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Making the best of both worlds: Can high-resolution agricultural administrative data support the assessment of High Nature Value farmlands across Europe?

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1 **Title:** Making the best of both worlds: can high-resolution agricultural administrative
2 data support the assessment of High Nature Value farmlands across Europe?

3

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23 **Abstract**

24 Worldwide, the role of farmlands for biodiversity conservation and the delivery of
25 multiple ecosystem services has been widely acknowledged. In the European Union
26 (EU), societal demands to include environmental conservation concerns within the
27 Common Agricultural Policy (CAP) has resulted in the recognition of the importance of
28 maintaining High Nature Value farmlands (HNVf).

29 HNVf constitute complex social-ecological systems, which owe their nature
30 conservation value to the maintenance of specific, mostly low-intensity farming
31 systems, supporting high levels of species and habitats dependent on agricultural
32 practices. Even though HNVf assessment in space and time is essential to evaluate the
33 effectiveness of Rural Development Programmes, the diversity of rural landscapes
34 across EU, the scarcity of data on farming systems, and the lack of common
35 methodological guidelines has hampered the implementation of HNVf mapping and
36 monitoring across Europe. Thus, there is a pressing need to develop and test
37 methodological approaches that may support HNVf assessment across the EU.

38 The Integrated Administration and Control System (IACS) which is mandatory for all
39 EU Member States constitutes a system for the management and control of CAP
40 payments to farmers. Essentially, IACS comprises high-resolution, spatially explicit
41 information on the type and intensity of agricultural land-use. Even though such data
42 has been referred as exhibiting high thematic, spatial and temporal resolution, IACS has
43 seldom been used, due to significant access restrictions. Here, the potential to use IACS
44 data to support the assessment of HNVf was evaluated within the German Federal State
45 of Lower Saxony by implementing a recently developed methodological framework.
46 Sets of indicators known to be essential for identifying potential HNVf and underlying
47 farming systems (expressing landscape structure and composition, farming systems, and

48 crop diversity), were derived from IACS. Spatial patterns of indicators were analyzed at
49 two different scales to delineate the potential distribution of HNVf across Lower
50 Saxony.

51 Results highlighted that most regions in Lower Saxony were characterized by intensive
52 farming practices including high livestock density, high share of intensive crops and
53 low density of linear elements. Only 3% of the Utilized Agricultural Area (UAA) of
54 Lower Saxony potentially constituted HNVf, with the majority of HNVf coinciding
55 with mosaics of arable and/or permanent crops and semi-natural features under less
56 intensive farming practices. Semi-natural grasslands, partially under agri-environment
57 scheme management contracts, covered roughly 1% of the UAA and were mostly
58 intermingled with other farmland habitats in extensively managed agricultural
59 landscapes.

60 In the context of the EU-wide HNVf assessment, IACS constitutes an important source
61 of data, characterized by a high spatial, thematic and temporal resolution of data
62 collected annually. Whilst having the potential for use in HNVf assessment, some
63 challenges remain, especially due to significant access restrictions. Nevertheless, IACS
64 constitutes a powerful tool to evaluate the extent and condition of HNVf across the EU
65 countryside. Making use of IACS data in such a way could provide a stepping-stone
66 towards achieving a more effective balance between the management and control of
67 CAP support payments and the growing societal demands related to the maintenance
68 and enhancement of farmland biodiversity and ecosystem services.

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72 **Keywords (4-6)**

73 Agro-biodiversity; Common Agricultural Policy (CAP); Indicators; Integrated

74 Administration and Control System (IACS); Land-sharing; Rural Development

75 Programmes (RDP)

76

77 **1. Introduction**

78 Globally, an expansion and intensification of agricultural land has occurred in the last
79 century (Wade *et al.*, 2008), with negative impacts on the environment and related
80 natural resources, such as biodiversity and ecosystem services (Millennium Ecosystem
81 Assessment, 2005; Aviron *et al.*, 2009; Shackelford *et al.*, 2015). Driven mainly by
82 economic, political and demographic processes, agricultural land in Europe has been
83 facing two opposite trajectories: either abandonment of economically marginal remote
84 and upland areas or the intensification of farming practices in the more productive
85 lowland areas (MacDonald *et al.*, 2000; Stoate *et al.*, 2009; Baudron and Giller, 2014;
86 Beilin *et al.*, 2014; van Vliet *et al.*, 2015).

87 While high yield farming is considered to be among the most damaging human-related
88 activities to wildlife (Balmford *et al.*, 2012; Shackelford *et al.*, 2015), the importance of
89 agricultural land for biodiversity maintenance and long-term conservation, and the
90 provision of ecosystem services, e.g., carbon sequestration, aesthetic landscapes, and
91 support of biodiversity, has also been acknowledged (Swinton *et al.*, 2007; Power,
92 2010). Farming has been shaping European landscapes for centuries or even millennia
93 and up to 50% of all species rely, to some extent, on agricultural ecosystems and
94 habitats, including endemic and threatened species (Bignal and McCracken, 1996;
95 Halada *et al.*, 2011; Lomba *et al.*, 2014). The role of traditional, low-intensity farmland
96 for the maintenance of natural capital and protection of the countryside has thus been
97 debated, ultimately developing into the ‘High Nature Value farmlands’ (HNVf) concept
98 (Egan and Mortensen, 2012; Plieninger and Bieling, 2013; Renwick *et al.*, 2013; Lomba
99 *et al.*, 2014; Lomba *et al.*, 2015; Strohbach *et al.*, 2015).

100 The HNVf concept was developed within the European Union (EU) to characterize
101 agriculture-dominated landscapes where high nature and/ or conservation value is

102 dependent on the continuation of specific low-intensity farming systems (Beaufoy *et al.*,
103 1994; Andersen *et al.*, 2003; Lomba *et al.*, 2014). These farming systems constitute
104 complex socio-ecological systems resulting from a long-term relationship between
105 human activity and the surrounding environment (Plieninger and Bieling, 2012;
106 Plieninger and Bieling, 2013). The intrinsic nature value of HNV farmlands is due to
107 the prevalence of low-intensity farming practices and either a high proportion of semi-
108 natural vegetation e.g. pastures and meadows (referred to as HNVf type 1; Oppermann
109 *et al.*, 2012), or the presence of small-scale elements in the agricultural landscapes, such
110 as field margins, hedgerows and tree lines (referred to a HNVf type 2; Andersen *et al.*,
111 2003). In addition, some intensively managed farmlands have been considered as HNVf
112 type 3 due to their importance for the maintenance and survival of some populations of
113 agriculture-dependent species with conservation interest (e.g. farmland birds and
114 reptiles; Andersen *et al.*, 2003).

115 In recognition of EU efforts towards sustainable rural development and land
116 stewardship (Plieninger and Bieling, 2013), HNVf was included in the Common
117 Monitoring and Evaluation Framework (CMEF) for the rural development policy within
118 the context of the EU Common Agricultural Policy (CAP; EC, 2006). Their role for
119 biodiversity conservation, provision of ecosystem services and public goods generated
120 has also been highlighted within the EU Biodiversity strategy to 2020 (EEA, 2015).
121 Nevertheless, the recent mid-term assessment of the EU Biodiversity strategy to 2020
122 reported that no relevant progress has been made towards the improvement of the
123 conservation status of most agriculture-dependent species and habitats. The assessment
124 recommended that greater and more effective efforts are urgently needed to increase the
125 contribution that farmlands, including HNVf, make to the maintenance and

126 enhancement of biodiversity in the European Union (EU) countryside (EC, 2015; EEA,
127 2015).

128 However, even though the assessment of HNVf indicators is mandatory across the EU,
129 the diversity of rural landscapes, the scarcity of (suitable) datasets on biodiversity, land
130 cover and land-use, and the lack of common guidelines and/or approaches for mapping
131 HNVf are important obstacles towards its successful implementation (see e.g. Peppiette,
132 2011; Lomba *et al.*, 2014; Strohbach *et al.*, 2015).

133 Following the EU guidelines (Paracchini *et al.*, 2008; EENRD, 2009), Lomba *et al.*
134 (2015) recently described a multi-step spatially-explicit framework to assess the extent
135 of HNV farmlands in the EU countryside. In short, such approach builds on the
136 spatially-explicit assemblage of indicators informing on the social-ecological
137 dimensions underlying the nature value of such farming systems, namely landscape
138 structure and composition, and the extensive character of farming practices (Lomba *et*
139 *al.*, 2014; Lomba *et al.*, 2015). In addition, it has been recommended that indicators
140 considered for HNVf assessment should be derived from the best spatial and/or
141 temporal resolution available for the target area (Lomba *et al.*, 2015). Whilst this
142 methodological framework has been shown to have a great potential to operationalize
143 the HNVf concept, some important challenges remain, such as its application to other
144 social-ecological contexts, datasets and scales across the EU countryside. Here, one of
145 such challenges is tackled through the implementation framework using indicators
146 derived from the Integrated Administration and Control System (IACS) database.

