgent need to address unanswered technical and social questions and develop rigorous audited testing in the field where the full elemental cycles can be closed, efficacy of CO₂ capture quantified, and the risks, benefits, socio-economics, techno-economics, and ethics assessed (**Table 1**).

References

1. Rosenzweig, C. et al. Assessing agricultural risks of climate change in the 21st century in a global gridded crop model intercomparison. *Proc. Nat. Acad. Sci., USA* **111**, 3268-3273, doi:10.1073/pnas.1222463110 (2014).
2. Intergovernmental Panel on Climate Change, *Climate Change 2014: Mitigation of Climate Change*, Edenhofer, O. et al. Eds. (Cambridge Univ. Press, New York, 2014).
3. Godfray, H.C.J. et al. Food security: the challenge of feeding 9 billion people. *Science* **327**, 810-818 (2010).
4. Amundson, J. et al. Soil and human security in the 21st Century. *Science* **348**, 1261071 (2015).
5. Paris Agreement: UNFCCC secretariat, available at http://unfccc.int/paris_agreement/items/9485.php
6. Hansen, J. et al. Young people's burden: requirement of negative CO₂ emissions. *Earth Syst. Dynam.* **8**, 577-616 (2017).
7. Gasser, T. et al. Negative emissions physically needed to keep global warming below 2 °C. *Nat. Comms.* **6**, doi:10.1038/ncomms8958 (2015).
8. Anderson, K. & Peters, G. The trouble with negative emissions. *Science* **354**, 182-183 (2016).
9. Rockstrom, J. et al. A roadmap for rapid decarbonisation. *Science* **355**, 1269-1271 (2017).
10. Lee, H. Turning the focus to solution. *Science* **350**, 1007 (2015).
11. Schuiling, R.D. & Krijgsman, P. Enhanced weathering: an effective and cheap tool to sequester CO₂. *Clim. Change* **74**, 349-354 (2006).
12. Kohler, P., Hartman, J. & Wolf-Gladrow, D.A. Geoengineering potential of artificially enhanced silicate weathering of olivine. *Proc. Natl. Acad. Sci. USA* **107**, 20228-20233 (2010).
13. Taylor, L.L. et al. Enhanced weathering strategies for stabilizing climate and averting ocean acidification. *Nat. Clim. Change* **6**, 402-406 (2016).
14. Ciais, P. et al. In: *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* (eds. Stocker, T.F. et al.). Cambridge University Press, Cambridge (2013).
15. Renforth, P. & Henderson, G. Assessing ocean alkalinity for carbon sequestration. *Rev. Geophys.* **55**, doi: 10.1002/2016RG000533 (2017).
16. West, T.O. & McBride, A.C. The contribution of agricultural lime to carbon dioxide emissions in the United States: dissolution, transport, and net emissions. *Agricult. Ecosys. Environ.* **108**, 145-154 (2005).
17. Thorley, R.M.S., Taylor, L.L., Banwart, S.A., Leake, J.R. & Beerling, D.J. The role of forest trees and their mycorrhizal fungi in carbonate rock weathering and its significance for global carbon cycling. *Plant, Cell Environ.* **38**, 1947-1961 (2015).
18. Hartmann, J. et al. Enhanced chemical weathering as a geoengineering strategy to reduce atmospheric carbon dioxide, supply nutrients, and mitigate ocean acidification. *Rev. Geophys.* **51**, 113-149 (2013).
19. Cripps, G., Widdicombe, S., Spicer, J.I. & Findlay, H.S. Biological impacts of enhanced alkalinity in *Carcinus maenas*. *Mar. Pollut. Bull.* **71**, 190-198 (2013).
20. Albright, R. et al. Reversal of ocean acidification enhances net coral reef calcification. *Nature* **531**, 362-365 (2016).
21. Edwards, D.P. et al. Climate change mitigation: potential benefits and pitfalls of enhanced rock weathering in tropical agriculture. *Biol. Letts.* **13**, 20160715 (2017).
22. Smith, P. et al. Biophysical and economic limits to negative CO₂ emissions. *Nat. Clim. Change* **6**, 42-50 (2016).
23. Field, C.B. & Mach, K.J. Rightsizing carbon dioxide removal. *Science* **356**, 706-707 (2017).
24. Peters, S.C., Blum, J.D., Driscoll, C.T. & Likens, G.E. Dissolution of wollastonite during the experimental manipulation of Hubbard Brook Watershed 1. *Biogeochem.* **67**, 309-329 (2004).
25. Shao, S. et al. Long-term responses in soil solution and stream-water chemistry at Hubbard Brook after experimental addition of wollastonite. *Environ. Chem.* **13**, 528-540 (2016).
26. Hartmann, J. & Kempe, S. What is the maximum potential for CO₂ sequestration by "simulated" weathering at the global scale? *Naturwissenschaften* **95**, 1159-1164 (2008).

