

Nature of air pollution, emission sources, and management in the Indian cities

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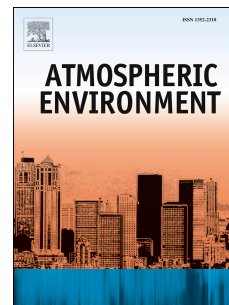
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Nature of Air Pollution, Emission Sources, and Management in the Indian Cities

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ABSTRACT

The global burden of disease study estimated 695,000 premature deaths in 2010 due to continued exposure to outdoor particulate matter and ozone pollution for India. By 2030, the expected growth in many of the sectors (industries, residential, transportation, power generation, and construction) will result in an increase in pollution related health impacts for the most cities. The available information on urban air pollution, their sources, and the potential of various interventions to control the pollution, should help us propose a cleaner path to 2030. In this paper, we present an overview of the emission sources and control options for better air quality in Indian cities, with a particular focus on interventions like urban public transportation facilities; travel demand management; emission regulations for power plants; clean technology for brick kilns; management of road dust; and waste management to control open waste burning. Also included is a broader discussion on key institutional measures, like public awareness and scientific studies, necessary for building an effective air quality management plan in Indian cities.

KEYWORDS

Emissions Control; Vehicular Emissions; Power Plants; Brick Kilns; Health Impacts

HIGHLIGHTS

- Air quality monitoring in the Indian cities
- Sources of air pollution in the Indian cities
- Health impacts of outdoor air pollution in India
- Review of air quality management options at the national and urban scale

1.0 INTRODUCTION

Air quality is a cause for concern in India, particularly in cities and air pollutants including particulate matter (PM), sulfur dioxide (SO₂), nitrogen oxides (NO_x), carbon monoxide (CO), and ozone (O₃) often exceed the National Ambient Air Quality Standards (NAAQS). According to the World Health Organization (WHO), 37 cities from India feature in the top 100 world cities with the worst PM₁₀ pollution, and the cities of Delhi, Raipur, Gwalior, and Lucknow are listed in the top 10 (WHO, 2014). A similar assessment by WHO, in 2011, listed 27 cities in the top 100. More than 100 cities under the national ambient monitoring program exceed the WHO guideline for PM₁₀.

In India, the national ambient standard for CO is better than the WHO guideline. The NO₂, SO₂, and O₃ standards are at par with the guidelines. However, the standards for PM₁₀ (Aerodynamic diameter <10 µm) and PM_{2.5} (aerodynamic diameter <2.5 µm) are lagging (comparative details in **Supplementary Material**).

As cities are increasing in size and population, there is a steady demand for motorized vehicles in both personal and public transport sectors. This puts substantial pressure on the city's infrastructure and environment, particularly since most Indian cities have mixed land use. For 40 cities highlighted in **Census-India (2012)**, the key urban characteristics are presented in **Table 1**. The urban population varies from 1.5 million to 17 million. The data shows that regardless of population size, 30 cities are densely populated with 100 persons per hectare or more. Another 30 cities have at least 30% of the households with a motorized two wheeler (MTW) and 19 cities have at least 10% households with a four-wheeler (a car or a utility vehicle). While most cities are supplied with liquefied petroleum gas (LPG) for domestic use, there is still a significant portion of households using other fuels - such as kerosene, biomass, and coal. Of the 40 cities in **Table 1**, 20 have at least 30% of households with a non-LPG cookstove.

In 2010, the Central Pollution Control Board (CPCB) developed the Comprehensive Environmental Pollution Index (CEPI), a methodology to assess air, water, and soil pollution at the industrial clusters in the country (CPCB, 2009). While industries typically rely on the grid electricity for operations and maintenance; frequent power cuts often necessitate the use of in-situ electricity generation (using coal, diesel, and heavy fuel oil), which adds to the industrial air pollution load. The study identified 43 clusters with a rating of more than 70, on a scale of 0 to 100, and listed them as critically polluted for further action. Most of these clusters are in and around major cities – most notably Korba (Chhattisgarh), Vapi (Gujarat), Faridabad and Ghaziabad (outside of Delhi), Ludhiana (Punjab), Kanpur and Agra (Uttar Pradesh), Vellore and Coimbatore (Tamil Nadu), Kochi (Kerala), Vishakhapatnam (Andhra Pradesh), Howrah (West Bengal), and Bhiwadi (Rajasthan). The CEPI ratings, where available, are listed by their ranking in **Table 1**.

The global burden of disease (GBD) assessments, listed outdoor air pollution among the top 10 health risks in India. The study estimated 695,000 premature deaths and loss of 18.2 million healthy life years due to outdoor PM_{2.5} and ozone pollution (IHME, 2013). Among the health risk factors studied, outdoor air pollution was ranked 5th in mortality and 7th in overall health burden in India. Household (indoor) air pollution from burning of solid fuels was

85 responsible for an additional one million premature deaths. A substantial increase was observed
86 in the cases of ischemic heart disease (which can lead to heart attacks), cerebrovascular disease
87 (which can lead to strokes), chronic obstructive pulmonary diseases, lower respiratory infections,
88 and cancers (in trachea, lungs, and bronchitis). Several other studies have estimated premature
89 mortality rates due to outdoor PM pollution for several Indian cities, using similar methodologies
90 and are summarized in **Table 2**.

91
92 While the field of air pollution and atmospheric science is gaining ground in India and
93 there has been a surge in the published research, much of the knowledge is widely scattered.
94 While reviews in the past have provided scientific recommendations (**Pant and Harrison, 2012;**
95 **Krishna, 2012**), there has been no concerted effort towards addressing the various aspects of the
96 air pollution (source to impacts), and providing a global summary as well as gaps in current
97 knowledge. Existing local (and international) knowledge can be leveraged in designing effective
98 interventions in India, where pollutant sources are often complex. In this paper, we aim to
99 present an overview of the emission sources and control options for air quality improvements in
100 Indian cities, with a particular focus on key sectors such as transportation, dust, power plants,
101 brick kilns, waste, industries, and residential. Also included is a broad discussion on key
102 institutional measures necessary for building an effective air quality management plan.

103 104 **2.0 AIR QUALITY IN INDIAN CITIES**

105 106 **2.1 Air Quality Data**

107
108 The national ambient monitoring program collects 24-hour averages of key air pollutants
109 2-3 times per week at 342 manual stations in 127 cities and is managed by CPCB. However, only
110 a limited number of cities operate continuous monitoring stations, measuring the full array of
111 criteria pollutants and as such, access to the monitoring data is limited. A summary of the
112 measurements for PM₁₀, SO₂, and NO₂ for 2009-10 is presented in **Table 1**. Delhi and Pune also
113 have citywide monitoring networks outside the national framework (**SAFAR, 2013**).

114
115 To supplement the data generated at on-ground monitoring stations, several studies have
116 utilized satellite data to derive global ground-level ambient PM_{2.5} concentrations (**Van**
117 **Donkelaar et al., 2010**). These were utilized for the GBD assessments (**IHME, 2013**). An
118 extract of this data covering India is presented in **Figure 1**. Since the satellite extractions are
119 available at 0.1° resolution (~10 km), there is some uncertainty associated with these derivatives
120 and these retrieval methods are being improved every year, to complement the on-ground
121 measurements. For example, most of southern India in **Figure 1** seems to comply with the WHO
122 guideline of 15 µg/m³. However, the urban pollution levels here are some of the highest in the
123 country including Chennai and Coimbatore (Tamil Nadu), Hyderabad and Vishakhapatnam
124 (Andhra Pradesh), Kochi (Kerala), and Bengaluru (Karnataka).

125
126 The Indo-Gangetic plain has the largest number of brick kilns, with old and inefficient
127 combustion technology, using a mix of biomass and coal for combustion needs (**Maithe et al.,**
128 **2012**). The states of Bihar, West Bengal, Jharkhand, Orissa, and Chhattisgarh harbor the largest
129 coal mines in the country, and a cluster of power plants around the mines (**Guttikunda and**
130 **Jawahar, 2014**). Several large power plants also exist in the states of Punjab, Haryana, Delhi,

131 and Uttar Pradesh, making the north and north-eastern belt the most polluted part of the country.
132 The cities in the north are also landlocked, which are also affected by the prevalent
133 meteorological conditions, unlike some of the Southern cities with the privilege of land-sea
134 breezes (Guttikunda and Gurjar, 2012).

