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The use of electromagnetic induction techniques in soils studies

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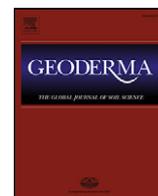
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Review

The use of electromagnetic induction techniques in soils studies

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ABSTRACT

Electromagnetic induction (EMI) has been used to characterize the spatial variability of soil properties since the late 1970s. Initially used to assess soil salinity, the use of EMI in soil studies has expanded to include: mapping soil types; characterizing soil water content and flow patterns; assessing variations in soil texture, compaction, organic matter content, and pH; and determining the depth to subsurface horizons, stratigraphic layers or bedrock, among other uses. In all cases the soil property being investigated must influence soil apparent electrical conductivity (EC_a) either directly or indirectly for EMI techniques to be effective. An increasing number and diversity of EMI sensors have been developed in response to users' needs and the availability of allied technologies, which have greatly improved the functionality of these tools. EMI investigations provide several benefits for soil studies. The large amount of georeferenced data that can be rapidly and inexpensively collected with EMI provides more complete characterization of the spatial variations in soil properties than traditional sampling techniques. In addition, compared to traditional soil survey methods, EMI can more effectively characterize diffuse soil boundaries and identify areas of dissimilar soils within mapped soil units, giving soil scientists greater confidence when collecting spatial soil information. EMI techniques do have limitations; results are site-specific and can vary depending on the complex interactions among multiple and variable soil properties. Despite this, EMI techniques are increasingly being used to investigate the spatial variability of soil properties at field and landscape scales.

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1. Introduction

It is widely recognized that there is considerable variability within soils (Brevik et al., 2003; Doolittle et al., 1996; Miller, 2012). Electro-magnetic induction (EMI) is widely used by soil scientists to better understand the spatial variability of soils and soil properties at field and landscape scales (Corwin, 2008; Tushmalani, 2010). Because of

its speed, ease of use, relatively low cost, and volume of data collected, EMI has immense advantages over traditional methods used to collect soil information. Recent improvements in instrumentation and integration with other technologies (global-positioning systems (GPS), data processing software, and surface mapping programs) have fostered the expanded use of EMI in soils applications. The impetus for this expanded use has been the need for more accurate soil maps than those provided by traditional mapping techniques (Batte, 2000; Brevik et al., 2003, 2012) and the demonstrated efficiency of EMI to improve the accuracy and reliability of soil maps and provide more detailed information on soils and soil properties.

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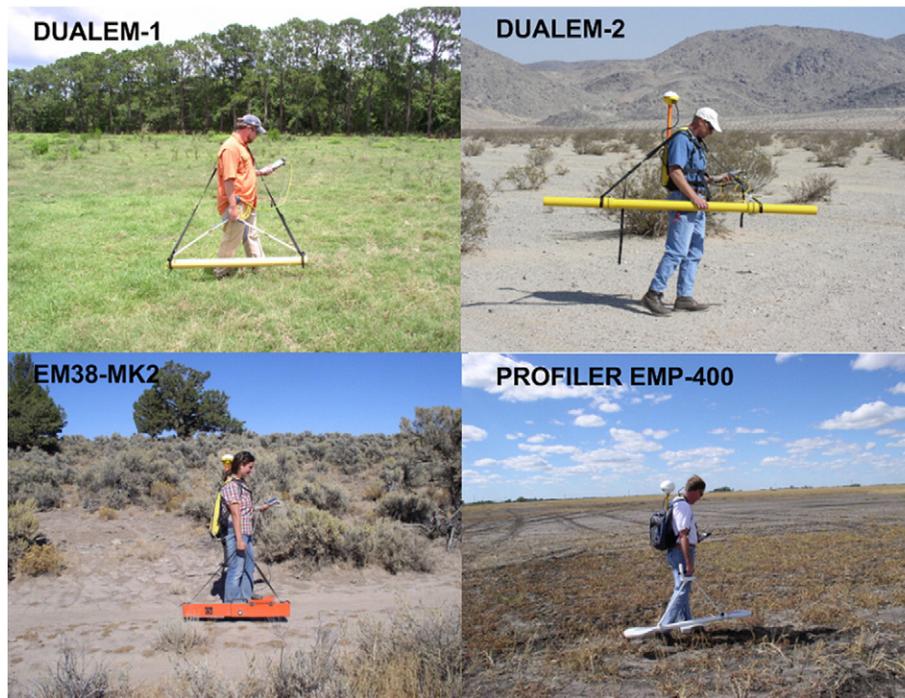


Fig. 1. Four EMI sensors commonly used in soil investigations are the DUALEM-1 meter, the DUALEM-2 meter, the EM38-MK2 meter, and the Profiler EMP-400.

Electromagnetic induction sensors measure changes in the apparent electrical conductivity (EC_a) of the subsurface without direct contact with the sampled volume (Allred et al., 2008; Daniels et al., 2003). Apparent electrical conductivity is a depth-weighted, average conductivity measurement for a column of earthen materials to a specific depth (Greenhouse and Slaine, 1983). Variations in EC_a are produced by changes in the electrical conductivity of earthen materials. Apparent electrical conductivity will increase with increases in soluble salt, water, clay contents, and temperature (Brevik and Fenton, 2002; Kachanoski et al., 1988; McNeill, 1980a; Rhoades et al., 1976).

2. EMI sensors

An increasing number of commercially available EMI sensors are available (Fig. 1). Electromagnetic induction sensors commonly used in agriculture and soil investigations include the DUALEM-1 and DUALEM-2 meters (Dualem, Inc., Milton, Ontario); the EM31, EM38, EM38-DD, and EM38-MK2 meters (Geonics Limited, Mississauga, Ontario), and the Profiler EMP-400 (Geophysical Survey Systems, Inc., Salem, New Hampshire).¹ These EMI sensors transmit a primary electromagnetic field, which induces electrical currents in the soil. These currents generate a secondary electromagnetic field, which is read by the sensor's receiver. Under conditions known as "operating under low induction numbers", the secondary field is proportional to the ground current and is used to calculate the "apparent" or "bulk" electrical conductivity (EC_a) for the volume of soil profiled. The dual-geometry configuration of the DUALEM-1 and DUALEM-2 meters, the dual orientation of the EM38-DD meter, and the dual receiver-transmitter spacings of the EM38-MK2 meter allow the simultaneous measurement of EC_a and/or apparent magnetic susceptibility (MS_a) over two distinct depths. The depth of investigation (DOI) for EC_a measurements made with sensors developed by Dualem, Inc. and Geonics Limited is commonly taken as the depth of 70% cumulative response. The Profiler EMP-400 is a multi-frequency sensor and its DOI is assumed to be "skin-depth"

¹ Manufacturer's names are provided for specific information; use does not constitute endorsement.

limited and dependent upon the frequency and the conductivity of the profiled materials. All of the aforementioned sensors support GPS communication, data loggers, and proprietary software. Some EMI sensors, such as the DUALEM-1, DUALEM-2S, and Profiler EMP-400, come with internal GPS receivers and display/keypads.