27. Monfreda, C., Ramankutty, N. & Foley, J.A. Farming the planet: 2. Geographic distribution of crop areas, yields, physiological types, and net primary production in the year 2000. *Global Biogeochem. Cycles* **22**, doi:10.1029/2007GB002947 (2008).
28. Fritz, S. et al. Downgrading recent estimates of land area available for biofuel production. *Env. Sci. Tech.* **47**, 1688-1694 (2013).
29. Kantola, I.B. et al. Potential of global croplands and bioenergy crops for climate change mitigation through deployment for enhanced weathering. *Biol. Letts.* **13**, 20160714 (2017).
30. Renforth P. The potential of enhanced weathering in the UK. *Int. J. Greenh. Gas Cont.* **10**, 229–243 (2012) (doi:10.1016/j.ijggc.2012.06.011).
31. Hangx, S.J.T. & Spiers, C.J. Coastal spreading of olivine to control atmospheric CO₂ concentrations: a critical analysis of viability. *Int. J. Greenh. Gas Cont.* **3**, 757–767 (2009).
32. Kohler, P. et al. Geoengineering impact of open ocean dissolution of olivine on atmospheric CO₂, surface ocean pH and marine biology. *Env. Res. Lett.* **8**, 014009 (2013).
33. Meysman, F.J.R. & Montserrat, F. Negative emissions via enhanced silicate weathering in coastal environments. *Biol. Letts.* **13**, 20160905 (2017).
34. Montserrat, F. et al. Olivine dissolution in seawater: implications for CO₂ sequestration through enhanced weathering in coastal environments. *Env. Sci. Tech.* **51**, 3980-3972 (2017).
35. ten Berge, H.F.M. et al. Olivine weathering in soil, and its effects on growth and nutrient uptake in ryegrass (*Lolium perenne* L.): a pot experiment. *PLoS One* **7**, e42098 (2012).
36. Renforth, P., von Strandmann, P.A.E. & Henderson, G.M. The dissolution of olivine added to soil: implications for enhanced weathering. *App. Geochem.* **61**, 109-118 (2015).
37. Shoji, S., Nanzyo, M. & Dahlgren, R.A. (eds) Volcanic ash soils. Genesis, properties and utilization. *Development in Soil Sciences* 21. Elsevier, Amsterdam, pp. 288 (1993).
38. Hinsinger, P. et al. Plant-induced weathering of basaltic rock: experimental evidence. *Geochim. Cosmochim. Acta* **65**, 137-152 (2001).
39. Brantley, S.L., Kubicki, J.D. & White, A.F. Kinetics of water-rock interaction. Springer, New York (2008).
40. Beerling, D.J., Taylor, L., Banwart, S.A., Kantzas, E.P., Lomas, M., Mueller, C., Ridgwell, A. & Quegan, S. Defining the ‘negative emission’ capacity of global agriculture deployed for enhanced rock weathering. American Geophysical Union, Fall General Assembly, abstract #GC21J-04 (2016).
41. Wollenberg, E. et al. Reducing emissions from agriculture to meet the 2 °C target. *Global Change Biol.* **22**, 3859-3864 (2016).
42. Taylor, L.L., Beerling, D.J., Quegan, S. & Banwart, S.A. Simulating carbon capture by enhanced weathering with croplands: an overview of key processes highlighting areas of future model development. *Biol. Letts.* **13**, 20160868 (2017).
43. Moosdorf, N., Renforth, P. & Hartmann, J. Carbon dioxide efficiency of terrestrial weathering. *Environ. Sci. Tech.* **48**, 4809-4816 (2014).
44. Harley, A.D. & Gilkes, R.J. Factors influencing the release of plant nutrient elements from silicate rock powders: a geochemical review. *Nutr. Cycling Agroecosyst.* **56**, 11-26 (2000).
45. Akter, M. & Akagi, T. Effect of fine root contact on plant-induced weathering of basalt. *Soil Sci. Plant Nutr.* **51**, 861-871 (2005).
46. Akter, M. & Akagi, T. Dependence of plant-induced weathering of basalt and andesite on nutrient conditions. *Geochem. J.* **44**, 137-150 (2010).
47. Burghel, C. et al. Mineral nutrient mobilization by plants from rock: influence of rock type and arbuscular mycorrhiza. *Biogeochem.* **124**, 187-203 (2015).
48. Goulding, K.W.T. Soil acidification and the importance of liming agricultural soils with particular reference to the United Kingdom. *Soil Use Manag.* **32**, 390-399 (2016).
49. Reichl, C., Schatz, M. & Zsak, G. World Mining Data. Volume 32. Minerals Production. International Organization Committee for the World Mining Congress, Vienna, pp. 255 (2017).
50. Popp, A. et al. Land-use futures in the shared socio-economic pathways. *Glob. Env. Change* **42**, 331-345, doi: <https://doi.org/10.1016/j.gloenvcha.2016.10.002> (2017).
51. Le Quéré, C. et al. Global carbon budget 2016. *Earth Sys. Sci. Data* **8**, 605-649 (2016).
52. Hansen, J. et al. Assessing “dangerous climate change”: required reduction of carbon emissions to protect young people, future generations and nature. *PLoS One* **8**, e81648 (2013).