135
136 Besides $PM_{2.5}$ concentrations, the satellite observations can also help estimate the
137 concentrations of SO_2 , NO_x , and CO, and help analyze the severity of on-ground anthropogenic
138 and natural emission sources (Streets et al., 2013).

139 140 **2.2 Sources of Air Pollution**

141
142 For city administrators, regulating air pollution is the primary concern and accurate
143 knowledge of the source contributions is vital to developing an effective air quality management
144 program. The contribution of various sources to the ambient PM pollution is typically assessed
145 via receptor modeling and this methodology has been applied in many Indian cities (CPCB,
146 2010; Pant and Harrison, 2012). However, between 2000 and 2013, 70% of the known studies
147 were conducted in five big cities - Delhi, Mumbai, Chennai, Kolkata, and Hyderabad and very
148 limited number in other cities, which are also listed as exceeding the ambient standards and
149 WHO guidelines (Table 1 and WHO, 2014). The number of studies in various cities is presented
150 in Table 3, with limited number of studies on $PM_{2.5}$ size fraction.

151
152 The most commonly identified sources are vehicles, manufacturing and electricity
153 generation industries, construction activities, road dust, waste burning, combustion of oil, coal,
154 and biomass in the households, and marine/sea salt. A multi-city study was conducted by CPCB
155 for six cities – Pune, Chennai, Delhi, Mumbai, Kanpur, and Bengaluru, at an approximate project
156 cost of US\$6 million (CPCB, 2010). A summary of this study is presented in Figure 2. Unlike
157 the popular belief that road transport is the biggest cause of urban air pollution, the CPCB
158 (2010) results showed that there are other sources which also need immediate attention and the
159 road transport is only one of the major contributors to the growing air quality problems in the
160 cities. In the six cities, the share of road transport ranged 7% in Pune to 43% in Chennai. Along
161 with Mumbai, Chennai harbors one of the largest commercial ports in India, which means a large
162 number of diesel fueled heavy duty trucks pass through the city, to and from the port, and thus
163 increasing the share of road transport in ambient PM pollution.

164
165 However, a vital limitation of the receptor modeling approach is the spatial representation
166 of the contributions, i.e., the results are representative of the sampling location and its close
167 vicinity (~2-3km radius from the sampling location). Hence, in order to understand the source
168 contributions in a city, it is important to conduct detailed analysis at multiple locations. The
169 receptor modeling studies also require detailed chemical characterization of emission sources
170 (either as an input or for validation). While several source profiles have been created in India,
171 there is scope for further addition and improvement (Pant and Harrison, 2012).

172 173 **2.3 Emissions Inventory & Dispersion Modeling**

174
175 While the receptor modeling approach is ideal to ascertain the source contributions, it is
176 an expensive method and has limited spatial coverage. This limitation can be overcome by

177 complementary source modeling. This relies on availability of data such as vehicle activity on
178 the roads, fuel consumption in the domestic, industrial, and electricity generation sectors, waste
179 collection and waste burning, silt loading on the roads for resuspension, and geography,
180 population, and meteorology of the city. Since the analysis includes spatial dispersion modeling,
181 there is a need for computational facilities to input, analyze, and output geo-referenced emissions
182 inventories and concentration fields. Several studies have been carried out in India since 2000
183 and are summarized together with known emission factor databases and key requirements to
184 conduct dispersion modeling in the **Supplementary Material**. An example of gridded vehicle
185 exhaust emissions from Pune, Chennai, and Ahmedabad at 1 km grid resolution is also presented
186 in the **Supplementary Material**. Most of the studies have focused on PM pollution and a very
187 few studies on ozone pollution.

188
189 The source modeling results are driven by user inputs and rely on measurements to test
190 their validity, before they can be used for any policy dialogue. It is also important to note that the
191 dispersion and receptor models should not be used as substitutes for each other and they provide
192 best results when used complementarily.

193 194 **3.0 POTENTIAL for AIR POLLUTION CONTROL**

195
196 In most Indian cities, SO₂ is the only pollutant that complies with NAAQS. Interventions
197 such as introduction of Bharat-4 diesel (with 50 ppm sulfur) in the cities and Bharat-3 diesel
198 (with 350 ppm sulfur) for the rest of the country and relocation or refurbishing of industries
199 consuming coal and diesel with better efficiency norms have led to this compliance. However,
200 the same is not true for PM₁₀, CO, and NO₂ concentrations, as their emissions from combustion
201 processes have significantly increased, regardless of the improvements in fuel quality and
202 technology. While most of the industrial equipment and the on-road vehicles individually adhere
203 to their respective emission norms; collectively, they emit enough to register ambient
204 concentrations beyond the standards (**Table 1**).

205
206 There is a vast potential for pollution control in the cities, which can be achieved through
207 the twin approach of stringent regulations technological interventions. The following sections
208 provide an overview of the potential measures, some under implementation and some which
209 need urgent attention.

210 211 **3.1 Vehicle Fuel Standards and Alternative Fuels**

212
213 Road transport plays a vital role in India's growing economy and the contribution of
214 vehicular emissions is only expected to increase (**Ghate and Sundar, 2013**). Fuel emission
215 standards in India lag behind the global emission standards (**Table 4**). It is essential to implement
216 and enforce Bharat-5 (equivalent of Euro-5) or higher standards nationwide by 2015 or sooner,
217 in order to maintain a balance between the energy demand and the growing emissions
218 (**Guttikunda and Mohan, 2014**). Any delay in implementation or staggered implementation (as
219 is the case currently), will result in a delayed response for improving air quality in Indian cities.

220
221 While the staggered introduction of the fuel standards is beneficial for the cities in the
222 short run, the overall benefits are lost in transition. For example, the heavy duty vehicles

operating on diesel contribute significantly to PM emissions and often run on lower grade fuel, which can lead to failure of catalytic converters. It is therefore imperative that “*one nation, one fuel standard*” norm is mandated for better air quality in the cities.

There is an increasing focus on shifting the public transport and para-transit vehicles to run on compressed natural gas (CNG). In Delhi, buses, three-wheeler rickshaws, and taxis were converted to operate on CNG and a steady supply of fuel coupled with lower prices is encouraging private car owners also to switch. The number of CNG outlets in Delhi increased from 30 in year 2000 to 300 in 2013, according to the Ministry of Petroleum & Natural Gas of India, there will be 200 cities with CNG network by 2015 (**PIB, 2013**).

3.3 *Urban Travel Demand Management*

As the cities are growing in geography and inhabitants, there is also a push to promote safe and clean public transport systems. In bigger cities like Delhi, Mumbai, Hyderabad, Kolkata, Chennai, Ahmedabad, and Bangalore, there is an established formal public transportation system and they also benefitted from Jawaharlal Nehru National Urban Renewal Mission (JNNURM) programs to better and increase the fleet (**MoUD, 2012**). Since 2009, more than 14,000 new buses were delivered under this program. However, most of these cities need to at least triple or quadruple the current fleets, in order to for the 4-wheeler and 2-wheeler passengers to shift public transport systems.

In urban India, implementation of dedicated bus corridors, known as “bus rapid transport (BRT) system” is among the priorities. International examples from Bogota (Colombia) and Curitiba (Brazil) serve as models with bus modal share of 62% and 45% respectively (**LTA Academy, 2011**). The cities of Delhi, Ahmedabad, Jaipur, Pune, and Indore have implemented BRT projects with varying corridor lengths and the cities of Rajkot, Surat, Bhopal, Vijayawada, and Visakhapatnam have approved BRT projects (**Mahadevia et al., 2013**). Since the projects are not fully integrated into the public transport systems, the results and the public response has been mixed.

For smaller cities, the definition of the public transport is changing. Most of these cities do not have an organized public transportation system; rather, they are supported by informal para-transit systems, mostly plying on the dominant corridors of the city. Among the para-transit systems, most common are the traditional three-wheeler auto-rickshaws (to seat up to 4 people), a larger version of the auto-rickshaws (to seat up to 10 people), and mini-buses. With their ability to negotiate the tiny by-lanes and weave through mixed traffic, these vehicles form an integral part of passenger and freight movement and in most cities is also a popular mode of mass transport for school children. An example is the city of Alwar (Rajasthan) where para-transit system “Alwar Vahini” was successfully formalized with regulations, as well as dedicated routes reaching various parts of the city. The para-transit systems have also benefited from the use of alternative fuels like CNG and LPG. In response to a Supreme Court mandate, the Delhi Government converted the entire three-wheeler fleet to CNG between 1998 and 2002, followed by similar initiatives in other cities.