Each of the aforementioned sensors has distinct operational advantages and disadvantages (Sudduth et al., 2003). Comparative studies have generally revealed close similarities between EC_a data collected with different sensors (Doolittle et al., 2001, 2002a; Saey et al., 2009a; Sudduth et al., 1999, 2003; Urdanoz and Aragüés, 2012). However, differences in sensor calibration, depth sensitivity and volume of soil material measured will affect measurements and result in slightly different EC_a values. In comparative studies using different sensors, the highest correlations in measured EC_a were obtained with sensors having similar depth sensitivities (Sudduth et al., 1999, 2003). Differences in EC_a data collected with different sensors have been attributed to differences in sensing depths and data collection modes (e.g., coil spacing, orientation, or geometry). In general, differences in EC_a data collected with different sensors have been more noticeable over soils with highly contrasting layers (Sudduth et al., 2003).

3. History

The first use of EMI in agriculture was for the assessment of soil salinity (Corwin and Rhoades, 1982; de Jong et al., 1979; Rhoades and Corwin, 1981; van der Leij, 1983; Williams and Baker, 1982). In 1976, Geonics Limited patented and manufactured the EM31 meter. The EM31 meter has a 3.66 m intercoil spacing and operates at a frequency of 9.8 kHz (Fig. 2). This meter provides DOI of 3 m and 6 m when operated in the horizontal (HDO) and vertical (VDO) dipole orientations, respectively. Consideration for near-surface applications in agronomy and soil science lead to the development of the EM38 meter in 1980. The EM38 meter is the most widely used EMI sensor in agriculture (Sudduth et al., 2001). The EM38 meter has a coil separation of 1 m and operates at a frequency 14.6 kHz. This meter provides DOI of 0.75 and 1.5 m when operated in the HDO and VDO, respectively.



Fig. 2. In the 1980s, pedestrian surveys were typically completed with EMI sensors across gridded areas. Here, an EM31 meter is operated in the station-to-station mode without a data logger (another person was required to record the data).

In the 1980s, EMI surveys were commonly completed in a station-to-station mode across gridded areas or along traverse lines. The establishment of a survey grid often took more time than the actual EMI survey. Pedestrian surveys were conducted by moving from one grid point to the next within the gridded area (Fig. 2). Typically, measurements were made in two dipole orientations at each grid intersection.

In the late 1980s, data loggers were first used with EMI meters. Data could now be quickly and accurately recorded in the field and later transferred to computers for processing and display. By the mid- to late-1990s, the maturation of GPS and its integration with EMI sensors and data loggers revolutionized the collection of EC_a data. The merger of these technologies allowed continuous sampling and made possible the rapid collection of geo-referenced EMI data using on-the-go or mobile platforms. This fusion of technologies allowed the rapid collection of spatially dense data sets and made intensive field-scale surveys practical and commonplace (Cannon et al., 1994; Carter et al., 1993; Freeland et al., 2002) (Fig. 3). However, when operated in the continuous recording mode, EMI sensors cannot be rotated back and forth between the two dipole orientations, and as a consequence, two separate surveys were required to collect data in both dipole orientations (two different depths). Many users wanted simultaneous measurements of EMI data in both dipole orientations to provide two DOI with one pass of the sensor. This led to the development of dual-dipole sensors, which



Fig. 3. The fusion of EMI and GPS technologies fostered the use of mobile EMI platforms, which facilitated the rapid collection of spatially dense data sets across large units of management. Here, a Profiler EMP-400 is being towed on sleds behind a 4WD all-terrain-vehicle.

permitted continuous, simultaneous measurements of EC_a and/or MSA in two dipole configurations.

In 1998, Dualem developed the DUALEM-4 sensor with dual 4-m arrays that provide DOI of 2 m and 6 m in the perpendicular (PRP) and horizontal co-planar (HCP) geometry, respectively. The PRP geometry is equivalent to HDO and the HCP geometry is equivalent to VDO as used in reference to the Geonics instruments. The DUALEM-4 was followed by the DUALEM 42, which could be assembled with either 4-m or 2-m arrays. The DOI for the 2 meter array is 1 m and 3 m in the PRP and HCP geometry, respectively. In 2004, Dualem introduced the DUALEM-1 and DUALEM-1S sensors that have 1-m arrays and provide DOI of 0.5 m and 1.5 m in the PRP and HCP geometry, respectively. The first Dualem sensor with dual arrays of multiple lengths was the DUALEM-21, introduced in 2007. This EMI sensor was followed by the DUALEM-421 in 2008 and the DUALEM-642 in 2011. These complex EMI sensors incorporate six arrays, which provide six DOI. Each DUALEM sensor operates at a fixed frequency of 9.0 kHz and provides simultaneous measurements of both EC_a and MSA. These sensors come with internal WAAS-enabled GPS receivers and a hand-held weather-proof display/keypad.

In 2000, Geonics Limited developed the EM38-DD meter, which consists of two EM38 meters bolted together and electronically coupled. One unit is positioned in the VDO and one unit is positioned in the HDO to provide simultaneous measurements of EC_a or MSA over two depth intervals. In 2008, Geonics Limited developed the EM38-MK2 meter, which operates at 14.6 kHz and has one transmitter coil and two receiver coils that are separated from the transmitter coil at distances of 1.0 and 0.5 m. This geometry results in DOI of 1.5 and 0.75 m when the meter is operated in the VDO, and 0.75 and 0.40 m when operated in the HDO. The EM38-MK2 meter provides simultaneous measurements of both EC_a and MSA.

The Profiler EMP-400 is a multi-frequency electromagnetic induction sensor developed by Geophysical Survey System, Inc. in 2007. The system's primary data output is the in-phase and quadrature phase components of the mutual coupling field ratio of the transmitted field to the induced field in parts per million (ppm) at all frequencies, and EC_a at 15 kHz. The Profiler can simultaneously collect both in-phase and quadrature phase component data at one to three frequencies. The Profiler has an intercoil spacing of 1.22 m and operates at frequencies of 1 to 16 kHz. The Profiler comes with an integrated GPS receiver and the sensor's electronics are controlled from a personal digital assistant (PDA) via a wireless Bluetooth communications interface.

Present EMI sensors are well suited to soil studies. The future will witness the expanded use of multiple-frequency and multiple-coil EMI sensors and various combinations of these instruments to more effectively assess the variability of soil properties (Triantafyllis et al., 2013). The multiple depth responses of EMI sensors will be increasingly exploited with multi-layer inversion modeling algorithms to improve the resolution of subsurface features and the assessment of soil properties (Meerschman et al., 2011b; Mester et al., 2011; Saey et al., 2012a; Triantafyllis and Monteiro Santos, 2013; Triantafyllis et al., 2013). Multi-layer inversion modeling of EMI data will advance the quantitative mapping of both the lateral and vertical variations in soil properties (Mester et al., 2011). As examples of this synergy, Saey et al. (2012b) and De Smedt et al. (2013a, 2013b) used multi-receiver EMI sensors and depth-slicing methods to improve the resolution of both archeological (drainage ditches) and pedological (soil horizons, stratigraphic layers, paleotopographical structures (paleochannels)) features.

