

Different photolysis kinetics at the surface of frozen freshwater vs. frozen salt solutions

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Abstract. Reactions at air-ice interfaces can proceed at very different rates than those in aqueous solution, due to the unique disordered region at the ice surface known as the quasi-liquid layer (QLL). The physical and chemical nature of the surfacial region of ice is greatly affected by solutes such as sodium halide salts. In this work, we studied the effects of sodium chloride and sodium bromide on the photolysis kinetics of harmine, an aromatic organic compound, in aqueous solution and at the surface of frozen salt solutions above the eutectic temperature. In common with other aromatic organic compounds we have studied, harmine photolysis is much faster on ice surfaces than in aqueous solution, but the presence of NaCl or NaBr – which does not affect photolysis kinetics in solution – reduces the photolysis rate on ice. The rate decreases monotonically with increasing salt concentration; at the concentrations found in seawater, harmine photolysis at the surface of frozen salt solutions proceeds at the same rate as in aqueous solution. These results suggest that the brine excluded to the surfaces of frozen salt solutions is a true aqueous solution, and so it may be possible to use aqueous-phase kinetics to predict photolysis rates at sea ice surfaces. This is in marked contrast to the result at the surface of frozen freshwater samples, where reaction kinetics are often not well-described by aqueous-phase processes.

through field, laboratory, and theoretical studies (ibid). However, there is not yet a fundamental understanding of the role of snow and ice as reaction media; thus, accurately predicting the chemical fate of species in snow and ice remains an elusive goal.

Ice is a heterogeneous and complex reaction medium. Veins, pockets, and grain boundaries within the crystal lattice contain liquid water. Experiments designed to measure reaction kinetics associated with ice typically determine reagent and product concentrations in the melt, following some reaction time in ice (we will signify these as freeze-react-melt, or FRM, experiments). The results of such measurements (which likely primarily reflect reactivity in the liquid inclusions, due to the larger volume fraction that these represent) have been reported to be similar to those in aqueous solution (Chu and Anastasio 2003, 2005, 2007; Grannas et al., 2007b; Jacobi et al., 2006; Kahan et al., 2010a, b; Ram and Anastasio 2009). In addition to liquid inclusions within the ice matrix, near the air-ice interface, a disordered region known as the quasi-liquid layer (QLL) exists whose physical and chemical properties are distinct from both bulk ice and liquid water (Kahan et al., 2007; Wei et al., 2001). Some recent laboratory studies (Dubowski and Hoffmann 2000; Grannas et al., 2007a; Kahan and Donaldson 2007, 2008, 2010; Kahan et al., 2010a, b; Matykiewiczova et al., 2007) suggest that reaction kinetics in the QLL can differ dramatically from those in aqueous solution. We have recently (Kahan et al., 2010a, b) compared FRM reaction kinetics with those obtained at air-ice interfaces, and observed very different reaction kinetics in the two cases. Direct photolysis of the polycyclic aromatic hydrocarbon (PAH) anthracene measured in an FRM experiment yields the same rate as in aqueous solution, but measurements made at ice surfaces show much faster kinetics (Kahan et al., 2010b). In other work we have determined that hydroxyl radicals react quickly with aromatic compounds both in aqueous solution and in ice samples measured using the FRM technique, but at ice surfaces

1 Introduction

Chemical reactions in snow and ice can affect pollutant fate and atmospheric composition (Domine and Shepson 2002; Grannas et al., 2007b) in urban and remote centres. Various aspects of snowpack chemistry have been investigated



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no reaction is observed (Kahan et al., 2010a). These results suggest that liquid inclusions in ice (such as those at grain boundaries) present similar reaction environments to liquid water, but that the QLL presents a unique reaction environment.

It is likely that many atmospheric reactions occur in the QLL, since reagents can both be excluded to the air-ice interface during freezing, and be deposited there from the gas phase. As discussed above, reactivity in the QLL can be quite different from that either in aqueous solution or liquid inclusions in bulk ice; therefore, measurements must selectively probe the air-ice interface in order to describe physical properties of the QLL, and to properly report reaction kinetics that occur there.

