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Connell, Sean Duncan; Russell, Bayden D.

[The direct effects of increasing CO₂ and temperature on non-calcifying organisms: increasing the potential for phase shifts in kelp forests](#) Proceedings of the Royal Society of London Series B - Biological Sciences, 2010; 277(1686):1409-1415.

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27th August, 2012

1 **Published in: *Proceedings of the Royal Society of London B*, 2010; 277(1686):1409-1415**

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4 **The direct effects of increasing CO₂ and temperature on non-calcifying**
5 **organisms: increasing the potential for phase-shifts in kelp forests**

6

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13

14 **Key words:** carbon dioxide, climate change, habitat resilience, phase-shift, turf-forming
15 algae

16

17 **Running headline:** direct effects of CO₂ and temperature on algal turf

18

19 **Summary**

20 Predications about the ecological consequences of oceanic uptake of CO₂ have been
21 preoccupied with the effects of ocean acidification on calcifying organisms, particularly those
22 critical to the formation of habitats (e.g. coral reefs) or their maintenance (e.g. grazing
23 echinoderms). This focus overlooks the direct effects of CO₂ on non-calcareous taxa,
24 particularly those that play critical roles in ecosystem-shifts. We used two experiments to
25 investigate whether increased CO₂ could exacerbate kelp loss by facilitating non-calcareous
26 alga that we hypothesised: (1) inhibit the recovery of kelp forests on an urbanised coast; and
27 (2) form more extensive covers and greater biomass under moderate future CO₂ and
28 associated temperature increases. Our experimental removal of turfs from a phase-shifted
29 system (i.e. kelp to turf-dominated), revealed that the number of kelp recruits increased,
30 thereby indicating that turfs can inhibit kelp recruitment. Future CO₂ and temperature
31 interacted synergistically to have a positive effect on the abundance of algal turfs, whereby
32 they had twice the biomass and occupied over four times more available space than under
33 current conditions. We suggest that the current preoccupation with the negative effects of
34 ocean acidification on marine calcifiers overlooks potentially profound effects of increasing
35 CO₂ and temperature on non-calcifying organisms.

36

37 **1. INTRODUCTION**

38 A vexing challenge to ecological research is to identify the perturbations that cause systems
39 undergo shifts from one state to another (Scheffer *et al.* 2001). Shifts in systems often occur
40 quite suddenly because their drivers can be insidious and combine to alter interactions or
41 competitive relationships between key species (Suding & Hobbs 2009). Factors that subtly
42 undermine the resilience of systems are generally unrecognised (Scheffer *et al.* 2001) and we
43 have an incomplete understanding of the effects of long-term perturbations (e.g. marine
44 eutrophication and switches in algal dominance; Smith & Schindler 2009). Nonetheless,
45 ecosystems continue to change, and the need to understand how future conditions (e.g.
46 climate) may contribute to this change has become a fundamental area of ecological research.

47
48 The role of global environmental change in driving habitat shifts in marine ecosystems has
49 received heightened attention (e.g. Hoegh-Guldberg *et al.* 2007; Hughes *et al.* 2007). Marine
50 waters currently absorb approximately 30 % of the anthropogenic derived CO₂ from the
51 earth's atmosphere and the resulting ocean acidification has been predicted to have drastic
52 effects over the next 100 years (Feely *et al.* 2004; Orr *et al.* 2005). Unsurprisingly, research
53 on the effects of climate change has a disproportionate focus on the effects of ocean
54 acidification on calcareous organisms that form habitats (i.e. coral reefs', Hoegh-Guldberg *et al.*
55 *al.* 2007; Anthony *et al.* 2008; Kuffner *et al.* 2008) or maintain habitats (e.g. grazers Fabry *et al.*
56 *al.* 2008; Byrne *et al.* 2009). However, research into the role of the changing climate in the
57 loss of marine habitats has been largely restricted to tropical waters (i.e. coral reefs) while in
58 temperate systems the focus has centred on individual organisms (e.g. Dupont *et al.* 2008;
59 Parker *et al.* 2009). This focus has, to date, overlooked historical and continuing deforestation
60 of algal canopies across the world's temperate coastline (Eriksson *et al.* 2002; Airoidi &
61 Beck 2007; Connell *et al.* 2008).

