

1	Title:
2	Key structural features of Boreal forests may be detected directly using L-moments from
3	airborne lidar data
4	
5	Authors:
6	Rubén Valbuena ^{*(1)} , Matti Maltamo ⁽¹⁾ , Lauri Mehtätalo ⁽²⁾ , Petteri Packalen ⁽¹⁾
7	
8	Affiliations:
9	(1) University of Eastern Finland, School of Forest Sciences, PO Box 111. Joensuu,
10	Finland; rubenval@uef.fi; matti.maltamo@uef.fi; petteri.packalen@uef.fi
11	(2) University of Eastern Finland. School of Computing. PO Box 111. Joensuu, Finland;
12	lauri.mehtatalo@uef.fi
13	*Corresponding author.
14	
15	
16	
17	
18	
19	
20	
21	
21	
22	

23 Abstract

24 This article introduces a novel methodology for automated classification of forest areas from airborne laser scanning (ALS) datasets based on two direct and simple rules: L-coefficient of 25 variation Lcv = 0.5 and L-skewness Lskew = 0, thresholds based on descriptors of the 26 mathematical properties of ALS height distributions. We observed that, while Lcv > 0.5 may 27 represent forests with large tree size inequality, Lskew > 0 can be an indicator for areas 28 29 lacking a closed dominant canopy. Lcv = 0.5 discriminated forests with trees of approximately equal sizes (even tree size classes) from those with large tree size inequality (uneven tree size 30 classes) with kappa $\kappa = 0.48$ and overall accuracy OA = 92.4%, while *Lskew* = 0 segregated 31 oligophotic and euphotic zones with $\kappa = 0.56$ and OA = 84.6%. We showed that a supervised 32 33 classification could only marginally improve some of these accuracy results. The rule-based approach presents a simple method for detecting structural properties key to tree competition 34 and potential for natural regeneration. The study was carried out with low-density datasets from 35 the national program on ALS surveying of Finland, which shows potential for replication with 36 the ALS datasets typically acquired at nation-wide scales. Since the presented method was 37 38 based on deductive mathematical rules for describing distributions, it stands out from inductive 39 supervised and unsupervised classification methods which are more commonly used in remote sensing. Therefore, it presents an opportunity for deducing physical relations which could 40 41 partly eliminate the need for supporting ALS applications with field plot data for training and 42 modelling, at least in Boreal forest ecosystems.

43 Key words

44 Airborne laser scanning; L-moments; Gini Coefficient; L-coefficient of variation; forest
45 structure; tree size inequality; shade-tolerance.

47 **1. Introduction**

Airborne laser scanning (ALS) can be a valuable tool for studying structural properties of 48 forests (Lefsky et al., 1999a; Drake et al., 2002; Frazer et al., 2005; Maltamo et al., 2005; 49 Valbuena et al., 2016a). The relationships of ALS to forest structure can be employed to analyse 50 asymmetric competition among trees (Kellner & Asner, 2009), and hence forest growth 51 conditions (Stark et al., 2010). In fully-stocked forests (Gove, 2004) light resource pre-emption 52 53 drives asymmetric competition processes, leading to mortality of the least competitive trees (Weiner, 1990). These are forests with closed canopies and structural properties yielding shady 54 areas, i.e. oligophotic zones (sensu Lefsky et al., 2002), under the dominant tree crowns. In 55 56 turn, detecting forest areas with light resource availability, which are characterized by large euphotic zones (sensu Lefsky et al., 2002), can be key to monitoring forest disturbance and 57 regeneration. Several metrics derived from ALS height distributions have potential for 58 59 describing these key characteristics related to forest structure (Zimble et al., 2003). For this reason, studies on ALS-based forest structure characterization by statistical inductive methods, 60 which relate ALS metrics to field attributes empirically, are commonplace (Hall et al., 2005; 61 Lefsky et al., 2005; Dalponte et al., 2008; Pascual et al., 2008; Disney et al., 2010; Jaskierniak 62 et al., 2011; Ozdemir & Donoghue, 2013; Valbuena et al., 2014). 63

Size hierarchy among trees growing in the vicinity influences competition processes in the forest community (Weiner, 1990; Valbuena et al., 2012). Knox et al. (1989) suggested the Gini coefficient (*GC*) (Gini, 1921) as a consistent descriptor of tree size inequality, and hence a reliable indicator of competition conditions in the forest (Cordonnier & Kunstler, 2015). For this reason, in the context of ALS estimation, the *GC* of tree sizes has been used as a basis for stratifying the forest area into homogeneous structural types (Bollandsås & Næsset, 2007; Valbuena et al., 2013a). Furthermore, Knox et al. (1989) also suggested the inclusion of skewness as a complement to the *GC* in describing forest structural properties. For this reason,
Valbuena et al. (2013a) included asymmetry in their analysis of forest structural properties, to
study relations of relative dominance between different strata in the forest vertical profile.

74 While Bollandsås & Næsset (2007) employed stand register data from previous inventories for carrying out their stratification, it would be advantageous if the same remote sensing material 75 could be used for wall-to-wall predictions of forest structure indicators and classifications into 76 forest structural types (Lefsky et al., 1999b; Drake et al., 2002). In particular, Ozdemir & 77 Donoghue (2013) and Valbuena et al. (2013b; 2016a) obtained predictions of the GC of tree 78 79 size inequality with reliable accuracy. As previous research has concentrated on the forest response (Lefsky et al., 1999a; Valbuena et al., 2013a), and on its analysis and estimation by a 80 wide range of different statistical methods – such as analysis of variance (Zimble et al., 2003), 81 82 canonical correlation (Lefsky et al., 2005), parametric (Hall et al., 2005) and non-parametric (Valbuena et al., 2014) modelling, histogram thresholding (Maltamo et al., 2005), or finite 83 mixtures (Jaskierniak et al., 2011) –, the next question to answer would be: do the ALS metrics 84 have, by themselves, capacity to discriminate among forest structural types, making no use of 85 statistical methods linking field data to ALS metrics?. 86

87 Moments are quantitative measurements of probability density distributions employed to summarize their properties. The most conventional are the product moments, expected values 88 of the powers of a random variable which lead to the use of mean, variance and skewness as 89 measures for location, scale and shape. These descriptors of ALS return height distributions 90 91 are metrics commonly employed as auxiliary variables in forest assessment (e.g., Næsset, 2002; 92 White et al., 2013). Alternatively, Frazer et al. (2011) and Ozdemir & Donoghue (2013) recently drew the attention towards the L-moments, a set of statistics known by their sample 93 efficiency (i.e., reliability at low sample sizes) and robustness to outliers, compared to 94

conventional moments (Hosking, 1990). Consider a sample order statistic $X_{k:r}$ – the k^{th} 95 smallest observation in a sample of size r –, which is a many-to-one transformation of a random 96 sample of size r, and therefore a random variable. The L-moments are based on its expected 97 values $E(X_{k:r})$ (Appendix A). Moreover, L-moment ratios have the advantage of being 98 bounded by finite intervals (Hosking 1989), making them comparable among ALS 99 distributions differing in their mean height. The L-coefficient of variation (Lcv) and the L-100 101 skewness (Lskew) are two types of L-moment ratios (Appendix A.2). Lcv is the ratio of the second (L2) to the first (L1) L-moments: 102

103 (1)
$$Lcv = \frac{L2}{L1} = \frac{E(X_{2:2}) - E(X_{1:2})}{2E(X)}$$

where E(X) is the expected value of X. In the case of ALS metrics, the variable X is the height of ALS returns. The *Lcv* is mathematically equivalent to the *GC* (Appendix A.3), and therefore the same properties apply to both of them. For instance, they are scale-invariant, and for positive random variables their values are bounded within the [0, 1] interval (Hosking, 1989). Also, Valbuena et al. (2012) showed that an asymptote at *GC* = 0.5 represents the case of maximum entropy among tree sizes in the forest. On the other hand, *Lskew* is the ratio of the third (*L*3) to the second (*L*2) L-moments:

111 (2)
$$Lskew = \frac{L_3}{L_2} = \frac{E(X_{3:3}) - 2E(X_{2:3}) + E(X_{1:3})}{E(X_{3:3}) - E(X_{1:3})}$$

In the case of *Lskew*, its theoretical bounds are [-1, 1] (Hosking, 1989). The value of *Lskew* = 0 corresponds to a symmetric distribution, while positive or negative values denote the type of asymmetry for the distribution of ALS heights. This article employs these mathematical properties of L-moments for describing ALS height distributions, in contrast to inductively researching explanatory potential in relation to field data attributes.