53. Renforth, P. et al. Silicate production and availability for mineral carbonation. *Env. Sci. Tech.* **45**, 2035-2041 (2011).
54. Nunes, J.M.G., Kautzmann, R.M. & Oliveira, C. Evaluation of the natural fertilizing potential of basalt dust wastes from the mining district of Nova Prata (Brazil). *J. Clean. Prod.* **84**, 649-656 (2014).
55. White, J.W., Holben, F.J. & Jeffries, C.D. The agricultural value of specially prepared blast-furnace slag. Penn. State Coll., School Agr. Exp. Stat. Bull. No. 341 (1937).
56. Yi, H. et al. An overview of utilization of steel slag. *Proc. Env. Sci.* **16**, 791-801 (2012).
57. Barry, G. A., Price, A. M., & Lynch, P. J. Some implications of the recycling of sugar industry by-products. *Proc. Aust. Soc. Sugar Cane Technol.* **20**, 52-55 (1998).
58. Kingston, G. A role for silicon, nitrogen and reduced bulk density in yield responses to mill ash and filter mud/ash mixtures. *Proc. Aust. Soc. Sugar Cane Technol.* **21**, 114-121 (1999).
59. Berthelsen, S., Hurney, A. H., Kingston, G., Rudd, A., Garside, A., & Noble, A. D. Plant cane responses to silicated products in the Mossman, Innisfail and Bundaberg districts. *Proc. Aust. Soc. Sugar Cane Technol.* **23**, 297-303 (2001).
60. Bryan, S.E. & Ernst, R.E. Revised definition of Large Igneous Provinces (LIPs). *Earth Sci. Rev.* **86**, 175-202 (2008).
61. Montgomery, D.R. Soil erosion and agricultural sustainability. *Proc. Natl. Acad. Sci., USA* **104**, 13268-13272 (2007).
62. Sanderman, J., Hengl, T. & Fiske, G.J. Soil carbon debt of 12,000 years of human land use. *Proc. Natl. Acad. Sci., USA* **114**, 9575-9580 (2017).
63. http://ec.europa.eu/eurostat/statistics-explained/index.php/Agri-environmental_indicator_-_soil_erosion
64. Gillman, G. P. The effect of crushed basalt scoria on the cation exchange properties of a highly weathered soil. *Soil Sci. Soc. Am. J.* **44**, 465-468 (1980).
65. Gillman, G. P., Burkett, D. C., & Coventry, R. J. A laboratory study of application of basalt dust to highly weathered soils: effect on soil cation chemistry. *Aust. J. Soil Res.* **39**, 799-811 (2001).
66. Wright, S.F. & Upadhyaya, A. A survey of soils for aggregate stability and glomalin, a glycoprotein produced by hyphae of arbuscular mycorrhizal fungi. *Plant Soil* **198**, 97-107 (1998).
67. Baldock, J.A., & Skjemstad, J. O. Role of the soil matrix and minerals in protecting natural organic materials against biological attack. *Organ. Geochem.* **31**, 697-710 (2000).
68. Lai, R. Beyond Copenhagen: mitigating climate change and achieving food security through soil carbon sequestration. *Food Sec.* **2**, 169-177 (2010).
69. Yu, G. et al. Mineral availability as a key regulator of soil carbon storage. *Env. Sci. Tech.* **51**, 4960-4949 (2017).
70. Minasny, B. et al. Soil carbon 4 per mille. *Geoderma* **292**, 59-86 (2017).
71. Shewry, P.R., Pellny, T.K. & Lovegrove, A. Is modern wheat bad for our health? *Nat. Plants* **2**, <http://dx.doi.org/10.1038/nplants.2016.97> (2016).
72. Guntzer, F., Keller, C., Meunier, J.-D. Benefits of plant silicon for crops: a review. *Agron. Sustainable Devel.* **32**, 201-213 (2012).
73. Guntzer, F. et al. Long-term removal of wheat straw decreases soil amorphous silica at Broadbalk, Rothamsted. *Plant Soil* **352**, 173-184 (2012).
74. Klotzbücher, T. et al. Plant-available silicon in paddy soils as a key factor for sustainable rice production in Southeast Asia. *Basic App. Ecol.* **16**, 665-673 (2015).
75. Tubana, B.S., Babu, T. & Datnoff, L.E. A review of silicon in soils and plants and its role in US agriculture: history and future perspectives. *Soil Sci.* **181**, 393-411 (2016).
76. Mecfel, J. et al. Effect of silicon fertilizers on silicon accumulation in wheat. *J. Plant Nutr. Soil Sci.* **170**, 769-772 (2007).
77. Marxen, A. et al. Interaction between silicon cycling and straw decomposition in a silicon deficient rice production system. *Plant Soil* **398**, 153-163 (2016).
78. Vandevenne, F.I. et al. Silicon pools in human impacted soils of temperate zones. *Global Biogeochem. Cycles* **29**, 1439-1450 (2015).
79. Leonardos, O.H., Fyfe, W.S. & Kronberg, B.I. The use of ground rocks in laterite systems: an improvement to the use of conventional soluble fertilizers? *Chem. Geol.* **60**, 361-370 (1987).

