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4 **ON THE ORIGIN OF AMS “COOKING ORGANIC**
5 **AEROSOL” AT A RURAL SITE**

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Abstract

A number of field observations employing aerosol mass spectrometers (AMS) have demonstrated that organic matter rich in monocarboxylic acids and aliphatic carbonyls originating from cooking activities (the COA factor) contributes significantly to ambient Organic Matter (OM) in urban environments. Little is known about the contribution and nature of COA in rural localities. We studied the correlation of COA with chemical tracers at a rural site in the Po Valley, Italy. Our statistical approach, based on Positive Matrix Factorization (PMF) shows that the COA factor was clearly linked to local emissions of chloride and methane sulphonic acid (MSA), chemical tracers not associated with cooking emissions, or with combustion sources. While the association with Cl is not understood at this stage, the emission of reduced sulphur compounds, aliphatic carbonyls and monocarboxylic acids is consistent with several agricultural practices (e.g., manure storage) and waste disposal systems (e.g., landfills) which characterize the sub-urban and rural areas of the Po Valley and of other many populated environments. It is concluded that the nature and origins of the AMS COA factor measured at a rural site are complex and include far more than the emissions from food cooking.

KEY WORDS: Aerosol mass spectrometer; cooking organic aerosol; COA; ATOFMS

69 **1. INTRODUCTION**

70
71 Urban air quality provides one of the main drivers for the study of atmospheric science,
72 especially due to growing urbanization. Although considerable progress in improving air
73 quality has been made, it is imperative to better understand the source attribution of
74 particles and health effects in the urban atmosphere¹. Among the constituents, organic
75 aerosol (OA) accounts for a large fraction of urban particulate matter². Primary OA (POA)
76 is directly emitted from fossil fuel combustion, biomass burning, and other sources, but the
77 atmospheric evolution of POA after emission remains poorly characterized³⁻⁵. Whilst road
78 traffic is often a major contributor to aerosol in urban areas next to major roads, OA in
79 urban background areas is not easily associated with a specific source. Few studies have
80 sought to characterise the sources of OA in rural areas, remote from major sources.

81
82 In recent years Mass Spectrometry of Atmospheric Aerosol (MSAA) has become one of
83 the fastest growing area of aerosol research⁶. Such techniques have greatly enhanced our
84 capacity of observing the atmospheric processes responsible for the formation and
85 evolution of airborne particles. In this regard, the Aerodyne Aerosol Mass Spectrometer
86 (AMS) is able not only to measure the OA concentration and size distribution but also to
87 provide its mass spectrum⁷. Analysis of AMS data collected at a large number of sites has
88 revealed that the oxidised OA component (OOA) tends to dominate everywhere, including
89 some heavily urbanized regions^{2,8}.

90
91 Many AMS studies have focused on the application of factor analysis to the organic mass
92 fraction in an attempt to deconvolve it into descriptive sub-components, namely a
93 hydrocarbon-like organic aerosol (HOA) factor, an oxygenated organic aerosol (OOA)
94 factor, and a semi-volatile factor (SV-OOA)^{4,9}. Such assignments are based primarily
95 upon the main mass spectral components and the diurnal profile. In some cases,

96 supporting evidence from chamber experiments is available. Recently, a previously
97 reported AMS spectrum^{10,11} has been firmly associated with primary emissions from
98 cooking activities. Such emissions have been claimed to account for up to half of the total
99 primary urban OA at sites in Europe¹²⁻¹⁴, Asia¹⁵ and America¹⁶. However, there is a
100 possibility for a factor to include a combination of factors associated with different
101 sources^{11,17}, and some recent studies report a COA factor lacking the diurnal profile
102 expected from cooking activities, although possible explanations have been given^{13,18}.
103 Other studies have also expressed the view that it is most reasonable to characterize the
104 COA component as "cooking influenced" but not purely from cooking sources¹⁹. In a
105 comparison of source apportionment by AMS-PMF with a Chemical Mass Balance model,
106 the AMS estimate of COA for a site in London exceeded the CMB-derived concentration of
107 cooking aerosol by a factor of ca 1.6²⁰.

108

109 Taken together, evidence for source attribution of the AMS COA factor is currently
110 incomplete and most importantly not well supported by other measurement techniques.
111 This lack of knowledge needs to be filled, and the goal of this paper is to use results of a
112 source apportionment study using six state-of-the-art spectrometric techniques deployed
113 at a rural site in the Po Valley (Italy) in 2009 to investigate the nature of the AMS COA
114 factor. Previous work²¹ provides a comprehensive picture of the nature of organic and
115 inorganic aerosols and aerosol precursors at a European rural site with an unprecedented
116 level of detail. Figure S1 shows an example of the high time resolution particle mass
117 spectrometry instruments deployed. In this study Positive Matrix Factorization (PMF) has
118 been applied to all the high time resolution data in order to better elucidate aerosol
119 sources not clearly identified when analyzing results from individual aerosol techniques on
120 their own. Particular attention is given to the AMS COA factor, but a complete overview of
121 the aerosol sources apportioned is also presented and discussed.

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2. EXPERIMENTAL

2.1 Study Site

The Po Valley is located in Northern Italy between two mountain ranges, the Alps in the north and west and the Apennines in the south. The Po Valley has 20 million inhabitants spread over an area of 48,000 km² (Fig S2). In the present study we used the EMEP Station of San Pietro Capofiume (SPC, 44°23'N 11°22'E, 11 m a.s.l.), a rural background site (distance from major pollution sources: 10 – 50 km²²). The measurements were conducted from 26 June to 15 July 2009.

2.2 Aerosol Mass Spectrometry Techniques

Four on-line aerosol spectrometers were used: the TSI Aerosol Time-Of-Flight Mass Spectrometer (ATOFMS), the Aerodyne High-Resolution Time-of-Flight Aerosol Mass Spectrometer (HR-TOF-AMS) and Soot Particle Aerosol Mass Spectrometer (SP-AMS), and finally a Thermal Desorption Aerosol Gas Chromatograph AMS (TAG-AMS). Twelve-hour resolution proton nuclear magnetic resonance (H-NMR) spectra were also obtained by off-line analysis. Finally, information on gas-phase precursors of secondary aerosols was acquired using Chemical Ionization Mass Spectrometry (CIMS). A set of other measurements are described in detail elsewhere²¹. Overall, the state-of-the-art instruments used for this analysis measure different species with different accuracy and precision, as discussed in great detail in other publications²³⁻²⁵.

2.3 Factor Analysis

Factor analysis involves a wide set of multivariate statistical techniques that have been extensively used in atmospheric science. Its aim in this field is to apportion aerosol sources on the basis of the internal correlations of observational data collected at a

155 measurement point, called the “receptor site”. Receptor modeling by factor analysis does
156 not need any detailed a prior knowledge of source profiles and it is therefore very useful
157 for determination of aerosol fractions of secondary origin. “Positive Matrix Factorization”
158 (PMF)^{26,27} has been used extensively for source apportionment of ambient particulate
159 matter (PM), and in particular is so far the most widespread tool for AMS spectral data
160 analysis²⁸. In this study the EPA open-source software EPA-PMF v3.0 was used for PMF
161 analysis. PMF was applied on the hourly data obtained by the analytical techniques
162 described above and also by the previous PMF analysis on spectroscopic data reported
163 elsewhere²¹.

164
165 The criteria adopted in this study for input data treatment, including the uncertainty matrix,
166 and for the evaluation of the PMF solutions are discussed extensively in the
167 supplementary material.

168

169 **3. RESULTS**

170 PMF analysis was performed for one to ten factors, and it was found that the solutions with
171 factor numbers greater than five provided no new meaningful information for both
172 datasets. Rotational constraint was explored with Fpeak. The solutions ultimately selected
173 was based on an Fpeak value of zero. Additional factors resulted in a splitting of the
174 existing factors (mainly splitting of secondary organic and inorganic components (Figures
175 S5 and S6; Table S2 and S3). Table S4 shows there is no clear improvement with the six
176 factor solutions from either of the two datasets. A full list of variables is found in Table S1.
177 The figures in brackets after the PMF input variable represent the percent contribution of
178 that atmospheric variable within the whole five factor solution (% of species). A description
179 of the strong agreement between the solutions from the two datasets is presented at the

180 end of this section. More details on the PMF diagnostics are presented in supplementary
181 material.

182

183

184 **3.1 Dataset 1 : Results from the Full Campaign (“ALL”)**

185

186 Five factors were found very well describing the long time series dataset (16 days). This
187 dataset included the first week of the study when the anticyclonic conditions favoured
188 transport of aerosol at the regional scale and the recirculation of pollutants in the Po
189 basin²¹. Figure 1 and table 1 show all the PMF factor profiles of this solution, whereas the
190 temporal trends and the diurnal profiles can be seen in Figures 2 and S9a, respectively.