62

63 Kelp forests occur along the majority of the world's temperate coastlines and are among the
64 most phylogenetically diverse and productive systems in the ocean (Mann 1973). On many coasts
65 where humans have altered chemical and biological conditions, however, canopies of algae
66 (e.g. kelp forests) have been replaced by mats of turf-forming algae (Eriksson *et al.* 2002;
67 Airoidi & Beck 2007; Connell *et al.* 2008). While kelp canopies inhibit turfs (Irving &
68 Connell 2006; Russell 2007), developing theory explains shifts from canopy to turf-
69 domination as a function of reduced water quality that enables the cover of turf to expand
70 spatially and persist beyond its seasonal limits (Gorman *et al.* 2009), subsequently inhibiting
71 the recruitment of kelp and regeneration of kelp forests. Unlike kelps, many turf-forming
72 species are ephemeral and require increased resource availability to enable their physiology
73 and life history to be competitively superior to perennial species (Airoidi *et al.* 2008). It is
74 critical, therefore, to identify future conditions that would have positive effects on turfs,
75 thereby exacerbating the loss of algal canopies.

76

77 Although recent studies have identified the effects of anticipated levels of acidification on
78 calcareous temperate algae (e.g. Martin & Gattuso 2009; Russell *et al.* 2009), none has
79 examined the effects of elevated CO₂ and temperature on non-calcareous species such as
80 algal turfs. Therefore, the purpose of our study was two-fold, to determine; (1) if turfs do in
81 fact inhibit the recruitment of kelp under human mediated conditions (i.e. on a metropolitan
82 coast), and if so; (2) to determine if future conditions could exacerbate the currently observed
83 shift from kelp to turf-dominated reefs. We tested the hypotheses that (1) the removal of turfs
84 on a metropolitan coast would cause greater recruitment of kelp, and (2) the abundance of
85 turfs would increase under combined future conditions (i.e. elevated CO₂ and temperature).

86

87 **2. METHODS**

88 **(a) Ability of turfs to inhibit kelp recruitment**

89 We first tested the prediction that the removal of algal turfs from turf-dominated ecosystems
90 (i.e. degraded systems; Connell *et al.* 2008) would enable recruitment of kelp (*Ecklonia*
91 *radiata*) to increase. Algal turf and associated sediment were removed from 12 replicate 1 m²
92 plots to expose the underlying substrate. These plots and 12 replicate controls (1 m²
93 untouched plots) were positioned within 5 m of remnant patches of canopy, which acted as a
94 source of recruits. This procedure was repeated at three sites (separated by > 1 km) that were
95 associated with both extensive covers of turfs and remnant patches of canopy on the Fleurieu
96 Peninsula, South Australia. The number of kelp recruits in plots was quantified in April 2008,
97 ~12 months after turfs were removed.

98

99 **(b) Effect of future conditions on turfs**

100 Algal turfs were exposed to current and predicted future concentrations of CO₂ (380 ppm and
101 550 ppm, respectively) in crossed combination with ambient and elevated temperatures (17°C
102 and 20°C, respectively) in a mesocosm experiment over 14 weeks from March – June 2008.
103 Both future CO₂ and temperatures were based on IS92a model predictions for the year 2050
104 (Meehl *et al.* 2007), with the ambient temperature being the summer maxima at the algal
105 collection site. There were two replicate mesocosms per combination of treatments ($n = 5$
106 replicate turf specimens per mesocosm).

107

108 The response of turfs to experimental conditions was assessed using three response variables;
109 percentage cover and dry mass of algae recruiting to initially unoccupied substrate (5 × 5 cm
110 fibreboard tiles), and effective quantum yield of algae on the original rock substrate. The
111 percentage cover of algae was visually estimated to the nearest 5 % at the end of the

112 experiment ($n = 5$ tiles per mesocosm) as suggested by Drummond and Connell (2005). Dry
113 mass of algae was measured by carefully scraping all algae from a standard area on each tile
114 (6.25 cm^2) into a pre-weighed aluminium tray, which was then rinsed with fresh water to
115 remove excess salt and dried at 60°C for 48 hours. Fibreboard tiles were used as unoccupied
116 substrate to remove confounding by any differences in either percentage cover or mass of
117 algal samples that were placed into the experiments. Further, the tiles were placed into
118 mesocosms with the rough side uppermost as turfs readily recruit to this surface (Irving &
119 Connell 2002), which has similar roughness to basalt rock at the collection site.