The aim of this research was to develop simple methods for explaining key features related to 117 forest structure from few L-moment ratios of ALS returns. Lcv and Lskew were used for 118 119 detecting tree size inequality and light availability, and they were utilized for an automated classification of forests from ALS datasets, which was applied directly without the use of field 120 data. The idea builds upon the hypothesis that two deductive mathematical rules, Lcv = 0.5121 and Lskew = 0, may be used to classify the forest area into two groups, based solely on the 122 ALS height distributions. We studied whether such classifications would be sound in terms of 123 124 explaining properties of size inequality among trees growing in vicinity (even or uneven tree sizes) and competitive conditions for light in the forest community (oligophotic or euphotic). 125 We compared the reliability of the rule-based method to results obtained from a supervised 126 127 classification. This article discusses suitable applications for this rule-based method.

128 2. Materials

129 2.1. Study area and ALS data

The research was conducted in a 252,000 ha study area including approximately 200,000 ha of 130 the Boreal forest ecosystems typically found in the region of North Karelia (Finland), which 131 consists of forests dominated by Scots pine (Pinus sylvestris L.) Norway spruce (Picea abies 132 (L.) Karst.) or Birch species (Betula ssp.) with various degrees of admixtures also with other 133 deciduous trees (such as Alnus ssp., Populus ssp. etc). The ALS data were acquired by Blom 134 Kartta Oy (Finland) during May 2012 with an ALS60 system from Leica Geosystems 135 (Switzerland). A flying height of 2,300 m above ground rendered an average density of 0.91 136 pulses per squared-meter. Country-wide laser data are being consistently acquired using 137 broadly similar parameters (National Land Survey of Finland; NLS, 2013). Methods may 138 therefore by consistently replicated throughout the country, bringing potential for upscaling the 139 results obtained at national-level. 140

Heights above ground for individual ALS returns were calculated by subtracting the digital 141 terrain model provided by the NLS. We considered that, as seedlings and saplings were 142 included in field mensuration (Valbuena et al., 2016b), their influence in laser pulse 143 interception had to be accounted for in ALS metric computation. Consequently, just a very 144 small height threshold of 0.1 m was used, only with the intention to mask out the influence of 145 the ground. Sample estimates of L-moments and their ratios (Wang, 1996) were computed from 146 147 the heights of all the ALS returns located within each cell over a regular grid covering the entire study area. The spatial resolution of this grid was $16 \text{ m} \times 16 \text{ m}$, a customary practice in Finland 148 149 that makes cell size roughly coincident in with the area of field plots operationally established and measured by Finnish Forest Centre (SMK, Suomen Metsäkeskus). 150

151 2.2. Field dataset used for validation

152 Field data for validation of the methods were partly acquired by University of Eastern Finland (UEF), partly provided by SMK. A total of N = 244 plots were acquired in a stratified random 153 sampling fashion with approximately equal per-stratum sample sizes (Valbuena et al., 2016b). 154 The strata employed were the forest development classes commonly used in operational 155 management in Finland (per-stratum sample sizes were n = 31, unless specified): Seedling, 156 Sapling, Young, Advanced, Mature, Shelterwood, Seed-tree (n = 29), and Multi-storied (n = 29)157 29). SMK's stand register data based on previous inventories was employed for the initial 158 randomization of field plot locations. Valbuena et al. (2016b) provides details about acquisition 159 protocol and processing of field data. Appendix B details the criteria used to assign a 160 development class to for each field plot, a task carried out independently by experienced SMK 161 personnel. 162

163 *3.* Methods

164 3.1. The rule-based method for stratifying forests based on ALS data

We used a deductive approach to thresholding using the L-moment ratios. The rules were deduced from their mathematical properties, as opposed to using inductive, supervised, datadriven optimization or classification:

- The value Lcv = 0.5 was used because it represents maximum entropy of tree sizes (Valbuena et al. 2012); also recall that Lcv = GC (see Appendix A.3). Since Lcvdescribes the relative dispersion of ALS heights, we postulated that Lcv could be used as descriptor for structural properties related to tree size inequality, and hypothesised that this threshold could be suitable for discriminating forests with trees of approximately equal sizes – even tree sizes – (Lcv < 0.5) from those with high tree size inequality – uneven tree sizes – (Lcv > 0.5).
- The value of *Lskew* = 0 was chosen because it represents a symmetric distribution of
 ALS heights, and distinguishes plots with positive or negative skewness (Hosking,
 1989). Being a descriptor of asymmetry, we postulated that *Lskew* could be used as
 descriptor for structural properties related to competitive dominance and light
 availability characteristics (Valbuena et al., 2013a), and hypothesised that this threshold
 could be useful for discriminating oligophotic zones (*Lskew* < 0) from euphotic ones
 (*Lskew* > 0).

We classified forests throughout the scanned area according to these rules directly, avoiding the use of field data in the training stage of the classification. The capacity of these rules to describe structural features of the forest was validated by comparing the classifications at field plot locations to the known development classes determined at the field plots. For that purpose, the development classes were aggregated into the target forest structural properties: even/uneven tree sizes and oligophotic/euphotic.

188 *3.2. Aggregation of development classes*

With the intention to study the hypothesised relationship between these thresholds of L-189 moment ratios for ALS height distribution and their related structural properties of forests, we 190 aggregated the forest development classes according to their structural properties. In even-aged 191 silviculture, the succession of development classes usually follows this a basic chronosequence 192 of even-sized forest types: Seedling, Sapling, Young, Advanced and Mature stands. Silviculture 193 based on natural regeneration yields more complex uneven-sized structural types: Shelterwood, 194 195 Seed-tree, and Multi-storied stands. In Finland, Shelterwood stands are forest areas attaining regeneration of shade-tolerant species under the shade casted by a closed dominant Mature 196 197 canopy (Appendix B). This is the oligophotic zone (Lefsky et al., 2002), which in the context of Eurasian Boreal forests corresponds to regeneration areas for Norway spruce (note: there are 198 many different types of shelterwood management systems and, although in Finland this term 199 200 is used specifically for shade-tolerant regeneration – Appendix B –, in other countries it may refer to regeneration of shade-intolerant species too, e.g. Valbuena et al., 2013a). Other 201 oligophotic areas are those which have reached the stem exclusion stage - Young, Advanced 202 and *Mature* stands –, limiting light availability under the dominant canopy (Zenner, 2005). On 203 the other hand, Seed-tree stands are areas where few parent trees provide seeds for natural 204 regeneration which recruits in the understorey generating *Multi-storied* stands (Appendix B). 205 These, as well as *Seedling* and *Sapling* stands, belong to the euphotic zone (Lefsky et al., 2002), 206 where the absence of a closed dominant canopy brings enough light to the ground as to allow 207 the growth of shade-intolerant species. Accordingly, to test the capacity of the Lcv = 0.5 and 208 Lskew = 0 rules to discriminate forest areas according to their respective hypotheses, the 209 210 development classes where aggregated as:

211 (1) First criterion. Inequality among tree sizes (Lcv = 0.5):