80. Van Straaten, P. Farming with rocks and minerals: challenges and opportunities. *Ann. Braz. Acad. Sci.* **78**, 731-747 (2006).
81. Anda, M., Shamshuddin, J. & Fauziah, C.I. Improving chemical properties of a highly weathered soil using finely ground basalt rocks. *Catena* **124**, 147-161 (2015).
82. Anda, M., Shamshuddin, J. & Fauziah, C.I. Increasing negative charge and nutrient contents of a highly weathered soil using basalt and rice husk to promote cocoa growth under field conditions. *Soil Till. Res.* **132**, 1-11 (2013).
83. de Villiers O.D. Soil rejuvenation with crushed basalt in Mauritius. Part I – consistent results of world-wide interest. *Int. Sugar J.* **63**, 363–364 (1961).
84. Albert, R. Untersuchungen über die Verwendbarkeit von Gesteinsabfällen verschiedener Herkunft und Art zur Verbesserung geringwertiger Waldöden. *Forstarchiv* **14**, 237-240 (1938).
85. Albert, R. Untersuchungen über Tiefenwirkung des Vollumbruches und der Basaltdüngung. *Forstarchiv* **16**, 231-232 (1940).
86. Basak, B.B., Sarkar, B., Biswas, D.R., Sarkar, S., Sanderson, P. & Naidu, R. Bio-intervention of naturally occurring silicate minerals for alternative source of potassium: challenges and opportunities. *Adv. Agron.* **141**, 115-145 (2017).
87. Ma, J.F. Role of silicon in enhancing the resistance of plants to biotic and abiotic stresses. *Soil Sci. Plant Nutr.* **50**, 11–18 (2004).
88. Rizwan, M., Meunier, J.-D., Miche, H. & Keller, C. Effect of silicon on reducing cadmium toxicity in durum wheat (*Triticum turgidum* L. cv. Claudio W.) grown in a soil with aged contamination. *J. Hazard. Mat.* **209**, 326-334 (2012).
89. Seyfferth, A.L. & Fendorf, S. Silicate mineral impacts on the uptake and storage of arsenic and plant nutrients in rice (*Oryza sativa* L.). *Env. Sci. Tech.* **46**, 13176-13183 (2012).
90. Ning, D. et al. Impacts of steel-slag-based silicate fertilizer on soil acidity and silicon availability and metals-immobilization in a paddy soil. *PLoS One* **11**, e0168163 (2016).
91. Bogdan, K. & Schenk, M.K. Arsenic in rice (*Oryza sativa* L.) related to dynamics of arsenic and silicic acid in paddy soils. *Env. Sci. Tech.* **42**, 7885-7890 (2008).
92. Greger, M., Kabir, A.H., Landberg, T., Maity, P.J. & Lindberg, S. Silicate reduces cadmium uptake into cells of wheat. *Environ. Poll.* **211**, 90-97 (2016.)
93. Ma, J.F. & Yamaji, N. A cooperative system of silicon transport in rice. *Trends Plant Sci.* **20**, 435-442 (2015).
94. Yamaji, N. et al. Orchestration of three transporters and distinct vascular structures in node for intervascular transfer of silicon in rice. *Proc. Natl. Acad. Sci. USA* **112**, 11401-11406 (2015).
95. J. Van Bockhaven, De Vleeschauwer, D. & Höfte, M. Towards establishing broad-spectrum disease resistance in plants: silicon leads the way. *J. Exp. Bot.* **64**, 1281-1291 (2013).
96. Alvarenga, R. et al. Induction of resistance of corn plants to *Spodoptera frugiperda* (J. E. Smith, 1797) (Lepidoptera: Noctuidae) by application of silicon and gibberellic acid. *Bull. Entomol. Res.* **107**, 527-533 (2017).
97. Yang, L. et al. Silicon amendment to rice plants impairs sucking behaviors and population growth in the phloem feeder *Nilaparvata lugens* (Hemiptera: Delphacidae). *Sci Rep.* **7**, Article number: 1101 (2017).
98. Swarbrick, P.J. et al. Global patterns of gene expression in rice cultivars undergoing a susceptible or resistant interaction with the parasitic plant *Striga hermonthica*. *New Phytol.* **179**, 515-529 (2008).
99. Mutuku, J.M. et al. The WRKY45-dependent signaling pathway is required for resistance against *Striga hermonthica* parasitism. *Plant Physiol.* **168**, 1153-1163 (2015).
100. Yoder, J.I. & Scholes, J.D. Host plant resistance to parasitic weeds; recent progress and bottlenecks. *Curr. Opin. Plant Biol.* **13**, 478-488 (2010).
101. Hamilton, S.K. et al. Evidence for carbon sequestration by agricultural liming. *Global Biogeochem. Cycles* **21**, doi:10.1029/2006GB002738 (2007).
102. Gibbons, J.M. et al. Sustainable nutrient management at field, farm and regional level: soil testing, nutrient budgets and the trade-off between lime application and greenhouse gas emissions. *Agri. Ecosys. Environ.* **188**, 48–56 (2014).