268 Most Indian cities have a majority share of trips by walk and cycle (**Mohan, 2013**). This
269 is because of low vehicle ownership (compared to the cities of the United States and the
270 European Union) as well as traditional mixed-use design of the cities, which leads to shorter
271 access to work, school, and other activities. In big cities with higher population density, in the
272 absence of dedicated NMT infrastructure, motorized vehicles also pose serious risk of injury,
273 because of which, people owning two-wheelers and cars are encouraged to use their vehicles,
274 even for walkable distances. In the context of growing cities, the measures to improve air quality
275 should include NMT policies as an integral part.

276
277 Some economic measures are also designed to force the use of public transport. One such
278 measure is the congestion pricing – where the motorists are charged to use a network of roads
279 during periods of the heaviest use. Its purpose is to reduce automobile (mostly car) use during
280 peak congestion periods, thereby easing traffic and encouraging commuters to walk, bike, or take
281 mass transit rail/bus as an alternative. Congestion pricing programs were successfully
282 implemented in Singapore, London, and Stockholm (**Eliasson, 2009; Menon and Guttikunda,**
283 **2010; Litman, 2011**). On average, in London, congestion pricing is estimated to have reduced
284 20-30% of the downtown passenger car traffic and promote the non-motorized transport,
285 whereas Stockholm experienced an immediate reduction of at least 20% in the daily car use. In
286 Singapore, the average traffic speeds increased by at least 15 km/h. In all three cities, 10-20%
287 reduction in eCO₂ emissions was estimated, along with health benefits of reducing air pollution.

288
289 A major reason for its success in Singapore, London, and Stockholm was the availability
290 of widely accessible public transport system (road and rail) which can support the shift to a car-
291 free environment. If implemented, there will be immediate benefits in big cities like Delhi,
292 Mumbai, and Chennai. However, the public transport system is still not at par with those in
293 Singapore, London, and Stockholm for effective implementation of this option.

294
295 While congestion pricing policies are difficult to replicate in the Indian context, at least
296 for the foreseeable future, there is an important lesson. With increasing costs for private vehicles
297 linked with their usage (fuel and other operational expenses), it is possible to achieve a shift to
298 public transport, if combined with the provision of an adequate, reliable, and safe public
299 transportation. One such measure is the increased parking cost. Currently, parking in most cities
300 is either free or priced very low. Increased parking cost, if coupled with the parking locations so
301 that they are as far as the bus and the rail stops, will make public transportation an attractive
302 option (**Barter, 2012; CSE, 2012**).

303
304 While the congestion pricing and parking policies target reduced vehicle usage, some
305 countries have used regulatory measures to reduce the growth of private vehicles. For instance, a
306 Chinese national regulation enacted in September, 2008, raised taxes on big cars and reduced on
307 smaller ones. Car owners with engines above 4-liters capacity have to pay a 40% tax; 15% to
308 25% for cars with engines above 3-liters capacity; and 1% to 3% for cars with engines below 1-
309 liter capacity (**Murad, 2008**). China also introduced a policy to limit the number of licenses
310 issued every year, where the license plates are auctioned in the cities of Beijing, Shanghai, and
311 Guangzhou. Similar to congestion pricing, for the time being, such measures are difficult to
312 implement under democratic political context of India.

313

3.4 *Regulations for Coal-fired Power Plants*

In 2011-12, there were 111 coal-fired power plants in India with a combined generation capacity of 121GW (CEA, 2012). The emissions and pollution analysis for these plants, resulted in an estimated 80,000 to 115,000 premature deaths and more than 20.0 million asthma cases from exposure to total PM_{2.5} pollution annually (Guttikunda and Jawahar, 2014). While Indian coal has a low sulfur content in comparison with other coals, ash levels are reported to be quite high and can contribute to coarse PM emissions (Pant and Harrison, 2012).

Despite the volume of coal used in the power generation sector, there are very few regulations in place to address these environmental and health costs. To date, for PM emissions, the emission standard in India lags to those implemented in China, Australia, the United States, and the European Union (Table 5). For other key pollutants, there are no prescribed emission standards despite the fact that India is a relatively dense country and several power plants are in close proximity to residential areas. Aggressive pollution control regulations such as mandating flue gas desulfurization, introduction of tighter emission standards for all criteria pollutants, and updating procedures for environment impact assessments are imperative for regional clean air and to reduce health impacts. For example, a mandate for installation of flue gas desulfurization systems could reduce the PM_{2.5} concentrations by 30-40% by eliminating the formation of the secondary sulfates and nitrates (Guttikunda and Jawahar, 2014).

Besides flue gas PM emissions, fugitive dust from coal-handling plants and ash ponds (after the disposal from the plants) is also a problem. According to Central Electrical Authority, after the combustion and application of control equipment, ash collection at the power plants range 70-80% of the total ash in the coal. In 2003, an amendment notification from MoEF mandated 25% bottom ash in all brick kilns within 100km radius of the power plant and all building construction within 100km to use 100% ash based bricks, blocks, and tiles. To date, percentage of ash utilized in the construction industry is low.

3.5 *Power Shortages and Diesel Generators*

In 2011, the peak electricity demand was approximately 122 GW and the peak electricity supply was approximately 110 GW (CEA, 2012). The gap between the supply and the demand is crucial to understand India's power generation sector. In India, a third of the population in rural areas do not have access to electricity and those areas on the grid are not assured of uninterrupted supply. The blackout in July, 2012, that paralyzed 600 million people in 22 states in the Northern, Eastern, and North-eastern India, is testament to how tenuous power situation in the country.

In the urban areas, power cuts are severe in the winter and the summer months, when heaters and air conditioners are in full service, respectively. These needs are usually supplemented by in-situ large, medium, and small diesel generator (DG) sets at hotels, hospitals, malls, markets, large institutions, apartment complexes, cinemas, and farm houses and form an additional source of air pollution to an already deteriorating air quality in the cities. Telecommunication towers stand out, which with more than 500 million mobile phones in the country, are heavily dependent on DG sets. According to the Telecom Regulatory Authority of

360 India (TRAI), 310,000 towers consumed over 2 billion liters of diesel in 2010 (TRAI, 2011).
361 The mobile subscriber base is expected to reach 800 million by the end of 2013 with an
362 additional 100,000 telecom towers in service.

363
364 For Chennai, Pune, Indore, Ahmedabad, Surat, and Rajkot, up to 10% of the modeled
365 PM₁₀ concentrations were found to originate from diesel generator sets (Guttikunda and
366 Jawahar, 2012). In case of Delhi, the contribution went up to 15% (Guttikunda and Calori,
367 2013; Guttikunda and Goel, 2013). The use of generator sets is even higher in the rural areas,
368 where they are mostly utilized for the pumping water in the agricultural lands.

369
370 An alternative to DG sets is not simple. One option is to increase the number of power
371 plants to meet the electricity demand and reduce the transmission losses or provide alternatives
372 like renewable energy. On the other hand, tightening of the emission standards for DG sets, at
373 par with the heavy and light duty vehicles can help control some emissions.

374 375 **3.6 Domestic Fuels**

376
377 The issue of indoor air pollution is also critical because of the high magnitude of
378 population getting exposed to such pollution every day (Venkatraman et al., 2010; Grieshop et
379 al., 2011; Smith et al., 2014). According to GBD 2010 assessments, household air pollution has
380 been a persistent health hazard in India and has retained its position as the 2nd highest risk factor
381 for disability life years lost and resulting in 1.0 million premature deaths annually (IHME,
382 2013). Several studies have characterized the impacts of household air pollution and health risks
383 due to indoor solid fuel use in India (Balakrishnan et al., 2011).

384
385 In 2011, Non-LPG fuels are used in 35% of urban and 89% of rural households in India,
386 compared to 52% and 94% respectively in 2001 (Census-India, 2012), with little improvement
387 for the rural households. Keeping in view the magnitude of the health risks, the Ministry of
388 Petroleum & Natural Gas of India aims to provide LPG connections to up to 75% of households
389 in India by year 2015 (PIB, 2013). For the rural areas, the program is also proposing to set up
390 community kitchens with LPG connections where users can pay as per usage. Alternatives like
391 electric (induction) stoves are cleaner (with pollution generated somewhere else), but their usage
392 is dependent on power supply in the rural communities. The use of electric stoves, in the urban
393 centers, is increasing and currently limited by the cost of stoves.