4. Applications

4.1. A surrogate measure for the assessment of soil properties

Considerable research is being conducted to better understand the soil properties that influence EC_a . The principal properties affecting

EC_a are the type and concentration of ions in solution, the amount and type of clays in the soil matrix, the water content, and the temperature and phase of the soil water (McNeill, 1980a). Apparent conductivity has also been associated with other ancillary soil properties such as bulk density, soil structure, ionic composition, CEC, pH, and soil organic carbon, nutrient, and $CaCO_3$ contents. These ancillary properties influence properties that determine soil EC_a , meaning they can indirectly influence EC_a and thus be investigated using EMI techniques (Heilig et al., 2011). The relationships between these interacting soil properties and EC_a are often complex and can vary over short distances (Bekele et al., 2005; Brevik and Fenton, 2004; Brevik et al., 2004; Carroll and Oliver, 2005; Farahani et al., 2005). As a result, the degree and in some cases the directions (\pm) of the relationship between EC_a and a specific soil property have varied. In general, stronger correlations are obtained where large differences in the measured soil property occur (horizontally and/or vertically), and all other soil properties that affect EC_a remain relatively constant. Weaker correlations and lower predictive accuracies occur where the measured soil property displays low variability in relation to several other interacting and more variable soil properties that affect EC_a . Complex interactions among different soil properties can confound interpretations and create ambiguous, inconsistent, and less significant results. Even with these challenges, EC_a has been increasingly used to infer and map the spatial variability of soil properties at field and landscape scales. Presently, EC_a mapping is recognized as one of the most valuable methods in agriculture for measuring the spatial variability of soil properties at field and landscape scales (Corwin, 2008; Lück et al., 2009).

4.1.1. Soil salinity

As noted by Corwin (2008), the adaptation of EMI to agriculture was largely motivated by the need for reliable, quick, and easy to take measurements of soil salinity at field and landscape scales. In soil surveys, the identification and mapping of salt-affected soils have been traditionally made by visual observations supported by limited laboratory measurements (Soil Survey Division Staff, 1993; United States Salinity Laboratory Staff, 1954). Visual observations, though adequate for general salinity mapping, provide only qualitative information and are dependent on the presence of plant cover, surface salts, and soil structural features and characteristics. Laboratory methods (e.g., electrical conductivity of the saturated paste extract and the saturated solution extract; sodium absorption ratio, exchangeable sodium percentage) are time-consuming and expensive to complete, and provide only a limited number of point measurements that may or may not be representative of the field or soil-landscape. Because of the high spatial variability of salt-affected soils, these traditional methods, while reasonably accurate, have limited values for the assessments of soil salinity and sodicity at field and regional scales (Corwin, 2008). A major advantage of EMI is its capacity to produce a large number of georeferenced, quantitative measurements that can be associated with the spatial variability of salinity and sodicity at field and landscape scales.

In areas of salt-affected soils, of all the physiochemical properties that influence EC_a , the concentration of soluble salts is the dominant contributing factor (Johnston et al., 1997; Mankin and Karthikeyan, 2002; van der Lelij, 1983; Williams et al., 2006). In these areas, EMI has been used to characterize unsaturated flow (Scanlon et al., 1999), estimate rates of groundwater recharge (Cook et al., 1989a, 1989b, 1992), map groundwater recharge and discharge zones (Richardson and Williams, 1994; Williams et al., 2006), and assess differences in soluble salt contents across landscapes (Cook et al., 1989a; Williams et al., 2006). Studies have confirmed that EMI provides reasonably accurate estimates of soil salinity at field scales (Diaz and Herrero, 1992; van der Lelij, 1983; Williams and Baker, 1982).

Early EMI research and field services in the late 1970s and early 1980s were principally directed towards the vertical profiling of salinity through the root zone (Corwin, 2008). A major challenge in using EMI to map soil salinity has been the conversion of apparent conductivity

(EC_a) into the conductivity of the saturated paste extract (EC_e); the most commonly used measure of soil salinity. A number of models were developed to predict EC_e from EC_a (Cook and Walker, 1992; Corwin and Rhoades, 1982, 1984, 1990; Johnston et al., 1997; Lesch et al., 1992, 1995a, 1995b; McKenzie et al., 1989; Rhoades et al., 1989a, 1989b; Slavich, 1990; Wollenhaupt et al., 1986). Unfortunately, models are imperfect and tend to be both time dependent and site specific (Lesch et al., 1998). As a consequence, calibration equations and modeled results usually cannot be extrapolated to other sites (Cassel et al., 2009). Another challenge to the use of EMI to map salinity occurs at high conductivity values, when the quadrature component of the received electromagnetic field is no longer linearly proportional to soil conductivity (breakdown of low induction number approximation); this occurs above conductivities of approximately 100 mS m^{-1} (McNeill, 1980b; Morris, 2009).

4.1.2. Subsurface water movement and soluble salts

The depth and movement of water through the subsurface have a direct effect on the physical and chemical properties and the morphology of soils (Richardson et al., 1992). Recharge processes remove soluble chemical constituents and translocate suspended colloids in soils. Discharge processes add soluble chemical constituents and suspended colloids to soils (Richardson et al., 1992). Because of upward leaching and evaporative processes, salts are concentrated near the soil surface in groundwater discharge sites (seeps) (Richardson and Williams, 1994). The higher concentration of soluble salts in surface layers results in higher EC_a and inverted salt profiles (EC_a is highest in surface layers and decreases with increasing depth). Conversely, groundwater recharge sites are characterized by the downward leaching and concentration of salts at greater soil depths. As a consequence, EC_a is relatively low in surface layers and increases with increasing depth (regular salt profile). In recharge areas, lower soluble salt and water contents are associated with lower EC_a (Mankin and Karthikeyan, 2002).

The aforementioned relationships have been used to map saline- and sodic-soils (Ammons et al., 1989; Doolittle et al., 2001; Ganjegunte and Braun, 2011; Heilig et al., 2011; Lesch et al., 1992; Nettleton et al., 1994; Thomas et al., 2009) and recharge and discharge areas (Hopkins and Richardson, 1999; Sherlock and McDonnell, 2003; Williams and Arunin, 1990). Fig. 4 shows spatial EC_a and interpreted salinity and sodicity data from a 1.5 ha field located in the Rolling Soft Shale Plain of southwestern North Dakota.² Within this site, soils are mapped as Janesburg–Dogtooth silt loams, 0 to 6% slopes. The moderately deep, well drained, Janesburg soils (fine, smectitic, frigid Typic Natrustolls) and the deep and very deep, moderately well and well drained Daglum soils (fine, smectitic, frigid Vertic Natrustolls) formed in clayey residuum. In Fig. 4, the plots on the left show the EC_a measured for two depth intervals with an EM38-MK2 meter. As evident in these plots, EC_a increases with increasing depth. The points shown on these plots are the sampling points identified and located using the ESAP software suite (Lesch et al., 2000). Based on limited sampling and stochastic equations contained in the ESAP program, levels of salinity and sodicity were predicted for each of the 1682 EC_a measurement points within this field. The presence of natric horizons and sodicity is recognized in the taxonomic classification and mapping of these soils. However, as shown in the right-hand plot of Fig. 4, for the 0 to 90 cm depth interval, saline non-sodic and saline-sodic conditions dominate (78%) this area, while non-saline and non-sodic soil conditions make up 22% of the site. The presence of these conditions should be recognized in this soil map unit.