191 The individual factors are as follows:

- 192 • Oxidised Organic Aerosol from photooxidation processes "Photox OOA" (30% of the
193 total, Figure 1a). It represents oxidised organic aerosols (accounting for a substantial
194 percentage of AMS_LV-OOA MO 48% and AMS_LV-OOA LO 49%), AMS_sulphate
195 (48%), Ozone (83%) and CIMS_MA (84%). This is associated with photochemically
196 aged secondary aerosols, and the temporal trends (Figure 2, S9a) show this aerosol
197 component occurring mainly during daytime and during the first week of the study.
- 198 • Nitrate "NIT-Regional" (18%, Figure 1b). This factor is strongly associated with
199 AMS_Nitrate (79%) and ATOFMS_NIT-Regional (32%). There is an important
200 component of AMS_Cl (38%) and AMS_SV-OOA (27%) associated with it. This factor
201 is associated with nitrate containing aerosol of regional origin²⁹, and related to high
202 RH (Figure 2), with time trends modulated by the diurnal variations of temperature and
203 relative humidity which regulate the concentration of semivolatile compounds in the
204 aerosol (ammonium nitrate and chloride, SV-OOA).
- 205 • Sulphate "SUL-Regional" (23%, Figure 1c). This aerosol source is mainly composed of
206 sulphate (ATOFMS_SUL-Reg 90%), ATOFMS_EC-Reg (72%), and AMS_OOA-MO

207 (32%). This aerosol source accounted for 35% of explained variation during the first
208 week of the study (stagnant air), but only 3% during the second week (Figure S9a).
209 Part of this component is thought to be the aerosol core from the evaporating regional
210 nitrate during daytime^{21,29}. This factor seems to be linked to the transport of sulphate
211 and BC during the days of anticyclonic conditions.

212 • "NO" (12%, Figure 1d). This represents the smallest aerosol source contribution as
213 expected given the rural location of San Pietro Capofiume. The factor is well described
214 primarily by NO (90%), followed by minor contributions from AMS_HOA (23%) and
215 NO₂ (25%). BC is only partially described by this factor (6%). Although the diurnal
216 profile (Figure 2) shows a main morning spike during the urban rush hour (7am),
217 processes other than traffic and certainly including photolysis of NO₂ and of HONO
218 most probably contributed to NO at the site. Other studies^{21,30,31} have already
219 demonstrated the presence of a traffic-influenced component at SPC station with
220 contributions ranging between 14-24% of total OA. This non-negligible contribution
221 and its diurnal trend could be explained both by the proximity of main transportation
222 routes and by the accumulation of the primary emissions overnight because of the
223 reduced atmospheric mixing and dispersion. However the NO apportioned in this study
224 could be also formed from photochemical activity (photolysis of NO₂ in the first hours
225 after the sunrise) and so we prefer to label this factor simply as "NO", the most
226 represented species, without assigning a specific source.

227 • "Cooking" Organic Aerosol "COA" (17%, Figure 1e). This factor accounts for a high
228 proportion of AMS COA (77%) and ATOFMS OC-SUL-NIT (77%). Other interesting
229 variables describing this factor includes CIMS MSA (41% of the total), AMS Cl (35%)
230 and NO₂ (31%). The temporal trends show a major occurrence during the second
231 week (24%) rather than the first regional pollution week (14%). The diurnal profile
232 shows a minor spike during evening time (7-9pm), followed by sustained

233 concentrations overnight. There is a lack of the midday peak seen associated with
234 COA factors in many studies^{13,18}.

235

236 The analysis of the correlation of the five factor concentrations with local wind direction
237 (Figure S10a) shows that the first three factors are associated with sustained winds from
238 either east (OOA), west (NIT-regional) or south-west (SUL-regional) directions, while the
239 COA factor is most concentrated in calm conditions. Therefore, the COA is more
240 influenced by local sources compared to the first three factors which are mainly
241 transported to the site. The NO factor shows an intermediate behaviour, with a component
242 from the west (possibly NO_x transport) and another local (possibly soil emissions³²).

243

244 **3.2 Dataset 2: Three Measurement Days Including SP-AMS Data (“HIGH”**

245 **Resolution)**

246

247 The factors of the dataset_2 (HIGH) PMF solution can be seen in Table 1 and Figure 3. In
248 the last days of the campaign, which were covered by the measurements described in this
249 section, the weather conditions were unfavourable for pollutant recirculation and the
250 aerosol concentrations were fairly low in day-time when the planetary boundary layer was
251 fully developed. Conversely, the concentrations at night-time and in the early morning
252 continued to be sustained by local sources in the Po Valley²¹. The analysis of this
253 campaign period is therefore particularly useful for the source apportionment of the aerosol
254 components originating from emissions in the Po Valley, such as cooking aerosols,
255 because the interference from the background particles is smaller with respect to the first
256 part of the campaign²¹. The diurnal profiles can be seen in Figure 2 and the temporal
257 trends in Figure S9b. Three of the five factors presented previously (Section 3.1, ALL)
258 were also found in this second solution, including:

- 259 • Oxygenated Organic Aerosol "Photox OOA" (38%, Figure 3a). This described
260 Secondary Organic aged aerosols accounting for a high percentage of AMS LV-OOA
261 MO (81%), SP-AMS "SV-OOA Day" (89%), ozone (83%) and sulphate from both HR-
262 AMS (68%) and SP-AMS (67%) instruments.
- 263 • Nitrate regional "NIT-Regional" (17%). A strong nitrate signal from all the on-line
264 spectrometers including HR-AMS (72%), SP-AMS (45%) and ATOFMS (NIT-Reg
265 71%) can be seen (Figures S9b and 2). Additionally, an hydrocarbon like contribution
266 of HOA (21%) and BC (20%) is notable.
- 267 • "NO" (17%). This shows a strong NO signature (93%) followed by HOA (22%) and BC
268 (11%) (Figure 3c). The diurnal profile seen in Figures 2 and S9b shows a morning
269 peak likely to be associated with traffic and/or photolysis.

270

271 Unlike the "ALL" solution, in this 3-day dataset_2 (AMS, ATOFMS, SP, CIMS, gas, 33
272 variables, last 3 days of the field study, dataset_2, HIGH) a factor related to sulphate of
273 regional origin was not identified. This is very likely due to the fact that during the second
274 week of the field study there was little contribution of regional pollution, so the solution was
275 not able to extract the SUL-Reg factor, mainly seen during the first week (Figure S9a). By
276 contrast, during nighttime the relatively higher concentration of nitrate allowed identification
277 of the NIT-Reg. factor. During daytime the SUL-Reg., if present, is likely to be incorporated
278 in the regional OOA factor.

279

280 However, the second solution with more variables (33 in total) was able to provide a better
281 description of the COA factor, because the statistical correlations of AMS COA
282 concentrations with the other variables are much more clear in the HIGH dataset than in
283 the ALL one (Table S6). The factor analysis of the HIGH set of variables actually provided
284 two different aerosol sources associated with the AMS "cooking aerosols" and specifically:

- 285 • "Cooking" organic aerosol "COA-MSA-Cl" (seen in Figure 3d). This factor presents a
286 strong AMS COA signal (54%), with CIMS_MSA (36%) and a strong inorganic AMS Cl
287 signature (70%). A strong semi-volatile component is also present (AMS SV-OOA and
288 SP-AMS "SV-OOA Night", both 58%). Additionally, the factor ATOFMS OC-SUL-NIT is
289 also associated with this factor (32%). In summary, this COA factor is associated with
290 MSA, chloride and semi-volatile OA components.
- 291 • "Cooking" organic aerosol "COA-MSA-HOA" (seen in Figure 3e). This factor shares
292 the part of the AMS_COA signal (46%) not identified by the previous factor (54%) and
293 a strong association with CIMS MSA (37%). However, this factor is more associated
294 with AMS HOA (38%) SP-AMS HOA (42%) and contains almost no semi-volatile
295 component. Interestingly, this factor is associated with nitrate of local origin (ATOFMS
296 NIT-Local, 42%). In support of this, it accounts for the highest percentage of NO₂
297 (33%) of all factors. In summary, this second "COA" factor is associated with
298 anthropogenic HOA and somehow related to nitrate locally formed, organic nitrate and
299 nitrogenous gaseous²⁹. This fingerprint could be characteristic of urban sources near
300 to the site.

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304 **3.3 Comparison Among the two PMF Solutions and External Correlations**

305 **Support**

306 The two PMF solution (ALL and HIGH) provided three common factors (OOA, NIT-Reg
307 and Traffic). When correlation of the overlapping temporal trends of the last three days are
308 examined, Table S4 shows a very good agreement between the two PMF solutions: OOA
309 (34-38%, $R^2=0.95$), Nitrate regional (14-17%, $R^2=0.94$) and NO (14-17%, $R^2=0.85$).
310 Sulphate Regional cannot be temporally compared, although the apportionment

311 contributions is similar for both solutions (0-3%). The "Cooking" factor solution for the
312 dataset_ALL is found to contribute less (25%) than the sum of the two "cooking" aerosol
313 factors of the dataset_HIGH (38%). However, it is important to note that the temporal
314 correlation between "COA" from dataset_ALL and the sum of COAs from dataset_HIGH
315 ("COA-MSA-CI" and "COA-MSA-HOA" together) is very high ($R^2=0.85$).

316

317 The factors presented are strongly supported by correlations with external measurements
318 presented elsewhere²¹. Table S5 shows that factor Photox. OOA correlates very well with
319 NMR_PMF_F4 (Aged humic-like substances, $R^2=0.90$), whereas factor SUL-Reg.
320 correlates well with NMR_PMF_F3 (organic sulphate and less aged humic-like
321 substances, $R^2=0.57$). Furthermore, whilst COA-MSA-CI correlates only with
322 NMR_PMF_F1 (aliphatic amines and unspciated aliphatic compounds, $R^2=0.77$), factor
323 COA-MSA-HOA correlates only with NMR_PMF_F2 (aliphatic alkanolic acids and oxo-
324 acids, $R^2=0.95$). These are robust external correlations which strongly support our PMF
325 solutions. Correlations reported in Table S4 are much stronger ($R^2=0.7-1.0$) than those
326 previously reported ($R^2=0.3-0.6$)²¹ because the chemical profiles of the PMF solutions
327 herein presented are less dependent upon specific markers, but rather show a
328 combination of markers that better describe an organic aerosol source.