Even tree size forest structural types: *Seedling*, *Sapling*, *Young*, *Advanced* and *Mature* 212 • 213 stands. Characterized by low relative dispersion in tree sizes (Valbuena et al., 2013a). Uneven tree size forest structural types: Shelterwood, Seed-tree and Multi-storied 214 • stands. Characterized by high relative dispersion in tree sizes (Valbuena et al., 2013a). 215 (2) Second criterion. Relative dominance of overstorey over the understorey (Lskew = 0): 216 • Oligophotic (forest structural types with a closed dominant canopy not allowing shade-217 intolerant regeneration): Young, Advanced, Mature and Shelterwood stands. 218 219 Characterized by negative asymmetries (Valbuena et al., 2013a). Euphotic (forest structural types with canopy openness allowing shade-intolerant 220 regeneration): Seedling, Sapling, Seed-tree, and Multi-storied stands. Characterized by 221 positive asymmetries (Valbuena et al., 2013a). 222

223 3.3. Comparison against supervised classification

In order to compare the rule-based method with more common data-driven methodologies 224 based on inductive statistical inference, we contrasted the results against those obtained by a 225 226 supervised classification. For that purpose, we employed the results obtained in Valbuena et al. (2016b) from a support vector machine (SVM) classification which employed the same field 227 plot dataset at the training stage as the one used for accuracy assessment in the present study. 228 SVM is becoming increasingly popular for classification of ALS data (Dalponte et al., 2008; 229 230 García et al., 2011), since it is suitable for operating with big datasets and complex relationships of covariance. SVM is a hard classifier which calculates hyperplanes between classes under a 231 cost function defined as a combination of maximizing distances from training samples to the 232 hyperplanes while minimizing the error of misclassified samples. Using package e1071 in R 233 234 statistical environment (Meyer et al., 2014a) and a SVM C-classification method, Valbuena et al. (2016b) computed predictions of all the above-mentioned development classes separately
which, in the present study, we aggregated into the established criteria: inequality (even and
uneven tree size classes) and dominance (oligophotic and euphotic), as detailed above. It may
be worth noting that, in contrast to the rule-based method which avoided the training stage, the
SMV predictions were obtained by an error minimization method using field data support and
the explanatory capacity of many more ALS metrics (Valbuena et al., 2016b: Table 2).

241 *3.4. Accuracy assessment.*

Field data plots were only used for assessing the accuracy of the rule-based method. 242 Relationships among L-moments of ALS heights were observed in scatterplots which depicted 243 the development class to which each plot belonged, observing the role of different development 244 classes in these relationships. Development classes were grouped as described above, and the 245 246 capacity of the Lcv = 0.5 and Lskew = 0 rules to describe those grouping characteristics was assessed with the help of contingency matrices. The degree of misclassification was evaluated 247 by the final overall accuracy (OA) and per-class user's (UA) and producer's (PA) accuracies, 248 which were all calculated following Olofsson et al.'s (2013) estimators for stratified random 249 sampling as: 250

251 (3)
$$OA = \sum p_{ii};$$

252 (4) $UA = \frac{p_{ii}}{p_{i}};$

253 (5)
$$PA = \frac{p_{jj}}{p_{\cdot j}},$$

calculated from the proportions of the total area for each predicted (i) and observed (j) class. Given the stratified random sampling design, and to adjust the accuracy estimates to account for the unequal sampling intensities for each class, these proportions were weighted according to the share of area for each class (A_j) with respect to the total (A_t) (Olofsson et al., 2013), as observed from the SMK's stand register dataset employed in the initial stratified random sampling (Appendix B):

$$260 \quad (6) \qquad \qquad p_{ij} = \frac{A_j}{A_t} \frac{n_{ij}}{N},$$

261 where n_{ij} was the number of plots observed for class j and predicted to be class i, and N the total number of plots. Similarly, Cohen's (1960) kappa coefficient (κ) was also calculated from 262 these weighted proportions p_{ij} , employing the sample estimator for stratified random sampling 263 suggested by Stehman (1996). Routines implemented in R-packages vcd (Meyer et al., 2014b) 264 265 and diffeR (Pontius & Santacruz, 2015) were employed for these tasks. Results were compared 266 with those resulting from grouping supervised SVM predictions, which were obtained in a leave-one-out fashion (Valbuena et al., 2016b). It is worth stating that the study design 267 complied with Westfall et al.'s (2011) recommendations for stratified estimation. 268

269 **4. Results**

270 *4.1. L-coefficient of variation of ALS heights*

First, we studied the relation between the *Lcv* of ALS heights and the forest development classes observed at field plots. From Eq. (1), the rule Lcv = 1/2 can be represented in the $L2 \sim$ *L*1 relation (dashed line in Fig. 1) as:

274 (7)
$$L2 = \frac{L1}{2}$$
.



276

Figure 1. Relationship between the first and the second L-moments of ALS heights (i.e, L-coefficient of variation).

The Lcv = 0.5 threshold in Eq. (7) is depicted in Fig. 1 with a dashed line. Thus, Fig. 1 shows how the different forest development classes distribute themselves at either side of this threshold, using ALS metrics only. We observed that *Seed-tree* and *Multi-storied* stands, which usually present large values of relative dispersion in tree sizes (*GC* > 0.5), also had wide

dispersion in their ALS returns being mainly above the threshold at Lcv > 0.5 as well. This 284 rule, however, failed to identify forest areas with regeneration of shade-tolerant species 285 286 recruited in the understorey under a closed dominant canopy. These correspond mainly to the Shelterwood development class, which fell largely under Lcv < 0.5. Fig. 1 shows that 287 Shelterwood areas were difficult to discriminate from Mature forests, and hence they were 288 likely to be misclassified by this rule as being even tree size forest types. Fig. 1 also shows the 289 lack of independence of L2 from L1, since the spread of L2 values is larger for increasing L1. 290 This demonstrates the advantage of the *Lcv* ratio, which normalizes the values of dispersion in 291 L2, making them comparable among distributions differing in the mean ALS height (L1, see 292 Eq. A3 in Appendix A). 293

294 Concerning the classification results, using the Lcv = 0.5 rule for discriminating even tree size (Seedling, Sapling, Young, Advanced and Mature) versus uneven tree size classes 295 296 (Shelterwood, Seed-tree and Multi-storied) (Table 1), obtained an overall accuracy of 92.4% and a coefficient of agreement $\kappa = 0.48$. A total of 92.7% of the even-sized plots were correctly 297 classified by this rule, with only few omission/commision errors. Most uncertainty was on the 298 identification of uneven tree size forests, due to the inability for the Lcv = 0.5 rule to identify 299 Shelterwood areas (Fig. 1), as this rule only classified 24.4% of those areas as being uneven-300 301 sized.

302

303

304

Table 1. Direct rule Lcv = 0.5. Contingency matrix of classification of even-sized versus uneven-sized development classes.

Observed			
Predicted	even-sized	uneven-sized	Totals
even-sized	139	48	187
uneven-sized	11	46	57
Totals	150	94	244

308

309 4.2. L-skewness of ALS heights

The next step was to observe the capacity of *Lskew* to incorporate additional information about forest structure with regards to the relationships of relative dominance among the trees. Using the rule *Lskew* = 0 in Eq. (2) gives

313 (8) *L*3=0.

Therefore the rule is demonstrated directly by the zero value on the y-axis of the $L3 \sim L2$ relation (horizontal dashed line in Fig. 2). In Fig. 2, we also observed a strong dependency of L3 on L2, since the spread of L3 values expands while L2 increases. This also illustrates the advantages of the *Lskew* ratio, which normalizes the L3 values of asymmetry, making them comparable among distributions of differing dispersion of ALS heights (hence, of different mean ALS height as well).



321

Figure 2. Relationship between the second and third L-moments of ALS heights (i.e., L-skewness).

The utility of analysing the asymmetry of the ALS height distributions was clear, as *Lskew* was associated with the capacity of penetration of the laser pulses, and therefore with the openness of the canopy. Positive skewness (*Lskew* > 0) was observed when there were large proportions of ALS returns with relatively lower heights, which indicates few dominant trees allow the laser beam to reach lower areas underneath an open upper canopy. On the other hand,

negative skewness (Lskew < 0) was observed when a closed dominant canopy backscatters most returns from the higher strata, and only few of them are returned from the understorey.