One would certainly expect that the physical properties of ice surfaces could affect chemical reactions which occur there. However, few studies have investigated these properties under environmentally relevant conditions; most investigations of QLL properties have used single-crystalline ice in pristine conditions in the absence of solutes. In the environment, ice is generally polycrystalline, and it is rarely pristine. Common solutes such as halide salts can significantly alter the molecular structure at air-ice interfaces, and thus possibly affect reaction kinetics there.

During freezing of a solution, ionic compounds such as halide salts are excluded from the ice matrix. Below the eutectic temperature of a salt-water mixture, a solid solution of ice and salt is formed. Above the eutectic temperature, halide exclusion results in the formation of concentrated brines in veins and pockets within bulk ice, as well as at the air-ice interface (Carignano et al., 2007; Cho et al., 2002; Eicken 1992). The exclusion of salts to the air-ice interface may also be inferred from their influence on the Raman spectrum of water molecules present there. Figure 1 shows glancing-angle Raman spectra (Kahan et al., 2007) of various aqueous interfaces acquired in this laboratory. The spectrum observed from the surface of a frozen freshwater sample (solid blue trace) is asymmetric, with a broad shoulder around 3100 cm^{-1} . This feature, which is not observed at the liquid water surface (shown as the dotted green trace) arises from a more extensive hydrogen bonded network present at the ice surface, compared to liquid water (Kahan et al., 2007). The shoulder is not observed at the surface of frozen NaCl solutions, as illustrated by the red dashed trace in Fig. 1. This observation implies that the surface of frozen salt solutions represent a different hydrogen bonding environment than that at the freshwater ice-air interface.

The decreased hydrogen bonding at the surface of frozen saltwater solutions has two possible explanations. First, the halide ions could disrupt the QLL by competing for hydrogen bonds with the water molecules present there. An alternate explanation is that the halide ions cause the formation of a true liquid brine at the ice surface. This is an important distinction because, as discussed previously, reaction kinetics in the QLL can be very different from kinetics in aqueous so-

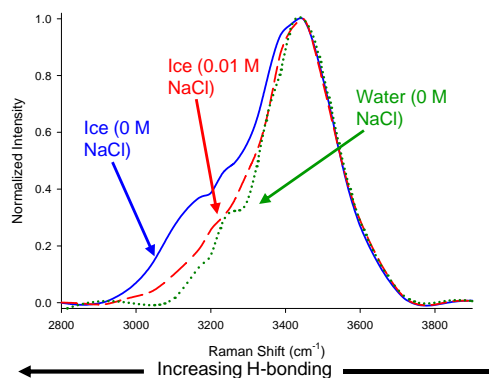


Fig. 1. Intensity-normalized glancing-angle Raman spectra of the surface of freshwater ice, of a frozen 0.01 mol L^{-1} NaCl solution, and of the surface of liquid freshwater. The lower relative intensity at lower wavenumbers for the frozen salt solution and the liquid freshwater is indicative of reduced hydrogen bonding there.

lution. Therefore, we might expect quite different kinetics at the surfaces of frozen salt solutions depending on which explanation is correct. If halide ions simply disrupt the QLL structure, reactions on sea ice surfaces could occur at similar rates as on frozen freshwater surfaces. However, if a concentrated liquid brine exists at the *surface* of frozen salt solutions, as well as within the liquid inclusions present in the iced matrix, one might expect the reaction kinetics to be similar to those in aqueous solution. In fact, recent measurements of the heterogeneous ozonation kinetics at the surface of frozen NaBr and NaI solutions (Wren et al., 2010) show kinetics which are consistent with the reaction occurring in concentrated aqueous solutions. In the present study, we measure photolysis kinetics of an organic probe molecule in aqueous solution and at the surfaces of frozen freshwater and salt solutions to determine whether the reaction environment at the surface of frozen salt solutions is more similar to a QLL or to aqueous solution.