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331 **4. DISCUSSION**

332

333 **4.1 Major Constituents**

334

335 The analysis of the correlations between the "regional" aerosol components (i.e., the
336 components expected to exhibit an extended source footprint) suggests that the OOA
337 atmospheric evolution over north Italy is completely distinct from that of nitrate of regional
338 origin. Factor "Photox. OOA" was found mainly associated with oxidised organic aerosols
339 (48-88%) and sulphate (48-68%). This is in line with previous studies³³ who reported aged

340 OOA detected above the Po Valley column to be secondary in nature and highly oxidized
341 in the regional aerosol, with much higher amounts of sulphate and organics relative to
342 nitrate. By contrast, NIT-Reg. is found to account for the majority of nitrate aerosol
343 detected (AMS_Nit 72-79%), of regional nature (ATOFMS NIT-Reg 32-71%). The organic
344 component in this factor is found to be of primary origin: AMS_HOA (21-45%). The
345 association of NIT-Reg. with HOA may suggest a larger size mode of nitrate during the
346 night acting as a coagulation sink for fresher finer local aerosols. Alternatively, the HOA
347 may in part be subjected to transport across the Po Valley basin together with nitrate. This
348 factor is associated with a specific mass spectrum and a characteristic temporal dynamic
349 already reported in London²⁹, Wales³⁴ and Barcelona³⁵. The LRT-NIT particle type is
350 volatile, going into the gas phase during the day and leaving a non-volatile internally
351 mixed core mostly composed of sulphate, elemental and organic carbon. The SUL-Reg is
352 mainly composed of EC-Sulphate (90%), BC (42%) and sulphate (27%). During this study
353 we find local, less aged nitrate of finer size mode of local origin (ATOFMS_NIT-local)
354 associated with COA-MSA-HOA.

355
356
357 HOA and BC are usually related to fresh traffic emissions in many AMS studies. This work
358 shows that traffic makes only a small contribution to BC (11%) and a modest one to HOA
359 (23%). By contrast, the majority of BC is related to the core of the regional nitrate, being
360 internally mixed with sulphate and BC (42% of total BC). As a result, it is important to
361 stress that the HOA and BC is mainly seen during nighttime and mainly associated with
362 nitrate.

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4.2 The "Cooking" Aerosol Factor(s)

370 The mass spectrum of the AMS COA factor obtained from San Pietro Capofiume was
371 converted in m/z unit mass resolution and compared with other factors in the literature. All
372 factors were compared at unit mass resolution. The COA factor did not correlate ($R^2 < 0.1$)
373 with any of the HOA, SV-OOA or LV-OOA in the literature. A good correlation (R^2) was
374 found with AMS COA reported in seven previous studies: 0.84 (laboratory studies¹¹); 0.74
375 and 0.83 (London¹²); 0.75 (Manchester¹²); 0.80 (Paris^{14,36}); 0.75 (Zurich¹⁰); 0.85 (Cork¹⁸).
376 Therefore, the same name "COA" was kept for this study.

377

378 The composition of the aerosol and of the air masses associated with the AMS COA
379 factors are key indicators of its origins. A first, unexpected finding is that the PMF analyses
380 from both datasets (ALL and HIGH) show the association of AMS COA with secondary
381 organic aerosol components (CIMS MSA and ATOFMS OC-SUL-NIT) rather than with
382 markers of primary combustion (or thermal) processes. More specifically, the analysis of
383 the HIGH dataset decomposes the "COA" factor into two, with one (COA-MSA-HOA)
384 retaining a partial fingerprint for primary sources (HOA, BC, NO_x) together with that of
385 secondary components (nitrate, MSA) while the other factor (COA-MSA-CI) shows very
386 small correlations with the primary tracers. Fig. S11 shows the main differences among the
387 two COA PMF factors. The low correlation of COA with tracers of urban pollution in this
388 last factor was unexpected because cooking aerosols should be transported in urban
389 plumes to the countryside along with the products of traffic emissions (BC, NO_x, HOA). It is
390 important to stress that very little is known on the transformation of COA in the
391 atmosphere³⁷. Preliminary findings of meat cooking aerosol ageing show an increase of
392 the AMS O:C ratio from 0.1 up to 0.3³⁸. By looking at the prevalent wind directions
393 associated with the two COA factors in the HIGH datasets, the COA-MSA-HOA shows a
394 transport component from the west, while the COA-MSA-CI has a more local footprint.
395 Therefore, it is very unlikely that the COA-MSA-CI originated from cooking activities in

396 urban areas. The local sources to be indicated could be cooking in rural houses and small
397 towns in the vicinity of SPC, but this does not explain the low correlation with the traffic
398 markers (as people who live there still use cars). Most importantly, the source profile
399 emerging from COA-MSA-Cl is inconsistent with the known composition of cooking
400 aerosol, in which particles are mainly organic in nature with inorganic ions being present
401 only in trace amounts³⁷, while here the aerosol mass AMS Cl/ AMS COA ratio is 1.63.
402 These findings indicate that either the published emission composition for cooking
403 emissions does not apply to this environment, or cooking is not a major source for the
404 organic aerosol included in the COA-MSA-Cl factor.

405

406 Explaining the correlation between COA and aerosol chloride is challenging. Aerosol
407 chloride in this study is mainly non-sea salt chloride, which may originate from various
408 anthropogenic sources including industry, combustion and incinerators^(39,45). There are no
409 industrial plants around SPC. Incinerators emit hydrochloric acid which comes from the
410 combustion of plastic material^(39,45). However, the single particle mass spectra associated
411 with this factor (ATOFMS_OC-NIT-SUL) does not have metals previously identified with
412 Cl-containing aerosols emitted by waste incinerators³⁹. Also the SP-AMS did not observe
413 any mass fragments from metals such as lead which typically characterize the emissions
414 of incinerators (Dr. J. Allan, personal communication).

415

416 Finally, the source profile of COA-MSA-Cl indicates that the same source is responsible for
417 the emission of reduced sulfur species (precursors of gas-phase MSA measured by CIMS)
418 which is inconsistent with the hypothesis of a high-temperature combustion process such
419 as in an incinerator. This is because in thermal oxidizers such as industrial incinerators
420 chemically reduced pollutants are generally destroyed via combustion forming more
421 oxidised species such as CO₂, SO₂ and H₂O. On the contrary, the production of reduced

422 sulfur species points to thermal processes occurring at low temperature (including some
423 cooking practices) or to emissions at ambient temperature⁴⁰. Dimethylsulfide emissions
424 from marine biota are believed to be largely responsible for global MSA production⁴¹;
425 however, in this study gas-phase MSA was not associated with marine air masses. On the
426 contrary the MSA concentrations reach a maximum at nighttime when the circulation is
427 mainly from the inland to the sea²¹. A recent review on cooking aerosol tracers actually
428 does not mention sulphur compounds as cooking tracers³⁷. However, it is worth
429 mentioning that roasting of coffee beans may be a source of reduced sulphur
430 compounds⁴² and also meat cooking can produce several thioethers⁴³. However, cooking
431 is only one of the possible anthropogenic sources of sulfur compounds. Urban sources of
432 DMS include aerobic composting of food wastes^{40,44-47}. More generally, volatile organic
433 sulphur compounds can readily form from the aerobic and anaerobic degradation of
434 organic matter, including solid wastes, wastewater, manure and livestock excreta and
435 feed⁴⁴⁻⁴⁸. If we quantify such sources using co-emitted compounds such as ammonia, then
436 emissions in the Po Valley are dominated by the agricultural practices and animal
437 husbandry activities. A prevalent source of MSA precursors in rural areas, with emissions
438 from manure management, livestock and agricultural land is consistent with the
439 appearance of the COA-MSA-CI factor as a local source at SPC unrelated to urban
440 sources. The origin of organic particulate matter associated with the aerobic/anaerobic
441 organic matter decomposition in agricultural land is poorly documented. Broadly, air
442 pollutants emitted from the agricultural sector are mainly methane (CH₄) and other VOCs,
443 nitrous oxide (N₂O) and ammonia (NH₃). Agriculture is also a main source of PM, both
444 primary and secondary in origin originating from livestock production, application of
445 fertilizers and pesticides, land preparation, harvesting and field burning of agricultural
446 waste⁴⁶⁻⁴⁸. It is known that some husbandry activities (poultry) are strong point sources of
447 PM₁₀ and PM_{2.5}, but this aerosol source has never been related to the AMS COA prior to

448 this study. By contrast, the emissions of VOCs from livestock and manure management
449 have been characterized in some detail and often associated with the production of
450 reduced sulfur species⁵⁰⁻⁵¹. Such VOCs comprise low-molecular weight organic acids, but
451 also C₆-C₁₀ aliphatic aldehydes and monocarboxylic acids. These chemical compounds
452 are analogous to the organic compounds believed to contribute to the AMS COA but
453 exhibit too low a molecular weight to exist in the particulate phase. Therefore, the
454 production of COA from these VOCs is possible only via a gas-phase oxidation step. C_{>8}
455 n-alkanals in particular can form SOA under low NO_x conditions⁵², while unsaturated
456 alkanals⁵³ are good SOA precursors also in NO_x-rich environments⁴⁹. According to this
457 hypothesis, the COA in the rural Po Valley may be contributed from secondary organic
458 compounds in the same manner that MSA is produced by the oxidation of reduced sulfur
459 species. It is worthy of note that MSA and COA concentrations peak at nighttime, when
460 OH concentrations are near zero while the nitrate radical NO₃ becomes the major
461 oxidant⁵²⁻⁵⁴. Unlike OH, the concentrations of NO₃ depend on the availability of NO_x, which
462 can explain why the COA-MSA factors show a component of westerly transport, i.e., from
463 the more polluted sector of the valley. It can also explain the correlation of COA with urban
464 tracers (including NO_x) in the factor COA-MSA-HOA. The correlation of COA with aerosol
465 chloride cannot be fully explained by this hypothesis. Possible point sources for
466 hydrochloric acid, reduced sulfur species and VOCs from organic matter degradation are
467 landfills, and there are actually two at ten-twenty km west of SPC.