Regarding the discrimination of oligophotic (*Young*, *Advanced Mature* and *Shelterwood*,) and euphotic (*Seedling*, *Sapling*, *Seed-tree* and *Multi-storied*) areas of the forest (Table 2), the overall accuracy obtained was 84.6% and $\kappa = 0.56$. These accuracies were quite large, considering a method making no use of field data, an indication that *Lskew* may be a good proxy for the degree of canopy closure.

337

Table 2. Direct rule *Lskew* = 0. Contingency matrix of classification of oligophotic (closed
canopies) versus euphotic (open canopies) areas.

Observed			
Predicted	oligophotic	Euphotic	Totals
oligophotic	102	17	119
euphotic	19	106	125
Totals	121	123	244

340

341 4.3. Comparing rule-based versus supervised method

Figure 3 shows a joint representation of both rules: Lcv = 0.5 and Lskew = 0, respectively represented by vertical dotted and horizontal dashed lines. It therefore illustrates how these measures of relative dispersion and asymmetry may be selected or combined in pursue of different objectives for classifying forest structure and development directly from the distribution of ALS returns. Furthermore, we also compared all results with those obtained by a supervised classification carried out with this same subsample dataset. Tables 3 and 4 are contingency matrices for the aggregation of development classes (according to section 3.2)
predicted by the supervised SVM classification. For direct comparison, Table 5 includes a
summary of results obtained by all the compared methods.

351



L-skewness vs. L-coefficient of variation

352

Figure 3. Relationship between the L-coefficient of variation and L-skewness of ALS heights.

	Observed		
Predicted	even-sized	uneven-sized	Totals
even-sized	131	15	146
uneven-sized	19	79	98
Totals	150	94	244

Table 3. Supervised classification. Aggregated classes from Valbuena et al. (2016b).
Contingency matrix of classification of even-sized versus uneven-sized development classes.

Table 4. Supervised classification. Aggregated classes from Valbuena et al. (2016b).
Contingency matrix of classification of oligophotic (closed canopies) versus euphotic (open canopies) areas.

	Observed			
	Predicted	oligophotic	Euphotic	Totals
	oligophotic	114	10	124
	euphotic	7	113	120
	Totals	121	123	244
361				
362				
363				
364				
365				
366				

367	Table 5.	Comparison	of accuracy	results.
			2	

	Rule-based	Supervised
Stratification	classification	classification*
Even vs. Uneven Tree Size	Lcv = 0.5	SVM
Overall accuracy (<i>OA</i>)	92.4%	87.3%
kappa (κ)	0.48	0.34
Even tree size omission (PA)	92.7%	87.3%
Even tree size commission (UA)	99.6%	99.8%
Uneven tree size omission (PA)	48.9%	84.0%
Uneven tree size commission (UA)	4.2%	4.1%
Oligophotic vs. Euphotic	Lskew = 0	SVM
Overall accuracy (<i>OA</i>)	84.6%	93.8%
kappa (κ)	0.56	0.80
Oligophotic omission (PA)	84.3%	94.2%
Oligophotic commission (UA)	96.8%	98.3%
Euphotic omission (PA)	86.2%	91.9%
Euphotic commission (UA)	52.9%	76.8%

³⁶⁸ *aggregated from Valbuena et al. (2016b).

369

Regarding the results obtained from the supervised classification, it can be observed that the classification of forest areas into even and uneven tree sizes (Table 3) reached an overall accuracy 87.3% and $\kappa = 0.34$, whereas oligophotic versus euphotic (Table 4) obtained overall accuracy of 93.8% and $\kappa = 0.80$. Differences between the rule-based method and the supervised approach were not so large if taking into account the simplicity and lack of involvement of

field data in the former one. User's accuracies obtained by the SVM classification were very 375 similar to those yielded by the rule-based method (Table 5), which demonstrates that they are 376 377 mainly due to differences in the proportions of area that each development class has from the population, and not differences between the two methods. The success of the Lcv = 0.5378 379 threshold in classifying the even and uneven tree size forests and Lskew = 0 for segregating the oligophotic and euphotic areas of forest was remarkably good if compared to the supervised 380 classification, which did not obtain much greater accuracies. The comparison of user's and 381 382 producer's accuracies against the supervised classification however highlighted the two major differences: the rule-based method increased the errors due to omission of uneven-sized areas 383 and commission of euphotic areas (Table 5). 384

385 **5. Discussion**

386 5.1. L-coefficient of variation may identify tree size inequality

Our prior presumption was that forests with trees of approximately equal sizes – i.e., even tree 387 size classes –, since they would backscatter most ALS returns from a single canopy stratum, 388 389 could be directly detected by low values of the Lcv of their ALS heights. Our results 390 corroborate this presumption, since 92.7% of the even tree size plots were correctly classified by this rule (blue colour in Fig. 4 examples). Fig. 3 shows that most uncertainty in even tree 391 size areas – those containing trees of approximately equal sizes – was due to Sapling stands, 392 whereas not one single plot belonging to either Advanced or Mature development classes 393 showed values of Lcv > 0.5. The low rate of omission errors implies that this rule could be 394 used as a rather conservative and simple method when the purpose is to predict even tree size 395 forest areas. 396



Figure 4. Examples of resulting maps of forests stratified with rule-based method. Left: canopy height model (CHM). Middle: areas with Lcv > 0.5 in yellow (uneven tree sizes) and Lcv <0.5 in blue (even tree sizes). Right: areas with Lskew > 0 in yellow (euphotic) and Lskew <0 in blue (oligophotic). The reference CHM was made from the same ALS dataset, courtesy of Aki Suvanto (Blom Kartta Oy).

404

On the other hand, it was also expected that in the presence of structurally heterogeneous forests 405 with more inequality of sizes among its trees, the ALS returns would also show a more spread 406 pattern as they backscatter along the full vertical profile of the canopy, showing higher values 407 of Lcv. In view of our results, that was the case for Seed-tree and most Multi-storied areas, 408 although not for *Shelterwood* stands. We therefore propose that the direct rule Lcv > 0.5 may 409 410 be used as an indicator of great tree size inequality only when regeneration is achieved by shade-intolerant species, and therefore it has been enabled by forest disturbance (Knox et al., 411 412 1989; Kellner & Asner, 2009). In other words, a correspondence between the GC of tree sizes (Valbuena et al. 2013a) and the *Lcv* of ALS heights may only happen when the large value of 413 414 GC is due to the presence of a gap in the canopy, which allows a large proportion of the laser 415 footprint to get through and disperse its corresponding returns along the vertical profile of the canopy (Stark et al., 2012). This highlighted the importance of employing an additional metric 416 discriminating areas with a large euphotic zone from those where regeneration occurs in the 417 oligophotic zone (Lefsky et al., 2002; Fig. 5). Whether or not more ALS metrics are required 418 419 for fully describing the structural properties of forests, it is worth noting the recurrence of *Lcv* as a variable selected by many different automated methods tested in our previous studies, and 420 therefore the role of Lcv in predicting structural attributes related to tree size inequality 421

- 422 (Valbuena et al., 2013b; 2014; 2016a) and forest development (Valbuena et al., 2013a; 2016b)
- 423 seems clear.

424



425

Figure 5. Schematic diagram representing the patterns of ALS return distribution that can be
found in different types of forest structures, and how they are described by ratios of L-moments:
L-coefficient of variation and L-skewness. Compare to Fig. 3 and Valbuena et al. (2013a: Fig.
429 4).