394 395 **3.7 Emerging Technologies for the Brick Kilns**

396
397 The traditional building construction in India believes in the fired clay bricks. The
398 construction is one of the fastest growing sectors, contributing approximately 10% of the
399 national GDP and has a growth rate of 9%. Overall, there are more 100,000 brick kilns in India,
400 producing 150-200 billion bricks annually, employing 10 million workers, and consuming 25
401 million tons of coal (Isebell et al., 2007; Maithel et al., 2012). Within a radius of 20-30km,
402 one can spot 500-1000 brick manufacturing kilns in the big cities of Delhi, Chennai, Kolkata,
403 Pune, Ahmedabad, Hyderabad, Kanpur, and Patna (Guttikunda and Jawahar, 2012;
404 Guttikunda and Calori, 2013).

405

406 Brick manufacturing includes land clearing for sand and clay, combustion of fuel for
407 burning, operation of diesel engines on site, and transport of the end product to various parts of
408 the city. Traditionally, the rectangle shaped clay bricks are sun dried and readied for firing in
409 “clamps” - a pile of bricks with intermittent layers of sealing mud, and fuel. A significant amount
410 of energy is lost during the cooling with no possibility of recycling. These are mostly located in
411 the Central and Northeast India. In the Indo-Gangetic plains, some advanced manufacturing
412 technologies with a 50m fixed chimney (FCK) are in practice. The chimney supports further
413 dispersion of the emissions and is noted for its low cost of construction. A major disadvantage of
414 these kilns is associated with weather – an open cast kiln means they can be operated only in the
415 non-monsoonal season.

416
417 The emerging technologies and their operational features are summarized in **Table 6**.
418 Between the available technologies, the newer technologies such as vertical shaft brick kilns
419 (VSBK), zigzag, and Hoffmann, could result in at least 40% reduction in the PM emission rates
420 compared to the currently in use FCK technology. VSBK design and construction details are
421 presented as a manual in **World Bank (2010)** with some pilots in India, Bangladesh, and Nepal.

422 **3.8 Construction**

423
424
425 Construction activities emit particles during various activities including block cutting,
426 excavation, demolition, mixing, road building, drilling, loading and unloading of debris etc. In
427 addition, movement of vehicles (especially trucks) in and around construction sites increases the
428 amount of particles by crushing and pulverizing the particles on the road surface. Several studies
429 have highlighted the importance of construction as a PM source, though in receptor modeling
430 studies, the source is combined with sources such as road dust (**Pant and Harrison, 2012**). For
431 six cities, construction accounted for up to 10% of the annual emissions (**Guttikunda and**
432 **Jawahar, 2012; CPCB, 2010**). There is a window of opportunity for the establishment of best
433 practices in the construction industry as there are no nation-wide norms and regulations.

434 **3.9 Resuspended/Road Dust**

435
436
437 Dust is a major concern for many cities in India (**Figure 2**) and comprises of particles
438 emitted due to wear and tear of tyres and brakes and materials from the roads, pavements, and
439 street furniture. Dust loadings increase even further in case the roads are not paved (**CPCB,**
440 **2010**).

441
442 Traditionally, all the streets, sidewalks and public areas are swept manually and
443 depending on the resources, poor or marginal areas receive reduced or inadequate service, or no
444 service at all, while wealthier, commercial, or tourist areas receive extensive service. However,
445 often in manual street sweeping, most of the dust swept is left on the side of the roads, which
446 gets re-entrained when the vehicle movement resumes during the day. A better alternative is
447 heavy-duty or light-duty trucks with vacuum cleaners to suck up dust from the roads and/or
448 water sprinklers, so that the resuspension of any leftover dust is suppressed. The operational
449 costs of mechanized sweeping could be similar to manual sweeping - given the latter is a labor
450 intensive exercise. Since the road dust accounts for up to 30-40% of the PM₁₀ pollution in most

451 cities (CPCB, 2010; Guttikunda and Jawahar, 2012), an immediate intervention for this
452 source is considered to be the lowest hanging fruit.

453 3.10 *Open Waste Burning*

454
455 There is no reliable national-level data on the technical or financial aspects of solid waste
456 management, and figures are therefore approximations. The country's annual generation of
457 municipal solid waste is in the range of 35 to 45 million tons; likely to quadruple by 2030 at
458 which time, the waste generation will be over 150 million tons a year (World-Bank, 2006;
459 Annepu, 2012). The scale is perhaps more comprehensible at the city level - the national capital
460 region of Delhi generates 10,000 tons/day and Mumbai 7000 tons/day. The waste collection
461 efficiency in the cities is 50-90%, depending on the commercial and residential activities in
462 various parts of the city. The waste not collected is eventually burnt. The best option for reducing
463 the waste burning is to implement an efficient solid waste management program across the
464 country. Currently, only a few cities operate landfills and manage waste collections. The waste
465 management falls under the jurisdiction of the municipal corporation and according to Toxics
466 Link (New Delhi, India), the municipalities spend an estimated INR 1500-2000 per ton of solid
467 waste for collection (60-70%), transportation (20-30%), and treatment and disposal (5-10%).
468

469
470 The open waste burning problems are particularly worse in the medium and small scale
471 cities, with very limited or no waste collection and no landfill facilities. Emissions of criteria
472 pollutants from garbage burning are hard to quantify and are accompanied by large uncertainty.
473 The toxic chemicals released during burning of paper, plastic, and biomass includes NO_x , SO_2 ,
474 volatile organic chemicals (VOCs), and dioxins.

475 4.0 POLICY IMPLICATIONS

476
477 According to the 2011 census, 2774 rural settlements are now reclassified as urban
478 settlements, pushing the total to 3894, primarily based on the definition of an urban settlement -
479 population exceeds 5000, population density is above 400 per km^2 , and more than 75% of the
480 male workforce is employed outside of agriculture (Census-India, 2012). The population in
481 urban areas is expected to grow from 30% to 50% by 2030 (MoUD, 2011). By 2030, the
482 expected growth in industrial, transportation, domestic, and power generation sectors will
483 consequently result in an increase in emissions and air pollution for almost all the cities listed in
484 **Table 1** and many more cities which are medium and small today. There are growing concerns
485 about the impact of air pollution on human health and general well-being, which requires a
486 multi-pronged approach for better air quality. Some key institutional measures which can led the
487 way are the following
488

489 4.1 *Monitoring and Data Dissemination*

490
491 Before the GBD 2010 assessments (IHME, 2013), the 2008 Olympic Games in Beijing
492 resulted in raising significant awareness on the urban air pollution problems. In Beijing, stringent
493 measures were introduced to achieve blue skies rating for the Games and this included shutting
494 down 50% of the on-road vehicles (based on their registration numbers) and shutting major coal-
495 fired industrial estates in the vicinity of the city. These measures not only raised the awareness
496

497 on the air pollution and benefits of pollution control measures, but also showed the level of
498 commitment necessary from the public and the government to achieve such a result.

499
500 The city of Delhi hosted the 2010 Commonwealth Games, which raised some awareness
501 on air pollution in the city. The number of official monitoring stations tripled during the Games.
502 The national bodies established variable messaging systems and also disseminated the
503 monitoring data on web (**Beig et al., 2013**). However, the emission control programs did not
504 involve any stringent measures like those practiced during the Beijing Olympics (**Streets et al.,**
505 **2007; Wang et al., 2010**) and the implemented measures were found to be ineffective during the
506 Games (**Beig et al., 2013**). While such mega events help raise awareness intermittently, more
507 efforts are necessary to prioritize air pollution and health into long term policy planning.

508
509 There is an acute need to not only increase the monitoring programs across all cities, but
510 also to publicly disseminate the information. While the monitoring stations can be established,
511 legislations amended, and standards improved, these efforts will be a waste, if the regular
512 dissemination of the information is not practiced to raise the awareness for pollution control

513 514 **4.2 Scientific Studies**

515
516 The array of air quality studies in India (listed in **Supplementary Material**) points to
517 significant need for information on spatial and temporal resolution of emissions inventories and
518 pollution dispersion characteristics in the cities. It is important to build the necessary capacity of
519 the state pollution control boards (SPCBs) to undertake focused analysis as well as scrutiny of
520 intervention programs to improve the air quality.