103. Kantola, I.B., Masters, M.D., Wolz, K. J. & DeLucia, E. H. Climate change mitigation through enhanced weathering in bioenergy crops. American Geophysical Union, Fall General Assembly, abstract #H13B-1358 (2016)
104. Reay, D.S. et al. Global agriculture and nitrous oxide emissions. *Nat. Clim. Change* **2**, doi: 10.1038/nclimate1458 (2012).
105. Sommer, U. et al. Pelagic food web configurations at different levels of nutrient richness and their implications for the ratio fish production : primary production. *Hydrobiol.* **484**, 11-20 (2002).
106. Ragueneau, O., Conley, D. J., Leynaert, A., Longphuir, S. N. & Slomp, C. P. In V. Ittekkot, D. Unger, C. Humborg, & N. T. An (Eds.), *The Silicon Cycle: Human Perturbations and Impacts on Aquatic Systems* (pp. 163-195). Washington, D.C.: Island Press (2006).
107. Griscom, B.W. et al. Natural climate solutions. *Proc. Natl. Acad. Sci. USA* **114**, 11645-11650 (2017).
108. UNEP The Emissions Gap Report 2017. United Nations Environment Programme (UNEP), Nairobi (2017).
109. Wright, M.J., Teagle, D.A.H. & Feetham, P.M. A quantitative evaluation of the public response to climate engineering. *Nat. Clim. Change* **4**, doi:10.1038/nclimate2087 (2014).
110. Pidgeon, N.F. & Spence, E. Perceptions of enhanced weathering as a biological negative emissions options. *Biol. Lett.* **13**, 2017002.
111. United Nations. Transforming our world: the 2030 agenda for sustainable development. A/Res/70/1.

Acknowledgements

We gratefully acknowledge funding from the Leverhulme Trust through a Leverhulme Research Centre Award (RC-2015-029). L.L.T. was supported by an ERC advanced grant awarded to D.J.B. (CDREG, 322998). Rachel Thorley is thanked for assistance with Figure 1. We dedicate this paper to the memory of Professor William (Bill) G. Chaloner FRS (1928-2016), a passionate scientific polymath and extraordinary mentor to generations of researchers.

Contributions

D.J.B. conceived the study and wrote the first draft of the manuscript, with input on sections and addition of appropriate references from all authors. E.K., L.L.T. and M.K. undertook data analysis.

Competing Interests

The authors declare no competing interests.

Reprints and permissions information is available at www.nature.com/reprints.

Correspondence and requests for materials should be addressed to D.J.B.

Figure legends

Figure 1. Summary of potential effects of applying crushed basalt or silicate-rich wastes, such as sugarcane mill ash, on crop productivity, nutrient use efficiency, water quality and CO₂ capture and sequestration, due to weathering.

Figure 2. Metal and phosphorus concentrations in a range of continental flood basalts (CFB) and ultra-basic rocks (peridotites). Values means \pm s.e., measurement n for each is given in the lower graph. Data obtained from GEOROC (Geochemistry of Rocks of the Oceans and Continents) database maintained by the Max Planck Institute for Chemistry in Mainz (<http://georoc.mpch-mainz.gwdg.de/georoc/>)

Figure 3. (a) Calculated 75% of the most productive annual croplands, based on a reanalysis of 10 km \times 10 km latitude-longitude resolution data for the year 2000, where crop net primary production (NPP) was calculated by converting FAO yield data²⁷, (b) ranked top 20 arable crop producing countries and (c) cumulative CO₂ emissions from all sources by country. CO₂ data from the Global Carbon Atlas: <http://globalcarbonatlas.org>

Figure 4. Enhanced weathering could address the 21st Century threats to society of climate, food and soil security.