120

121 Chlorophyll fluorescence, a relative measure of the photochemistry of Photosystem II (Genty
122 *et al.* 1989), was measured under the experimental light conditions using a Pulse Amplitude
123 Modulated (PAM) fluorometer (Walz, Germany). Effective quantum yield (Y) was calculated
124 using the equation $Y = (F'_m - F)/F'_m$ (Genty *et al.* 1989), where F'_m is the maximal
125 fluorescence, and F the minimal fluorescence, under illuminated conditions (van Kooten &
126 Snel 1990). F was measured by holding the fiberoptics of the PAM fluorometer in contact
127 with the algal sample (*in situ* in mesocosms) and exposing it to a pulsed measuring beam of
128 weak red light ($0.15 \mu\text{mol m}^{-2} \text{s}^{-1}$, 650 nm) followed immediately by a pulse of saturating
129 actinic light (0.8 s, $6000 \mu\text{mol m}^{-2} \text{s}^{-1}$) to measure F'_m (Beer *et al.* 1998). Each yield value
130 used in the analyses was a mean of three replicate measurements taken on different parts of
131 each algal sample so that yield was not underestimated due to recovery of the photosystems
132 from repeated saturating light pulses.

133

134 Turf specimens used in experiments were collected from a rocky reef at Victor Harbour,
135 South Australia (35.57126°S 138.61221°E) at 2 m – 4 m depth. The turf assemblages used
136 were comprised mainly of *Feldmannia* spp., which form densely packed mats of filaments up

137 to 2 cm in height. Turfs were collected still attached to their rocky substrate (approximately
138 the same size; ~ 5 cm × 5 cm) and allowed to acclimate in holding mesocosms for two weeks
139 before the experiment commenced. During acclimation, physical conditions in the
140 mesocosms were similar to those at the collection site (i.e. 17°C and current atmospheric CO₂
141 concentrations). Algae were then randomly re-assigned to mesocosms in which experimental
142 conditions were gradually increased over a further 2 week period until they reached their pre-
143 designated levels. All mesocosms were aerated at 10 L min⁻¹, with either current atmospheric
144 air or air enriched with CO₂. Future concentrations of CO₂ in water were maintained at 550
145 ppm CO₂ (pH 7.95, based on the IS92a model for 2050; Meehl *et al.* 2007) using pH probes
146 attached to automatic solenoids (Sera, Heinsberg, Germany) and CO₂ regulators. Probes were
147 temperature compensated and calibrated using NBS calibration buffers on a daily basis.
148 Elevated temperature was achieved by using heaters in the 20°C treatment mesocosms. Total
149 alkalinity (TA) of the seawater in mesocosms was measured on a weekly basis to monitor
150 CO₂ and bicarbonate (HCO₃⁻) concentrations (see online supplementary material for more
151 detail).

152

153 Each mesocosm system consisted of a 40 L experimental aquarium connected to a 200 L
154 reservoir tank with water recirculated in a closed loop, ensuring that all replicate mesocosms
155 were independent of each other. To ensure quality of the growing conditions in mesocosms,
156 one-third of the water was removed from reservoir tanks and replaced with fresh seawater
157 weekly (see Russell *et al.* 2009). Lighting was supplied in a 12:12 light:dark cycle by pairs of
158 fluorescent lights directly above each mesocosm (see online supplementary material for more
159 detail).

160

161 (c) *Statistical analyses*

162 The effect of turf on kelp recruitment was analysed using a two-factor ANOVA, with factors
163 of turf (turf present *v.* turf removed) and site (three sites). Both factors were treated as
164 orthogonal, ‘turf’ as fixed, ‘site’ as random ($n = 12$ replicate plots). Data were $\text{Ln}(X + 1)$
165 transformed before analysis to conform to assumptions of homogeneity.