521
522 Much of the past and current work focuses on PM_{10} and there is a growing trend towards
523 analysis of $PM_{2.5}$ (**Table 3**). Due to the negative health impacts, there is a growing need for
524 assessment of particles in the small size ranges (PM_1 and less). Very little is known in terms of
525 spatial and temporal distribution of air pollutants and parameters such as particle size distribution
526 (mass/number/volume). Semi-volatile and volatile organic species have also not been analyzed in
527 much detail (**Krishna, 2012; Ram et al., 2014**). Some recent work focused on unique sources
528 such as funeral pyres (**Chakrobarty et al., 2013**) and festive biomass burning (**Deka and**
529 **Hoque, 2014**) and specific chemical components of PM such as brown carbon (**Srinivas and**
530 **Sarin, 2014**).

531
532 One of the key missing pieces is the personal exposure and toxicological analysis which
533 is an important input for the assessment of health impacts of outdoor and household air pollution
534 (**Clark et al., 2013**). This includes assessment of indoor and outdoor concentrations of various
535 pollutants and integrated physio-chemical characterization (**Wong et al., 2008; Massey et al.,**
536 **2012; Balakrishnan et al., 2013; Chithra and Shiva Nagendra, 2013; Habil et al., 2013**).

537 538 **4.3 Conceptualizing Combined-Benefits**

539
540 All the combustion sectors are interlinked, which leads to the concept of co-benefits for
541 better air quality in the cities. There are many definitions for co-benefits and one of them is “a
542 single activity or policy that can generate multiple benefits across varying sectors or fields”. For

543 example (a) a program in the solid waste management where the wet waste is collected and used
544 to generate biogas, can support the domestic energy needs and thus reduce the load at the local
545 power plant (b) a program designed to increase the energy efficiency of industries can lead to the
546 supply of the excess power to areas plagued with frequent outages, and thus resulting in lesser
547 usage of DG sets.

548
549 The definition of co-benefits can also be extended to pollutants. For example, programs
550 to control pollutants like PM, SO₂, NO_x, and CO, can also reduce greenhouse gas emissions, and
551 vice versa. For example (a) a program to improve the fuel economy of the vehicles by increasing
552 vehicle speeds by reducing the number of cars and motorcycles on the roads, will result in a
553 systematic reduction of all pollutant emissions (b) a program to encourage LPG and electricity
554 usage for domestic cooking and heating, will result in lesser usage of the conventional fuels like
555 coal, biomass, and kerosene, and can result in reduction of all pollutant emissions.

556 **5.0 CONCLUSIONS**

557
558
559 Ground measurements, computational studies, and satellite measurements, are all
560 pointing towards changing pollution trends in India. Today, India has the unique opportunity to
561 further the air pollution abatement measures at the urban and the regional scale, but these depend
562 on effective inter-sector and inter-ministerial collaboration. This is primarily due to the fact that
563 all the sources contributing to the air pollution are interlinked. Among the sectors, transportation
564 is the most critical and most connected.

565
566 Following the review, two challenges have emerged for better air quality in the Indian
567 cities (a) the need to secure greater public awareness of the problems and commitment to action
568 at civic, commercial, and political levels (b) to ensure that action to tackle air pollution is seen in
569 the context of wider social and economic development policies. For example, how much can
570 these interventions help reduce the local challenges, like providing safer and reliable public
571 transportation systems; cleaner and efficient waste management; dust free roads; and pollution
572 free industries and power plants.

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574
575
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733

1 **Table 1: Cities at a glance**

2

City	AR	Pop	A	B	C	D	E	F	PM10 ($\mu\text{g}/\text{m}^3$)	SO2 ($\mu\text{g}/\text{m}^3$)	NO2 ($\mu\text{g}/\text{m}^3$)
Hyderabad	500	7,749,334	155	50%	14%	32%	70.82 (40)	No	81.2 \pm 34.0	5.0 \pm 2.4	22. \pm 7.0
Vijayawada	79	1,491,202	189	26%	4%	21%		No	79. \pm 14.9	4.6 \pm 0.5	13.5 \pm 3.1
Vishakhapatnam	159	1,730,320	109	36%	8%	21%		Yes	91.2 \pm 34.8	11.9 \pm 12.7	29.1 \pm 13.8
Guwahati	145	968,549	67	10%	3%	80%		No	132.6 \pm 89.9	8.1 \pm 3.3	16.6 \pm 5.3
Patna	86	2,046,652	238	32%	10%	29%		No	138.8 \pm 84.4	5.3 \pm 2.8	32.9 \pm 18.8
Korba	39	365,073	94	43%	8%	56%		No	116.9 \pm 17.	13.3 \pm 0.7	21.3 \pm 0.8
Raipur	95	1,122,555	118	38%	9%	48%	65.85 (63)	No	272.2 \pm 43.3	17.8 \pm 3.7	45.9 \pm 2.7
Delhi	669	16,314,838	244	39%	21%	9%		No	260.1 \pm 117.1	6.5 \pm 4.2	51.1 \pm 17.2
Ahmedabad	275	6,352,254	231	51%	13%	24%	75.28 (22)	No	94.3 \pm 21.8	15.9 \pm 3.5	20.9 \pm 4.0
Rajkot	86	1,390,933	162	60%	10%	33%	66.76 (59)	No	105.6 \pm 27.	11.3 \pm 2.1	15.4 \pm 2.6
Surat	155	4,585,367	296	44%	9%	28%	57.9 (79)	No	89.1 \pm 13.1	18.6 \pm 3.9	26.3 \pm 3.2
Vadodhara	145	1,817,191	125	60%	14%	20%	66.91 (57)	No	86. \pm 34.6	16.2 \pm 5.8	30.2 \pm 13.1
Vapi	37	163,605	44	44%	11%	32%	88.09 (2)	No	78.3 \pm 8.1	16.4 \pm 1.9	23.9 \pm 1.7
Yamuna Nagar	41	383,318	93	42%	13%	24%		No	281.5 \pm 132.3	12.7 \pm 2.7	27.1 \pm 3.3
Dhanbad	45	1,195,298	266	31%	5%	72%	78.63 (13)	No	164. \pm 95.5	16.6 \pm 3.5	41. \pm 8.9
Jamshedpur	119	1,337,131	112	49%	12%	38%	66.06 (61)	No	171.7 \pm 13.4	36.4 \pm 2.2	49.3 \pm 3.9
Ranchi	106	1,126,741	106	43%	13%	36%		No	178.9 \pm 67.9	18.1 \pm 2.2	31.6 \pm 3.0
Bangalore	556	8,499,399	153	46%	18%	20%		No	109.4 \pm 92.6	15. \pm 3.1	37.5 \pm 6.0
Jammu	123	651,826	53	48%	25%	13%		No	118.2 \pm 37.4	8.2 \pm 4.4	12.7 \pm 3.4
Trivandrum	108	1,687,406	156	34%	17%	43%		Yes	62.9 \pm 17.8	9.7 \pm 5.2	26.1 \pm 5.2
Bhopal	178	1,883,381	106	48%	15%	30%		No	118.5 \pm 73.2	7.1 \pm 2.4	17.5 \pm 5.9
Gwalior	78	1,101,981	141	45%	8%	29%	54.63 (83)	No	227.7 \pm 84.6	8.6 \pm 1.9	16.8 \pm 4.1
Indore	102	2,167,447	212	50%	13%	17%	71.68 (38)	No	160.6 \pm 73.4	9.4 \pm 4.3	16.4 \pm 6.5
Jabalpur	104	1,267,564	122	46%	8%	34%		No	135.7 \pm 13.0		24.3 \pm 2.1
Ujjain	33	515,215	156	40%	6%	26%		No	78.4 \pm 42.0	10.9 \pm 3.4	11.9 \pm 3.1
Shillong	46	354,325	77	9%	16%	42%		No	78.8 \pm 31.0	19.4 \pm 19.0	12.5 \pm 5.4
Amritsar	90	1,183,705	132	50%	15%	21%		No	188.7 \pm 24.2	14.8 \pm 2.2	35.1 \pm 3.1
Chandigarh	115	1,025,682	89	47%	26%	27%		No	79.9 \pm 32.6	5.8 \pm 0.5	15.4 \pm 7.8
Ludhiana	167	1,613,878	97	50%	19%	19%	81.66 (10)	No	251.2 \pm 21.9	8.4 \pm 2.3	36.2 \pm 7.0
Chennai	426	8,917,749	210	47%	13%	17%		Yes	121.5 \pm 45.5	12.1 \pm 3.5	20.8 \pm 7.0
Agra	129	1,746,467	135	48%	12%	27%	76.48 (19)	No	184.1 \pm 95.9	6.6 \pm 3.5	20.8 \pm 12.1
Allahabad	71	1,216,719	171	54%	11%	26%		No	165.3 \pm 70.7	3.6 \pm 1.0	23.7 \pm 15.9
Firozabad	21	603,797	288	25%	4%	40%	60.51 (75)	No	195.6 \pm 78.2	21.6 \pm 4.8	32.1 \pm 4.9
Kanpur	150	2,920,067	195	11%	3%	42%	78.09 (15)	No	211.5 \pm 25.3	7.5 \pm 1.2	31.3 \pm 4.9
Lucknow	240	2,901,474	121	52%	15%	20%		No	200.4 \pm 28.4	8.4 \pm 1.0	36.1 \pm 2.6
Varanasi	102	1,435,113	141	40%	7%	29%	73.79 (29)	No	125.3 \pm 8.4	17.2 \pm 0.7	19.6 \pm 0.7
Asansol	49	1,243,008	254	27%	4%	61%	70.2 (42)	No	162.7 \pm 98.7	9.4 \pm 3.1	61.8 \pm 18.5
Durgapur	56	581,409	104	27%	4%	61%	68.26 (52)	No	172.5 \pm 107.1	9.8 \pm 3.2	63.9 \pm 18.6
Kolkata	727	14,112,536	194	12%	9%	34%		No	160.8 \pm 109.3	17.3 \pm 15.4	59.7 \pm 27.8