468

469 In conclusion, the origin of factor COA-MSA-Cl (21% of total aerosol, 54% of AMS COA)
470 can be explained by hypothesizing emissions from agricultural/husbandry activities with a
471 potential additional contribution from waste disposal at urban sites. A second AMS COA
472 source (COA-MSA-HOA, 17% of the total aerosol, 46% of the remaining AMS COA) was

473 also linked with secondary components (nitrate, MSA), although retaining a partial
474 fingerprint for primary traffic sources (HOA, BC, NO_x).

475

476 Quantification of such sources is challenging and calls for more research. However, the
477 results of this study suggests that the current estimates of OA sources from organic matter
478 degradation processes in agricultural and waste systems in populated areas may be
479 substantially underestimated. The COA factor as determined by the AMS includes far
480 more sources and processes than solely primary emissions from cooking activities.

481

482 **Associated content**

483 **Supporting Information**

484 A description and details of the field campaign discussed in the manuscript; PMF analysis,
485 temporal trends of the PMF solutions, external correlation between this PMF study and
486 previous ones are discussed. This material is available free of charge via the Internet at
487 <http://pubs.acs.org>.

488

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493 **Notes**

494 The authors declare no competing financial interest.

495

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507 **References**

508

509

510 (1) Pope, C. A. and Dockery, D. W.: Health effects of fine particulate air pollution: Lines
511 that connect, *J. Air Waste Manage.*, **2006**, 56, 709–742.

512

513 (2) Zhang, Q., Jimenez, J.L., Canagaratna, M.R., Allan, J.D., Coe, H., Ulbrich, I.,
514 Alfarra, M.R., Takami, A., Middlebrook, A.M., Sun, Y.L., et al. Ubiquity and dominance of
515 oxygenated species in organic aerosols in anthropogenically influenced northern
516 hemisphere mid-latitudes. *Geophys. Res. Lett.*, **2007**, 34 L13801.

517

518 (3) Robinson, A. L., Donahue, N. M., Shrivastava, M. K., Weitkamp, E. A., Sage, A. M.,
519 Grieshop, A. P., Lane, T. E., Pierce, J. R., and Pandis, S. N.: Rethinking organic aerosols:
520 Semivolatile emissions and photochemical aging, *Science*, **2007**, 315, 1259–1262.

521

522 (4) Jimenez, J. L., Canagaratna, M. R., Donahue, N. M., Prevot, A. S. H., Zhang, Q., Kroll,
523 J. H., DeCarlo, P. F., Allan, J. D., Coe, H., Ng, N. L., et al.: Evolution of organic aerosols in
524 the atmosphere, *Science*, **2009**, 326, 1525–1529.

525

526 (5) Dall'Osto, M., Thorpe, A., Beddows, D. C. S., Harrison, R. M., Barlow, J. F., Dunbar, T.,
527 Williams, P. I., and Coe, H.: Remarkable dynamics of nanoparticles in the urban
528 atmosphere, *Atmos. Chem. Phys.*, **2011**, 11, 6623-6637, doi:10.5194/acp-11-6623-2011

529

530 (6) Laskin A, Laskin J, Nizkorodov SA. Mass spectrometric approaches for chemical
531 characterization of atmospheric aerosols: critical review of the most recent advances.
532 *Environ Chem* **2012**;9:163–89.

533

534 (7) Jayne, J. T., Leard, D. C., Zhang, X., Davidovits, P., Smith, K. A., Kolb, C. E., and
535 Worsnop, D. R.: Development of an aerosol mass spectrometer for size and composition
536 analysis of submicron particles, *Aerosol Sci. Technol.*, **2000**, 33, 49–70.

537

538 (8) Donahue, N.M., Robinson, A.L. and Pandis, S.N. Atmospheric Organic Particulate
539 Matter: From Smoke to Secondary Organic Aerosol. *Atmos. Environ.*, **2009**, 43: 94–106.

540

541 (9) Ulbrich, I. M., Canagaratna, M. R., Zhang, Q., Worsnop, D. R., and Jimenez, J. L.:
542 Interpretation of organic components from Positive Matrix Factorization of aerosol mass
543 spectrometric data, *Atmos. Chem. Phys.*, **2009**, 9, 2891–2918, doi:10.5194/acp-9-2891-
544 2009.

545

- 546 (10) Lanz, V. A., Alfarra, M. R., Baltensperger, U., Buchmann, B., Hueglin, C., and Prevot,
547 A. S. H.: Source apportionment of submicron organic aerosols at an urban site by factor
548 analytical modelling of aerosol mass spectra, *Atmos. Chem. Phys.*, **2007**, 7, 1503–1522,
549 doi:10.5194/acp-7-1503-2007.
550
- 551 (11) Mohr, C., Huffman, J. A., Cubison, M. J., Aiken, A. C., Docherty, K. S., Kimmel, J. R.,
552 Ulbrich, I. M., Hannigan, M., and Jimenez, J. L.: Characterization of primary organic
553 aerosol emissions from meat cooking, trash burning, and motor vehicles with High-
554 Resolution Aerosol Mass Spectrometry and comparison with ambient and chamber
555 observations, *Environ. Sci. Technol.*, **2009**, 43, 2443–2449, doi:10.1021/es8011518.
556
- 557 (12) Allan, J. D., Williams, P. I., Morgan, W. T., Martin, C. L., Flynn, M. J., Lee, J., Nemitz,
558 E., Phillips, G. J., Gallagher, M. W., and Coe, H.: Contributions from transport, solid fuel
559 burning and cooking to primary organic aerosols in two UK cities, *Atmos. Chem. Phys.*,
560 **2010**, 10, 647-668, doi:10.5194/acp-10-647-2010.
561
- 562 (13) Mohr, C., DeCarlo, P. F., Heringa, M. F., Chirico, R., Slowik, J.G., Richter, R., Reche,
563 C., Alastuey, A., Querol, X., Seco, et al.: Identification and quantification of organic aerosol
564 from cooking and other sources in Barcelona using aerosol mass spectrometer data,
565 *Atmos. Chem. Phys.*, **2012**, 12, 1649–1665, doi:10.5194/acp-12-1649-2012.
566
- 567 (14) Crippa, M., DeCarlo, P. F., Slowik, J. G., Mohr, C., Heringa, M.F., Chirico, R., Poulain,
568 L., Freutel, F., Sciare, J., Cozic, J., et al.: Wintertime aerosol chemical composition and
569 source apportionment of the organic fraction in the metropolitan area of Paris, *Atmos.*
570 *Chem. Phys.*, 2013, 13, 961–981, doi:10.5194/acp-13-961-2013.
571
- 572 (15) Huang, X.-F., He, L.-Y., Hu, M., Canagaratna, M. R., Sun, Y., Zhang, Q., Zhu, T., Xue,
573 L., Zeng, L.-W., Liu, X.-G., Zhang, Y.-H., et al.: Highly time-resolved chemical
574 characterization of atmospheric submicron particles during 2008 Beijing Olympic Games
575 using an Aerodyne High-Resolution Aerosol Mass Spectrometer, *Atmos. Chem. Phys.*,
576 **2010**, 10, 8933–8945, doi:10.5194/acp-10-8933-2010.
577
- 578 (16) Sun, Y. L., Zhang, Q., Schwab, J. J., Yang, T., Ng, N. L., and Demerjian, K. L.: Factor
579 analysis of combined organic and inorganic aerosol mass spectra from high resolution
580 aerosol mass spectrometer measurements, *Atmos. Chem. Phys.*, **2012**, 12, 8537-8551,
581 doi:10.5194/acp-12-8537-2012.
582
- 583 (17) He, L. Y., Huang, X. F., Xue, L., Hu, M., Lin, Y., Zheng, J., Zhang, R., and Zhang, Y.
584 H. L.: Submicron aerosol analysis and organic source apportionment in an urban
585 atmosphere in Pearl River Delta of China using high-resolution aerosol mass
586 spectrometry, *J. Geophys. Res.*, **2011**, 116, D12304, doi:10.1029/2010JD014566.
587
- 588 (18) Dall'Osto M, Ovadnevaite J, Ceburnis D, Martin D, Healy RM, O'Connor IP, et al.
589 Characterization of urban aerosol in Cork city (Ireland) using aerosol mass spectrometry.
590 *Atmos Chem Phys*, **2013**;13:4997–5015. <http://dx.doi.org/10.5194/acp-13-4997-2013>.
591
592
- 593 (19) Hayes, P. L., A. M. Ortega, M. J. Cubison, K. D. Froyd, Y. Zhao, S. S. Cliff, W. W. Hu,
594 D. W. Toohey, J. H. Flynn, B. L. Lefer, et al., Organic aerosol composition and sources in
595 Pasadena, California, during the 2010 CalNex campaign, *J. Geophys. Res.*, **2013**,
596 doi:10.1002/jgrd.50530.
597

598 (20) Yin, J., Cumberland, S. A., Harrison, R. M., Allan, J., Young, D. E., Williams, P. I., and
599 Coe, H.: Receptor modelling of fine particles in southern England using CMB including
600 comparison with AMS-PMF factors, *Atmos. Chem. Phys.*, **2015**, 15, 2139-2158,
601 doi:10.5194/acp-15-2139-2015.
602