Exploring the reasons why only 24.4% of *Shelterwood* stands were classified by the Lcv > 0.5431 rule as being uneven-sized, it could be taken into account that this development class was also 432 433 the one showing most error in the SVM classification (Valbuena et al., 2016b). The fact that a supervised method, which used the explanatory potential of many other metrics as well, still 434 435 failed to reliably identify *Shelterwood* areas may be an indication that the limitation is due not 436 to the metrics but rather to the original ALS data. Due to the low-density nature of this national dataset (NLS, 2013), the laser footprint probably detects very infrequently the presence of 437 understory under closed dominant canopies. In that case, scan density would need to be 438 439 increased for this task. We considered the advantages of testing the rule-based method with this type of ALS dataset since, due to its simplicity, could have potential for replication at 440 national scales. Further research should, however, employ datasets of larger densities to clarify 441 whether Lcv could then show better capacity for detecting regeneration of shade-tolerant 442 species. If direct replication of the rule-based method is to be envisaged, the effect of other 443 flight parameters in these L-moment ratios, such as scanner device or maximum scanning angle 444 (Næsset, 2004; Disney et al., 2010), should also be object of future investigations. 445

446 5.2. L-skewness may identify fully closed canopies

The threshold derived from the asymmetry measure of L-moments, Lskew = 0, was 447 448 demonstrably practical with regards to discriminating oligophotic from euphotic areas. *Lskew* < 0 denotes areas where most ALS returns were backscattered from a closed dominant 449 450 canopy which only allows small proportions of the laser footprint – and the light resource – to reach the understorey. Conversely, Lskew > 0 was observed whenever there were large 451 proportions of ALS heights with relatively lower heights, and it was therefore related to the 452 presence of only few returns backscattered from upper areas in the canopy, which indicates 453 that the dominant trees allow the laser beam – and thereby the light resource – to reach lower 454

areas underneath an open canopy. This can be relevant with regards to findings by Drake et al.
(2002) and Lefsky et al. (2005), who found the degree of canopy closure to be one of the most
relevant covariates in the relation between biomass and ALS heights.

It may be worth noting that the Lskew > 0 rule was capable for practically delineating 458 Seedling, Sampling and Seed-tree stands directly (Fig. 4). Although the method was carried out 459 at pixel-level, the resulting maps identified entire stands sharply. The rule-based stratification 460 by Lskew > 0 was therefore fairly insensitive to the within-stand variation that usually makes 461 difficult to discriminate stands, especially *Seed-tree* areas, by standard area-based procedures 462 463 in remote sensing. These type of problems usually require more complex analyses at objectlevel – representing stands –, which involve segmentation procedures with subjective steps, 464 parameters determined by trial-and-error, or manual delineation (e.g., Pascual et al., 2008). In 465 466 contrast, the rule based method offers a simple procedure to determine Seedling, Sampling and Seed-tree stands directly. 467

468 *5.3. Synergies between the rules*

469 Overall accuracies obtained by the rule-based methods were, respectively, 92.4% and 84.6% which we considered a remarkable achievement for a rule-based method not requiring field 470 support for training and that they were comparable to the results obtained by the supervised 471 classification (87.3% and 93.8%, respectively; Table 5). As a rule of thumb, it may be affirmed 472 473 that Lskew > 0 characterizes canopies not fully closed (areas not having reached stem exclusion), whereas those areas which also had values of Lcv > 0.5 presented high inequality 474 among tree sizes driven by forest disturbance (Fig. 5). In our results in Fig. 3, values of wide 475 dispersion Lcv > 0.5 occurred only in the presence of positive skewness Lskew > 0. This was 476 also corroborated out of the sample, as pixels with Lcv > 0.5 also had Lskew > 0 as well (Fig. 477 4). This demonstrates that, in these low-density datasets, the variance of ALS heights only 478

increases as a cause of openness in the canopy and an increase of the euphotic zone (Lefsky et 479 al., 2002), possibly due to forest disturbance, which leads to positive skewness in the 480 481 distribution. As a consequence, the maps obtained with Lcv > 0.5 were expanded by the Lskew > 0 rule (Fig. 4), extending the areas of large tree size inequality towards those simply 482 presenting potential for growth with no limitation from light resource. In turn, negatively 483 skewed Lskew < 0 ALS height distributions (Fig. 2) are indicative of forests with large 484 oligophotic zone (Lefsky et al., 2002) and therefore can only allow the regeneration of shade-485 486 tolerant species. It is worth commenting that uneven tree size and euphotic forest areas stand out of a general relationship between first moments of ALS heights and forest attributes related 487 to mean diameter (Lefsky et al., 2002, 2005), and therefore we suggest that one potential use 488 489 of the rule-based method could to decrease the signal-to-noise ratio when obtaining ALSassisted estimations in heterogeneous forest areas. 490

491 *5.4. Practical benefits and further research needs*

In this article, we applied deductive science (Appendix A) to infer that L-moments from the 492 distribution of ALS returns can have a direct relationship to forest structural characteristics at 493 the community level, namely tree size inequality and canopy closure (Fig. 5), in addition to the 494 already well-known fact that ALS height relates to tree height (e.g., Lefsky et al., 2005; 495 496 Maltamo et al., 2005; Miura & Jones, 2010). The main benefit of these research findings is on increasing our understanding (Fig. 5) of how ALS explains key structural features related to 497 forest structure (Gove, 2004; Valbuena et al., 2012) and tree competition (Weiner, 1990; 498 499 Cordonnier & Kunstler, 2015). These can be relevant to enhance the potential of ALS for describing light availability conditions (Lefsky et al., 2002), forest disturbance characteristics 500 (Kellner & Asner, 2009), or tree growth (Stark et al., 2010) and regeneration (Valbuena et al., 501 502 2013a). Further research should clarify the role of different flight configurations, scanners

systems or scanning density (Næsset, 2004; Disney et al., 2010) in the relationships between
ALS L-moments and forest structural characteristics.

505 The resulting classification could be used e.g. in stratification of a forest area for the field data collection of an ALS inventory campaign, since Hawbaker et al. (2009), Maltamo et al. (2011) 506 507 and Gobakken et al. (2013) demonstrated that a field sampling strategy based on a priori knowledge extracted from the ALS itself may be advantageous. In the presence of within-stand 508 heterogeneity (e.g., Valbuena et al., 2013a), L-moments could be valuable for delineating 509 microstands (van Aardt et al., 2006). There are potential applications in guiding future forest 510 management operations directly from ALS datasets, once unveiling the relationship between 511 GC and silvicultural alternatives (Pukkala et al., 2016) and thereby to L-moments of ALS 512 513 returns. For ecosystem studies, there is potential for studying canopy structure, e.g., discrimination of single- and multi-layered forests, and other traits relevant to old-growth 514 forests (Lefsky et al., 2002; Miura & Jones, 2010). We encourage further research to exploit 515 the potential of L-moments in forest estimation and other applications. 516

517

518 **6.** Conclusions

We developed a rule-based classification deduced from L-moments summarizing the relative 519 dispersion and skewness of ALS heights. Classification by two simple deductive mathematical 520 rules, L-coefficient of variation Lcv > 0.5 and L-skewness Lskew > 0, was carried out 521 directly on the ALS return cloud, omitting training stages making use of field plot data. Lcv 522 was related to tree size inequality, while *Lskew* provided information on the degree of closure 523 524 of the dominant canopy. These provide relevant information about competition conditions in different areas of the forest, which can be deduced directly from ALS datasets. Our 525 conclusions, however, may apply only to Boreal ecosystems, were light availability and its 526

527 interception by the dominant canopy is the competitive process that limits forest growth. Some 528 of the accuracies obtained were remarkably large, being a direct classification using no field 529 data support, and they were comparable to those obtained by a supervised classification. Two 530 flaws of the rule-based method were the omission of uneven-sized forest with shade-tolerant 531 regeneration and commission errors for the euphotic areas, to be solved by further research 532 perhaps making use of datasets with higher density. These rules can be executed directly over 533 ALS datasets, providing an unambiguous procedure with multiple applications.