Notes:

AR = build-up area (in km^2) is estimated from Google Earth maps.; A = population density (per hectare); B = % households with a motorized two wheelers; C = % households with a four wheeler; D = % households with a non-gas cookstove; E = CEPI rating (rank); F = is the city coastal;

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4

5 **Table 2: Estimated premature mortality due to outdoor air pollution in India**

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City/Region	Study year	Pollutant	Premature mortality	Reference
All India	1990	PM ₁₀	438,000	IHME (2013)
Delhi	1990	Total PM	5070	Cropper et al. (1997)
Mumbai	1991	PM ₁₀	2800	Shah and Nagpal (1997)
Delhi	1993	PM ₁₀	3800-6200	Kandlikar and Ramachandran (2000)
Mumbai	1993	PM ₁₀	5000-8000	Kandlikar and Ramachandran (2000)
Delhi	2001	PM ₁₀	5000	Nema and Goyal (2010)
Kolkata	2001	PM ₁₀	4300	Nema and Goyal (2010)
Mumbai	2001	PM ₁₀	2000	Nema and Goyal (2010)
Chennai	2001	PM ₁₀	1300	Nema and Goyal (2010)
Ahmedabad	2001	PM ₁₀	4300	Nema and Goyal (2010)
Kanpur	2001	PM ₁₀	3200	Nema and Goyal (2010)
Surat	2001	PM ₁₀	1900	Nema and Goyal (2010)
Pune	2001	PM ₁₀	1400	Nema and Goyal (2010)
Bhopal	2001	PM ₁₀	1800	Nema and Goyal (2010)
Pune	2010	PM ₁₀	3600	Guttikunda and Jawahar (2012)
Chennai	2010	PM ₁₀	3950	Guttikunda and Jawahar (2012)
Indore	2010	PM ₁₀	1800	Guttikunda and Jawahar (2012)
Ahmedabad	2010	PM ₁₀	4950	Guttikunda and Jawahar (2012)
Surat	2010	PM ₁₀	1250	Guttikunda and Jawahar (2012)
Rajkot	2010	PM ₁₀	300	Guttikunda and Jawahar (2012)
All India	2010	PM _{2.5} + ozone	695,000	IHME (2013)
Delhi	2010	PM _{2.5}	7350 to 16,200	Guttikunda and Goel (2013)
Delhi	2030	PM _{2.5}	22,000	Dholakia et al. (2013)

7 **Table 3: Number of receptor modeling studies conducted between 2000 and 2013 in India**

8

9

City	A	B	C	A&B	D	Total
Delhi	4	1		1	5	11
Mumbai	3	1		1	2	7
Kolkata	2	1			1	4
Chennai	1	1		2		4
Hyderabad		1		1		2
Agra				1	1	2
Kanpur			1			1
Ahmedabad				1		1
Chandigarh					1	1
Tirupati	1					1
Talcher	1					1
Dhanbad	1					1
Jorhat		1				1
Virudhanagar				1		1
Mithapur					1	1
Bhubaneshwar					1	1
Multi-city		1		1		2
Raipur	1				5	1

Notes:
A = PM₁₀; B = PM_{2.5}; C = PM₁; D = mixed size fractions;

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11 **Table 4: Chronology of Bharat emission standards**

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Standard	Date	Region
India 2000	2000	Nationwide
Bharat-2 (Ref: Euro-2)	2001	NCR, Mumbai, Kolkata, and Chennai
	2003.04	NCR + 13 cities
	2005.04	Nationwide
Bharat-3 (Ref: Euro-3)	2005.04	NCR + 13 cities
	2010.04	Nationwide
Bharat-4 (Ref: Euro-4)	2010.04	NCR + 13 cities
	2012.03	NCR+ 13 cities + 7 cities
	2015	50+ cities
* NCR is the national capital region of Delhi, including Delhi and its satellite cities		
** 13 cities are Mumbai, Kolkata, Chennai, Bengaluru, Hyderabad, Ahmedabad, Pune, Surat, Kanpur, Lucknow, Sholapur, Jamshedpur and Agra		
*** 7 cities are Puducherry, Mathura, Vapi, Jamnagar, Ankaleshwar, Hissar and Bharatpur		

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14 **Table 5: Summary of emission standards for coal-fired power plants**

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Country ^a	PM	SO ₂	NO ₂	Mercury
India	350mg/Nm ³ for <210MW 150mg/Nm ³ for >210MW	None	None	None
China	30mg/Nm ³ (proposed all) 20mg/Nm ³ for key regions	100mg/Nm ³ for new 200mg/Nm ³ for old 50mg/Nm ³ for key regions	100mg/Nm ³	None
Australia	100mg/Nm ³ for 1997-2005 50mg/Nm ³ after 2005	None	800mg/Nm ³ for 1997-2005 500mg/Nm ³ after 2005	In discussion based on USA standards
European Union	Pre-2003 100mg/Nm ³ for <500MW 50mg/Nm ³ for >500MW Post 2003 50mg/Nm ³ for <100MW 30mg/Nm ³ for >100MW	Pre-2003 Scaled for <500MW 400mg/Nm ³ for >500MW Post 2003 850mg/Nm ³ for <100MW 200mg/Nm ³ for >100MW	Pre-2003 600mg/Nm ³ for <500MW 500mg/Nm ³ for >500MW Post 2003 400mg/Nm ³ for <100MW 200mg/Nm ³ for >100MW	In discussion
USA ^b	6.4 gm/GJ	640 gm/MWh	720 gm/MWh for old 450 gm/MWh for new	0.08 gm/MWh for lignite 0.01 gm/MWh for IGCC
a – Source: Guttikunda and Jawahar (2014); b – in official units; mercury as 12 month rolling average				