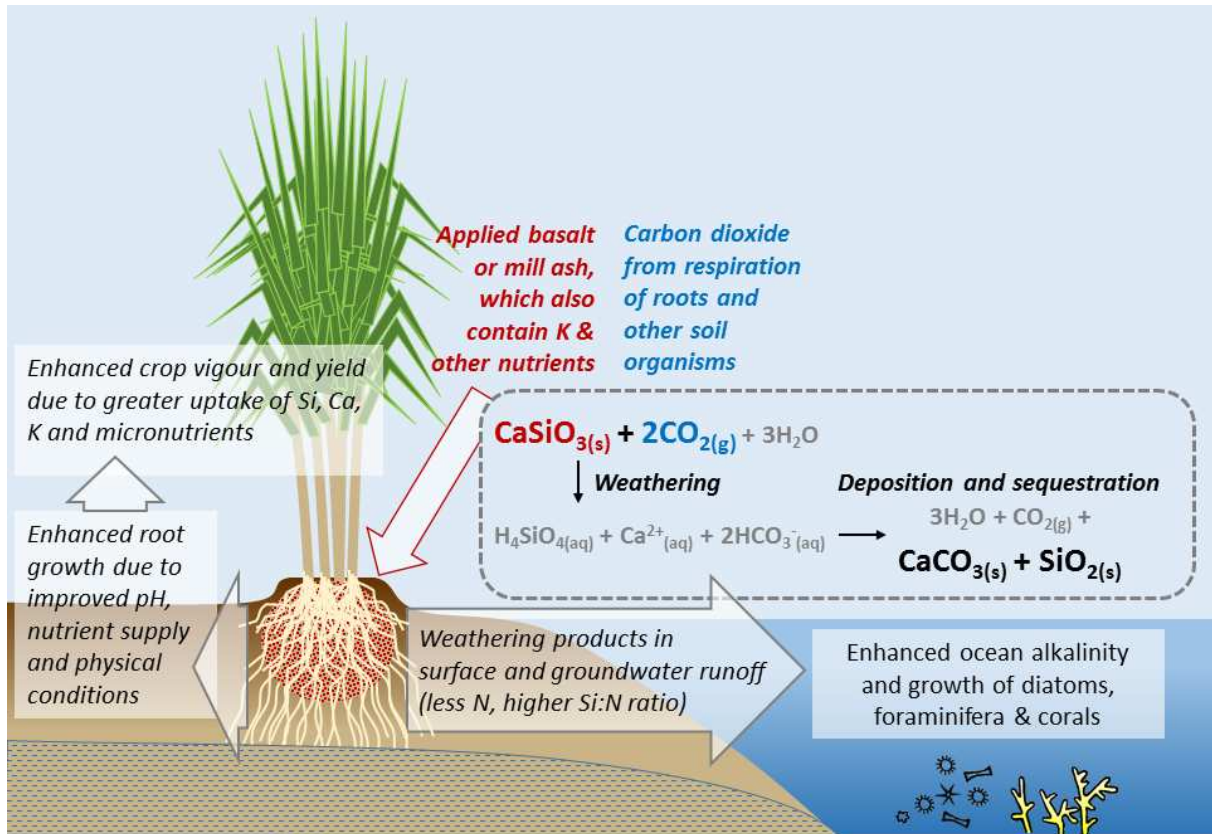


Figure 1
(Beerling et al.)