603 (21) Decesari, S., Allan, J., Plass-Duelmer, C., Williams, B. J., Paglione, M., Facchini, M.
604 C., O'Dowd, C., Harrison, R. M., Gietl, J. K., Coe, H., et al.: Measurements of the aerosol
605 chemical composition and mixing state in the Po Valley using multiple spectroscopic
606 techniques, *Atmos. Chem. Phys.*, **2014**, 14, 12109-12132, doi:10.5194/acp-14-12109-
607 2014.
608

609 (22) Larssen, S., R. Sluyter, C. Helmis, Criteria for EUROAIRNET, the EEA Air Quality
610 Monitoring and Information Network, **1999**.
611

612 (23) Berresheim, H., Elste, T., Plass-Dülmer, C., Eisele, F. L., and Tanner, D. J.: Chemical
613 ionization mass spectrometer for long-term measurements of atmospheric OH and
614 H₂SO₄, *Int. J. Mass Spectrom.*,
615 202, 91–109, 2000.
616

617 (24) Song, X. H., Hopke, P. K., Fergenson, D. P., and Prather, K. A.: Classification of single
618 particles analyzed by ATOFMS using an artificial neural network, ART-2A, *Anal. Chem.*,
619 71, 860–865, 1999.
620

621 (25) Onasch, T. B., Trimborn, A., Fortner, E. C., Jayne, J. T., Kok, G.L., Williams, L. R.,
622 Davidovits, P., and Worsnop, D. R.: Soot particle aerosol mass spectrometer:
623 development, validation, and initial application, *Aerosol Sci. Tech.*, 46, 804–817, 2012.
624

625 (26) Paatero, P. and Tapper, U.: Positive matrix factorization: A nonnegative factor model
626 with optimal utilization of error estimates of data values, *Environmetrics*, **1994**, 5, 111–126
627

628 (27) Paatero, P. and Tapper, U.: Positive matrix factorization: A non-negative factor model
629 with optimal utilization of error estimates of data values, *Environmetrics*, **1994**, 5, 111–126
630

631 (28) Paatero, P.: Least squares formulation of robust non-negative factor analysis,
632 *Chemometrics and Intelligent Laboratory Systems*, **1997**, 37, 23–35.
633

634 (29) Dall'Osto, M., Harrison, R. M., Coe, H., Williams, P. I., and Allan, J. D.: Real time
635 chemical characterization of local and regional nitrate aerosols, *Atmos. Chem. Phys.*,
636 **2009**, 9, 3709-3720.
637

638 (30) Gilardoni, S., Massoli, P., Giulianelli, L., Rinaldi, M., Paglione, M., Pollini, F.,
639 Lanconelli, C., Poluzzi, V., Carbone, S., Hillamo, R., et al.: Fog scavenging of organic and
640 inorganic aerosol in the Po Valley, *Atmos. Chem. Phys.*, 14, 6967-6981, doi:10.5194/acp-
641 14-6967-2014, 2014.
642

643 (31) Saarikoski, S., Carbone, S., Decesari, S., Giulianelli, L., Angelini, F., Canagaratna,
644 M., Ng, N. L., Trimborn, A., Facchini, M. C., Fuzzi, et al.: Chemical characterization of
645 springtime submicrometer aerosol in Po Valley, Italy, *Atmos. Chem. Phys.*, 12, 8401–
646 8421, doi:10.5194/acp-12-8401-
647 2012, 2012.
648

- 649 (32) Yienger, J. J., and H. Levy II. 1994. Empirical model of global soil biogenic NO_x
650 emissions. *Journal of Geophysical Research* 100:11 447–11 464.
651
- 652 (33) Crosier, J. Allan, J.D., Coe, H., Bower, K. N., Formenti P. and Williams P. I.,
653 Chemical composition of summertime aerosol in the Po Valley (Italy), northern Adriatic and
654 Black Sea, *Q. J. R. Meteorol. Soc.*, **2007**, 133: (S1) 61–75
655
- 656 (34) Dall'Osto, M., M. J. Booth, W. Smith, R. Fisher and R. M. Harrison. A study of the size
657 distributions and the chemical characterization of airborne particles in the vicinity of a large
658 integrated steelworks. *Aerosol Science And Technology*, **2008**, 42(12): 981-991.
659
- 660 (35) Dall'Osto, M., Querol, X., Alastuey, A., Minguillon, M. C., Alier, M., Amato, F., Brines,
661 M., Cusak, M., Grimalt, J. O., Karanasiou, et al.: Presenting SAPUSS: Solving Aerosol
662 Problem by using synergistic strategies at Barcelona, Spain, *Atmos. Chem. Phys.*, **2013**,
663 12, 18741-18815,
664
- 665 (36) Crippa, M., El Haddad, I., Slowik, J. G., DeCarlo, P. F., Mohr, C., Heringa, M., Chirico,
666 R., Marchand, N., Sciare, J., Baltensperger, U., et al.: Identification of
667 marine and continental aerosol sources in Paris using high resolution aerosol mass
668 spectrometry, *J. Geophys. Res.*, **2013**, 118, 1950–1963, doi:1910.1002/jgrd.50151.
669
- 670 (37) Abdullahi X. and Harrison Roy. M. Emissions and indoor concentrations of particulate
671 matter and its specific chemical components from cooking: A review., *Atmospheric*
672 *Environment*, **2013**, 71, 260-294
673
- 674 (38) Kaltsonoudis, C., Kostenidou, E., Louvaris, E., Psichoudaki, M., Tsiligiannis, E.,
675 Florou, K. , and S.N. Pandis, Characterization of fresh and aged organic aerosol
676 emissions from meat charbroiling. EAC 2015, Milan 6-11 September 2015
677
- 678 (39) Moffet RC, Desyaterik Y, Hopkins RJ, Tivanski AV, Gilles MK, Wang Y, et al.
679 Characterization of aerosols containing Zn, Pb, and Cl from an industrial region of Mexico
680 City. *Environ Sci Technol* **2008**;42:7091–7. <http://dx.doi.org/10.1021/es7030483>
681
- 682 (40) Zhang H., Schuchardt F., Li G., Yang Y., Yang Q. Emission of volatile sulfur
683 compounds during composting of municipal solid waste (MSW). *Waste Management*,
684 **2013**, 33, 957–963.
685
- 686 (41) Rinaldi, M., S. Decesari, E. Finessi, L. Giulianelli, C. Carbone, S. Fuzzi, C. D. O'Dowd,
687 D. Ceburnis, and M. C. Facchini. Primary and secondary organic marine aerosol and
688 oceanic biological activity: Recent results and new perspectives for future studies, *Adv.*
689 *Meteorol.*, **2010**, 310682, doi:10.1155/2010/310682.
690
- 691 (42) Kabir and Kim. An investigation on hazardous and odorous pollutant emission during
692 cooking activities. *Journal of Hazardous Materials*, **2013**, 188 (2011) 443–454
693
- 694 (43) Byrne, D. V., Bredie, W. L. P., Mottram, D. S., and Martens, M. Sensory and Chemical
695 Investigations on the Effect of Oven Cooking on Warmed-Over Flavour Development in
696 Chicken Meat. *Meat Sci.*, **2001**, 61:127–139.
697
- 698 (44) Wu, T., Wang, X.M., Li, D.J., Yi, Z.G., Emission of volatile organic sulfur compounds
699 (VOSCs) during aerobic decomposition of food wastes. *Atmos. Environ.*, **2010**, 44, 5065–
700 5071.

701
702 (45) Komilis, D.P., Ham, R.K., Park, J.K., Emission of volatile organic compounds
703 during composting of municipal solid wastes. *Water Res.*, **2004**, 38, 1707–1714.
704
705 (46) Trabue, S., K. Scoggin, F. Mitloehner, H. Li, R. Burns, H. Xin, Field sampling method
706 for quantifying volatile sulphur compounds from animal feeding operations, *Atmos.*
707 *Environ.*, **2008** 42, 3332–3341.
708
709 (47) Song, S.-K., Z.-H. Shon, K.-H. Kim, Photochemical oxidation and dispersion of
710 gaseous sulfur compounds from natural and anthropogenic sources around a coastal
711 location, *Atmos. Environ.*, **2009**, 43, 3015–3023.
712
713 (48) Kim, K.-H., R. Pal, J.-W. Ahn, Y.-H. Kim, Food decay and offensive odorants: A
714 comparative analysis among three types of food, *Waste Management*, **2009**, 29 1265–
715 1273.
716
717 (49) Chan, A. W. H., Chan, M. N., Surratt, J. D., Chhabra, P. S., Loza, C. L., Crouse, J.
718 D., Yee, L. D., Flagan, R. C., Wennberg, P. O., and Seinfeld, J. H.: Role of aldehyde
719 chemistry and NO_x concentrations in secondary organic aerosol formation, *Atmos. Chem.*
720 *Phys.*, **2010**, 10, 7169-7188, doi:10.5194/acp-10-7169-2010.
721
722 (50) Alanis, P., M. Sorenson, M. Beene, C. Krauter, B. Shamp, A. S. Hasson,
723 Measurement of non-enteric emission fluxes of volatile fatty acids from a California dairy
724 by solid phase micro-extraction with gas chromatography/mass spectrometry, *Atmos.*
725 *Environ.*, **2008**, 42, 6417–6424.
726
727 (51) Howard et al., Direct Measurements of the Ozone Formation Potential from Livestock
728 and Poultry Waste Emissions, *Environ. Sci. Technol.* **2010**, 44, 2292–2298.
729
730 (52) Aimanant, S., and P. J. Ziemann, Chemical Mechanisms of Aging of Aerosol Formed
731 from the Reaction of n-Pentadecane with OH Radicals in the Presence of NO_x, *Aerosol*
732 *Science and Technology*, **2013**, 47:979–990
733
734 (53) Kim, K.-H., and S.-Y. Park, A comparative analysis of malodor samples between
735 direct (olfactometry) and indirect (instrumental) methods, *Atmos. Environ.*, **2008**, 42
736 5061–5070.
737
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746 **TABLE LEGEND**

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748 **Table 1.** Percentages of factors over the same time interval (9-12/07/2009) for the
749 dataset_1 (all field study) and dataset_2 (high, only second part).