Fig. 5 shows the spatial variability of EC_a across a 65 ha field that contains saline seeps in the *Brown Glaciated Plain* of north central Montana (Doolittle, 2013). In Fig. 5, the saline seeps are identified by their high EC_a ($>150 \text{ mS/m}$). These seeps appear to be arranged in a discontinuous, sinuous pattern that meanders across the field from

² Names for all Major Land Resource Areas taken from United States Department of Agriculture, Natural Resources Conservation Service (2006).

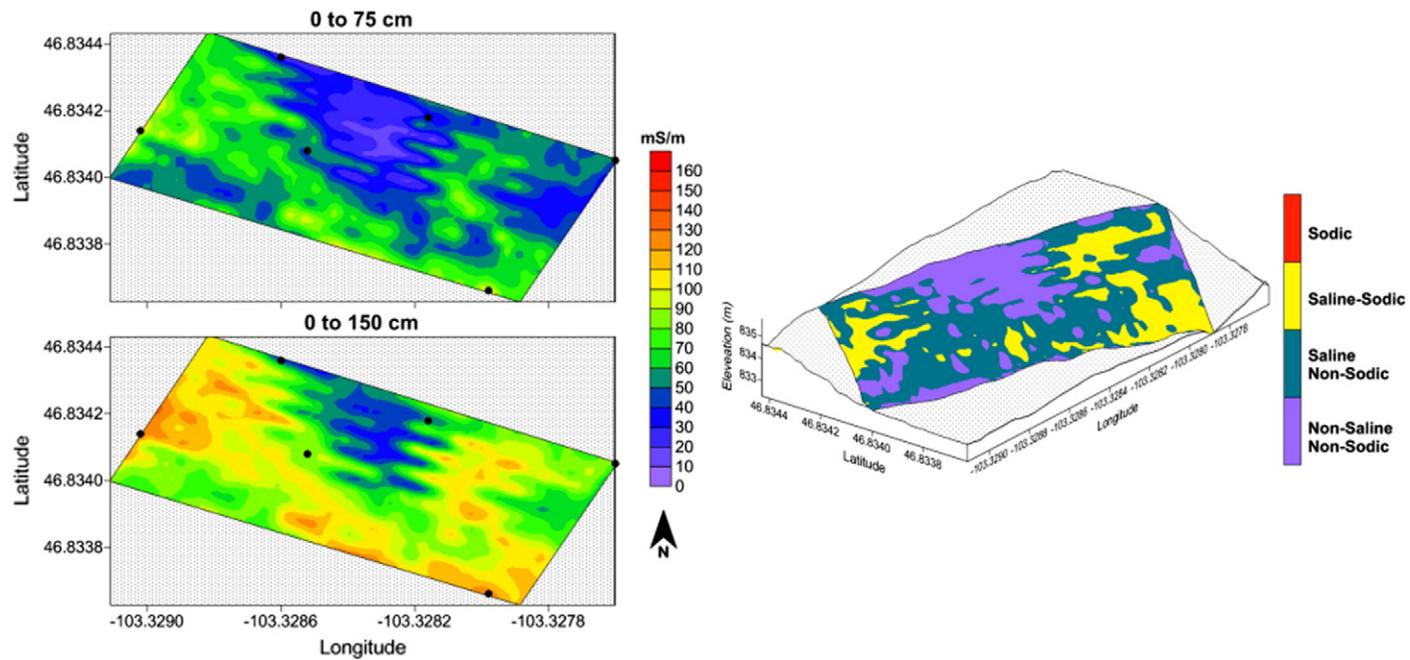


Fig. 4. These images show the spatial distribution of EC_3 for two depth intervals (left-hand plots) and classes of salt-affected soils (right-hand plot) for a field in southwestern North Dakota. The classes of salt-affected soils are based on both SAR and soil salinity levels for the 0 to 90 cm soil column as predicted by stochastic models of EC_3 (Heilig et al., 2011).

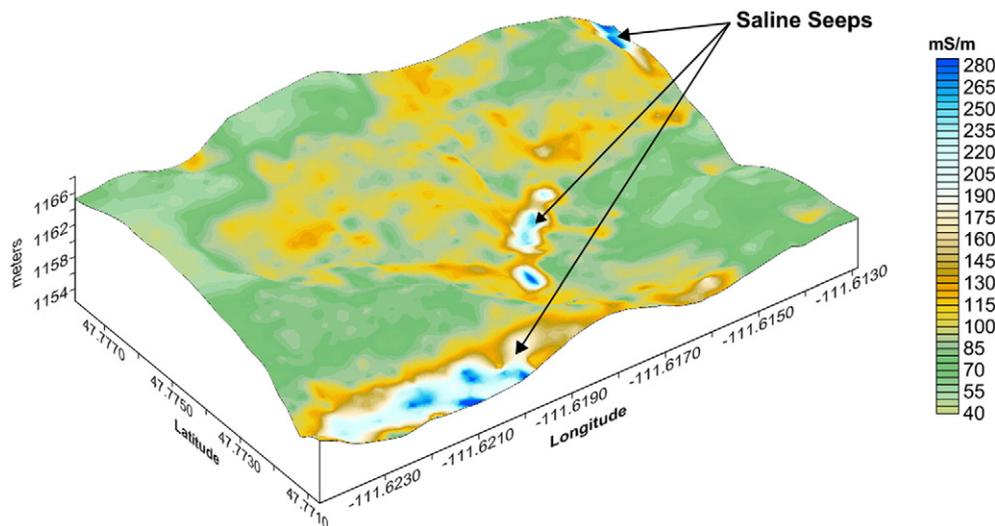


Fig. 5. This 3D simulation shows the spatial distribution of EC_a and the locations of saline seeps within an area of dryland farming in north central Montana (Doolittle, 2013).

the southwest to the northeast corner. Also evident on this plot are lines of moderate EC_a values that extend in a west-northwesterly and upslope direction away from these seeps. These delineations are believed to represent potential subsurface flow paths that have relatively higher concentrations of soluble salts. These lines may represent preferential channels for excess water to drain from recharge areas (located on higher-lying areas to the west and north). In Fig. 5, suspected recharge areas have relatively low EC_a values.