166

167 Analysis of the mesocosm experiment proceeded in two steps. First, three-factor ANOVAs
168 were used to identify if there was any difference in experimental effects between replicate
169 mesocosms for all measures (percentage cover, dry mass and effective quantum yield). Both
170 CO_2 and temperature were treated as fixed and orthogonal, with two levels in each factor, and
171 two replicate mesocosms were nested within both CO_2 and temperature ($n = 5$ replicate
172 samples of algae per mesocosm). No differences were detected between replicate mesocosms
173 within treatments (i.e. no “tank” effects). Therefore, to avoid pseudoreplication within
174 mesocosms, data for the five algal specimens within each mesocosm were averaged, and data
175 reanalysed using two factor ANOVAs; CO_2 and temperature were again treated as fixed and
176 orthogonal, with mesocosms as replicates. Where significant treatment effects were detected,
177 Student-Newman-Keuls (SNK) post-hoc comparison of means was used to determine which
178 factors differed. Percentage cover data were arcsin transformed prior to analysis to remove
179 heterogeneity (Underwood 1981).

180

181 **3. RESULTS**

182 **(a) *Ability of turfs to inhibit kelp recruitment***

183 The removal of turfs resulted in the greater recruitment of kelp at all three phase-shifted sites
184 (Fig. 1a). While there was significant difference in the number of kelp recruits among the
185 three replicate sites, the number of kelp recruits was consistently greater in removal plots

186 than plots where turfs were left intact (turf removal \times site interaction: $F_{2,66} = 6.10$, $p = 0.0037$;
187 SNK: turf removal $>$ turf intact at all three sites).

188

189 **(b) Effect of future conditions on turfs**

190 CO₂ and temperature had an interactive effect on the percentage cover of turf-forming algae
191 that recruited to available space (Fig. 1b; CO₂ \times temperature interaction, $F_{1,4} = 7.73$, $p =$
192 0.0498). Under current CO₂ concentrations, temperature had a positive effect on the
193 percentage cover of turfs that recruited to available space (Fig. 1b; SNK test of CO₂ \times
194 temperature interaction). In contrast, future CO₂ had no effect on the cover of turfs at ambient
195 temperatures (17°C). When future CO₂ and elevated temperature were present in
196 combination, however, turfs occupied $> 80\%$ of available space (Fig. 1b). Importantly, this
197 represented a synergistic effect whereby turfs occupied 25% more space than would be
198 predicted by the independent effects of CO₂ and temperature.

199

200 Both elevated CO₂ and temperature had positive effects on the dry mass of turfs (Fig. 1c; $F_{1,4}$
201 = 19.20, $p = 0.0119$ and $F_{1,4} = 11.39$, $p < 0.0279$, respectively). There is no graphical
202 evidence of an interaction between these factors (Fig. 1c) as the increase in mass by CO₂ is
203 proportionally similar between the CO₂ treatments, and vice versa. This interpretation is
204 supported by the lack of a significant interaction term between these factors ($F_{1,4} = 0.41$, $p =$
205 0.5558; power = 0.08) as shown by the effect of temperature in each CO₂ treatment
206 (approximately double the mass), and the effect of CO₂ in each temperature treatment
207 (approximately double the mass) (Fig. 1c). Hence, the combined effects of CO₂ and
208 temperature are approximately four times greater than ambient conditions.

209

210 The effects of these factors on quantum yield were relatively small; yield of turfs was 5 %
211 greater under future CO₂ concentrations (Fig. 1d, ANOVA: $F_{1,4} = 14.11$, $p = 0.0198$) but 3 %
212 less under elevated temperature (Fig. 1d, $F_{1,4} = 16.73$, $p = 0.0150$). Again, the proportional
213 influence of each factor was similar within each level of the crossed factor (Fig 1d) as also
214 indicated by the lack of a significant interaction term between these factors ($F_{1,4} = 1.24$,
215 $p = 0.3276$; power = 0.14). Whilst we report low power for non-significant interactions, we
216 consider the combined effects of temperature and CO₂ are indeed additive rather than
217 multiplicative.