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753 **FIGURE LEGENDS**

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755 **Figure 1.** PMF factors of the dataset 1 (ALL field study).

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757 **Figure 2.** Diurnal profiles of dataset_1 (ALL field study) and dataset_2 (second part,
758 HIGH). Figure 3 (1-2) shows again the COA factors coming from dataset_1
759 (COA) and dataset_2 (COA-MSA-CI and COA-MSA-HOA).

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761 **Figure 3.** PMF factors of the dataset_2 for the second part of the field study (HIGH).

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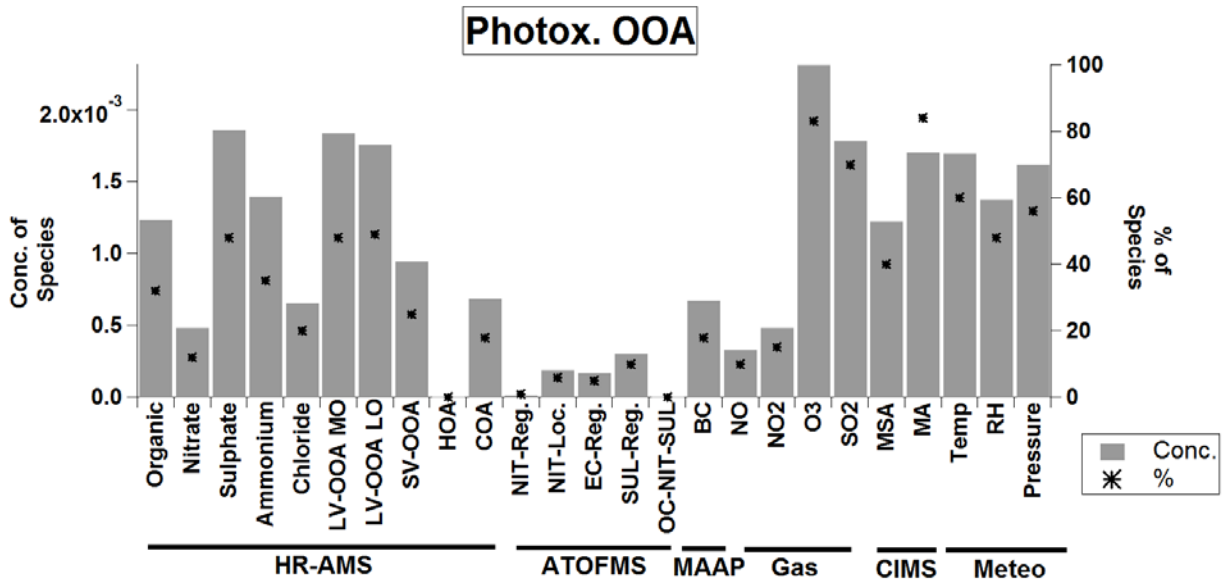
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772 **Table 1.** Percentages of factors over the same time interval (9-12/07/2009) for the
 773 dataset_1 (all field study) and dataset_2 (high, only second part)
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Time period	dataset type	OOA-Reg.	NIT-Reg.	SUL-Reg.	NO	COA-MSA-CI (from ALL)	COA-MSA-CI (from HIGH)	COA-MSA-HOA (from HIGH)
Whole study	ALL	30	18	23	12	17		
Only 9-12/07/2009	ALL	38	17	3	17	25		
Only 9-12/07/2009	HIGH	34	14	0	14		21	17

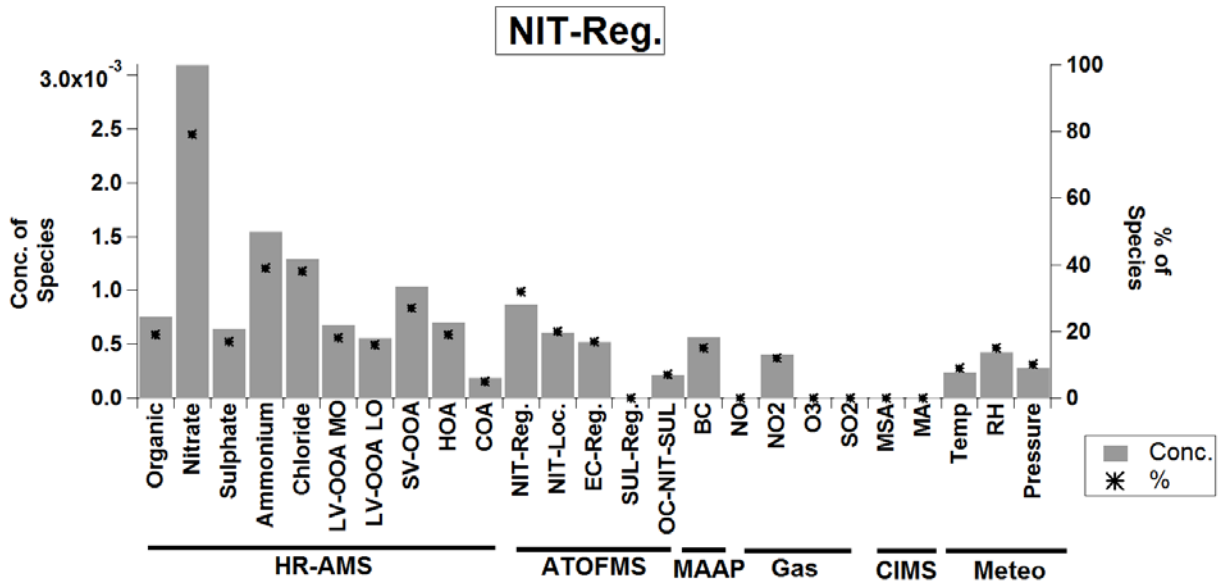
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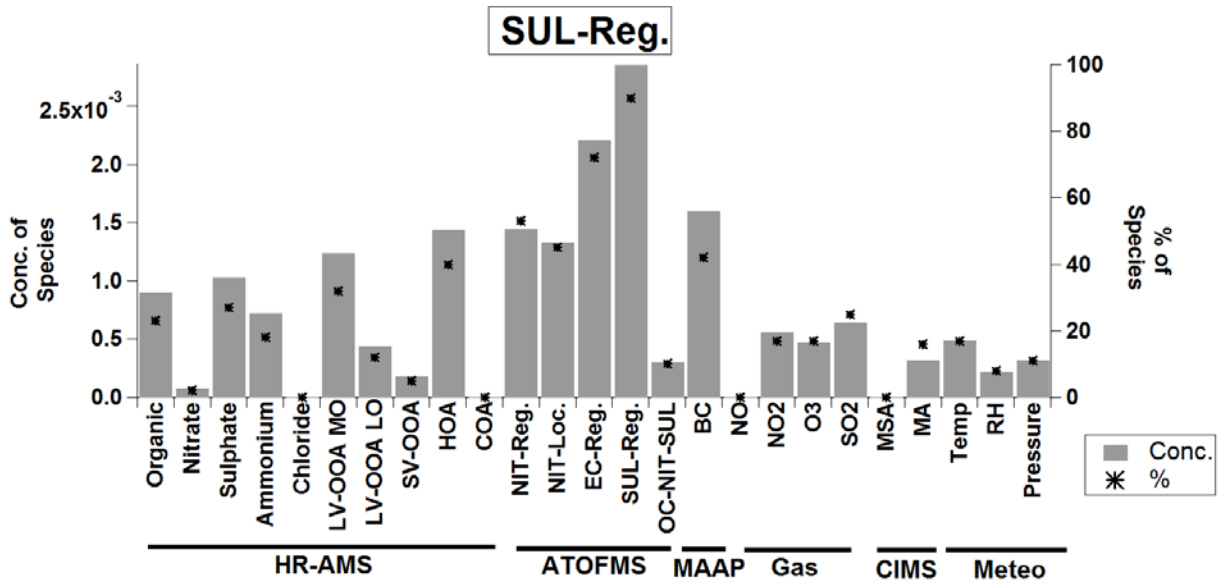
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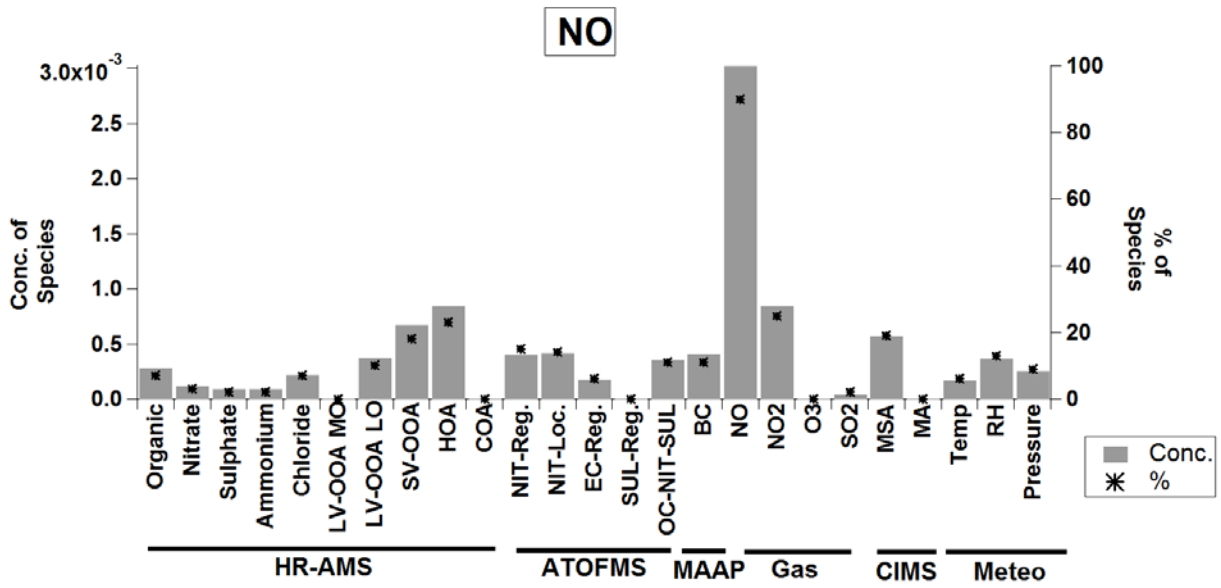