2 Experimental

The aromatic compound harmine (7-methoxy-1-methyl-9H-pyrido[3,4-b]indole) was chosen as a molecular probe because its fluorescence is not strongly quenched in high ionic strength environments (Clifford et al., 2007). Solutions containing $1 \times 10^{-7}\text{ mol L}^{-1}$ harmine (Aldrich, 98%) and a known concentration of either NaCl or NaBr (both ACP Chemicals, 99%) were prepared daily in $18\text{ M}\Omega\text{cm}$ deionized water. Samples were irradiated by the output of a 75 W xenon arc lamp which passed through a 295 nm longpass cutoff filter. A first-surface mirror with maximum reflectivity at 355 nm directed the lamp's output downward onto the sample at an angle of approximately 45° .

Liquid-phase samples were photolyzed in quartz cuvettes at room temperature. Ice samples were contained in a

reaction chamber which has been described previously (Kahan and Donaldson, 2007). Briefly, the chamber consists of a Teflon box with cooling tubes running beneath the copper floor of the chamber; the floor was maintained at ~ 253 K for experiments on ice. Approximately 1 mL of solution was pipetted onto a small copper plate on the chamber floor and allowed to freeze from bottom to top. The ice samples, which froze over the course of ~ 10 min, resembled small pancakes with diameters of approximately 1.5 cm and heights of 1–2 mm. The ice that was formed was transparent, with bubbles visible throughout the sample.

Kinetics were determined by monitoring harmine fluorescence intensity during irradiation. Fluorescence was induced by the unfocussed 337 nm output of a nitrogen laser, and collected by a liquid light guide (LLG). For aqueous samples in cuvettes, the LLG was positioned to the side of the cuvette, perpendicular to the laser beam. For experiments on ice, the laser beam entered the reaction chamber through a quartz window at a shallow angle ($>85^\circ$ from the surface normal), such that it impinged upon the ice surface. The LLG was positioned approximately 7 mm above the sample. This geometry affords us surface sensitivity (Kahan et al., 2007), and allows us to measure kinetics at air-ice interfaces (Kahan and Donaldson 2007, 2008, 2010; Kahan et al., 2010a, b; Wren et al., 2010).

The emission collected by the LLG passed through a monochromator set to 430 nm, where harmine emits strongly, then was detected by a photomultiplier tube whose output was collected on a digital oscilloscope. The fluorescence decay was averaged over 64 laser shots; a 500 ns slice of the decay was captured and stored for analysis. Fluorescence intensity was measured in the dark for several minutes at the beginning of each experiment to ensure that no dark loss was occurring. The sample was then irradiated for up to 4 hours, with fluorescence intensity acquired at known time intervals ranging from 1–10 min. The xenon arc lamp was blocked during data acquisition.

3 Results and discussion

Figure 2a shows representative semi-logarithmic results for the loss of harmine, measured via its fluorescence intensity, observed during irradiation in aqueous solution and at an air-ice interface. In aqueous solution, the fluorescence decay is well-described by a single exponential fit, indicating that the reaction is first-order in harmine. At the air-ice interface, some experiments showed an initial very fast loss of harmine followed by a slower loss, which was several times more rapid than seen in the liquid. For these results, both single and double exponentials provide good fits to the data, with the double exponential generally providing a slightly better fit. In this work we present the first-order rate constants based on single-exponential fits to the data. To check that the photolytic loss was not due a self-reaction on the ice