218

219 **4. DISCUSSION**

220 A substantial part of research into global environmental change centres on the negative
221 effects of ocean acidification and increasing temperature on organisms that form calcareous
222 structures (e.g. Fabry *et al.* 2008; Jokiel *et al.* 2008; Kuffner *et al.* 2008). While elevated CO₂
223 can be beneficial to plants in terrestrial systems (Ainsworth & Long 2005), there is little
224 recognition of the positive effects on some non-calcareous marine species. Here, we show
225 that predicted moderate concentrations of CO₂ and temperature had a synergistic positive
226 effect on the abundance of non-calcareous algal turfs. Yet, it is important to recognise that
227 such positive effects could act as perturbations in ecological systems. Turfs form a natural
228 component of the early successional stages of kelp-dominated landscapes. Under natural
229 conditions algal canopies inhibit these algae (Irving & Connell 2006; Eriksson *et al.* 2007;
230 Russell 2007), but under altered environmental conditions turfs expand (Connell 2007;
231 Russell & Connell 2007) by inhibiting kelp recruitment (i.e. eroding resilience of forests).
232 Our results indicate that kelp loss may be exacerbated on human-dominated coasts, by the
233 positive effects of increasing CO₂ and temperature on kelp inhibitors, motivating the need to
234 assess such switches on coasts that are currently considered unaffected by human activity.

235

236 Recruitment that replenishes lost habitat-forming individuals is key to resilience against
237 phase shifts in ecosystems founded on habitat-forming species (Pickett & White 1985).

238 Disturbance is part of the dynamics of kelp forests which would otherwise fully occupy space
239 (e.g. storms Dayton *et al.* 1984). We recognise that it is not so much the direct effects of
240 climate stressors on kelp forests that may affect their future abundance, but rather the indirect
241 loss of kelp via their competitors or inhibitors. Altering global (i.e. CO₂) and local (i.e.
242 eutrophication) stressors in combination can allow turfs to expand to more rapidly occupy
243 available space (Russell *et al.* 2009). It is noteworthy that our experimentally increased CO₂
244 and temperature, two inherently linked global stressors, enabled turfs to occupy nearly five
245 times more space than under current conditions. While it may be possible to mitigate the
246 effects of climate-driven environmental change by removing nutrient inputs (e.g. recycling
247 wastewater and sewage, Russell & Connell 2009), such actions would not be possible in the
248 case of synergistic effects between multiple global stressors. Indeed, understanding the
249 degree to which these factors will combine to accelerate and expand ecosystem-shifts is of
250 key concern (Scheffer *et al.* 2001; Suding & Hobbs 2009).

251

252 Increasing temperatures are commonly predicted to result in changes in marine communities
253 because of a shift in the geographic ranges of species (e.g. Fields *et al.* 1993; Poloczanska *et*
254 *al.* 2008). While community shifts have been observed, local conditions and competitive
255 interactions may alter the outcomes (Helmuth *et al.* 2002; Poloczanska *et al.* 2008). In such
256 cases, taxa that are natural components of a system may play substantially altered roles in
257 their maintenance and disruption (Suding & Hobbs 2009). In Australia, *Ecklonia radiata*
258 canopies have high rates of natural turnover, and their maintenance relies on rapid
259 recruitment and replenishment into canopy-gaps in the winter months (Kennelly 1987b).

260 While turfs are a natural component of these kelp-dominated systems (Irving *et al.* 2004) they
261 are ephemeral, rapidly occupying available space in summer but declining in cover and
262 biomass over the colder months (Russell 2007; S.D. Connell, B.D. Russell, D. Gorman, A.
263 Airoidi, unpubl. data). Importantly, *Ecklonia radiata* produce gametophytes, the smallest and
264 therefore more susceptible stage of the life cycle, in the colder months when turfs are at their
265 lowest abundance. Yet, we show that turfs increased in abundance under elevated
266 temperatures, suggesting that future increases in temperature could allow turfs to be
267 increasingly abundant throughout periods of naturally low abundance (i.e. winter). Similarly,
268 turfs exhibit a phenological shift due to elevated nutrients (S.D. Connell, B.D. Russell, D.
269 Gorman, A. Airoidi, unpubl. data), possibly leading to habitat shifts on urbanised coasts
270 (Gorman *et al.* 2009). As algal turfs can inhibit kelp recruitment (Kennelly 1987a; this study),
271 any phenological shift that allows turfs to persist though periods of kelp recruitment is likely
272 to reduce the resilience of kelp forests to disturbance. While it is accepted that such habitat
273 shifts are common on human-dominated coasts (Airoidi 2003; Connell 2007), temperature,
274 unlike nutrients, will increase even on “pristine” coasts, potentially causing habitat shifts in
275 the absence of local human populations.