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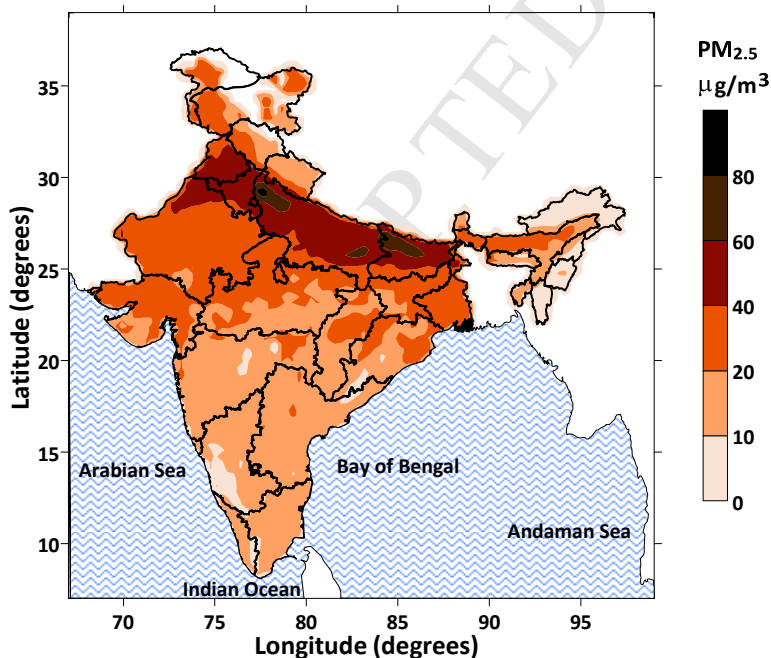
19 **Table 6: Comparison of technical and operational benefits and constraints of current**
 20 **and alternative brick manufacturing technologies**
 21

Technology	Fuel consumed per 100,000 bricks	Investment and operational costs (million USD) ^f	Brick production capacity (million/kiln)	Number of kilns required to produce 3.5 billion bricks	Average tons of CO ₂ produced per 100,000 bricks	Average reduction in PM emissions compared to FCK
FCK	20-22 tons coal	1.7	4.0	1000	50	
Zigzag ^b	16-20 tons coal	1.6	4.0	1000	40	40%
Hoffmann ^c	15,000-17,000 m ³ NG	5.7	15.0	270	30	90%
Hoffmann ^d	12-14 tons coal	5.7	15.0	270	30	60%
VSBK ^e	10-12 tons coal	1.6	5.0	800	25	60%

a. FCK = fixed chimney bull trench kiln; NG = natural gas; VSBK = vertical shaft brick kiln
 b. Some zigzag pilot kilns are in operation, listed as medium performance. Any improvement in the efficiency of operations can lead to further reductions in coal consumption
 c. Manufacturing period for Hoffmann kilns is round the year, compared to the current non-monsoonal month operations for the other kilns; thus increasing the land and raw material requirements; Link to natural gas grid and continuous fuel supply is a major constraint
 d. Initial investments are higher for Hoffmann kilns
 e. Operational models are available in India and Kathmandu
 f. Costs include initial investment, land, building, operational, and taxes estimates (World-Bank, 2010)

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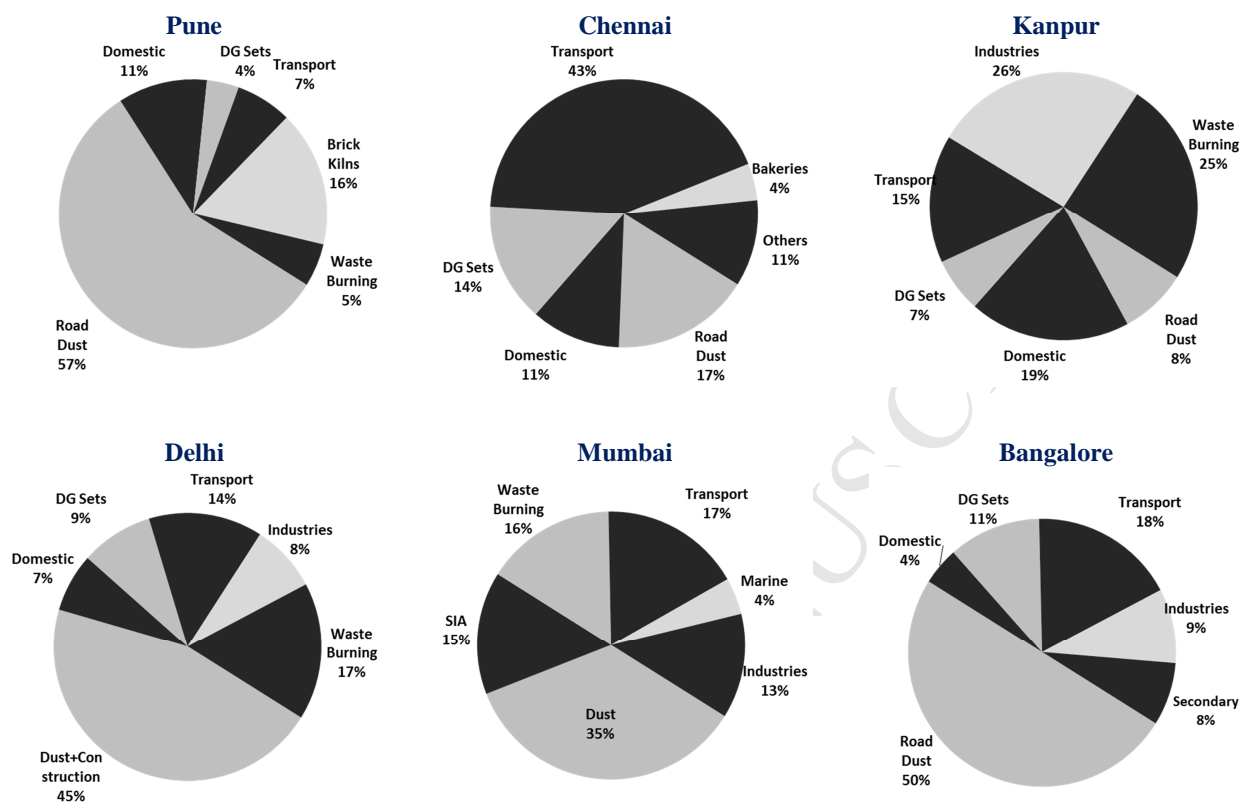
Figure 1: Ambient PM_{2.5} concentrations derived from the satellite observations



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29 **Figure 2: Average percent contributions of major sources to PM₁₀ pollution (CPCB, 2010)**

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Nature of Air Pollution, Emission Sources, and Management in the Indian Cities

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HIGHLIGHTS

- Air quality monitoring in the Indian cities
- Sources of air pollution in the Indian cities
- Health impacts of outdoor air pollution in India
- Review of air quality management options at the national and urban scale

Nature of Air Pollution, Emission Sources, and Management in the Indian Cities

SUPPLEMENTARY MATERIAL

Table 1: Summary of national ambient standards and the WHO guidelines for air quality

Pollutant	Averaging Time	India	China	EU	USA	USA **	WHO
PM ₁₀	Annual	60	40	40	-	-	20
	24-hour	100	50	50	150	150	50
PM _{2.5}	Annual	40	15	25	15	15	10
	24-hour	60	35	-	35	35	25
SO ₂	Annual	50	20	-	-	-	-
	24-hour	80	50	125	-	-	20
	1-hour	-	150	350	75 (ppb)	196	-
NO ₂	Annual	40	40	40	53 (ppb)	107	40
	24-hour	80	80	-	-	-	-
	1-hour	-	200	200	100 (ppb)	203	200
CO	24-hour	-	4000	-	-	-	-
	8-hour	2000	-	10000	9 (ppb)	11,100	-
	1-hour	4000	10000	-	35 (ppb)	43,170	30,000
Ozone	8-hour	100	100	120	0.075 (ppm)	159	100
	1-hour	180	160	-	-	-	-

All the values are in $\mu\text{g}/\text{m}^3$, except for those with units listed in brackets
 ** Numbers converted to $\mu\text{g}/\text{m}^3$ for comparison

12 **Table 2: An inventory of emissions and dispersion modeling studies in India**