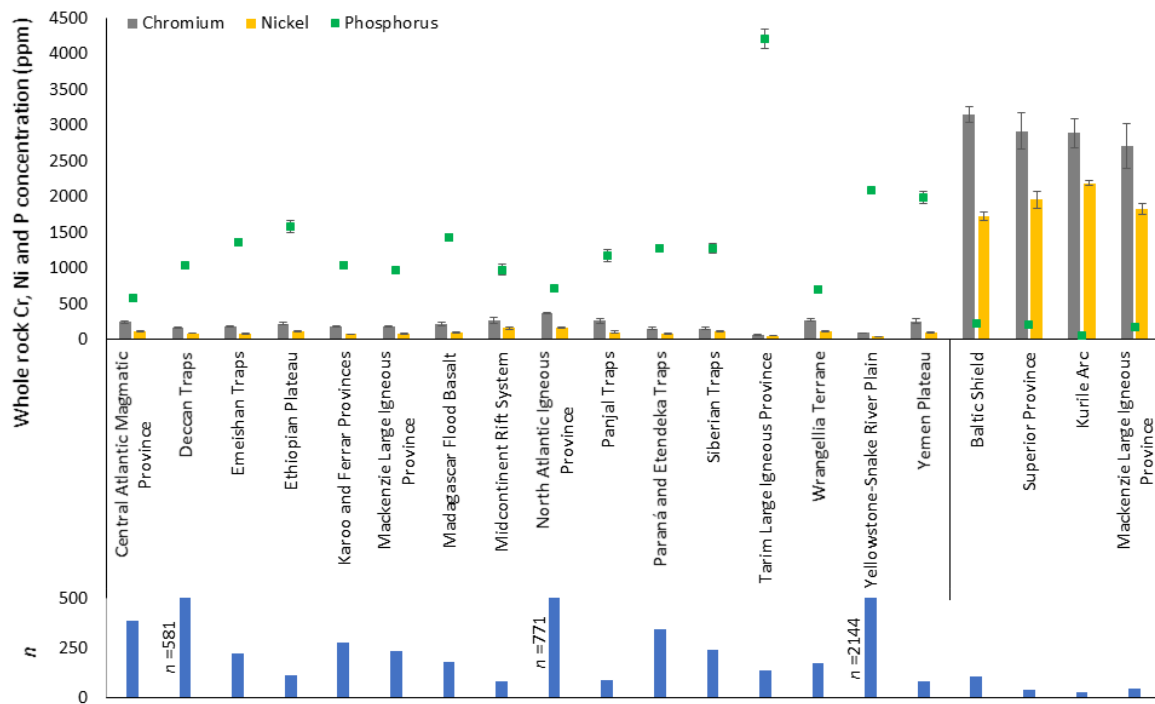
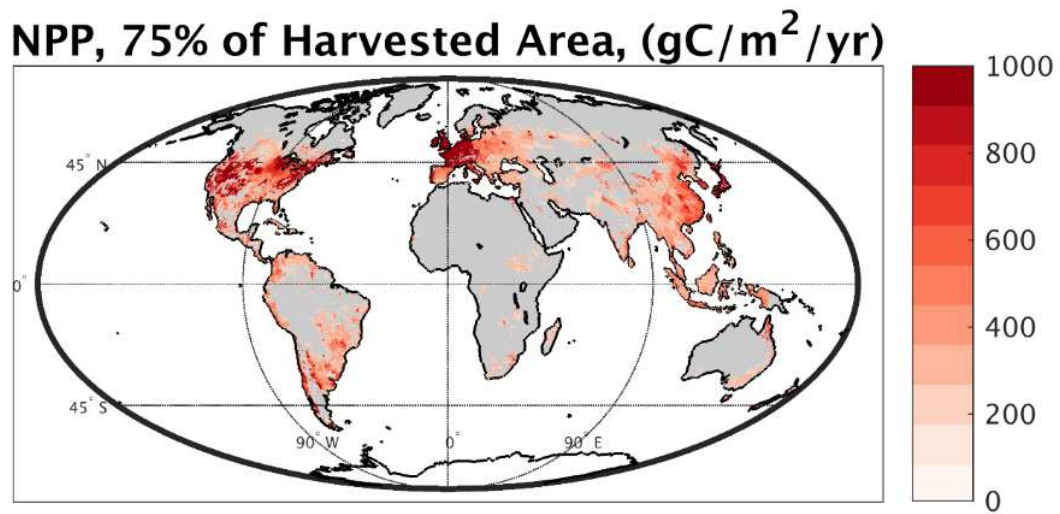


Figure 2
(Beerling et al.)

(a) Net primary production (NPP) of 75% of the harvestable area of global cropland



(b) Agricultural production by nation

(c) Cumulative CO₂ emissions by nation

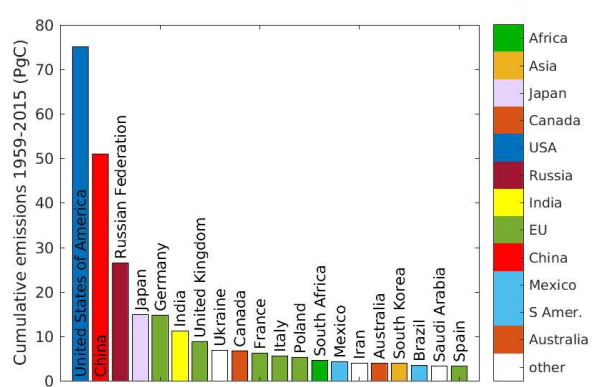
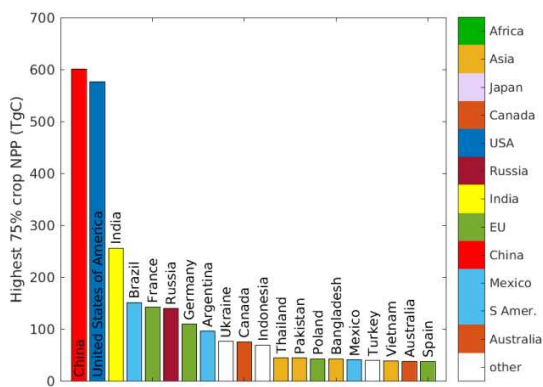


Figure 3

(Beerling et al.)