147 IACS, which was established in the early 1990s (EEC, 1992), mainly consists of high
148 resolution, annually-updated farm-level information (e.g. livestock) and parcel-level
149 information (e.g. crop type; Keenleyside *et al.*, 2014). Despite its availability across EU
150 Member States, IACS data has seldom been used as a spatially-explicit dataset for

151 indication and monitoring of HNVf, mainly due to access restrictions put in place to
152 protect land manager privacy (Lomba *et al.*, 2014; Strohbach *et al.*, 2015). To-date,
153 Steinmann and Dobers (2013) have analyzed patterns of crop rotation and sequence
154 across the federal state of Lower Saxony based on the German IACS. The same
155 database has been used by Nitsch *et al.* (2012) to assess land-use change between
156 grasslands and arable land. In addition, Ribeiro *et al.* (2014) used IACS data to model
157 HNV farming systems dynamics as response to policy change for a region in southern
158 Portugal, the Austrian HNVf indicator relies on IACS data to incorporate information
159 on land-use and agri-environment schemes, and the Scottish Government uses IACS
160 data to estimate annual changes in extent and distribution of HNVf (AES; Bartel *et al.*,
161 2011; Scottish Government, 2011).

162 Here, we tested the potential of IACS data to support the assessment of HNVf using the
163 German Federal State of Lower Saxony as case study. HNV farmlands assessment
164 followed the multi-criteria framework recently described by Lomba *et al.* (2015),
165 targeting HNVf types 1 and 2. Overall, HNVf assessment was built on a multi-criteria
166 analysis of spatially-explicit indicators expressing landscape structure and composition,
167 farming systems, and crop diversity (Landscape Elements, Extensive Practices and Crop
168 Diversity sets of indicators; Lomba *et al.*, 2015), known to inform on relevant social-
169 ecological components of HNV farming systems. Results are discussed in the context of
170 Lower Saxony assessment and monitoring of HNVf and implications drawn with
171 respect to High Nature Value farmlands assessment and monitoring across the EU.

172

173 **2. Methods**

174 2.1. Study area

175 The study area is situated in north-western Germany and covers the federal state of
176 Lower Saxony (Fig. 1). Lower Saxony is Germany's second-largest federal state in
177 terms of its area (ca. 47787 km²) and includes 1041 local administrative units (LAU 2)
178 corresponding to municipalities, the lowest level in the administrative structure of
179 Germany. Total land area covered by each municipality ranges between 0.2 km² and
180 401.7 km² (mean: 45.9 km²).

181 About 56% of total land area within Lower Saxony is used for agricultural production,
182 consisting of arable land (including temporary grassland), permanent grassland and land
183 under permanent crops, and ca. 22% is covered by forests. Less than 10% of the total
184 land area is covered by urban areas. Arable land accounts for ca. 71% of the Utilized
185 Agricultural Area (UAA), permanent grasslands make up 28% of the UAA and less
186 than 1% is used for the production of permanent crops. There are, however, large
187 regional differences in agricultural land-use and major crop type distributions across
188 Lower Saxony.

189 Lower Saxony is located in a transition zone between a more maritime climate in the
190 North-West and a more continental climate in the South and East. According to
191 climatic, geomorphological, hydrological and soil characteristics Lower Saxony can
192 roughly be divided into three major biogeographical regions. The coastal regions (Fig.
193 1C; 1) along the North Sea include the East Frisian Islands, mudflats and salt marshes
194 with an average annual temperature of 9.4 °C and a mean precipitation of 814 mm in the
195 period of 1981-2010 (DWD 2015). The lowlands (Fig. 1C; 2) are dominated by
196 agricultural land-use with an average annual temperature of 9.4 °C and a mean

197 precipitation of 769 mm. In the northwestern part of the lowlands more than half of the
198 UAA is covered by highly productive permanent grassland (Smit *et al.*, 2008), with
199 intensive dairy farming and zero-grazing ('stall feeding') dominating (Klimek *et al.*,
200 2014). The eastern part of the lowlands is characterized by cultivation of cereals and
201 fragments of heathland. Most of these lowland heathland areas are designated as
202 European special areas of conservation (SAC; e.g. 'Lüneburger Heide') and are mainly
203 grazed by sheep with financial support from agri-environment schemes. Particularly in
204 the western part of the lowlands, production of energy crops has emerged as a new
205 agricultural activity in recent years, mainly at the expense of permanent grasslands. It
206 has been shown that maize for bioenergy production was the dominant crop after the
207 conversion of permanent grassland to arable land (Nitsch *et al.*, 2012). This region of
208 the lowlands is dominated by intensive livestock production, particularly pig and
209 poultry. The uplands (Fig. 1C; 3) in the southern part of Lower Saxony are
210 characterized by a large proportion of arable land and forest, interposed with patchily
211 distributed fragments of permanent grassland (Klimek *et al.*, 2007). The average
212 precipitation is 832 mm, with a mean annual temperature of 9.0 °C. In the south, the
213 Harz is the highest mountain range (up to 1141 m a.s.l.). The northern part of the
214 uplands is characterized by fertile loess soils and highly-productive large-scale cereal-
215 based farming systems.

216

217 #Fig. 1 approximately here#

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219

220

221 2.2 Integrated Administration and Control System (IACS)

222 IACS consists of a number of interconnected databases, which provide an identification
223 system for farmers and their payment entitlements, an identification and registration
224 system for livestock, and an identification system for agricultural areas called the Land
225 Parcel Identification System (LPIS) (EC, 1996; Sagris and Devos, 2009). LPIS is
226 implemented differently across and even within EU member states (see Sagris and
227 Devos, 2009 for details), and in the case of Lower Saxony, the so-called physical block
228 representation, was adopted. According to this approach, agricultural parcels sharing a
229 common boundary e.g. a ditch, footpath or forest edge, are grouped as a physical block.
230 Agricultural parcels within each physical block may be property of one or more farmers,
231 and thus be under different distinct land-use.

232 In the specific case of Lower Saxony, LPIS is available as polygons and was thus
233 provided as an ESRI shapefile. Through a unique ID attributed to each physical block, it
234 is possible to map IACS data across Lower Saxony. IACS includes two main levels of
235 information: 1) type of land-use and size of agricultural parcels for which financial
236 support was claimed; 2) farm-level information including the number and type of
237 animals and farm type (organic/ conventional, fulltime /second income). Through the
238 unique physical block ID, agricultural parcels can be linked to a single physical block
239 and to a farm, even though the exact location of each agricultural parcel is not disclosed
240 in this dataset to ensure land managers/owners' privacy and data confidentiality.

241

242

243

244 2.3. Assessment of HNV farmlands using IACS data: spatially-explicit indicators and
245 statistical analysis

246 HNVf assessment followed the methodological framework described by Lomba *et al.*
247 (2015), and targeted specifically HNVf1, i.e. farmlands with high proportion of semi-
248 natural habitats; and, HNVf2, i.e. landscape mosaics where small crop fields are
249 intermingled with small-scale features. As HNVf3 nature value derives from the
250 occurrence of individual species with high conservation interest (e.g. farmland bird
251 species), often in intensively managed agricultural landscapes, such farmlands were not
252 targeted in the context of this landscape-level research (Andersen *et al.*, 2003; Lomba *et*
253 *al.*, 2014).

254 A multi-criteria approach was implemented using the three sets of spatially-explicit
255 indicators defined by (Lomba *et al.* 2014, 2015): i) landscape elements, ii) extensive
256 practices; and, iii) crop diversity. Whilst landscape elements set of indicators depict
257 landscape structure and composition; indicators included within the extensive practices
258 set reflect the intensity of farming practices; and, indicators on crop diversity, aim to
259 inform the diversity of farming practices (Table 1; for detailed information regarding
260 the approach implemented to calculate spatially-explicit indicators from IACS and LPIS
261 data). All indicators related to farming systems were ascertained from IACS data for the
262 year 2005 (Table 1). As IACS is restricted to agricultural land for which payments have
263 been claimed, ancillary indicators derived from the German digital basic landscape
264 model (Basis-DLM) were considered to enhance the accuracy of HNVf assessment
265 whenever required (cf. Table 1).

266 Spatially-explicit patterns of targeted indicators were analyzed at two different scales: i)
267 municipality (coincident with the local administrative unit, LAU 2;
268 <http://epp.eurostat.ec.europa.eu>); and, ii) physical block. Analysis followed a sequential

269 three step framework described below, which enabled the integration of high spatial
270 resolution IACS data, while pursuing a landscape-level outcome (Fig. 2).

271 #Fig. 2 approximately here#

272

273 According to EU guidelines, the area of agricultural land covered by HNVf (i.e. HNVf
274 baseline indicator, EENRD, 2009) should be expressed in relation to the utilized
275 agricultural area (UAA). The Land Parcel Identification System (LPIS) provides highly
276 disaggregated information on most agricultural landscapes (Nitsch *et al.*, 2012) and was
277 used to comply with this EU recommendation. UAA was therefore determined as the
278 sum of the area covered by physical blocks at the municipality level (coincident with the
279 local administrative unit, LAU 2).