4.1.3. Other soil properties

Apparent conductivity has also been used as a surrogate measure of soil water content (Allred et al., 2005; Brevik et al., 2006; Hezarjaribi and Sourell, 2007; Huth and Poulton, 2007; Kachanoski et al., 1988, 1990; Khakural et al., 1998; Korsath et al., 2008; Mueller et al., 2003; Sheets and Hendrickx, 1995; Tromp-van Meerveld and McDonnell, 2009; Waive et al., 2000), soil texture (Heil and Schmidhalter, 2012; James et al., 2003; Saey et al., 2012a; White et al., 2012); and clay content (Cockx et al., 2009; Harvey and Morgan, 2009; King et al., 2005; Mueller et al., 2003; Sommer et al., 2003; Weller et al., 2007; Wienhold and Doran, 2008; Williams and Hoey, 1987). Electromagnetic induction has been used to assess difference in lithology and mineralogy (Bourgault and Rabenhorst, 2012; Doolittle et al., 2005, 2013), soil compaction (Al-Gaadi, 2012; Brevik and Fenton, 2004; Sudduth et al., 2010); CEC (Korsath et al., 2008; Triantafilis et al., 2009;), exchangeable Ca and Mg (McBride et al., 1990), $CaCO_3$ (Vitharana et al., 2008b); soil pH (Bianchini and Mallarino, 2002; Dunn and Beecher, 2007; Van Meirvenne et al., 2013; Vitharana et al., 2008b; Wienhold and Doran, 2008), soil organic carbon (Jaynes, 1996b; Johnson et al., 2001; Korsath et al., 2008; Martinez et al., 2009; Vitharana et al., 2008b), field-scale leaching rates of solutes (Slavich and Yang, 1990), herbicide partition coefficients (Jaynes et al., 1994), and available N (Eigenberg et al., 2002; Wienhold and Doran, 2008). In these studies, EC_a was either directly related to a soil property or the property (such as soil organic carbon) was associated with changes in a property (e.g., moisture and/or clay content) that affects EC_a .

4.2. Refine and improve the quality of soil maps

Electromagnetic induction has been increasingly used to support soil surveys and site-specific management (Brevik, 2012). At field and landscape scales, EC_a maps have the potential to provide higher levels of resolution and greater distinction of soil types than soil maps prepared with traditional tools and survey methods provided there is significant variation in at least one of the factors that affects soil EC_a (James et al.,

2003; Jaynes, 1995, 1996a; Shaner et al., 2008). However, EC_a surveys can also be used to confirm highly uniform soil properties throughout a field (Brevik et al., 2012).

Interpretations of EC_a maps are based on the identification of spatial patterns within data sets. Though seldom diagnostic in themselves, lateral and vertical variations in EC_a are used to infer changes in soil types and properties (Corwin, 2008; Daniels et al., 2003; Doolittle et al., 1994, 1996; Jaynes et al., 1993; Kravchenko et al., 2002; Sudduth et al., 1995). The effectiveness of EMI as a soil mapping tool will depend upon the degree to which differences in the physical and chemical properties that affect EC_a correspond to differences in soils. Where strong and meaningful relationships can be established between soils and EC_a , field-scale EC_a mapping has been used to identify areas of reasonably homogenous soils and soil properties (Doolittle et al., 1996; Frogbrook and Oliver, 2007; Johnson et al., 2001), and to improve existing soil maps (Doolittle et al., 2008; Hedley et al., 2004; Vitharana et al., 2008a). A major contribution of EMI surveys has been the identification and delineation of small included areas of dissimilar soils within soil polygons that have been mapped on second-order soil maps (Fenton and Lauterbach, 1999), although some common EMI sampling methods tend to represent soils that occupy a large percentage of the area within a field more representatively than soils that occupy small areas (Brevik, 2012).

In many areas, spatial EC_a patterns correspond well with soil patterns shown on soil maps. Fig. 6a and b show high-intensity EC_a maps for two fields located within the *Mississippi Alluvial Plain* of south-eastern Missouri. These maps were prepared from measurements made with an EM38 meter operated in the VDO (DOI of 1.5 m). On each map, soil names and map unit boundary lines have been imported from the Web Soil Survey.³ Although the same color ramp has been used, differences in scales make the EC_a not directly comparable between these maps.

Each of the surveyed fields shown in Fig. 6a and b has been land-leveled and agricultural drainage pipes have been installed to improve soil drainage and crop yields. The taxonomic classifications of the soils identified in these fields are listed in Table 1. Soil–landform relationships, which were once more obvious in these landscapes, have been obscured by land leveling. Land leveling has not only disturbed the soils, but has eliminated many “topographic breaks” that are used by soil scientists to identify soil boundaries and map soils using soil–

³ Soil Survey Staff, Natural Resources Conservation Service, United States Department of Agriculture. Web Soil Survey. Available online at <http://websoilsurvey.nrcs.usda.gov/>. Accessed [08/28/2013].