(b)

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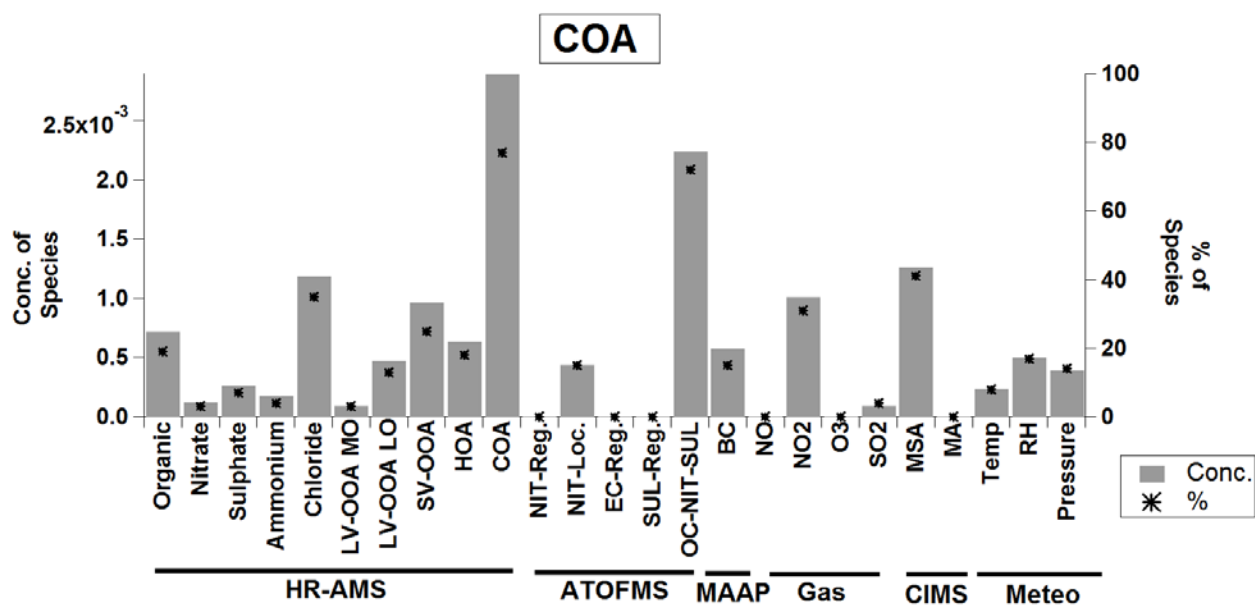
(c)



(d)

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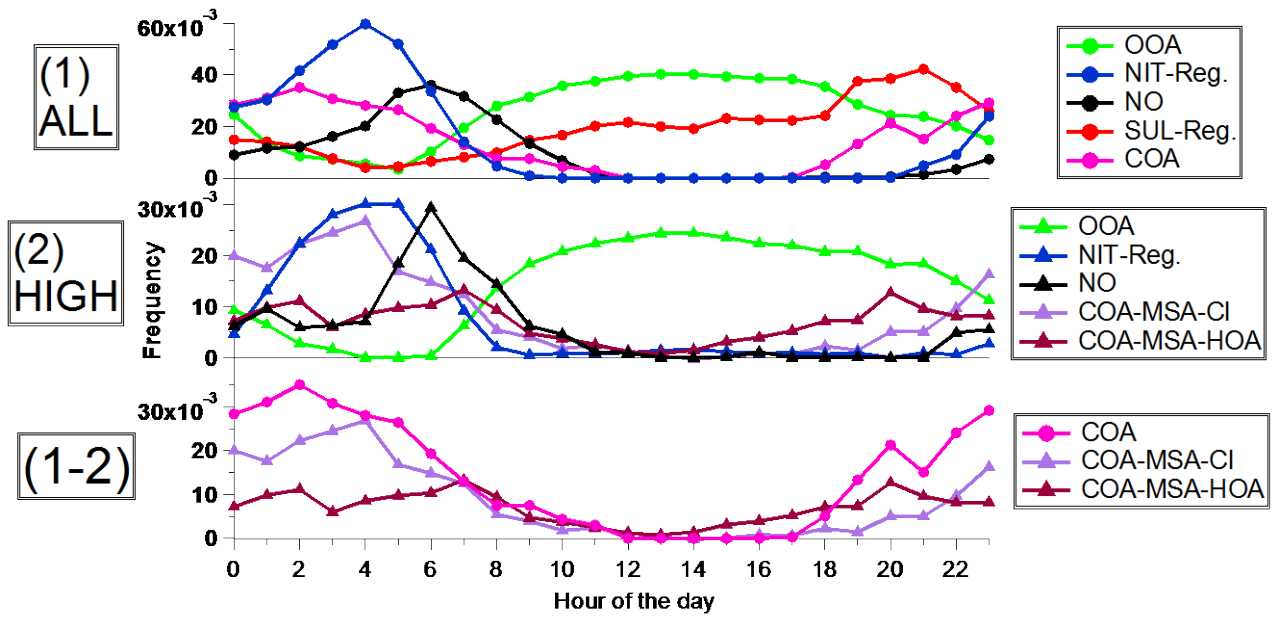


(e)

Figure 1. PMF factors of the dataset 1 (ALL field study)

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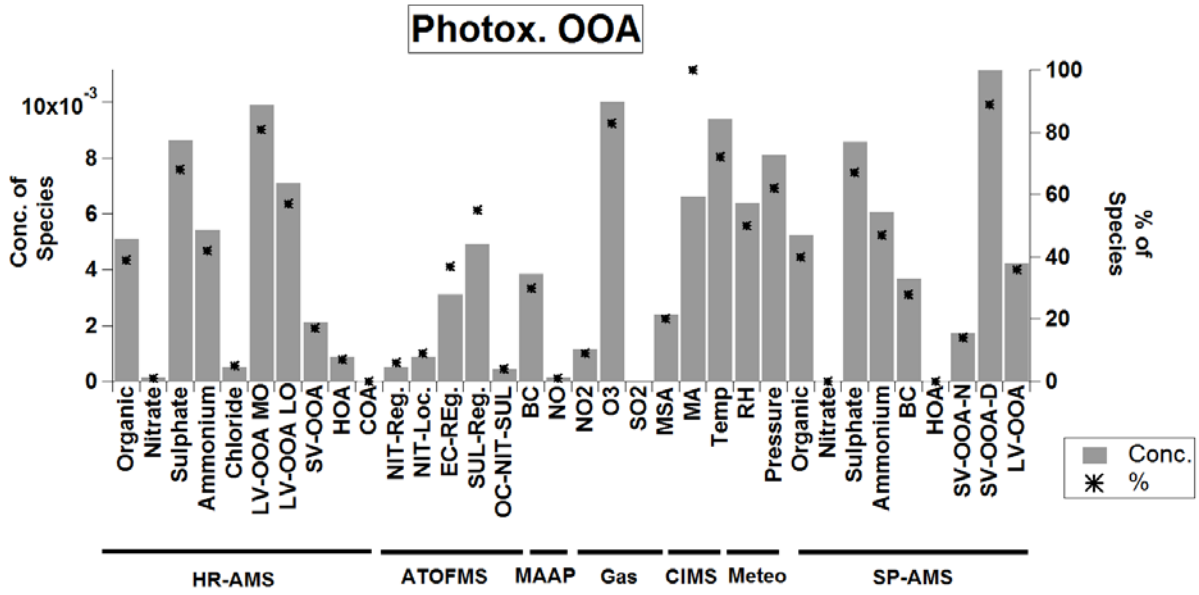
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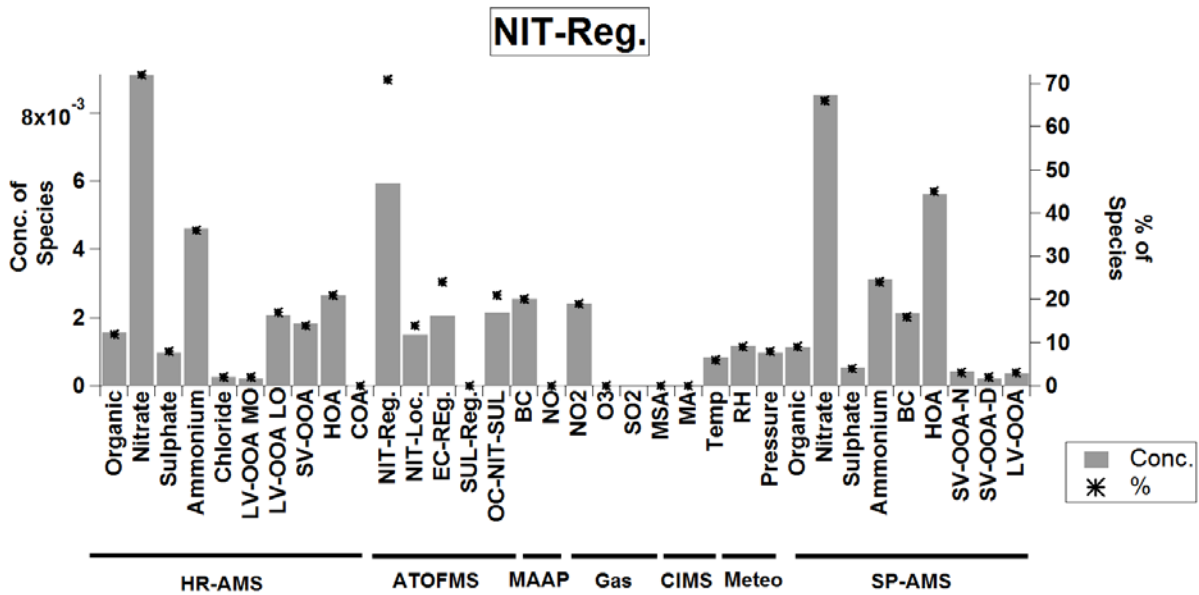
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Figure 2. Diurnal profiles of dataset_1 (ALL field study) and dataset_2 (second part, HIGH). Figure (1-2) shows again the COA factors coming from dataset_1 (COA) and dataset_2 (COA-MSA-CI and COA-MSA-HOA)