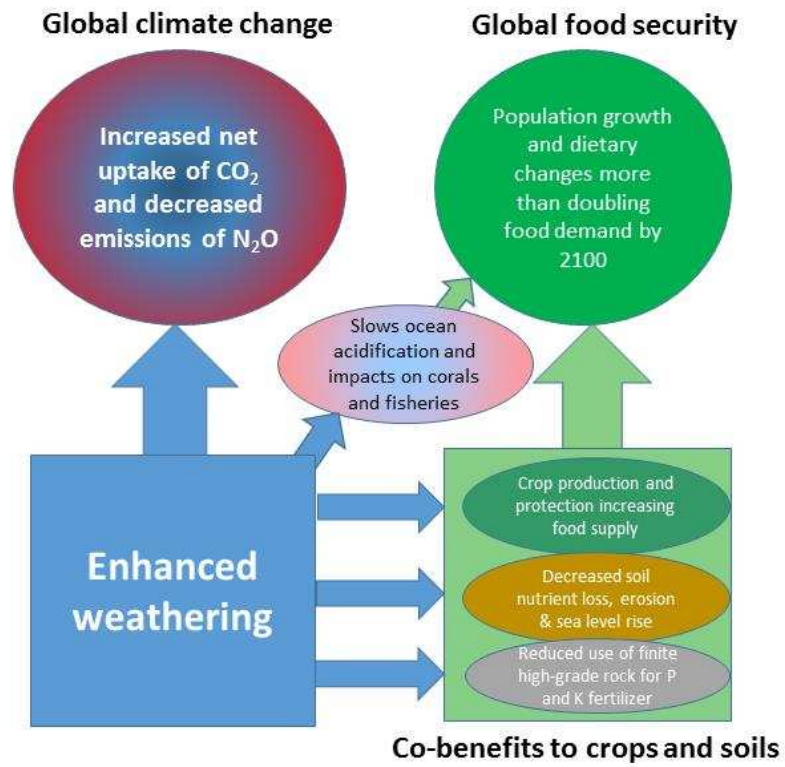


Figure 4
(Beerling et al.)

Table 1. Critical R&D needs for assessing the viability and effectiveness of enhanced weathering (EW) for CO₂ capture via silicate application to agricultural soils at scale.

	Approach	Goal
1	Sites over different crops and major soil types within major global production areas, equipped with eddy-covariance to measure year round GHG emissions and instrumented field drains to measure drainage water chemistry and flux, allowing full budgets and environmental impact assessments.	Quantify net CO ₂ capture and sequestration, silicate weathering rates and fertilization of crop performance (yield, water-use) under natural climate conditions that could reduce fertilizer application, costs and conserve finite P resources.
2	Field crop trials with different major silicate sources, ideally in conjunction with 1 above.	Assessment of relative merits of different types of silicate rocks for CO ₂ capture (e.g., basalt, dunite)
3	Controlled environment tests and replicated field trials of anticipated benefits of silicate applications on crop pest and disease resistance.	Determine translational opportunities for increasing crop protection and reducing pesticide usage and costs.
4	Genetic selection for high-weathering crops through a combination of enhanced exudation of weathering-enhancing root exudates and recruitment/associations with weathering-enhancing soil microbes.	Identification of weathering-controlling genetic traits and select for crop varieties with an enhanced capacity for weathering and releasing Si(OH) ₄ .
5	Genetic selection for crop varieties that are better capable of expressing Si-induced resistance, through a combination of Si-uptake mechanisms (i.e. Si transporters) and Si-responsive priming of JA-dependent immunity.	Characterization of the genetic basis of Si uptake, Si-induced cell wall defence and Si-induced immune priming to select for crop varieties with an increased capacity for Si-induced resistance.
6	Assessment of regional farm services capability to store, handle and spread silicates, coupled with past agronomic experience in spreading lime and silicate rich slags	Determine practicalities of deployment on croplands.
7	A full life-cycle economic/energy analysis of the cost-benefits of mining, grinding and spreading silicates, with and without carbon credits	Quantify costs and energy penalty of deployment across different scales.
8	Geographic land use assessment to determine where application of silicates would be most economically and environmentally viable	Optimize EW cost-benefits with respect to individual regions.
9	Linkage of the above into a full system model from biogeochemistry and crop yields, capable of intergration with Earth system models.	Develop realistic simulation capability for understanding Earth system response to EW
10	Investigate and reflect wider public views on EW strategies to mitigate climate change.	Understand ethical and moral concerns underlying risk perceptions of EW science