280 From the concept, a landscape-level approach seems to be appropriate to assess HNV
281 farmlands (Beaufoy *et al.*, 1994; Andersen *et al.*, 2003; Lomba *et al.*, 2014).

282 Accordingly, Step 1 comprised the definition of farmland dominance at the landscape
283 level (Fig. 2). Farmland dominance was considered when two conditions were fulfilled:
284 i) a value of 40% agricultural area (P.UAA) per municipality (rule of thumb proposed
285 by Lomba *et al.*, 2015), and, ii) higher values for the share of agricultural cover
286 (P.UAA_m) in relation to shares of urban (P.Urban_m) and forest areas, respectively
287 (P.Forest_m; Table 1).

288 In Step 2, spatially-explicit patterns of indicators expressing the intensity of farming
289 practices, specifically the livestock density index (LSI_m) and the share of intensive
290 crops (P.ICrops_m), were analyzed (cf. Step 2, Fig. 2 and Table 1). Statistical results
291 highlighted the most significant gradients of intensification underlying distinct farming

292 systems and discriminated farmlands more likely to support high nature value in Lower
293 Saxony, which were then considered for further analysis in Step 3 (Fig. 2, Table 1).

294 Finally, Step 3 aimed at targeting High Nature Value farmlands of types 1 and 2 in areas
295 highlighted potential HNVf in Step 2. To accomplish this goal, analyses were
296 performed at the municipality (Step 3a) and at the physical block (Step 3b) level (cf.
297 Fig. 2, Step 3, and Table 1). Step 3a consisted on a detailed analysis of spatially-explicit
298 patterns for the three sets of indicators (extensive farming practices, landscape elements
299 and crop diversity sets of indicators; cf. Fig. 2, Step 3a, and Table 1). In Step 3b, a fine-
300 scale analysis was implemented, combining results from Step 3a with high-resolution
301 IACS data at the physical block level (Fig. 2, Step 3b). The fine-scale assessment of
302 HNVf1 (Step 3b) was built on the assumption that in Lower Saxony, HNVf1 mainly
303 consist of species-rich permanent grasslands (Oppermann *et al.*, 2012). Thus indicators
304 informing on the share of permanent grasslands and the share of grasslands under agri-
305 environment scheme were considered (cf. Table 1, Fig. 2). As for HNVf2 (step 3b), the
306 share of intensive crops and crop diversity at the level of the physical block were
307 considered to be suitable indicators (cf. Table 1 and Fig. 2).

308 All spatially-explicit analyses were conducted using the Spatial Statistics Toolbox for
309 ArcGIS 10.3.1 Desktop (ESRI, 1999-2015). Spatial Autocorrelation analysis (Global
310 Moran's I; ESRI, 1999-2015) was first applied to evaluate patterns (clustered, dispersed
311 or random) exhibited by each of the considered spatially-explicit indicators at the
312 municipality and physical block level (Table 1). For subsequent analyses only indicators
313 found to exhibit clustered patterns, expressed as statistically-significant positive
314 Moran's I index values, were considered. Spatially-explicit analysis of patterns for
315 targeted indicators, in each of the steps formerly described (Fig. 2), except Step 3b, was
316 performed using the Mapping Clusters toolset (ESRI, 1999-2015). To ensure that all

317 groups include members that have natural neighbours, the Grouping Analysis tool was
318 implemented with K-Nearest Neighbours as spatial constraints parameters. Outcomes
319 included overall and within resulting group statistics, the discrimination ability of each
320 indicator considered for analysis (expressed as higher R^2 values), and an evaluation of
321 the optimal number of groups. Optimal number of groups outcomes are expressed as
322 higher values for the pseudo F-statistic, and reflect a trade-off between the numbers of
323 groups and indicators used in the analysis (Calinski-Harabasz pseudo F-statistic,
324 hereafter F-statistic, assesses grouping effectiveness, and reflects within-group
325 similarity and between-groups differences (c.f. ESRI, 1999-2015).

326 Step 3b (cf. Fig. 2) consisted on a spatially-explicit selection of physical blocks based
327 on the analysis of variation of each indicator across targeted areas, whilst ensuring that
328 the physical blocks identified as HNVf1 and 2 exhibited connectivity at the landscape-
329 level. Thresholds were defined from spatially-explicit variation within each of the
330 indicators in order to identify semi-natural grasslands embedded within farmlands under
331 more extensive farming systems (HNVf1), or mosaics with high crop diversity and
332 small-scale landscape features (HNVf2). The extent of HNVf corresponding to types 1
333 and 2 was addressed individually, assuming that both can coexist, depending on the
334 characteristics of agricultural landscapes and underlying farming systems (Lomba *et al.*,
335 2014). Although analyses were performed at the physical block level, geometries are not
336 disclosed to ensure privacy of managers and land owners, and protection of IACS data
337 provided.

338 Indicators considered were tested for collinearity by Spearmans's rho index (ρ) using
339 STATISTICA software (Statsoft, 2013), and a value of 0.7 was established as a
340 maximum threshold for indicators to be included in the analysis (Dormann *et al.*, 2013).
341 All values shown for indicators are expressed as mean \pm standard deviation (SD).

342

343 #Table 1 approximately here#

344

345 **3. Results**

346 **3.1. Dominance of farmlands across Lower Saxony**

347 Lower Saxony has roughly ~2.7 million hectares of Utilized Agricultural Area,
348 distributed across 1041 municipalities. Overall, agriculture constitutes a major land-use
349 ($62.84 \pm 12.11\%$, $P.UAA_m$) followed by forest and urban areas, with median values of
350 cover of $16.83 \pm 12.14\%$ and $9.27 \pm 5.99\%$ respectively. Step 1 (for detailed description
351 see section 2.3.), highlighted the dominance of farmlands on ~85% of Lower Saxony
352 municipalities ($n = 887$). Table 2 shows the variation observed within spatially-explicit
353 indicators considered for farmland-dominated municipalities.

354

355 #Table 2 approximately here#

356

357 **3.2. High Nature Value farmlands across Lower Saxony**

358 In Step 2, spatially-explicit indicators reflecting the extensive character of farming
359 practices, LSI_m and $P.ICrops_m$, supported the discrimination of farmlands with high
360 potential to support high nature value from non-HNV farmland (Fig. 3, Table 2).
361 Grouping analysis resulted in the delineation of clusters A and B with 496 and 391
362 municipalities, respectively. Evaluation of the optimal number of groups was
363 implemented within the Grouping Analysis toolbox, and '2' highlighted as the most

364 effective number of groups in relation to the indicators considered (expressed as the
365 highest F-statistic value: 677.63; cf. methods section 2.4). Further analysis was based on
366 R^2 values for the targeted indicators, which were found to be higher for the Livestock
367 density index (R^2 : 0.66) than for the Share of Intensive Crops (R^2 : 0.19).

368 Variation of spatially-explicit indicators considered (Step 2) within resulting clusters
369 depicted divergent patterns (cf. Table 3). Cluster (A) was found to exhibit higher values
370 for the Livestock density index (LSI_m : 1.50 ± 0.48) and lower values for the share of
371 intensive crops ($P.ICrops_m$: 0.31 ± 0.13). Conversely, lower values for LSI_m and higher
372 values for the share of intensive were observed for cluster (B) (cf. Table 3). Similar
373 trends were observed for the share of values within each cluster (cf. Table 3).

374

375 #Table 3 and Fig. 3 approximately here#

376

377 The joint analysis of values of R^2 observed for LSI_m and the internal variation of within-
378 clusters indicators showed that the cluster (B) had greater potential to support farmlands
379 with High Nature Value. Accordingly, Step 3a of the analysis focused on cluster (B)
380 farmlands.

381 Step 3a was built upon grouping analysis of indicators expressing extensive farming
382 practices, landscape elements and crop diversity. Supported by F-statistic values (F-
383 statistic value: 120.57; cf. Table 4), three groups were considered to better discriminate
384 variation of indicators across the targeted area (cluster (B) from Step 2; cf. Table 3). As
385 presented in Table 4, R^2 values highlighted the share of intensive crops ($P.ICrops_m$; R^2 =
386 0.66) and crop diversity (expressed as SEI_m ; R^2 =0.47) as indicators contributing most to
387 within-clusters discrimination, followed by the Livestock density Index (LSI_m ; R^2 =

388 0.23) and the density of tree lines and hedgerows (DT_m ; $R^2=0.18$). Variation observed
389 for indicators revealed distinct patterns across clusters (a), (b) and (c), cf. Table 4.
390 Overall, cluster (a) showed the highest values for the share of intensive crops
391 ($P.ICrops_m$: 0.69 ± 0.15) and lowest values for crop diversity (SEI_m : 0.39 ± 0.057).
392 Conversely, cluster (b) was characterized by the lowest value for the share of intensive
393 crops ($P.ICrops_m$: 0.30 ± 0.11) while showing the highest values for crop diversity
394 (SEI_m : 0.50 ± 0.056) and density of linear elements (DT_m : 5.59 ± 2.71). As for the
395 livestock density index (LSI_m) average values lower than 0.50 LSU per ha/UAA were
396 observed within clusters. Considering both LSI_m variation and discrimination ability (cf.
397 R^2 value, Table 4), the share of intensive crops was considered most important
398 regarding the intensity of farming practices across clusters (a), (b) and (c). Thus,
399 differences observed across clusters supported their classification according to HNVf
400 types. Cluster (a) was considered to include farmlands under intensive farming practices
401 and therefore classified as non-HNVf. Due to lower shares of intensive crops and higher
402 crop diversity and density of linear elements, cluster (b) was considered most likely to
403 support both HNVf types 1 and 2. Finally, cluster (c) was classified as potentially
404 supporting HNVf type 2.