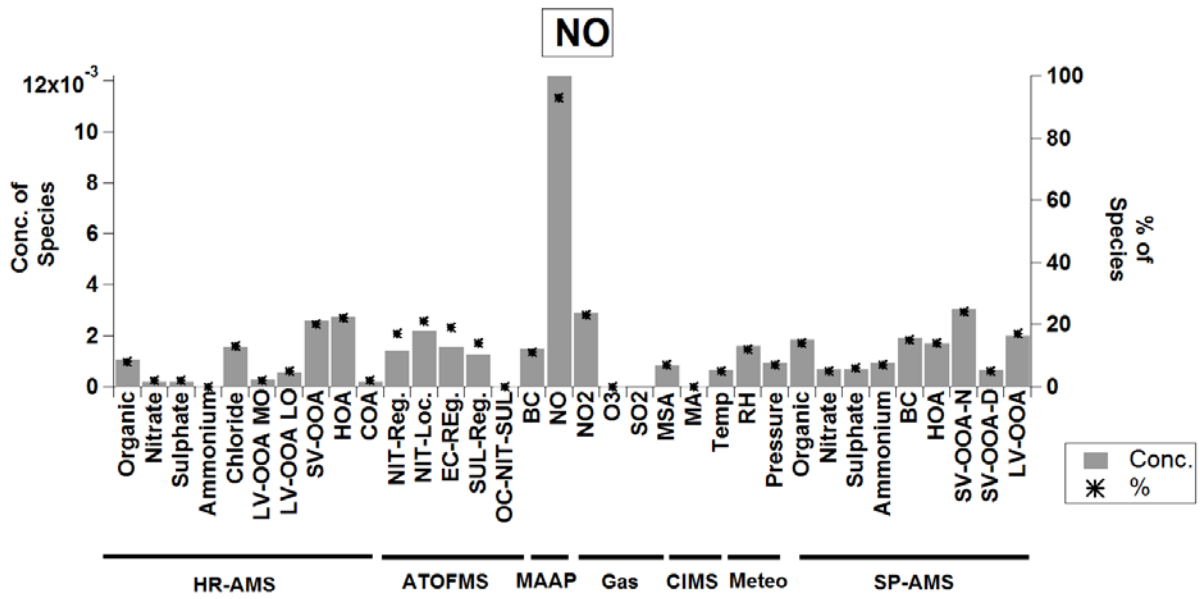
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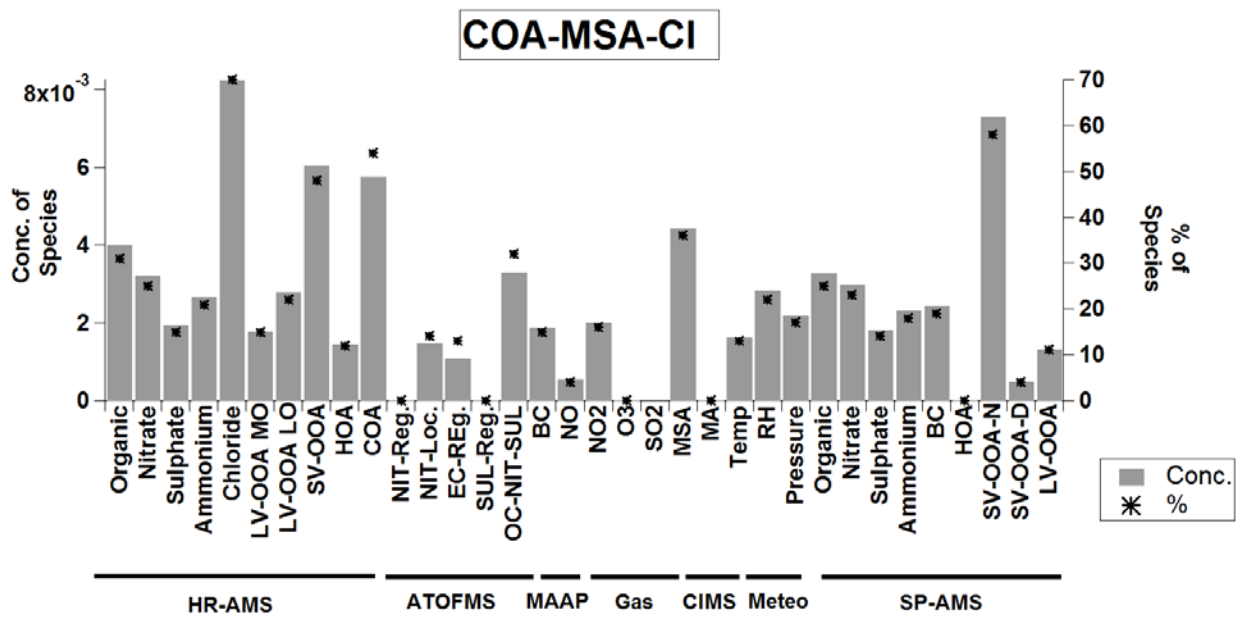
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910 (a)
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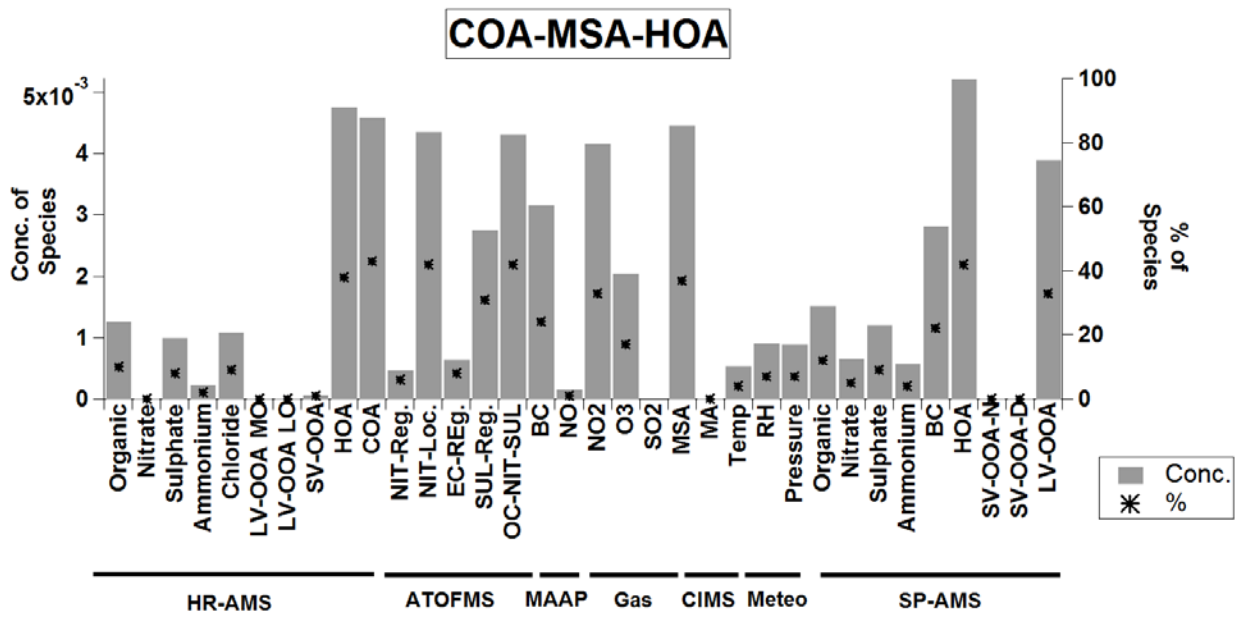
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915 (b)
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919 (c)
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925 (d)
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929 (e)

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Figure 3. PMF factors of the dataset_2 for the second part of the field study (HIGH)