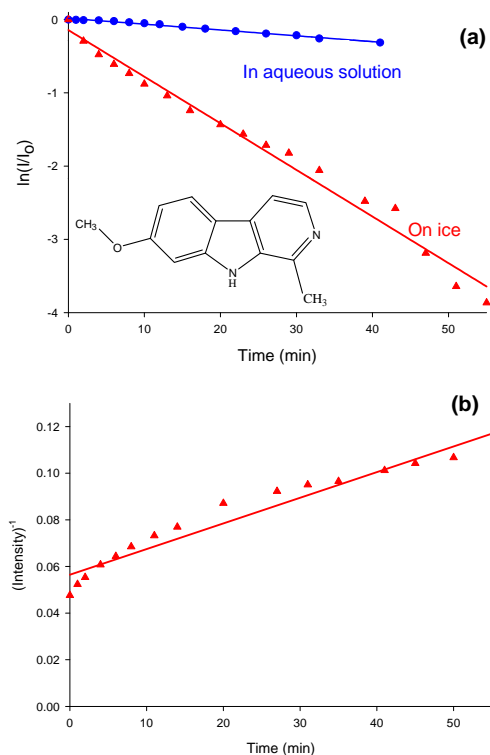


Fig. 2. (a) Loss of harmine emission intensity as a function of irradiation time in aqueous solution and at an air-ice interface shown as semi-log plots (first-order kinetics). The solid traces are linear fits to the data. The inset displays the chemical structure of harmine; (b) the change in $(\text{harmine emission intensity})^{-1}$ as a function of irradiation time at an air-ice interface (second-order kinetics). The solid trace is a linear fit to the data.

surface, a second-order kinetic analysis was carried out, illustrated in Fig. 2b. Second-order kinetic analysis provides a poor fit to the data, implying that the reaction is not bimolecular.

The measured photolysis rate is approximately 4 times greater on ice than in aqueous solution. Many bimolecular reactions show enhanced rates in frozen solution, due to solute exclusion yielding enhanced concentrations in liquid regions within the ice matrix (Pincock, 1969). This freeze-concentration effect has recently been demonstrated for a bimolecular reaction in ice under atmospheric conditions (Grannas et al., 2007a). However, solute exclusion (harmine in the present case) can not explain the enhanced reaction rate we observe at the air-ice interface: The kinetics we observe are first-order with respect to harmine, which means that the reaction is unimolecular, and thus the rate constant is independent of concentration.

We have previously reported higher photolysis rates for other aromatic species such as anthracene, naphthalene, and benzene on ice surfaces than in aqueous solution, or at the air-water interface (Kahan and Donaldson 2007, 2010; Kahan et al., 2010b). In the case of anthracene (Kahan and

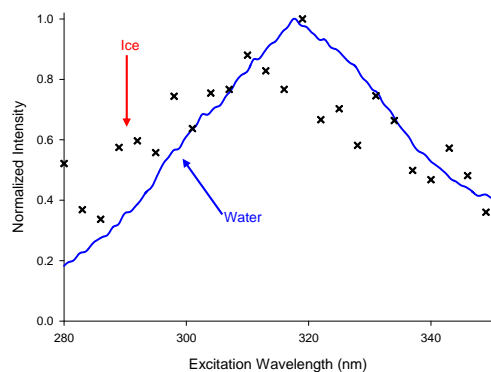


Fig. 3. The solid trace shows an absorption spectrum of harmine in room temperature aqueous solution; the crosses indicate an excitation spectrum of harmine at an air-ice interface at 253 K.

Donaldson, 2007), we measured identical first order rate constants at the air-ice interface whether anthracene was frozen from solution (in which case freeze-concentration effects could influence the kinetics) or whether it was deposited from the gas phase. The kinetics were also insensitive to the amount of anthracene deposited on the surface, further arguing against enhanced concentrations on ice being responsible for the faster photolysis kinetics there.

There are several possibilities other than reagent concentration which could explain the different reaction rates measured in aqueous solution and at air-ice interfaces. We have discussed these at length previously (Kahan and Donaldson 2007, 2010; Kahan et al., 2010b). Aromatic compounds tend to self-associate at air-ice interfaces, even at low surface coverages (Ardura et al., 2009; Kahan and Donaldson 2007, 2008, 2010). Even in the absence of self-reaction, such association can lead to red-shifts in the absorption spectra. For species which absorb primarily at wavelengths below 295 nm in aqueous solution, even a slight red-shift can significantly enhance the photolysis kinetics, as we reported for the case of benzene on ice (Kahan and Donaldson 2010). In the present case, there appears to be no such red shift, however. Figure 3 compares the absorption spectrum of harmine in aqueous solution with the excitation spectrum measured at the air-ice interface. There is significant spectral overlap, with the majority of light absorption occurring at wavelengths above 295 nm in both media. Therefore, a red-shift in harmine absorbance on ice does not explain the enhanced reaction rates we observe there.