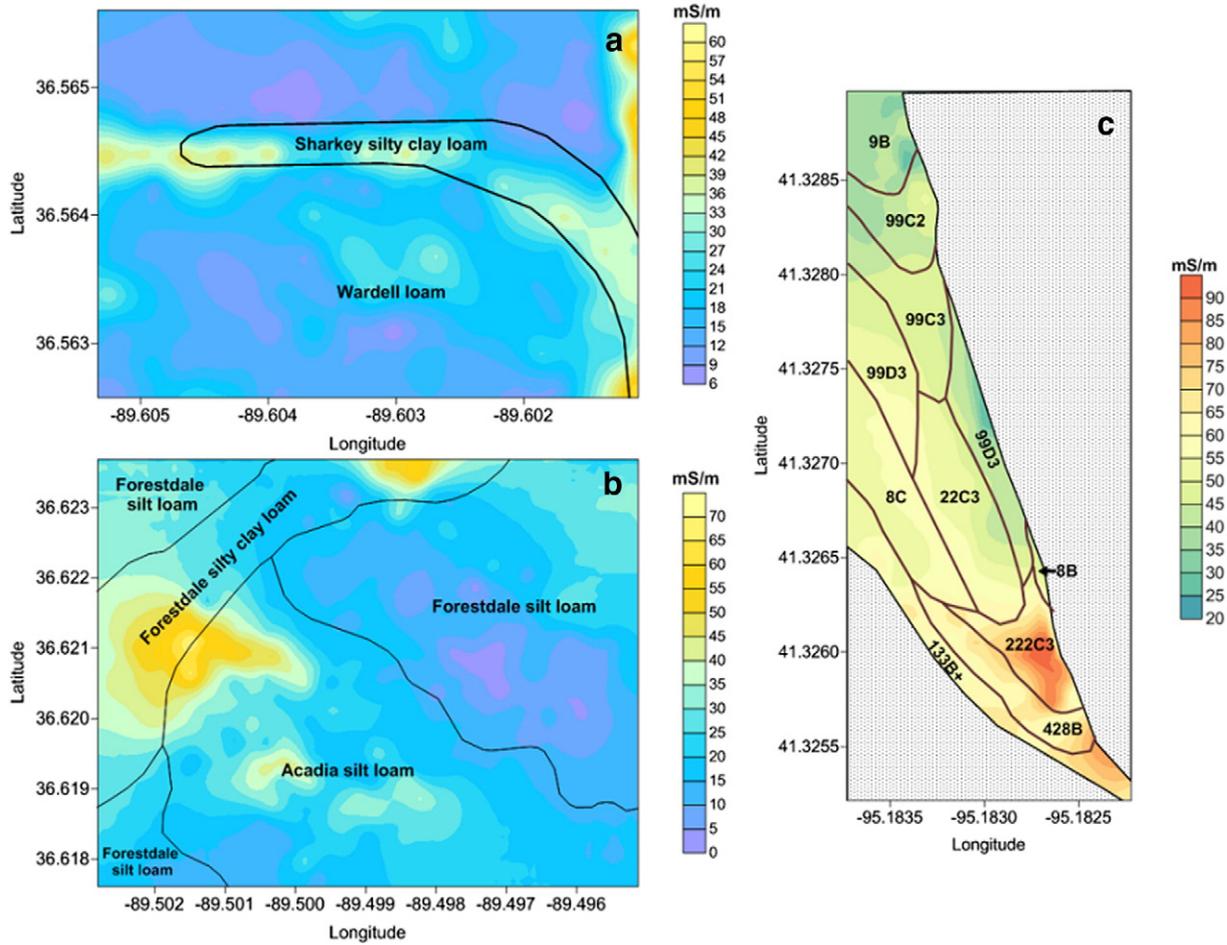


Fig. 6. a & b. These EC_a maps were prepared from data collected with an EM38 meter in southeastern Missouri. The soil lines have been imported from the Web Soil Survey. Each delineation is identified by its dominant soil(s) and surface texture (Doolittle et al., 2002). c. The soil EC_a map for a field in southwestern Iowa with the first-order soil survey overlaid on it. Note the high EC_a values in the area mapped as Clarinda (222C3), a paleosol-derived soil that can have as much as 60% clay in the upper 1.5 m (Brevik and Fenton, 2003).

landform relationships. Lacking noticeable slope breaks, the task of identifying and mapping the soils would be more time-consuming and laborious using traditional soil mapping field procedures (Brevik and Fenton, 1999) than with EMI.

Soils with higher clay contents have higher EC_a. In Fig. 6a and b, areas of clayey Arcadia, Forestdale, and Sharkey soils have the highest EC_a (>35 mS/m). In Fig. 6b, included areas of Wiville soils have the lowest EC_a (<6 mS/m). The Wiville soils formed on former dunes, and have lower clay contents and sandy C horizons. These former dunes are no longer present on this land-leveled field and relatively extensive areas of Wiville soil were overlooked in mapping. In Fig. 6a, areas of Wardell soils have intermediate clay contents and EC_a.

The delineation of soils with paleosol parent materials can be important for agricultural management in the *Deep Loess Hills* of southwestern Iowa due to the high clay-content of these soils. However, it is not always easy to pick out small areas of such soils during field mapping. Brevik and Fenton (2003) investigated the use of the EM38 to identify paleosol-derived soils and compared the EC_a patterns to first-order

soil surveys that had been prepared for several fields in southwestern Iowa. One of the resulting maps is shown in Fig. 6c. In this example, the EM38 did a good job of differentiating between paleosol-derived and other soils, with the paleosol-derived soils often having EC_a values in excess of 80 mS/m while other soils in these fields rarely exceeded 70 mS/m (92% of the EC_a readings were 70 mS/m or less). Taxonomic classification information for the dominant soils in the Iowa field is given in Table 2.

In Fig. 6c, the delineation of Clarinda soils (222C2) has clay contents of as much as 60% in the upper 1.5 m of the profile. The paleosol-derived Clarinda soil clearly stands out from the coarser-textured soils around it.

On each of the maps shown in Fig. 6, spatial EC_a patterns conform to the general soil patterns of the first- or second-order soil survey.

Table 1
Taxonomic classification of soils identified in Fig. 6a and b.

Soil series	Taxonomic classification
Arcadia	Fine, smectitic, thermic Aerie Epiaqualfs
Forestdale	Fine, smectitic, thermic Typic Endoaqualfs
Sharkey	Very-fine, smectitic, thermic Chromic Epiaquerts
Tuckerman	Fine-loamy, mixed, active, thermic Typic Endoaqualfs
Wardell	Fine-loamy, mixed, superactive, thermic Mollic Epiaqualfs
Wiville	Fine-loamy, siliceous, active, thermic Ultic Hapludalfs

Table 2
Taxonomic classification of soils identified in Fig. 6c.

Soil series	Soil map unit(s)	Taxonomic classification
Judson	8B and C	Fine-silty, mixed, superactive, mesic Cumulic Hapludolls
Marshall	9B	Fine-silty, mixed, superactive, mesic Typic Hapludolls
Dow	22C2	Fine-silty, mixed, superactive, calcareous, mesic Typic Udorthents
Exira	99C2, C3, and D2	Fine-silty, mixed, superactive, mesic Typic Hapludolls
Colo	133B	Fine-silty, mixed, superactive, mesic Cumulic Endoaquolls
Clarinda	222C2	Fine, smectitic, mesic Vertic Argiaquolls
Ely	428B	Fine-silty, mixed, superactive, mesic Aquic Cumulic Hapludolls

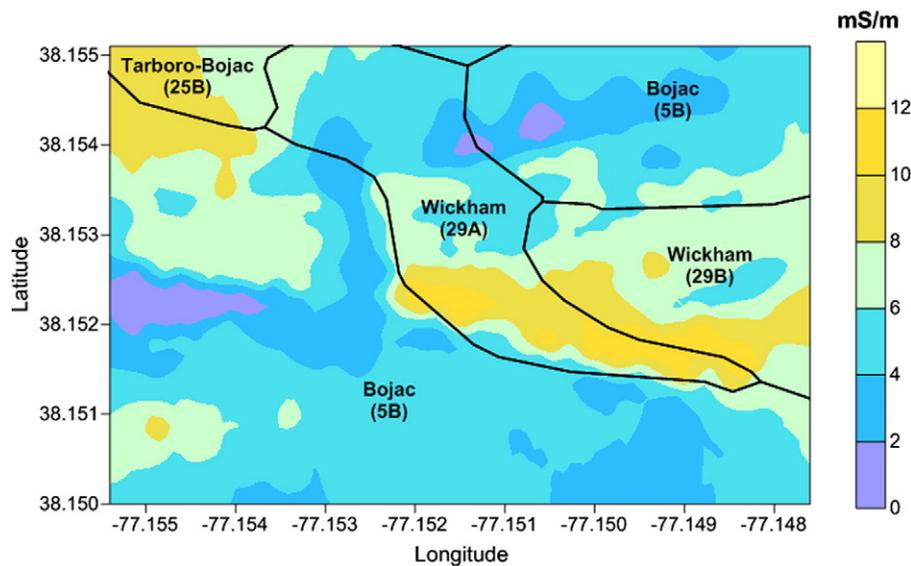


Fig. 7. This EC_a map was prepared from data collected with an EM38 meter on the Coastal Plain of Virginia. Soil delineations have been imported from the Web Soil Survey and are identified by their dominant soil(s) and map unit symbol (Anderson-Cook et al., 2002).