534 Acknowledgements

This research was funded by Suomen Metsäkeskus (SMK Finnish Forest centre). Special thanks to Juho Heikkilä and Jussi Lappalainen (SMK), Heli Laaksonen (NLS), and Aki Suvanto (Blom Kartta Oy) for their support at different stages of this study. Rubén Valbuena would like to thank The Finish Society of Forest Sciences for awarding a IUFRO Grant which sponsored his travel to present this work at Silvilaser 2015 Conference in La Grande Motte (France). Special issue editors are thanked for inviting this communication and three anonymous reviewers for their helpful and constructive comments.

542 **References**.

- Asner, G.P., & Mascaro J. (2014). Mapping tropical forest carbon: Calibrating plot estimates
 to a simple LiDAR metric. *Remote Sensing of Environment, 140*, 614-624
- Bollandsås, O.M., & Næsset, E. (2007). Estimating percentile-based diameter distributions in
- 546 uneven-sized Norway spruce stands using airborne laser scanner data. *Scandinavian Journal*
- 547 *of Forest Research*, 22, 33–47.
- 548 Cohen, J. (1960). A coefficient of agreement for nominal scales. *Educational and*
- 549 Psychological Measurement, 20 (1): 37–46

- 550 Cordonnier, T., & Kunstler, G. (2015). The Gini index brings asymmetric competition to
- 1551 light. *Perspectives in Plant Ecology, Evolution and Systematics, 17* (2), 107-115.
- 552 Dalponte, M. Bruzzone, L., Gianelle, D. 2008. Fusion of hyperspectral and LIDAR remote
- sensing data for classification of complex forest areas. *IEEE Transactions in Geosciences*
- *and Remote Sensing*, *46* (5), 1416-1427.
- David, H.A., & Nagaraja, H.N. (2003). *Order Statistics* (third edition). Wiley Series in
 Probability and Statistics. New York: John Wiley.
- 557 Disney, M.I., Kalogirou, V., Lewis, P., Prieto-Blanco, A., Hancock, S., & Pfeifer, M. (2010).
- 558 Simulating the impact of discrete-return lidar system and survey characteristics over young
- conifer and broadleaf forests. *Remote Sensing of Environment, 114*, 1546-1560
- Drake, J.B., Dubayah, R.O., Knox, R.G., Clark, D.B., & Blair, J.B. (2002) Sensitivity of largefootprint lidar to canopy structure and biomass in a neotropical rainforest, *Remote Sensing of Environment*, 81 (2–3), 378-392.
- Frazer, G.W., Wulder, M.A., & Niemann, K.O. (2005). Simulation and quantification of the
 fine-scale spatial pattern and heterogeneity of forest canopy structure: A lacunarity-based
 method designed for analysis of continuous canopy heights, *Forest Ecology and Management*, *214* (1–3), 65-90.
- Frazer G.W., Magnussen S., Wulder M.A., & Niemann K.O. (2011). Simulated impact of
 sample plot size and co-registration error on the accuracy and uncertainty of LiDAR-derived
 estimates of forest stand biomass. *Remote Sensing of Environment*, 115, 636-649.
- 570 García, M., Riaño, D., Chuvieco, E., Salas, J., Danson, F.M. 2011. Multispectral and LIDAR
- 571 Data Fusion for Fuel Type Mapping Using Support Vector Machine and Decision Rules.
- 572 *Remote Sensing of Environment, 115* (6), 1369-1379.

- 573 Gini, C. (1921) Measurement of inequality of incomes. *Economic Journal*, 31, 124–26.
- Gobakken, T., Korhonen, L. & Næsset, E. (2013). Laser-assisted selection of field plots for
- an area-based forest inventory. *Silva Fennica*, 47 (5), 943.
- Gove, J.H. (2004) Structural stocking guides: a new look at an old friend. *Canadian Journal of Forest Research*, *34*,1044–1056.
- Hall, S. A., Burke, I. C., Box, D. O., Kaufmann, M. R. & Stoker, J. M. (2005). Estimating
- 579 stand structure using discrete-return lidar: An example from low density, fire prone

ponderosa pine forests. *Forest Ecology and Management, 208* (1-3), 189-209.

- Hawbaker, T.J., Keuler, N.S., Lesak, A.A., Gobakken, T., Contrucci, K., & Radeloff, V.C.
- 582 (2009). Improved estimates of forest vegetation structure and biomass with a LiDAR-
- 583 optimized sampling design. *Journal of Geophysical Research*, *114*, G00E04.
- Helmert, F.R. (1876) Die Berechnung des wahrscheinlichen Beobachtungsfehlers aus den
 ersten Potenzen der Differenzen gleichgenauer director Beobachtungen. *Astronomische Nachrichten*, 88, 127–132.
- Hosking, J.R.M. (1989). Some theoretical results concerning L-Moments. *Research Report RC14492.* IBM research, Yorktown Heights.
- Hosking, J.R.M. (1990). L-Moments: Analysis and Estimation of Distributions Using Linear
- 590 Combinations of Order Statistics. Journal of the Royal Statistical Society. Series B
- 591 (*Methodological*), 52, 105–124.
- Jaskierniak, D., Lane, P.N.J., Robinson, A., & Lucieer, A. (2011). Extracting LiDAR indices
- 593 to characterise multilayered forest structure using mixture distribution functions. *Remote*
- *Sensing of Environment, 115, 573-585*

- 595 Kellner, J.R. & Asner, G.P. (2009) Convergent structural responses of tropical forests to
- 596 diverse disturbance regimes. *Ecology Letters*, *12*, 887–897
- 597 Kleiber, C. (2005) The Lorenz curve in economics and econometrics. Invited paper, Gini-
- 598 Lorenz Centennial Conference, Siena (Italy).
- 599 Knox, R.G., Peet, R.K. and Christensen, N.L. 1989. Population dynamics in loblolly pine
- stands: Changes in skewness and size inequality. *Ecology*, 70, 1153-1167
- Lefsky, M. A., Cohen, W. B., Acker, S. A., Spies, T. A., Parker, G. G., & Harding, D.
- 602 (1999a). Lidar remote sensing of biophysical properties and canopy structure of forest of
- 603 Douglas-fir and western hemlock. *Remote Sensing of Environment*, 70, 339–361.
- Lefsky, M.A., Harding, D., Cohen, W.B., Parker, G., & Shugart, H.H. (1999b). Surface Lidar
- Remote Sensing of Basal Area and Biomass in Deciduous Forests of Eastern Maryland, USA.
- 606 *Remote Sensing of Environment*, 67, 83-98.
- Lefsky, M.A., Cohen, W.B., Parker, G., & Harding, D., (2002). Lidar remote sensing for
- ecosystem studies. *Bioscience*, *52*, 19-30.
- 609 Lefsky, M.A., Hudak, A.T., Cohen, W.B., and Acker, S.A. (2005). Patterns of covariance
- 610 between forest stand and canopy structure in the Pacific Northwest. *Remote Sensing of*
- 611 *Environment*, *95*, 517-531
- Maltamo, M., Packalen, P., Yu, X., Eerikainen, K., Hyyppa, J., & Pitkanen, J. (2005).
- 613 Identifying and quantifying structural characteristics of heterogeneous boreal forests using
- 614 laser scanner data. *Forest Ecology and Management*, 216, 41–50.