405

406 #Table 4 and Fig. 4 approximately here#

407

408 **3.3. Fine-scale assessment of High Nature Value farmlands in Lower Saxony**

409 The results of the fine-scale assessment of HNVf (Step 3b) are presented in Table 5.

410 Overall, farmlands considered suitable for a fine-scale assessment of HNVf (including

411 clusters (b) and (c); Step 3a, cf. Fig 4), totaled 429,471.16 ha of UAA (86,253 physical
412 blocks).

413

414 #Table 5 approximately here#

415

416 From the area identified as potentially supporting HNVf1 and 2 (cluster (b), Fig. 4),
417 5,024 physical blocks showed the highest values for both grasslands ($P.G_{pb} \geq 83.41\%$)
418 and grasslands under agri-environment schemes (cf. $P.GAES_{pb} \geq 9.93$; cf. Table 5), and
419 were thus considered likely to correspond to HNVf1. In addition, 9,884 physical blocks
420 showed the highest values for crop diversity (cf. Table 5, $SEI_{pb} \geq 0.58$) while having the
421 lowest shares of intensive crops ($P.ICrops_{pb} \leq 0.24$; cf. Table 5) HNVf2 potential area (cf.
422 cluster (c) Fig. 4) included 28,445 physical blocks, corresponding to 129,435.48 ha of
423 UAA. Analysis of variation within indicators at the physical block level (cf. Table 5)
424 resulted in a final HNVf2 area including ca. 19 % of total physical blocks (5542,
425 corresponding to 17,841.68 ha of UAA). Those physical blocks showed the highest
426 values of crop diversity, reflected as $SEI_{pb} \geq 0.49$, and the lowest share of intensive
427 crops $P.ICrops_{pb} < 0.88$ (cf. Table 5).

428 Spatially-explicit representation of HNVf1 and HNVf2 fine-scale assessment (Step 3b)
429 is presented in Fig. 5.

430

431 #Fig. 5 approximately here#

432

433 **4. Discussion**

434 4.1. IACS data and the assessment of HNVf in Lower Saxony

435 Overall, spatially explicit indicators derived from IACS allowed the assessment of
436 HNVf in Lower Saxony, following the methodological approach recently described by
437 (Lomba *et al.*, 2015). Indicators reflecting landscape structure and composition,
438 extensive farming practices, and crop diversity used within each step of the described
439 approach (cf. Table 1), were considered to be the best trade-off between IACS thematic,
440 spatial and temporal resolution, and available information regarding farming systems
441 and their variation across the study-area. As the definition of hard thresholds for
442 indicators for HNVf assessment is far from being consensual, our analysis was assumed
443 to be region (or administrative unit) specific (e.g. see Boyle *et al.*, 2015). Whilst such
444 assumption may hamper our ability to cope with HNVf assessment when continuous
445 regions are assessed individually, it may also be an added-value. In fact, such an
446 approach has higher potential to be implemented on broader scales, thus allowing the
447 validation of regional assessments, and the harmonization of criteria at the national
448 level. Moreover, in order to provide and target cost-effective support under the CAP,
449 detailed knowledge on the distribution and extent of HNVf is needed (Keenleyside *et*
450 *al.*, 2014). In this regard, our results may contribute to improved targeting, monitoring
451 and evaluating the impact of CAP support for HNVf.

452 The link between IACS and the Land Parcel Identification Systems (LPIS), which
453 covers most agricultural areas, allowed HNVf extent to be ascertained in relation to the
454 Utilized Agricultural Area, and thus comply with EU standards for reporting HNV
455 indicators (EENRD, 2009; Lomba *et al.*, 2015). Due to its high spatial and thematic
456 high-resolution (Nitsch *et al.*, 2012; Strohbach *et al.*, 2015), IACS data allowed the
457 assessment of HNVf at two different spatial scales, the municipality, coincident with the
458 local decision level (LAU 2), and the physical block, corresponding to a fine-scale level

459 (cf. Fig. 2). In addition, the approach adopted in this study also allowed discrimination
460 between HNV farmlands with a high proportion of agriculture dependent habitats
461 (HNVf 1; Lomba *et al.*, 2014) and mosaics of semi-natural and small-scale landscape
462 features (HNVf2; Lomba *et al.*, 2014).

463 In Lower Saxony, as across the majority of Germany, most agricultural landscapes are
464 currently under intensive farming practices (Nitsch *et al.*, 2012; Oppermann *et al.*,
465 2012; Klimek *et al.*, 2014), which makes the assessment of High Nature Value
466 farmlands and targeting of CAP support measures quite a challenge. Whilst extensive
467 farming practices are not widely adopted across Lower Saxony, some landscapes have
468 been described as farmlands with high nature value and thus under HNV farming
469 systems, particularly semi-natural grasslands (Oppermann *et al.*, 2012).

470 Indicators expressing the extensive character of farming practices, including the
471 livestock density index and the share of intensive crops, were analyzed to achieve
472 deeper insights on differences between farming systems across the study area. Spatial-
473 statistics analysis resulted in the delineation of two farmland-dominated regions,
474 coincident with western and eastern areas of Lower Saxony (clusters (A) and (B); cf.
475 Fig. 3). Importantly, the livestock density index had the best discriminating ability
476 across Lower Saxony (cf. Table 3). As a result, the Western part of Lower Saxony
477 (cluster (A), Fig. 3), where farmland-dominated municipalities were generally
478 intensively managed in terms of livestock density, was considered as non-HNV
479 farmland. This corresponds to recent studies that have demonstrated that some districts
480 in western Lower Saxony (e.g. 'Vechta' and 'Cloppenburg') are characterized by the
481 highest densities of livestock in Germany and Europe, respectively (Deblitz *et al.*, 2008;
482 Neumann *et al.*, 2009). Smit *et al.* (2008) further highlighted that estimated grassland
483 productivity (dt ha^{-1}) and milk production per ha UAA is comparatively high throughout

484 the northwestern part of Lower Saxony. Moreover, it has been shown that the western
485 part of Lower Saxony is characterized by widespread and intensive cultivation of maize
486 for use as fodder or corn and for biogas production (Nitsch *et al.*, 2012; Steinmann and
487 Dobers, 2013). The eastern region of Lower Saxony was found to exhibit lower
488 livestock density values (cluster (B), Fig. 3; cf. Table 3), and was thus considered more
489 likely to support farmlands with high nature value (cf. Table 3). As a result, only
490 farmland-dominated municipalities within the Eastern region of Lower were considered
491 for detailed analysis.

492 Subsequent analysis was built on three sets of indicators, expressing the landscape
493 structure and composition, the extensive character of farming practices and crop
494 diversity. Spatial-statistics analysis performed in Step 3a allowed the selection of
495 municipalities under intensive farming systems, reflected as higher shares for
496 intensively managed crops and low crop diversity (cf. Table 4; cluster (a) Fig. 4), thus
497 assumed as non-HNVf (cf. Fig. 4). Cluster (a) (cf. Fig. 4), coincident with the
498 ‘Braunschweig-Hildesheimer Lössbörde’ area is, in fact, characterized by fertile loess
499 soils and large-scale production of intensive crops such as sugar beet and winter wheat.
500 As for clusters (b) and (c) (cf. Fig. 4), variation of extensive farming practices, crop
501 diversity indicators and landscape elements sets of indicators (cf. Table 4) supported
502 their classification as farmlands with potential to support both HNVf1 and 2 (North
503 region, Eastern Lower Saxony) or only HNVf2 (South region, Eastern Lower Saxony),
504 respectively. Cluster (b) was found to be under more extensive farming practices
505 (expressed as lower values for both the share of intensively managed crops and
506 livestock density index; cf. Table 4), while showing higher crop diversity and density of
507 linear elements. Accordingly, cluster (b) was considered to be a mosaic of semi-natural

508 vegetation (HNVf1) and low intensity cultivated land intermingled with small-scale
509 landscape elements (HNVf2) (Boyle *et al.*, 2015; Lomba *et al.*, 2015)

510 A fine-scale analysis, implemented at the physical block level (cf. Table 5 and Fig. 5),
511 allowed the refinement of the initial HNVf1 and HNVf2 assessments. Building upon a
512 thorough analysis of variation of all sets of indicators at the scale of the physical-block
513 (high spatial resolution), such refinement contributed to a more detailed delineation of
514 farmlands with the potential for high nature value in the study area.