However, the more intensive EMI sampling has resulted in more intricate patterns and the depiction of greater spatial variability than displayed by the imported soil boundary lines alone. In these soil-landscapes spatial EC_a patterns are principally associated with differences in clay content. In the Missouri fields, the correlation between EC_a and clay content at different soil depths ranged from 0.545 to 0.903 (Doolittle et al., 2002a); while in the Iowa field, soils with the highest clay content as well as the highest smectite clay mineral content displayed the highest EC_a values.

Soil variability and the transition from one soil type to another are well expressed on the soil EC_a maps shown in Fig. 6. The imported map unit boundary lines have a fixed width and cannot accurately portray the spatial rate of change in soils and soil properties. This can result in prediction errors, especially in areas where fairly broad transitional zones exist. As evident on the maps shown in Fig. 6, spatial EC_a data can be used to improve the placement and representation of soil boundaries (Adamchuk et al., 2004; Greve and Greve, 2004; James et al., 2003). On these maps, the spatial rates of change in EC_a provide measures of transition zone widths and improve the precision of map unit positioning and the representation of soil variability (Greve and Greve, 2004; Kweon, 2012).

Though widely used in site-specific management, the use of EMI by state and federal agencies to create or refine soil maps has been very limited (Kitchen et al., 1998). However, in several documented studies, EC_a data were used to improve the quality of second-order soil maps (Harvey and Morgan, 2009; Lobell et al., 2010; Vitharana et al., 2008a). In addition, Saey et al. (2009b) prepared an EC_a map of Belgium's East Flanders province (3000 km²) that was based on 4887 topsoil samples. This map has been useful in evaluating field-measured EC_a patterns and measurements.

High-intensity or first-order soil mapping based on EMI is offered commercially in many countries to provide information on the distribution and variability of soils at field and landscape scales (Brevik and Fenton, 2003; Khakural et al., 1998; King et al., 2005; Korsaeath et al., 2008; Kravchenko, 2008; Weller et al., 2007). Recently, EMI has also been used to improve the quality of several first-order soil maps (Anderson-Cook et al., 2002; Doolittle et al., 2008, 2009; Farahani and Flynn, 2007; White et al., 2012). White et al. (2012) and Doolittle et al. (2008, 2009) used EC_a data with geographical information system (GIS) to improve first-order soil maps that were constructed using traditional methods on the *Gulf Coastal Plain* of Alabama and the *Till Plains Section* of Illinois,

respectively. In both physiographic areas, soil maps prepared with EMI and conventional, high-intensity soil survey methods produced similar results. In these studies, a significant contribution of spatial EC_a data was the increased confidence of soil scientists in their mapping decisions. The information provided by EC_a maps led soil scientists to reevaluate soil mapping decisions and conceptual soil landscape models, recognize different soils, and modify soil maps.

Anderson-Cook et al. (2002) used EC_a to distinguish soil types and significant differences in subsoil texture in an area dominated by very deep, well drained Bojac (coarse-loamy, mixed, semiactive, thermic Typic Hapludults) and Wickham (fine-loamy, mixed, semiactive, thermic Typic Hapludults) soils on the *Atlantic Coastal Plain* of Virginia (Fig. 7). Using the EC_a data, Anderson-Cook et al. (2002) were able to correctly classify the soil type with an accuracy of greater than 85%. Fig. 7 shows the distribution of EC_a across this study site in relationship to the second-order soil map, which was prepared using traditional soil survey methods. Compared with the second-order soil map, the EC_a map shows greater variability and more intricate spatial patterns.

In several field-scale and landscape-scale special soil research projects, EMI has been used to map small areas and assess the depths to argillic horizons, claypans, fragipans, hardpans, and petrocalcic horizons (Boettinger et al., 1997; Chen et al., 2000; Doolittle et al., 1994; Mueller et al., 2003; Saey et al., 2012a; Stroh et al., 1993; Sudduth and Kitchen, 1993; Sudduth et al., 1995, 2009), reconstruct buried landscapes (De Smedt et al., 2013a, 2013b; Saey et al., 2008, 2013) and periglacial features (Meerschman et al., 2011a, Saey et al., 2012a); estimate depths to bedrock (Bork et al., 1998; Doolittle et al., 1998, 2002b; Palacky and Stephens, 1990; Zalasiewicz et al., 1985), and assess differences in soil drainage (Kravchenko et al., 2002).

4.3. A tool for soil-hydrologic studies

Soil-hydrology relationships vary across landscapes and are often exceedingly complex. Characterizing these complex relationships at different spatial-temporal scales and assessing their impacts on subsurface flow and transport is a major challenge to hydrologic modelers. Soil-hydrology-landscape relationships have been traditionally measured and inferred from point-based pedologic observations. Point-sampling methods (such as soil pits, monitoring wells, core samples, and soil moisture probes) provide detailed, but highly site-specific soil and hydrologic data. As the collection of point data is time-consuming, labor-intensive, costly, and generally destructive (Brevik