276

277 Loss of canopy-forming algae can be a consequence of overgrazing by increasing urchin
278 populations (Estes *et al.* 1998), but in many parts of the world, including most of southern
279 Australia, such deforestation is not possible because of the types and sparse densities of
280 herbivores (Connell & Vanderklift 2007; Connell & Irving 2008). Nevertheless, canopy-
281 forming algae has long been disappearing from human dominated coasts lacking strong
282 herbivory, but experiencing strong water pollution (Eriksson *et al.* 2002; Airoidi *et al.* 2008;
283 Connell *et al.* 2008), yet the specific mechanisms underlying this loss are often a point of
284 conjecture and contention. Previous studies have demonstrated that some more erect forms of

285 turf-forming and foliose algae can dominate available space and inhibit canopy recruitment
286 (Kennelly 1987a; Airoidi 2003) but, to our knowledge, ours is the first study to show that the
287 removal of filamentous turfs can enhance the recruitment of kelp. By removing turfs from the
288 substrate we created more available space for kelp recruits to settle and become established.
289 As the creation of new space is a prerequisite for community change (Pickett & White 1985;
290 Airoidi & Virgilio 1998), it is unlikely that these phase-shifted reefs (i.e. from kelp to turf
291 dominated; Connell *et al.* 2008) will be able to return to domination by kelp canopies until
292 the environmental conditions on these coasts revert to their more natural state (e.g. nutrients;
293 Gorman *et al.* 2009). Nevertheless, we demonstrate that the removal of turfs can create the
294 space necessary for the recruitment and recovery of kelp and that the observed phase-shift
295 (Connell *et al.* 2008) may not be permanent.

296

297 While the productivity of terrestrial plants stands to increase with predicted future CO₂,
298 especially in plants which utilise C3 photosynthesis (Ainsworth & Long 2005), there is still
299 debate on whether this will be the case in marine algae. Most marine algae have carbon
300 concentrating mechanisms (CCMs) which allow them to use bicarbonate for photosynthesis,
301 meaning that photosynthesis is carbon saturated at current concentrations (Gao & McKinley
302 1994; Beardall *et al.* 1998). Experiments have so far been inconclusive, with some species
303 showing carbon saturation at current CO₂ (e.g. Beer & Koch 1996; Israel & Hophy 2002),
304 others demonstrating increased photosynthetic production with increasing CO₂ (e.g.
305 Holbrook *et al.* 1988), and yet others switching the source of carbon with greater CO₂
306 availability (e.g. Johnston & Raven 1990; Schmid *et al.* 1992). Yet, general consensus within
307 the literature seems to be that algae with CCMs will not increase productivity under future
308 conditions (see review by Beardall *et al.* 1998). It is no surprise, then, that the positive effects
309 of CO₂ on algae have not been a substantial part of the climate change literature; if

310 productivity is not enhanced by elevated CO₂, then why look for ecological effects? Our
311 experiments do not clarify this issue with respect to photosynthetic activity of algae; we
312 found a small increase (~ 5 %) in the effective quantum yield of turfs under future
313 concentrations of CO₂, but this seemed to be counteracted by elevated temperature.
314 Nevertheless, it seems that elevated CO₂ conditions can cause an increase in the growth
315 (Kubler *et al.* 1999) and abundance (Andersen & Andersen 2006; Kuffner *et al.* 2008; Russell
316 *et al.* 2009) of non-calcareous algae and this deserves more attention. We propose that
317 elevated inorganic carbon has positive effects on some taxa, and that the non-uniform effects
318 among alternate taxa (review by Gao & McKinley 1994) have relatively unexplored
319 ecological consequences, particularly if growth is limited by sources of inorganic carbon.

320

321 **ACKNOWLEDGEMENTS**

322 Thanks go to J. Thompson and I. Bunker for assistance in the laboratory and D. Gorman in
323 the field. Financial support for this research was provided by an ARC grant to B.D.R. and
324 S.D.C. and an ARC Fellowship to S.D.C.