515 By definition, HNVf1 owe their nature value to the presence of semi-natural
516 agriculture-related habitats (e.g. Andersen *et al.*, 2003 and Lomba *et al.*, 2014). It has
517 been demonstrated that the area covered by semi-natural habitats at the farm or
518 landscape scale could be used as an effective proxy for farmland biodiversity (e.g.
519 Boyle *et al.*, 2015, but see also Billeter *et al.*, 2008). Even though essential e.g. for
520 HNVf1 assessment, comprehensive spatial-explicit surveys of semi-natural grasslands
521 are seldom available at the regional or national scales (Veen *et al.*, 2009). Here, this
522 limitation was mitigated by deriving the share of the physical block covered by
523 permanent grasslands and the share of grasslands under agri-environment scheme
524 management contracts from IACS data, and considering both indicators as proxies for
525 the occurrence of semi-natural grasslands (cf. Table 1 and Table 5; Boyle *et al.*, 2015).

526 In fact, while permanent grasslands are not included within crop rotations for more than
527 five consecutive years (contrary to temporary grasslands, (Huyghe *et al.*, 2014),
528 permanent grasslands under agri-environment scheme management contracts are further
529 characterized by restrictions on farming intensity such as limited stocking rates,
530 restricted use of fertilizers and chemical pesticides, and can therefore be assumed to be
531 more extensively managed (Boyle *et al.*, 2015). Incorporating information on the
532 occurrence of habitats and species known to be associated with HNV farmlands would

533 constitute an important contribution to our results, by providing data for testing the
534 sensitivity of the methodological approach (Lomba *et al.*, 2015) and validation of
535 results (Doxa *et al.*, 2010).

536 Overall, our results indicated that only 3% of the UAA of Lower Saxony constituted
537 High Nature Value farmlands. Also, whilst ca.1% of the total UAA of Lower Saxony
538 corresponded to semi-natural grasslands under extensive farming practices (HNVf1; cf.
539 section 3.3 and Table 5), more than 70% of all HNV farmlands were found to be HNVf
540 type 2. Even though HNVf types 1 and 2 were considered to co-occur, spatial patterns
541 pinpointed the floodplains of the Elbe River as one hot-spot for HNVf1 (cf. Fig. 5). As
542 described by (Ludewig *et al.*, 2014), these species-rich floodplain meadows are among
543 the most threatened plant communities in Europe. Moreover, our delineation of areas
544 under HNVf is generally in agreement with those of Oppermann *et al.*, 2012.

545 Discrimination of the nature value farmlands as HNVf1 and HNVf2, according to
546 Andersen *et al.*, 2003, may constitute in the future a tool for the optimization of
547 monitoring programs and for establishing priority areas to be supported under agri-
548 environmental schemes.

549

550 4.2. IACS data and the assessment of HNVf across Europe

551 Our study demonstrated the potential of IACS as a tool for HNVf assessment across the
552 EU countryside. IACS was used as data source to derive sets of spatially-explicit
553 indicators characterizing landscape elements, extensive practices and crop diversity.

554 However, even though IACS has high potential for being used for HNVf assessment
555 across EU, some challenges concerning the use of IACS remain to be addressed (e.g.
556 see Lomba *et al.*, 2014; Strohbach *et al.*, 2015).

557 Due to its thematic resolution, IACS data allows for the discrimination between crop
558 types and some agriculture-related habitats such as permanent and temporary
559 grasslands, which is an added-value when compared with broad-scale land-cover and/or
560 land-use data such as Corine Land Cover maps. However, in some cases, habitats likely
561 to be HNVf type 1 may not be accounted for as eligible agricultural area within IACS,
562 due to CAP-related subsidy regulations, namely grasslands constituting common
563 grazing land or woody pastures (Jakobsson and Lindborg, 2015). Overcoming such
564 limitations, incorporating semi-natural habitats that constitute HNVf1, would most
565 likely require ancillary data, such as vegetation/habitat maps or ground-based surveys.
566 However, even though spatially-explicit data on vegetation and ecosystem types, and
567 vegetation databases are available for some European Member States (Veen *et al.*, 2009;
568 Chytrý, 2015), their potential to be used to complement IACS data will depend, to a
569 large extent, on their thematic, spatial and temporal resolution, and thus will have to be
570 assessed whenever its required.

571 Currently, efforts are been made towards the simplification of EU agricultural control
572 procedures, including the amount of information that farmers and land managers are
573 asked to report each year. Even though such a review of IACS is regarded as being
574 essential by the EU and Member States, we argue that the benefits to the wider society
575 from the use of such data could far outweigh the requirements being put on the
576 beneficiaries of CAP payments. In addition, a thorough analysis of the full potential to
577 use pre-existing IACS spatial and temporal data for such wider purposes than originally
578 intended may provide key information and evidence to the ongoing IACS review.
579 Nevertheless, the potential applicability of IACS as data source for HNVf assessment
580 would be enhanced if other information e.g. reflecting fertilizer and pesticide inputs
581 would be recorded and incorporated. Such information would be essential to further

582 inform on the intensity of farming practices underlying farming systems, and thus
583 enable a more precise identification of extensively managed farmlands likely to be
584 HNVf. While requesting more information under IACS surveys may imply more short-
585 time costs, long-term trade-offs between costs and benefits should be considered e.g.
586 IACS data usage for improved targeting, monitoring and evaluating of agri-environment
587 schemes.

588 HNVf assessment using IACS time-series data for other territories within the EU would
589 allow progress on the evaluation of farmlands with high nature value in space and time.
590 Such progress would benefit from research focusing specifically the evaluation and
591 eventually the re-definition of spatially-explicit indicators that could support the
592 selection of a set of informative and cost-effective indicators for that purpose. IACS
593 data collection does, however, differ between Member States in terms of spatial and
594 thematic resolution. It is therefore of utmost importance to test its use in other social-
595 ecological contexts, to evaluate how widespread such data could be used across Europe.
596 Extending HNVf assessment using IACS data to other EU Member States would also
597 allow testing the ability of the approach to cope with the variation of HNV landscapes
598 across the EU. At the same time, it would allow the assessment of issues relating to
599 differences in data quality and compatibility between and within Member States, that
600 may hamper our ability to implement similar approaches across EU.

601

602 **5. Conclusions**

603 Our results highlight the potential of IACS as spatially and thematic high-resolution
604 data source for assessing and monitoring the extent of High Nature Value farmlands in
605 the EU. Even though trade-offs between thematic and spatial resolution of available

606 IACS data need to be weighted for each targeted area, the conceptual and
607 methodological approach is, from our point of view, flexible enough to allow an
608 effective assessment of HNVf in space and time across the EU countryside.

609 Overall, making use of IACS data in such a way could provide a stepping-stone towards
610 achieving a more effective balance between the management and control of CAP
611 support payments and the growing societal demands related to the maintenance and
612 enhancement of farmland biodiversity and ecosystem services.