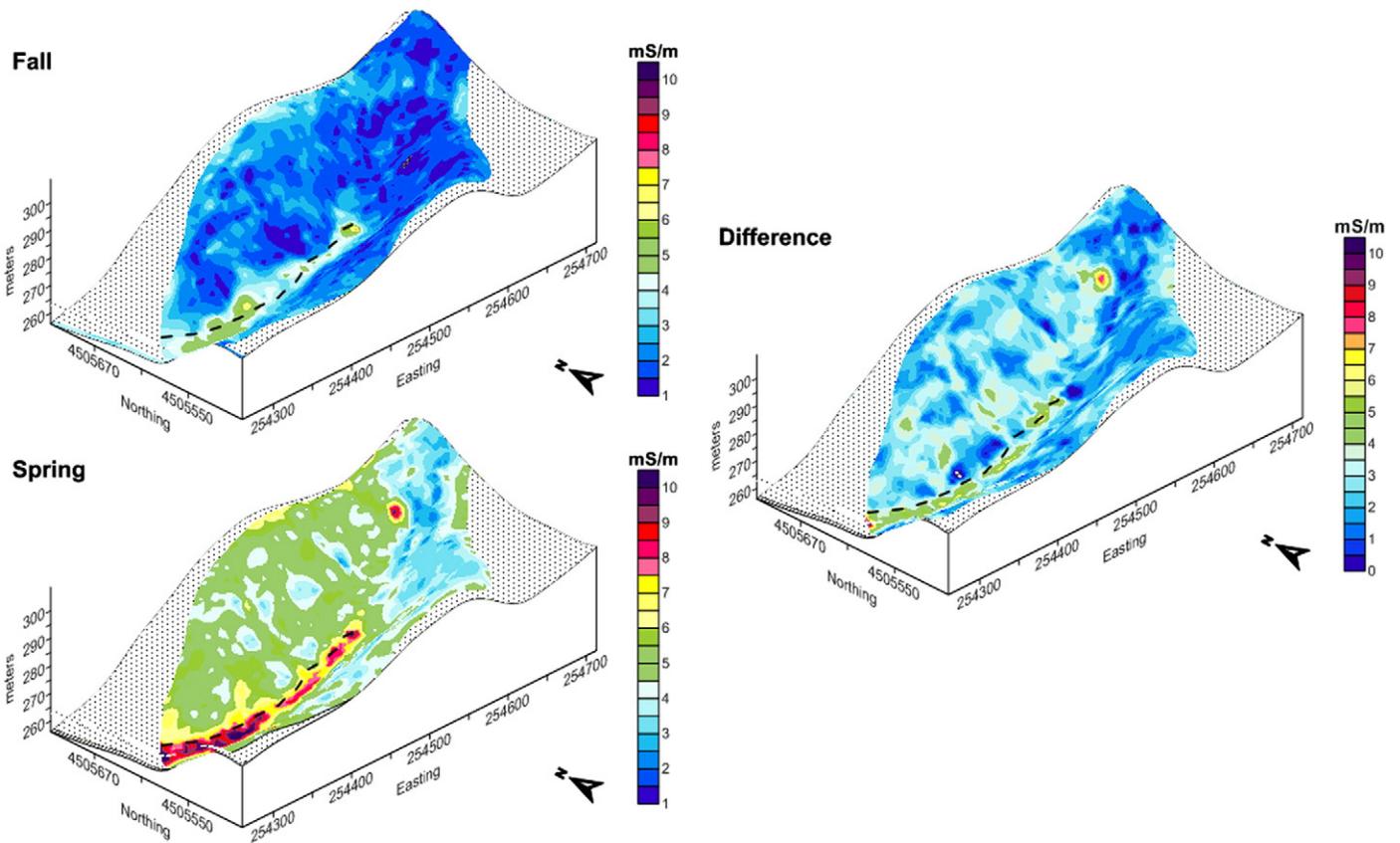


Fig. 8. Results of two time-lapsed EMI surveys that were conducted within a small catchment located in central Pennsylvania in the fall (dry conditions) and spring (wet conditions) using an EM38 meter with a density of about 750 measurements per ha (Doolittle et al., 2012).

and Batten, 2012; Brevik et al., 2003; Zhu et al., 2013), it is confined to a limited number of sampling points. Because of these limitations, soil and hydrologic properties and processes for the larger areas among the widely-spaced sampling points must be inferred.

Electromagnetic induction has been effectively used to reveal the complexity of soil-landscape architectures and their impact on subsurface flow at field, hillslope, and catchment scales, and to fill in data gaps caused by the limitations of point-sampling methods (Doolittle et al., 2012; Zhu et al., 2010a, 2010b). In hydrogeological studies, EMI has been used to indirectly measure and characterize soil water content, subsurface flow, depth to water table, and soil drainage classes (Allred et al., 2005; Doolittle et al., 2000; Kachanoski et al., 1990; Khakural et al., 1998; Kravchenko et al., 2002; Robinson et al., 2008; Scanlon et al., 1999; Schumann and Zaman, 2003; Sheets and Hendrickx, 1995; Williams et al., 2006; Zhu et al., 2010a, 2010b). Studies have revealed that the relative difference in soil EC_a across most landscapes remained relatively stable over time (Brevik et al., 2006; Zhu et al., 2010a). These relatively stable spatial EC_a patterns correspond to soil-landform units. In addition, changes in the magnitude and spatial extent of EC_a patterns over seasons within the same landscape indicated active zones of subsurface flow, which corresponded with simulated water flow paths and observed soil morphology.

In a study of a small (7.9-ha) catchment located in the *Northern Appalachian Ridges and Valleys* of central Pennsylvania, EC_a data were collected under relatively dry (fall) and wet (spring) conditions (Doolittle et al., 2012). The catchment is incised into a ridge composed of thinly bedded, highly fractured, and folded acid shale. Seven well-defined, linear swales of varying dimensions extend down slope and onto the valley floor; five along the south-facing slopes, and two along the north-facing slopes. Soils are dominantly shallow (0 to 51 cm) and moderately deep (51 to 102 cm) to bedrock on side slopes and summit areas. However, soils are very deep (>152 cm) along the valley floor.

The catchment is characterized by exceedingly low and relatively invariable EC_a (Fig. 8). Within this catchment, the very low EC_a reflects the electrically resistive nature of the soil and bedrock, and the low ionic concentration of the soil solution. Reconnaissance EMI surveys conducted in fall and spring months revealed that EC_a ranged from about 0 to 24 mS/m. However, over most of the catchment, EC_a did not vary by more than 4 mS/m. In spite of the low and relatively invariable EC_a , temporal differences in EC_a were observed in this landscape: higher and more variable EC_a data were collected in the wet spring than in the dry fall (Fig. 8; left-hand plots). On the EC_a difference map (Fig. 8; right-hand plot), the EC_a recorded in the fall has been subtracted from the EC_a recorded in the spring. On this map, EC_a increases throughout the catchment, but the change is most noticeable along the valley floor. Several weakly-expressed, linear patterns of relatively higher EC_a extend up slope from the stream channel and identify the general locations of swales.

In soils that have low salt contents, EC_a is strongly influenced by variations in clay and moisture contents (Brevik and Fenton, 2002; Carroll and Oliver, 2005; Johnson et al., 2001; Kachanoski et al., 1990). As a consequence, King et al. (2005) associated changes in EC_a with changes in soil type and hydrology. Although absolute EC_a values respond to temporal changes in soil moisture, most spatial EC_a patterns remain temporally stable (Johnson et al., 2001; King et al., 2005; Sudduth et al., 2000). Collectively, the spatial EC_a patterns shown in Fig. 8 suggest two major, temporally-stable soil-landscape units within the catchment: the valley floor and higher-lying slope components. Lower and less variable EC_a values were consistently recorded on the side slopes and summit areas where well-drained, shallow and moderately deep to shale bedrock soils are dominant; while higher and more variable EC_a values were measured along the valley floor where somewhat poorly-drained, very deep soils are dominant. For the higher-lying slope components, the south-facing slopes displayed a

greater variability in EC_a and are known to be more hydrologically active than the north-facing slopes. Along the valley floor, the persistently higher EC_a is attributed to higher clay content and wetter soil conditions.

5. Summary

Present EMI systems are suitable for use in soils investigations. In recent years, electromagnetic induction sensors have experienced a rapid succession of design improvements and have been successfully integrated with new technologies (e.g., field computers, PDAs, GPS receivers and DGPS technologies, Bluetooth) to become even more versatile and useful tools in soils research. The use of EMI to quickly and easily identify, characterize and map spatially-varying soil types and properties offers distinct advantages over traditional methods. This noninvasive geophysical tool can help facilitate the collection of large volumes of moderate to high resolution data, provide more comprehensive coverage of sites, and greater confidence in site assessments. However, results are site-specific and can vary depending on the complex interaction among multiple, interacting and variable soil properties. It is also important to note that EC_a readings are a composite of soil properties; they cannot replace the detail provided by sampling and describing soils in the field. For this reason, ground-truthing of EMI data will remain important (Brevik and Hartemink, 2010) and EMI technologies will not completely replace trained field specialists (Brevik et al., 2006). Instead, traditional soil sampling and EMI techniques can be used together to provide even more information about the soils at a given site than is possible using either approach alone. Even with these challenges, EC_a has been increasingly used to infer and map the spatial variability of soil properties at field and landscape scales.

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