325

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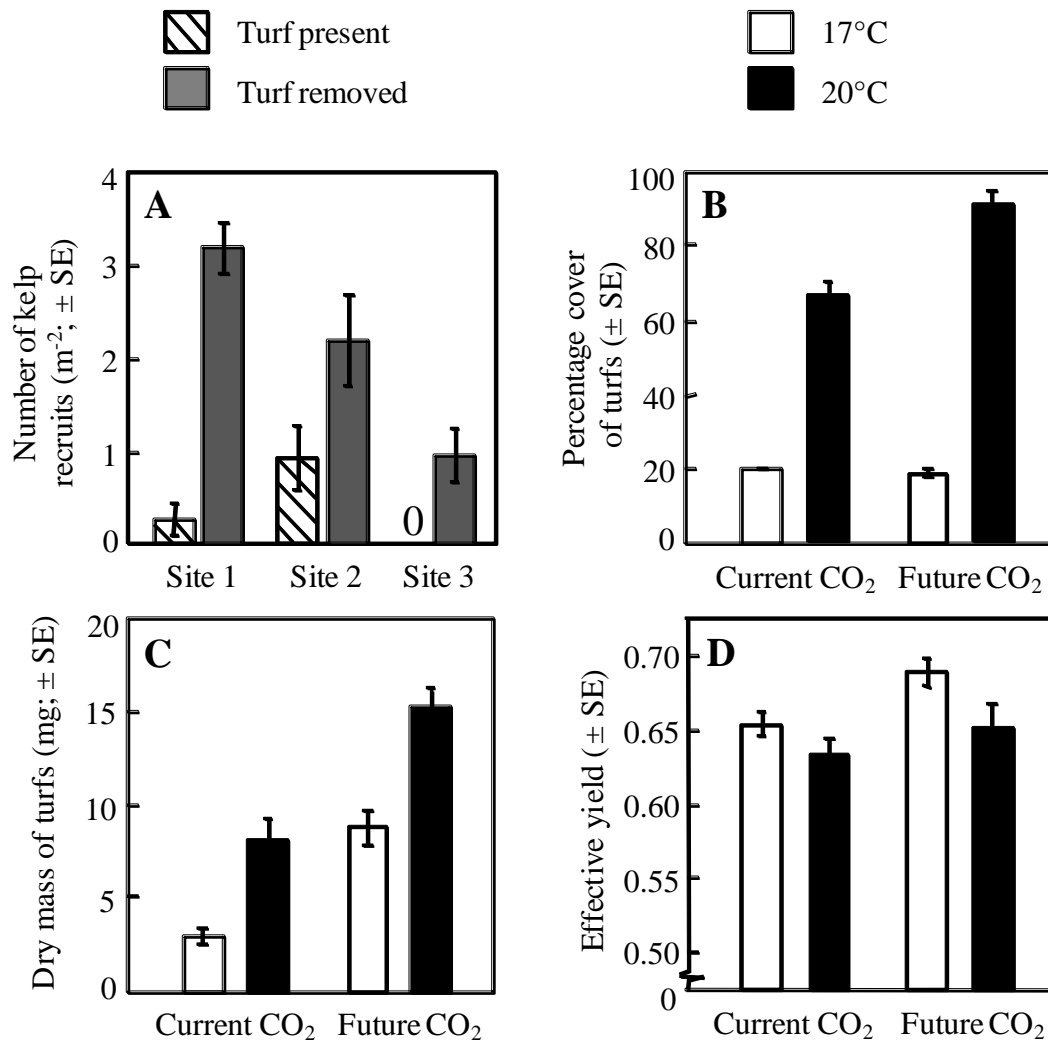
492 **Figure Legend**

493 **Figure 1.** (A) The inhibitory effect of turf on recruitment of kelp at three phase-shifted sites
494 (i.e. kelp-domination to turf-domination) with treatments of turf presence and turf removal,
495 and the effect of forecasted CO₂ and temperature on turfs as observed by (B) recruitment to
496 available space (percentage cover), (C) biomass (dry mass), and (D) effective quantum yield.
497 Note: “0” in A signifies no kelp recruits.

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500 **Fig. 1.**



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