Another factor that can affect photochemical reaction rates is photon flux. It is quite likely that scattering by the ice gives rise to a different light intensity at the ice surface than that occurring in aqueous solution. We have recently investigated the effect of varying light intensity on the photolysis kinetics of strongly absorbing aromatic compounds in aqueous solution and at ice surfaces (Kahan and Donaldson 2010; Kahan et al., 2010b). Surprisingly, both in solution and on ice surfaces we found a saturation in the first order

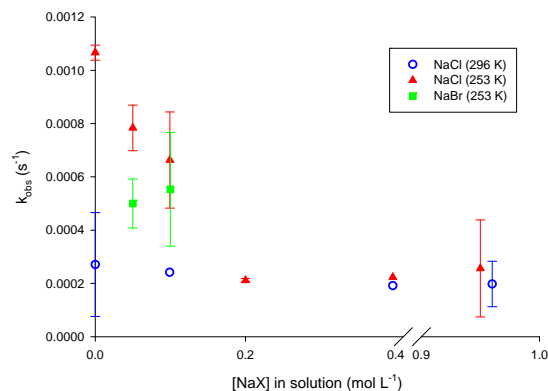


Fig. 4. Observed rate constants for harmine photolysis as a function of salt concentration: In aqueous solution at room temperature (hollow blue circles); at the surface of frozen aqueous NaCl solutions at 253 K (solid red triangles) and at the surface of frozen aqueous NaBr solutions at 253 K (solid green squares). Error bars represent the uncertainty about the mean for at least three trials.

photolysis rates of anthracene above an irradiance threshold of about 1×10^{13} photon $\text{cm}^{-2} \text{s}^{-1}$. The lamp power used in the present experiments corresponds to photon fluxes of $\sim 3 \times 10^{13}$ photon $\text{cm}^{-2} \text{s}^{-1}$. Given the similar absorption strengths for harmine and anthracene in the actinic spectral region, it seems unlikely that the faster harmine photolysis kinetics on ice are due to differences in light intensity there.

The discussion above outlines why we are confident that the enhanced photolysis rates on ice are due to differences in the physical nature of the QLL at the air-ice interface compared to that of liquid water. The situation becomes different at the surface of frozen salt solutions. Figure 4 shows the rate constants obtained from the slopes of plots like those shown in Fig. 2 for harmine photolysis in aqueous solution and at air-ice interfaces under a range of salt concentrations. As displayed by the hollow symbols, the presence of salt does not affect harmine photolysis kinetics in aqueous solution up to 1 mol L^{-1} NaCl, which roughly corresponds to seawater concentrations. At air-ice interfaces at $T = 253 \text{ K}$, however, increasing the salt concentration of the solution prior to freezing decreases the photolysis rate, indicated by the red triangles, until NaCl concentrations exceed $\sim 0.2 \text{ mol L}^{-1}$, at which point the observed rate at the ice surface becomes identical to that in aqueous solution. Experiments were also performed on frozen solutions of NaBr at 253 K having initial $[\text{NaBr}]$ of 0.05 and 0.1 mol L^{-1} , as indicated by the green squares in Fig. 4. The photolysis rate is more suppressed in the presence of NaBr than NaCl for a similar solution salt concentration.