615	Maltamo, M., Bollandsås, O.M., Næsset, E., Gobakken, T., & Packalén, P. (2011). Different
616	plot selection strategies for field training data in ALS-assisted forest inventory. Forestry, 84,
617	23-31

- 618 Miura, N. & Jones, S.D. (2010). Characterizing forest ecological structure using pulse types
- and heights of airborne laser scanning, *Remote Sensing of Environment*, 114 (5), 1069-1076.
- 620 Meyer, D., Dimitriadou, E., Hornik, K., Weingessel, A. & Leisch, F. (2014a) e1071: Misc
- Functions of the Department of Statistics, TU Wien. R package version 1.6-4.
- 622 <u>http://ugrad.stat.ubc.ca/R/library/e1071/html/00Index.html</u>. Visited in Jan. 2014.
- 623 Meyer, D., Zeileis, A. and Hornik, K. (2014b) vcd: Visualizing Categorical Data. R package
- 624 version 1.3-2. <u>https://cran.r-project.org/web/packages/vcd/index.html</u> Visited in Jan. 2014.
- NLS National Land Survey of Finland 2013. Laser scanning data (available online at
 maanmittauslaitos.fi). Visited in Sep. 2013.
- 627 Næsset, E. (2002). Predicting forest stand characteristics with airborne scanning laser using a
- 628 practical two-stage procedure and field data. *Remote Sensing of Environment*, 80, 88-99.
- 629 Næsset, E. (2004). Effects of different flying altitudes on biophysical stand properties
- estimated from canopy height and density measured with a small-footprint airborne scanning
- 631 laser. Remote Sensing of Environment, 91 (2), 243-255.
- 632 Ozdemir, I., & Donoghue, D.N.M. (2013). Modelling tree size diversity from airborne laser
- 633 scanning using canopy height models with image texture measures. *Forest Ecology and*
- 634 *Management*, 295, 28–37.
- 635 Pascual, C., García-Abril, A., García-Montero, L.G., Martín-Fernández, S., & Cohen, W.B.
- 636 (2008) Object-based semi-automatic approach for forest structure characterization using lidar

- data in heterogeneous *Pinus sylvestris* stands. *Forest Ecology and Management*, 255 (11),
 3677-3685.
- Pontius, R.G., & Santacruz, A. (2015). diffeR: Metrics of Difference for Comparing Pairs of
 Maps. R package version 0.0-4. <u>http://CRAN.R-project.org/package=diffeR</u>. Visited in Apr.
 2016.
- Pukkala, T., Laiho, O., & Lähde, E. (2016). Continuous cover management reduces wind
 damage. *Forest Ecology and Management*, *372*, 120-127
- Robbins, H.E. (1944). On the Expected Values of Two Statistics. *The Annals of Mathematical Statistics*, *15*, 321-323.
- 646 Stark, S.C., Leitold, V., Wu, J.L., Hunter, M.O., de Castilho, C.V., Costa, F.R., McMahon,
- 647 S.M., Parker, G.G., Shimabukuro. M,T., Lefsky, M.A., Keller, M., Alves, L.F., Schietti, J.,
- 648 Shimabukuro, Y.E., Brandão, D.O., Woodcock, T.K., Higuchi, N., de Camargo, P.B., de
- 649 Oliveira, R.C., Saleska, S.R., & Chave, J. (2012). Amazon forest carbon dynamics predicted
- by profiles of canopy leaf area and light environment. *Ecology Letters*, *15* (12), 1406-1414.
- Olofsson, P., Foody, G.M., Stehman, S.V., & Woodcock, C.E. (2013) Making better use of
 accuracy data in land change studies: Estimating accuracy and area and quantifying uncertainty
 using stratified estimation, *Remote Sensing of Environment*, *129*, 122–131.
- Stehman, S.V. (1996) Estimating the Kappa coefficient and its variance under stratified random
 sampling. *Photogrammetric Engineering & Remote Sensing*, 62 (4), 401–402.
- Valbuena, R., Packalen, P., Martín-Fernández, S., & Maltamo, M. (2012). Diversity and
 equitability ordering profiles applied to study forest structure. *Forest Ecology and Management*, 276, 185–195.

- 659 Valbuena, R., Packalen, P., Mehtätalo, L., García-Abril, A., & Maltamo, M. (2013a).
- 660 Characterizing forest structural types and shelterwood dynamics from Lorenz-based
- 661 indicators predicted by airborne laser scanning. Canadian Journal of Forest Research, 43,
- 662 1063–1074.
- Valbuena, R., Maltamo, M., Martín-Fernández, S., Packalen, P., Pascual, C., & Nabuurs, G.J.
- 664 (2013b). Patterns of covariance between airborne laser scanning metrics and Lorenz curve
- descriptors of tree size inequality. *Canadian Journal of Remote Sensing*, 39, S18–S31.
- Valbuena, R., Vauhkonen, J., Packalen, P., Pitkänen, J., Maltamo, M. (2014) Comparison of
- 667 airborne laser scanning methods for estimating forest structure indicators based on Lorenz
- 668 curves. ISPRS Journal of Photogrammetry & Remote Sensing, 95, 23–33
- 669 Valbuena, R., Eerikäinen, K., Packalen, P. & Maltamo, M. (2016a). Gini coefficient
- 670 predictions from airborne lidar remote sensing display the effect of management intensity on
- 671 forest structure. *Ecological Indicators*, 60, 574–585.
- Valbuena, R., Maltamo, M. & Packalen, P. (2016b). Classification of multi-layered forest
- 673 development classes from low-density national airborne lidar datasets. *Forestry*, 89, 392–401.
- van Aardt, J.A.N., R.H. Wynne, & Oderwald, R.G. (2006). Forest volume and biomass
 estimation using small-footprint LiDAR-distributional parameters on a per-segment basis. *Forest Science*, 52 (6), 636-649.
- Wang, Q. J. (1996). Direct sample estimators of L moments. *Water Resources Research*, 32
 (12), 3617–3619.
- Weiner, J. (1990) Asymmetric competition in plant populations. *Trends in Ecology and Evolution*, 5, 360–364.

- Westfall, J. A., Patterson, P. L., & Coulston, J. W. (2011). Post-stratified estimation: withinstrata and total sample size recommendations. *Canadian Journal of Forest Research*, *41* (5),
 1130–1139.
- 684 White, J.C.; Wulder, M.A.; Varhola, A.; Vastaranta, M.; Coops, N.C.; Cook, B.D.; Pitt, D.;
- 685 Woods, M. (2013). A best practices guide for generating forest inventory attributes from
- 686 airborne laser scanning data using an area-based approach. Natural Resources Canada,
- 687 Information Report FI-X-010. Canadian Forest Service, Victoria, BC.
- Zenner, E.K. (2005). Development of tree size distributions in Douglas-fir forests under
 differing disturbance regimes. *Ecological Applications*, *15*, 701-714
- Evans, D.L., Carlson, G.C., Parker, R.C., Grado, S.C., & Gerard, P.D. (2003).
- 691 Characterizing vertical forest structure using small-footprint airborne LiDAR. *Remote Sensing*692 *of Environment*, 87, 171-182.

693 Appendix A. L-moments and their relationship to Gini Coefficient

694 A.1. L-moments for describing a distribution

Let an order statistic $X_{k:r}$ be the *k*-th smallest observation in a sample of size *r* of the random variable *X* (e.g. ALS return heights), and let $E(X_{k:r})$ be its expected value. For example, consider $E(X_{1:2})$ in the following population of size 3: {12,16,14}. There are three possible samples of size r = 2, with sample minima (k = 1): {12,12,14}. The expected value is the mean over these, i.e., $E(X_{1:2}) = 12.67$. In the analysis of this paper, the population is the unknown infinite set of all possible ALS returns over the primary calculation unit (sample plot or grid cell). The expected value is estimated using the observed sample of returns. 702 L-moments describe the distribution of a scalar random variable *X* through weighted sums of 703 $E(X_{k;r})$. Hosking (1990) defined the L-moments as:

704 (A1)
$$Lr = r^{-1} \sum_{k=0}^{r-1} (-1)^k \cdot \binom{r-1}{k} \cdot E(X_{r-k:r}).$$

The first L-moment (L1) is obtained by substituting r = 1 in equation (A1) to get:

706 (A2)
$$L1 = E(X_{1:1}) = E(X),$$

which is thus equivalent to the first product-moment (expectation) of *X*. Hence, *L*1 is the Lmeasure for the location or central tendency of the distribution. If observations of *X* are available, *L*1 can be estimated as the arithmetic mean:

710 (A3)
$$\widehat{L1} = \overline{X}$$

711 The second L-moment (*L*2), follows the case for r = 2:

712 (A4)
$$L2 = \frac{1}{2}E(X_{2:2}) - \frac{1}{2}E(X_{1:2}) = \frac{1}{2}E[X_{2:2} - X_{1:2}],$$

which is the expected value of half difference between minimum $(X_{1:2})$ and maximum $(X_{2:2})$ in a sample of size two. It therefore provides the mean of half differences, and thus it is the Lmeasure for the dispersion of the distribution.