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621

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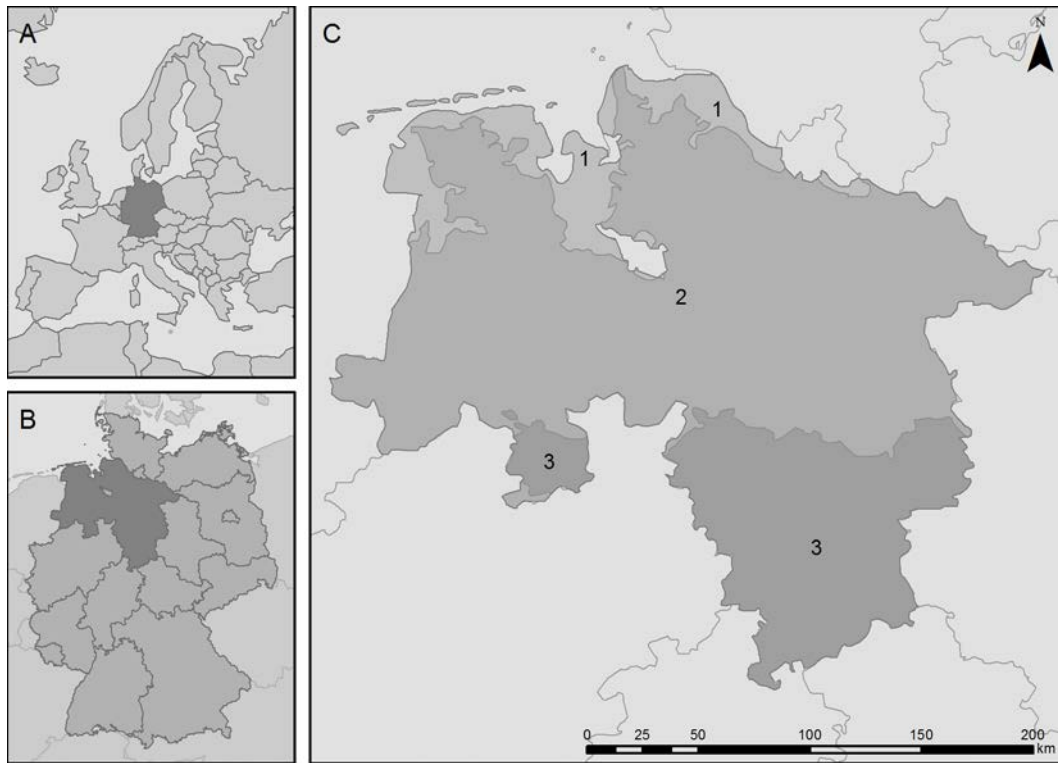
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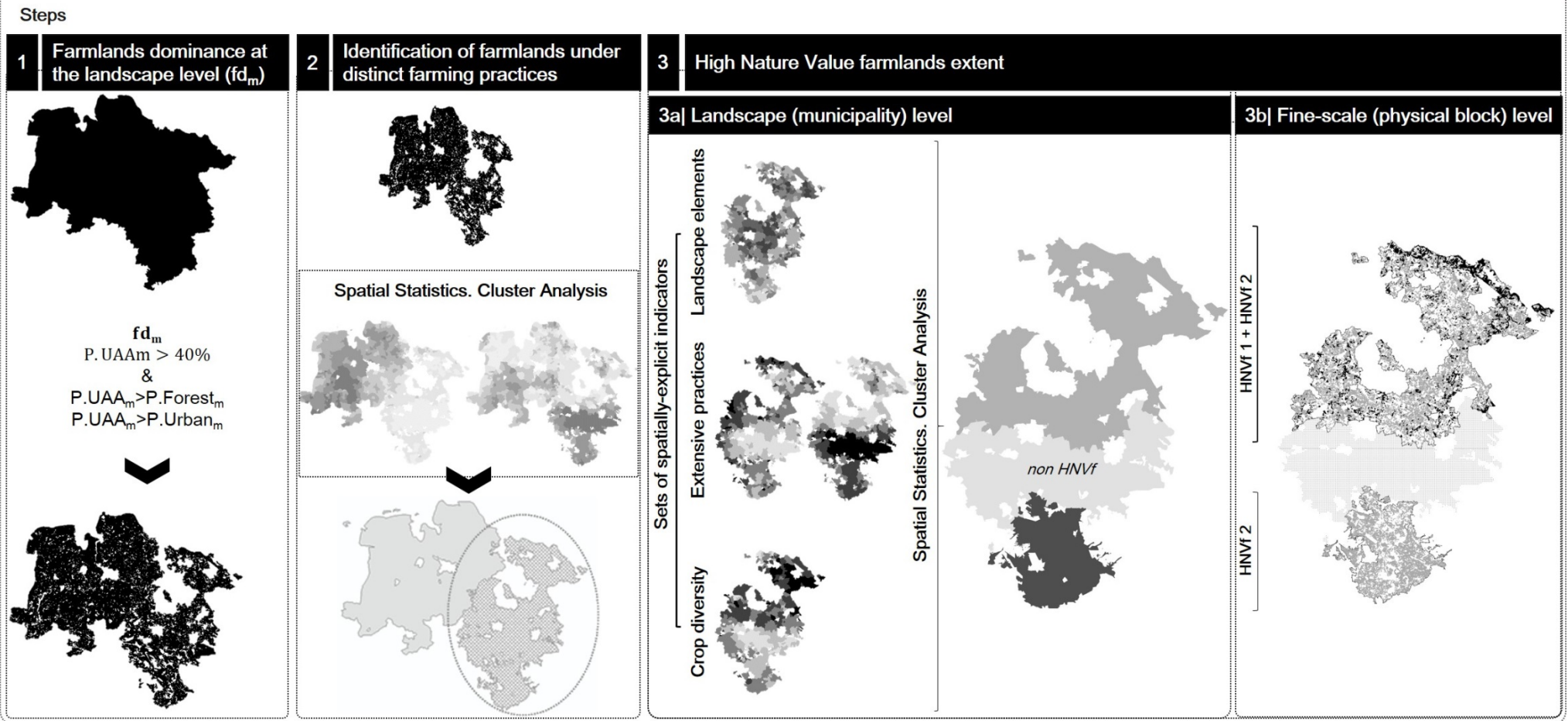
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- 1 **Title:** Making the best of both worlds: can high-resolution agricultural administrative
2 data support the assessment of High Nature Value farmlands across Europe?
3 Figures and tables



4
5 Fig. 1. The geographic location of the study area, the federal state of Lower Saxony,
6 within Europe (A) and Germany (B), and the three major biogeographical regions
7 dividing the study area (C): coastal areas (1), the lowlands (2) and the uplands (3).

Framework for High Nature Value farmlands assessment using high-resolution agricultural administrative data (IACS)



8

9 **Fig. 2.** Framework implemented to assess the extent of High Nature Value farmlands (HNVf) in Lower Saxony using high-resolution
 10 administrative agricultural data. Overall, the presented framework consists on a three-step, sequential, multi-criteria analysis of spatially-explicit

11 indicators. In Step 1, indicators informing on the composition of the landscape, including the percentage of utilized agricultural ($P.UAA_m$), forest
12 ($P.Forest_m$), and urban areas ($P.Urban_m$), were used to highlight municipalities where farmlands were dominant at the landscape level (fd_m).
13 Farmland dominance was defined considering a minimum threshold value of 40% of $P.UAA_m$, and the prevalence of higher values for the
14 percentage cover of Utilized Agricultura Area ($P.UAA_m$) in relation to forest ($P.Forest_m$) and urban areas ($P.Urban_m$). Step 2, implemented on
15 farmland-dominated municipalities selected in Step 1, consisted of a spatially-explicit analysis of patterns of indicators expressing the intensity of
16 farming practices aiming to discriminate areas under distinct farming practices. By discriminating such areas, we aimed to focus on extensively
17 managed farmlands, as those more likely to support farmlands with high nature value. In Step 3, areas of HNVf types 1 and 2 were , by first
18 implementing a landscape-level analysis of spatially-explicit patterns for indicators reflecting extensive farming practices, landscape elements
19 and crop diversity (Step 3a) and, combining outcomes from such analysis with a fine-scale analysis, built on high-resolution data at the physical
20 block level, in Step 3b.

21 **Table 1.** Sets of spatially-explicit indicators used to assess the extent of High Nature Value farmlands in Lower Saxony, Germany. Steps of
22 analysis: Step 1 consisted on the definition of farmlands dominance at the landscape level (municipality); Step 2 highlighted areas under farming
23 practices more likely to support farmlands with high nature value; and, Step 3 targeted areas of HNVf types 1 and 2 by first implementing a
24 landscape-level detailed analysis (Step 3a), combined with a fine-scale analysis built on high-resolution data at the physical block level (Step 3b).
25 (%), percentage; (n.a.), not applicable; (ha), hectares; HNVf1, High Nature Value farmlands type 1; HNVf2, High Nature Value farmlands type
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Designation	Code(s) and units	Rationale	Source and determination	Step of analysis	HNVf types
Landscape elements					
Farmlands dominance in the landscape	P.UAA _m (%)	Areas where the percentage (%) cover of farmlands (P.UAA _m) is dominant in relation to forests (P.Forest _m) and urban areas (P.Urban _m), at the municipality level.	IACS and associated LPIS	Step 1	Types 1 and 2
	P.Forest _m (%)	Values were calculated from the area covered by each class (farmland, forest and urban) in relation to the total municipality area, and are thus expressed as the percentage. A threshold value of 40%	Forest, including closed deciduous and coniferous forests classes.		
	P.Urban _m (%)	P.UAA _m was considered to define farmland dominance at the municipality level, following the rule of thumb proposed by	Urban and built-up areas covered by buildings, streets and other urban		

		Lomba <i>et al</i> (2015).	land-uses.		
				Basis-DLM	
Density of tree lines and hedgerows	DT _m (m/ha)	Density of tree lines and hedgerows within each municipality.	Areas covered by linear elements, including rows of trees, hedges and hedge-banks.	Step 3a	Type 2
Share of grasslands in each physical block	P.G _{pb} (%)	Determined as the share of the physical block covered by grasslands, highlights areas predominantly covered by grasslands.	IACS	Step 3b	Type 1
Extensive practices					
Livestock Density Index	LSI _m (total LSU per ha/UAA)	Indicator expressing the pressure of livestock on the environment, measured as livestock units (LSU) per hectare of UAA	IACS	Step 2 Step 3a	Types 1 and 2