13

City/Region	Study year	Emissions							Dispersion	Reference
		PM	SO ₂	NO _x	CO	VOC	BC/OC	GHGs		
India	1985-2005	*	*	*	*			*		Garg et al (2006)
Delhi	1987-88	*	*	*					*	Singh et al (1990)
Delhi	1990-2000	*	*	*				*	*	Mohan et al (2007)
Delhi	1990-2000	*	*	*	*	*		*		Gurjar et al (2004)
Delhi, Calcutta	1990-2015	*		*		*	*			Sharma et al (2002)
Delhi	1990	*	*	*	*	*		*		Bose and Anandalingam (1996)
India	1990, 1995		*	*						Garg et al (2001)
Mumbai	1992-93	*	*	*					*	Shah and Nagpal (1997)
Jamshedpur	1993		*	*					*	Bhanarkar et al (2005)
India	1996-97	*	*							Reddy and Venkataraman (2002)
India	2000-2008	*	*	*	*	*	*	*		Kurokawa et al (2013)
Pune	2003	*	*	*	*	*				Barth et al (2007)
Delhi, Mumbai, Chennai, Kolkata	2003-04					*				Srivastava and Majumdar (2010)
India	2003-04	*	*	*	*	*		*		Ramachandra and Shwetmala (2009)
Nagpur	2004		*					*		Majumdar and Gajghate (2011)
Delhi	2004	*								Mohan et al (2011)
Kanpur	2004		*				*			Mehta et al (2009)
Pune	2005	*							*	Kesarkar et al (2007)
Kanpur	2006	*							*	Behera et al (2011)
Hyderabad	2006-07	*	*	*					*	IES (2008)
Delhi	2010	*								Sahu et al (2011)
Delhi	2010	*	*	*	*	*			*	Guttikunda and Calori (2013)
Pune, Chennai, Ahmedabad, Indore, Surat, Rajkot	2010	*	*	*	*	*			*	Guttikunda and Jawahar (2012)

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15

16 When developing emissions inventories, emission factor is a very important input, in
 17 addition to data on consumption patterns and spatial information of sources. An emission factor
 18 can be defined as the mass of pollutant emitted up on combustion of various fuels in various
 19 sectors, under varying technical and operational conditions. It is assumed that a series of tests are
 20 conducted in the cities for as many sectors as possible and a database of emission factors is
 21 available for use. However, this is also a very expensive exercise and a very limited number of
 22 studies are conducted in India to ascertain these values. **CPCB (2010)** provides one such
 23 database for transport, industrial, domestic, and electricity generation, based on the tests
 24 conducted during the project period (2006-08). Other useful regional and global databases
 25 include **DIESEL (2008)**, **GAINS (2012)**, **SEI (2012)**, and **EEA (2013)**.

26
 27 The best available regional and global emissions inventories commonly utilized for the
 28 bottom-up source modeling are REAS (Regional Emission inventory in ASia); GAINS
 29 (Greenhouse Gas and Air Pollution Interactions and Synergies); EDGAR (Emission Database for
 30 Global Atmospheric Research) and GEIA (Global Emissions Initiative).

31
 32 The atmospheric dispersion models utilize physics from the meteorology and chemistry
 33 from the emissions inventories to predict concentrations of various air pollutants. Using such
 34 models, the contributions of different sources to PM, SO₂, NO_x, CO, and ozone pollution can be
 35 estimated. A wide range of dispersion models has been published in scientific papers and even a
 36 larger number of unpublished models and special model versions exist depending on end-use
 37 such as regulatory purposes, policy support, public information, and scientific research. A
 38 subjective assessment of the level of requirements, ease of operations, and possible detail of
 39 results for various model types is presented in **Table 3**. Most of these models have established
 40 forums where users can connect with others and discuss operational and data analysis queries.

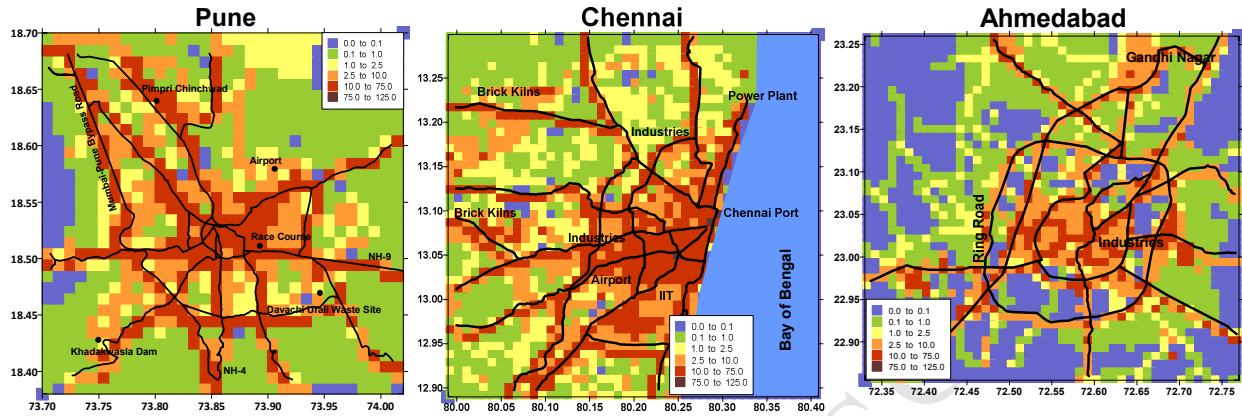
41
 42 **Table 3: What to expect from different types of atmospheric dispersion models?**

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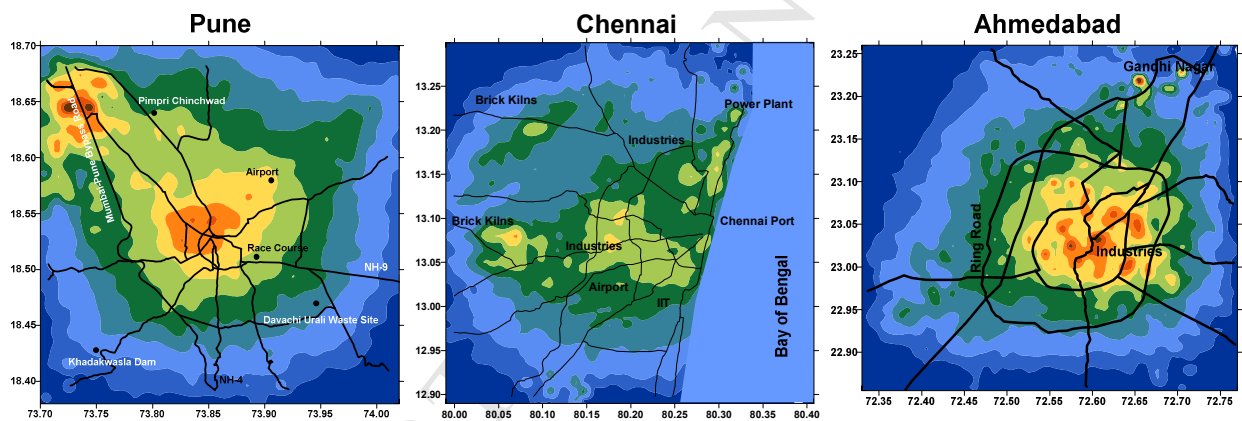
	Box	Gaussian/Plume	Eulerian
Example		ISC3 ATMoS AERMOD	WRF-Chem CMAQ CAMx
Model formulation and complexity	*	**	****
Data requirements	*	**	****
Advection of pollutants	-	**	****
Pollution chemistry	***	*	****
Ease of operations	****	**	*
Computational requirements	*	**	****
Computational times	*	**	****
Results – level of detail	*	**	****
Source apportionment	****	****	**
* = input requirements/quality of results is less			
**** = input requirements/quality of results is the highest			

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47 **Figure 1: Estimated on-road vehicle exhaust PM₁₀ emissions (Units: tons/km²) in 2010**
 48 **(Guttikunda and Jawahar, 2012)**
 49



50
 51 **Figure 2: Modeled PM₁₀ concentrations (Units: µg/m³) in 2010**
 52 **(Guttikunda and Jawahar, 2012)**
 53



54
 55

56 **EMISSION INVENTORIES and DISPERSION MODELS**

57 Website last accessed on April 25th, 2014

58

59 **REAS**

60 @ <http://www.jamstec.go.jp/frsgc/research/d4/emission.htm>

61

62 **GAINS**

63 @ <http://gains.iiasa.ac.at/models/>

64

65 **EDGAR**

66 @ <http://edgar.jrc.ec.europa.eu/>

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68 **GEIA**

69 @ <http://www.geiacenter.org/>

70

71 **ISC3**

72 @ http://www.epa.gov/ttn/scram/dispersion_alt.htm#isc3

73

74 **AERMOD**

75 @ http://www.epa.gov/ttn/scram/dispersion_prefrec.htm#aermod

76

77 **ATMoS**

78 @ <http://www.urbanemissions.info/model-tools/atmos-40.html>

79

80 **WRF-Chem**

81 @ <http://www.acd.ucar.edu/wrf-chem/>

82

83 **CMAQ**

84 @ <http://cmascener.org/cmaq/>

85

86 **CAMx**

87 @ <http://www.camx.com/home.aspx>

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