Following a similar logic for the third L-moment (L3), substituting r = 3 in (A1) yields:

717 (A5)
$$L3 = \frac{1}{3}E(X_{3:3}) - \frac{2}{3}E(X_{2:3}) + \frac{1}{3}E(X_{1:3}),$$

which is a weighted sum of minimum, $(X_{1:3})$, median $(X_{2:3})$, and maximum $(X_{3:3})$ of a sample with size three. It can further be written as:

720 (A6)
$$L3 = \frac{1}{3}E[(X_{3:3} - X_{2:3}) - (X_{2:3} - X_{1:3})],$$

to show that *L*3 expresses the expected difference between the maximum-median and medianminimum differences in a sample of size three, which provides a L-measure for the asymmetry of the distribution of *X*. Hence, L3 = 0 corresponds to a symmetric distribution, L3 > 0describes positive asymmetry (left-skewed distribution) and L3 < 0 describes negative asymmetry (right-skewed distribution).

726 A.2. L-moment ratios

Hosking (1990) also defined the ratios for L-moments. They have the advantage of being
bounded by finite intervals (Hosking 1989), yielding comparable relative descriptions for the
distribution of *X*.

The second L-moment ratio is obtained as the ratio of the second to the first L-moments. It is called the L-coefficient of variation (Lcv) for its comparison to conventional moments. From equations (A2) and (A4) it can be observed that Lcv equals:

733 (A7)
$$Lcv = \frac{L2}{L1} = \frac{E(X_{2:2}) - E(X_{1:2})}{2E(X)}$$

734 For positive random variables, the values for the second L-moment ratio are bounded by the [0, 1] range (Hosking, 1989). Just like the coefficient of variation of conventional moments, 735 Lev is a descriptor of dispersion relative to central tendency; that is to say, concentration. This 736 brings the advantage that concentration measures are comparable among distributions differing 737 in their location or central tendency (L1), and also independently of the units of measure. It is 738 739 worthwhile to note that Hosking never defined a second L-moment ratio, as their generalized definition stands only for r = 3, 4 ... (Hosking 1990: 108), and the L-coefficient of variation 740 was simply presented alongside. It was only later that many authors have regarded *Lcv* to be 741 the second L-moment ratio. 742

The third L-moment ratio is obtained by division between the third and the second L-moments. It is called the L-skewness (*Lskew*), as it has been found to be a robust descriptor for the asymmetry of the distribution of *X*. From equations (A4) and (A6), and using the equivalence $E(X_{3:3} - X_{1:3}) = \frac{3}{2}E(X_{2:2} - X_{1:2})$ (Robbins, 1944: Eq. 22; David & Nagaraja, 2003: 44, 56) it yields:

748 (A8)
$$Lskew = \frac{L3}{L2} = \frac{E(X_{3:3}) - 2E(X_{2:3}) + E(X_{1:3})}{E(X_{3:3}) - E(X_{1:3})}$$

As explained for *L*3, *Lskew* = 0 corresponds to a symmetric distribution, while positive or negative values denote the type of asymmetry for the distribution. Additionally, *Lskew* has the advantage of presenting theoretical bounds within the [-1, 1] interval (Hosking 1989). Consequently, *Lskew* is a descriptor of asymmetry relative to dispersion, and therefore independent of the units of measure and the dispersion of the distribution of *X*.

A.3. Equivalence between the Gini coefficient and the L-coefficient of variation

The Gini coefficient of a scalar random variable *X* (*GC*) is the ratio of the area comprisedbetween the Lorenz curve and the diagonal line of equality (Gini, 1921):

757 (A9)
$$GC = 1 - 2 \int_0^1 L(X) dX,$$

Where L(X) is the Lorenz curve: the relative cumulative distribution of a variable against the cumulative frequency distribution of the proportion of individuals in the population. From Eq. (A9), Kleiber (2005: Eq. 6) showed that:

761 (A10)
$$GC = 1 - \frac{E(X_{1:2})}{E(X)}$$

On the other hand, the Lcv gives also the GC. From Eq. (A7) it derives:

763 (A11a)
$$Lcv = \frac{E(X_{2:2}) - E(X_{1:2})}{2E(X)}$$

764 (A11b)
$$= \frac{E(X_{2:2} - X_{1:2}) + 2E(X_{1:2}) - 2E(X_{1:2})}{2E(X)}$$

765 (A11c)
$$= \frac{E(X_{2:2} + X_{1:2}) - 2E(X_{1:2})}{2E(X)}$$

766 (A11d)
$$= \frac{2E(X) - 2E(X_{1:2})}{2E(X)}$$

767 (A11e)
$$= 1 - \frac{E(X_{1:2})}{E(X)}$$

Equation (A11d) results from (A11c) because $X_{1:2} + X_{2:2}$ is the sum of two independent and identically distributed samples, and it is therefore equivalent to $X_1 + X_2$. Consequently, (A10) and (A11e) demonstrate:

771 (A12)
$$GC = Lcv$$

The result in Eq. (A12) is essentially a special case of a 140-years-old result (Helmert, 1876; as cited in David and Nagaraja, 2003: 249) presented in equation 9.4.2 of David and Nagaraja (2003), which might even provide interesting extensions using expectations of order statistics in sample sizes larger than r = 1, 2, 3.

776

777 Appendix B. Criteria for determining forest development classes

Silvicultural development classes are used in Finland to classify forest stands and assist in
decision-making for forest management planning. It was possible to apply stratified sampling
using the stand register dataset employed by the Finnish Forest Centre (SMK, Suomen
Metsäkeskus) for their operational management planning, since a development class has been

explicitly assigned to each stand from previous inventories. The development class to which each sample plot belonged to was nevertheless ultimately corroborated in the field, being the criteria used *in-situ* prevalent over the stand register data. Minor differences in per-stratum sample sizes were simply caused by such type of discrepancies found in few plots. The criteria that segregated forest areas into different forest classes were:

787

788

•

Seedling: stands with average tree height lower than 1.3 m, and absence of mature trees (overstorey).

- *Sapling*: stands with average tree height greater than 1.3 m, and average diameter at
 breast height (DBH) smaller than 8 cm, and absence of mature trees (overstorey).
- *Young*: stands with average DBH ranging 8-16 cm and average tree height ranging 7-9
 m high.
- *Advanced*: stands with average DBH greater than 16 cm.
- *Mature*: stands reaching a quadratic mean DBH (QMD) greater than 18 cm.
- Shelterwood: stands including a dense overstorey of mature trees (DBH > 16 cm) which
 reaches at least 100-300 stems ha⁻¹, and also a dense understorey of seedlings (height <
- 797 1.3 m) of shade-tolerant species, usually Norway spruce $(1500-1800 \text{ stems} \cdot \text{ha}^{-1})$.
- Seed-tree: stands including a sparse overstorey of mature trees (DBH > 16 cm) of only
 50-100 stems·ha⁻¹, and also a dense understorey of seedlings (height < 1.3 m) of shade-
 intolerant species, usually Scots pine (1500-2200 stems·ha⁻¹) or Birch species (1100 1600 stems·ha⁻¹).
- *Multi-storied*: stands including a dense understorey (above-mentioned densities) of
 seedlings (height < 1.3 m) and saplings (height > 1.3 m, DBH < 8 cm) of any species,
 usually deciduous but also Scots pine or Norway spruce. The size of trees in the
 overstorey is not a determinant criterion, but trees in the understory must reach their
 sapling stage.