	(LSU/ha) at the municipality level. Lowest values of LSI_m are usually observed in landscapes where semi-natural forage (i.e. permanent grasslands) predominates.			
Share of intensive crops	Determined as the share of UAA covered by intensive types of crops (winter wheat, sugar beet, maize and oilseed rape), highlights municipalities under intensive farming practices.	IACS	Step 2	Types 1 and 2
	Determined as the share of UAA covered by intensive crops (winter wheat, sugar beet, maize and oilseed rape) at the level of the physical block.	IACS	Step 3b	Type 2

Share of grasslands under agri-environment schemes in each physical block	P.GAES _{pb} (%)	Determined as the share of the physical block covered by permanent grasslands under agri-environment scheme management contracts.	IACS	Step 3b	Type 1
Crop diversity					
Cropping patterns	SEI _m (n.a.)	Cropping patterns expressed as the Shannon Evenness Index. The index accounts for the diversity of crops and the evenness of their distribution and was calculated at the municipality level using the shares of crops registered in the IACS database. Varies between 0 and 1.	IACS	Step 3a	Type 2
	SEI _{pb} (n.a.)	Cropping patterns expressed as the Shannon Evenness Index on farm level, averaged at the physical block level.	IACS	Step 3b	Type 2

34 Table 2. Variation of spatially-explicit sets of indicators (landscape elements, extensive
 35 practices and crop diversity) across municipalities dominated by farmlands in Lower
 36 Saxony. (%) stands for percentage; n.a., not applicable. Mean, minimum (Min),
 37 maximum (Max) and standard deviation (SD) values are presented.

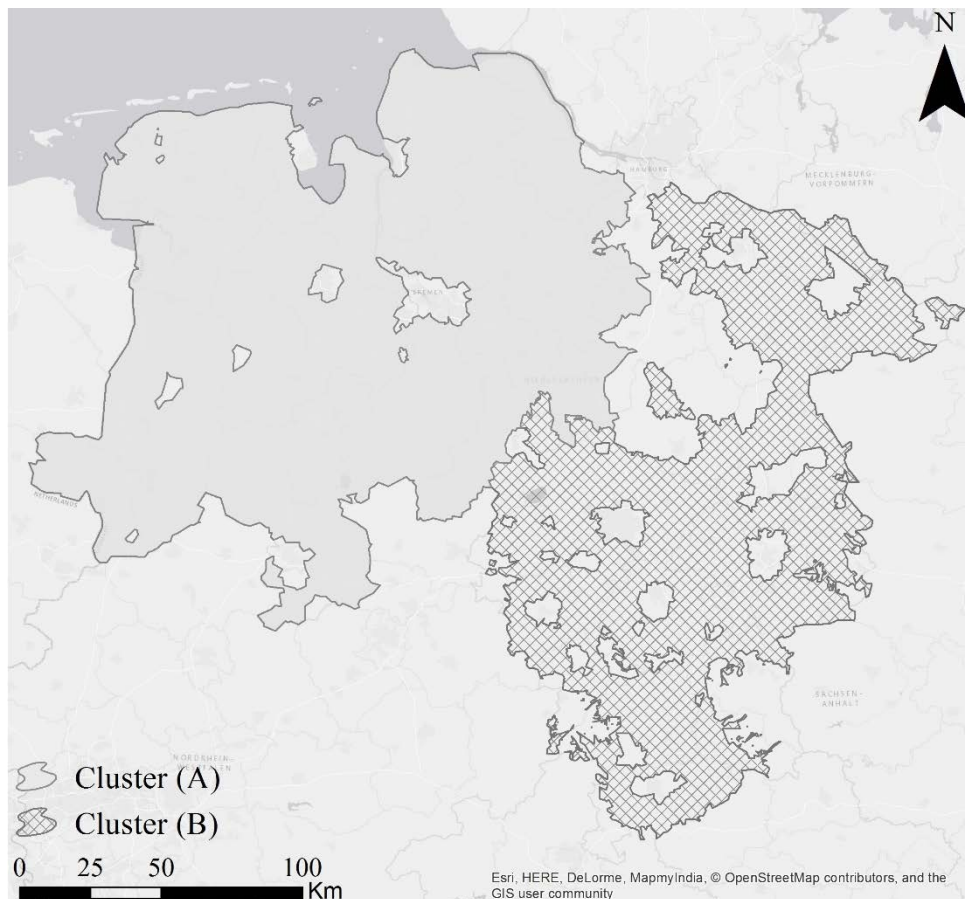
Sets of indicators	Code and units	Mean	SD	Min	Max
Landscape elements	DT _m (ha)	4.14	3.24	0.00	23.10
Extensive practices	LSI _m (LSU per ha/UAA)	0.99	0.69	0.00	3.20
	P.ICrops _m (%)	0.39	0.20	0.00	0.99
Crop diversity	SEI _m (n.a.)	0.38	0.10	0.04	0.63

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39 Table 3. Clusters resulting from the spatially-explicit analysis of indicators expressing
 40 the extensive character of farming practices across farmland dominated municipalities
 41 in Lower Saxony. *n*, stands for the number of municipalities. Mean, minimum (Min),
 42 maximum (Max) and standard deviation (SD) values are presented. Share values depict
 43 the ratio between the range of values observed within groups (A, B) and the full range
 44 of values observed for each indicator.

Cluster	<i>n</i>	Code and units	Mean	SD	Min	Max	Share (%)
A	496	LSI _m (LSU per ha/UAA)	1.50	0.48	0.080	3.20	0.98
		P.ICrops _m (%)	0.31	0.13	0.01	0.66	0.66
B	391	LSI _m (LSU per ha/UAA)	0.37	0.24	0.00	1.18	0.37
		P.ICrops _m (%)	0.50	0.22	0.050	0.99	0.96

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47 Fig. 3. Outcomes of grouping analysis implemented on the extensive farming practices
 48 set of indicators. Groups (A and B) resulted from the cluster analysis of patterns for
 49 both the livestock density index and the share of intensive crops across farmland-
 50 dominated municipalities across Lower Saxony. Clusters reflected either high values of
 51 LSI_m and moderate levels of $P.ICrops_m$ in the western area of Lower Saxony, cluster
 52 (A); or lower global values for LSI_m and slightly higher values of $P.ICrops_m$ in the
 53 eastern cluster of municipalities, cluster (B). LSI_m , livestock density expressed as
 54 livestock units per hectares of Utilized Agricultural Area (ha.UAA); $P.ICrops_m$, share of
 55 intensive crops per municipality, expressed as the percentage (%) of total area occupied
 56 by crops more prone to be under intensive farming practices.

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58 **Table 4.** Results from the grouping analysis targeting farmlands more likely to be High
59 Nature Value farmlands (HNVf) in Lower Saxony. *n*, stands for the number of
60 municipalities; *n.a.*, not applicable. Mean, minimum (Min), maximum (Max) and
61 standard deviation (SD) values are presented. R^2 , reflects the discriminating ability of
62 each individual variable, which is higher for larger values. Share values depict the ratio
63 between groups (HNVf1 and HNVf2; HNVf2, and non-HNVf) and full area (Eastern
64 Lower Saxony) range values for indicators. HNVf type refers to the classification of
65 areas delineated within each cluster according to their overall characteristics. HNVf1
66 and HNVf2, stand for High Nature Value farmlands types 1 or 2, respectively. Non-
67 HNVf refer to farmlands that do not exhibit characteristics that convey a high nature
68 value.

	<i>n</i>	Code and units	Mean	SD	Min	Max	R^2	
Full area		P.ICrops _m (%)	0.50	0.22	0.050	0.99	0.66	
		SEI _m (<i>n.a.</i>)	0.44	0.074	0.20	0.63	0.47	
		LSI _m (LSU per ha/UAA)	0.37	0.24	0.00	1.18	0.23	
		DT _m (ha)	4.39	2.90	0.00	13.30	0.18	
Grouping	n	Code and Units	Mean	SD	Min	Max	Share (%)	HNVf type
(a)	163	P.ICrops _m (%)	0.69	0.15	0.15	0.99	0.89	Non HNVf
		SEI _m (<i>n.a.</i>)	0.39	0.057	0.23	0.55	0.73	
		LSI _m (LSU per ha/UAA)	0.23	0.21	0.00	0.94	0.80	
		DT _m (ha)	4.10	2.83	0.00	13.3	1.00	
(b)	165	P.ICrops _m (%)	0.30	0.11	0.030	0.64	0.63	HNVf1 and
		SEI _m (<i>n.a.</i>)	0.50	0.056	0.28	0.62	0.80	HNVf2

		LSI _m (LSU per ha/UAA)	0.46	0.23	0.050	1.18	0.96	
		DT _m (ha)	5.59	2.71	0.33	12.17	0.89	
		P.ICrops _m (%)	0.52	0.13	0.080	0.71	0.67	
(c)	63	SEI _m (<i>n.a.</i>)	0.42	0.042	0.20	0.49	0.68	HNVf2
		LSI _m (LSU per ha/UAA)	0.49	0.18	0.15	0.98	0.70	
		DT _m (ha)	2.05	1.64	0.00	7.72	0.58	