The equilibrium phase diagrams for the NaCl-water (Akiniev et al., 2001), and NaBr-water (Tang et al., 2003) systems are useful tools for interpreting these results. The temperature of these experiments, 253 K, lies just above the eutectic temperature of NaCl, meaning that the overall

Table 1. Predicted photochemical behaviour of aromatic compounds in and at the surface of liquid and frozen aqueous media, relative to those in bulk liquid water

Medium	Bulk	Surface
Liquid water	n/a	Same as bulk liquid water (Kahan and Donaldson 2007)
Liquid brine (up to 1 mol L ⁻¹ NaCl)	Same as liquid water (this work)	Not studied
Frozen freshwater (ice)	Same as liquid water (Kahan et al., 2010a; Kahan et al., 2010b; Ram and Anastasio 2009) 3× faster than liquid water (Dubowski and Hoffmann 2000)	Up to 10× faster than liquid water (Kahan and Donaldson 2007, 2010; Kahan et al., 2010b), this work)*
Frozen brine (≤0.2 mol L ⁻¹ NaCl in solution)	Not studied	4× faster than liquid water in the absence of salt, but decreases with increasing [NaCl] (this work)
Frozen brine (>0.2 mol L ⁻¹ NaCl in solution)	Not studied	Same as liquid water up to 1 mol L ⁻¹ (this work)

* The rate enhancements for benzene and naphthalene are due at least in part to red-shifts in the absorption spectra. It is possible that other aromatic species display greater rate enhancements on ice than the factor of 10 reported here.

sample at equilibrium should consist of a mixture of solid ice and NaCl solution, having a concentration of ~4.5 mol L⁻¹. The overall amount of this brine is dependent on the overall amount of NaCl present; however, how this is distributed among the air-ice interface, liquid inclusions and grain boundaries is not predicted. The results of Cho et al. (Cho et al., 2002) indicate that brine exists in all of these locations. Nevertheless, comparison of the NaCl and NaBr phase diagrams suggests that under our experimental conditions, the NaBr system should contain a greater volume of brine at 253 K than does frozen NaCl having the same overall salt concentration.

With this in mind, we interpret the dependence of the ice surface photolysis rate on initial salt concentration to arise from increasing amounts of liquid brine being excluded to the air-ice interface as the salt concentration increases. Thus the observed kinetics at low initial [NaX] represent a combination of “true” QLL kinetics (fast), with those appropriate for the aqueous surface (slow); as the overall [NaX] increases the fraction of the air-ice interface which is well described as a true brine increases, giving rise to a decrease in the observed rate. Eventually (at [NaCl] >0.2 mol L⁻¹) the surface is completely coated in brine, so that the kinetics are well described by those seen on the liquid surface.

This interpretation, based on equilibrium phase behaviour, should be used with caution, however, and is unlikely to give quantitative predictions. One study of NMR spectra of frozen brine solutions (Cho et al., 2002) suggested that salt concentrations in frozen brines above the eutectic temperature are well-described by thermodynamic phase diagrams even with more rapid cooling than was used in our experiments. However, Thibert and Domine show that environmental ice and snow is not equilibrated with potential solutes (in their

case gas phase HCl) (Thibert and Domine 1997). Given the further uncertainties in whether the same brine concentration exists at the interface as in liquid inclusions within the ice matrix, over reliance on thermodynamic phase diagrams seems unwarranted.

The results of this study expand our understanding of environmental ice surfaces as reaction substrates. The observed harmonic photolysis kinetics suggest that the liquid brine formed at the surface of frozen seawater has physical properties more similar to liquid water than to the QLL at air-ice interfaces. Halide concentrations in remote inland polar regions are orders of magnitude lower than those in seawater (Grannas et al., 2007b), and should not form liquid brines. In urban snows, however, chloride concentrations of ~0.02 mol L⁻¹ have been reported (Tatarniuk et al., 2009), due primarily to road salt. At this concentration, reaction kinetics may differ from those at the surface of the relatively pristine ice found for example in polar regions.

To conclude, we present a summary (shown in Table 1) of the photochemical transformation behaviour observed for aromatic compounds in and on ice in this study and others (Dubowski and Hoffmann 2000; Kahan and Donaldson 2007, 2010; Kahan et al., 2010b; Ram and Anastasio 2009). It is clear that in order to accurately predict the fate of species in snowpacks, we must understand both the composition of the ice and the partitioning of the reagents (i.e. sequestered within the bulk versus present at the air-ice interface).

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