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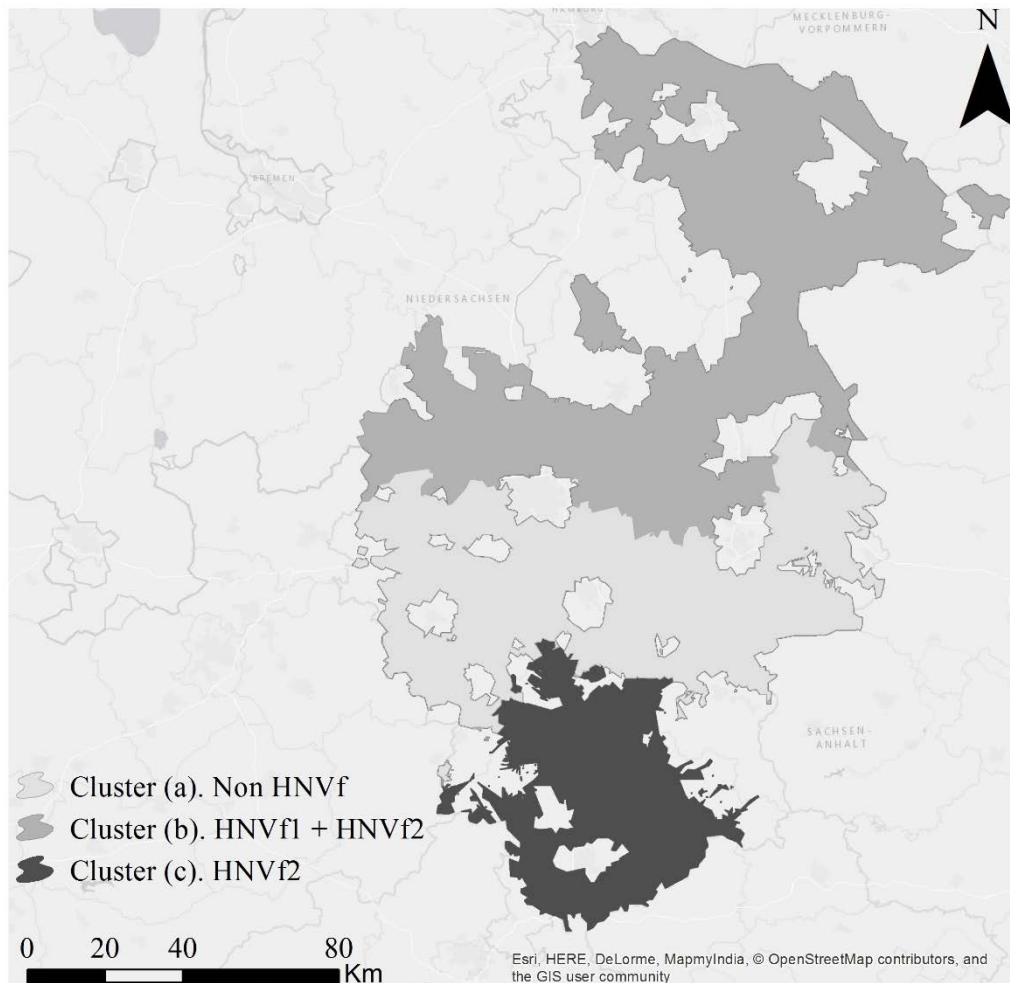
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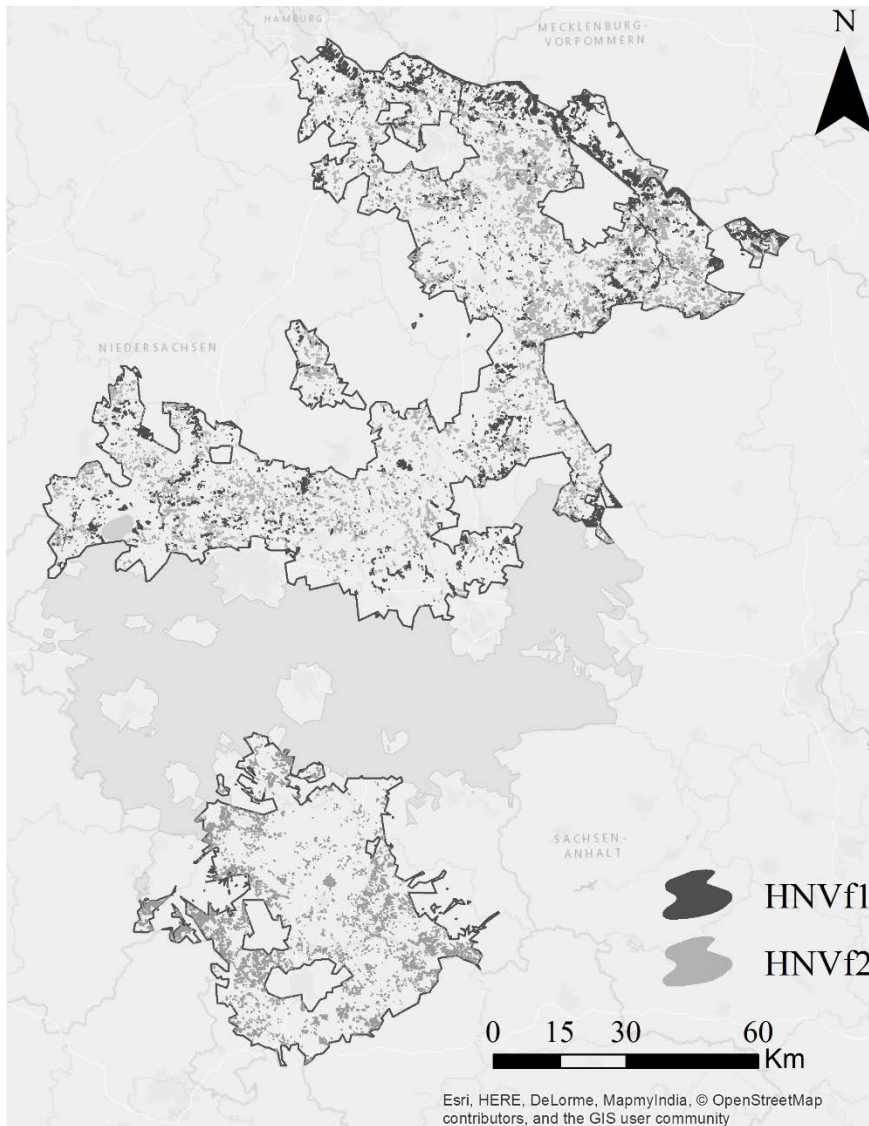
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93 Fig. 4. Delineation of High Nature Value farmlands (HNvf) potential areas in Eastern
 94 Lower Saxony, based on grouping analysis on spatially-explicit indicators informing on
 95 the intensity of farming practices, landscape elements and crop diversity. HNvf, High
 96 Nature Value farmlands; HNvf1, High Nature Value farmlands type 1; HNvf2, High
 97 Nature Value farmlands type 2.

98 **Table 5.** Variation observed within each indicator considered at the physical block level for farmlands potentially supporting both High Nature
 99 Value farmlands types 1 and 2 (HNVf1 + 2), or only High Nature Value farmlands type 2 (HNVf2). *pb*, stands for physical block; *pb*, physical
 100 blocks targeted for each HNVf type; *n.a.*, not applicable; SD, standard deviation; Q₁, refers to the first quartile; Q₃, third quartile; IQR,
 101 interquartile range; ha UAA, refers to area in hectares of the Utilized Agriculture Area. Values applied as thresholds to include or exclude
 102 physical blocks are highlighted as bold.

		Indicators _{pb} /units	Mean	SD	Q ₁	IQR	Q ₃	<i>t</i> _{pb}	ha UAA
HNVf1 + 2	HNVf1	P.G _{pb} (%)	85.77	25.68	83.41	16.59	100.00	5024	27,214.94
		P.GAES _{pb} (%)	19.47	36.595	0.00	9.93	9.93		
	HNVf2	SEI _{pb} (<i>n.a.</i>)	0.45	0.19	0.39	0.19	0.58	9884	47,912.48
		P.ICrops _{pb} (%)	0.19	0.34	0.00	0.24	0.24		
HNVf2	SEI _{pb} (<i>n.a.</i>)		0.42	0.11	0.38	0.11	0.49	5542	17,841.68
	P.ICrops _{pb} (%)		0.37	0.43	0.00	0.88	0.88		
Total HNVf									92,969.10



103

104 **Fig. 5.** Fine-scale mapping of High Nature Value farmlands (HNVf) in Lower Saxony
 105 at the physical block level. To assure privacy and protection of land owners and
 106 managers geometries are not disclosed. HNVf1, High Nature Value farmlands type 1;
 107 HNVf2, High Nature Value farmlands type 2.

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