1	Anthropogenic disturbance of deep-sea megabenthic assemblages: a study with						
2	Re	motely-Operated Vehicles in the Faroe-Shetland Chanel, NE Atlantic					
3							
4		Jones, D. O. B. ¹ , Wigham, B. D. ² , Hudson, I. R. ¹ and Bett, B. J. ¹					
5							
6	1	DEEPSEAS Group, George Deacon Division, National Oceanography					
7		Centre, Southampton. SO14 3ZH. UK.					
8	2	Dove Marine Laboratory, School of Marine Science and Technology,					
9		Newcastle University, Cullercoats. NE30 4PZ. UK.					
10							
11	Correspor	ading Author:					
12		Daniel O. B. Jones					
13		e-mail: dj1@noc.soton.ac.uk					
14		tel: +44 (0) 2380 596 357					
15		fax: +44 (0) 2380 596 247					
16							
17							

1 ABSTRACT

2

3 The effects of local-scale anthropogenic disturbance from active drilling platforms on 4 epibenthic megafaunal abundance, diversity and assemblage pattern were examined in 5 two West of Shetland hydrocarbon fields at 420 m and 508 m water depth. These 6 areas were selected to include a range of disturbance regimes and contrasting faunal 7 assemblages associated with different temperature regimes. Remotely Operated 8 Vehicle (ROV) video provided high-resolution megafaunal abundance and diversity 9 data, which were related to the extent of visible disturbance from drilling spoil. These 10 data, in conjunction with a study deeper in the Faroe-Shetland Channel, have allowed 11 comparison of the effects of disturbance on megabenthos across a range of sites. 12 Disturbance to megafaunal assemblages was found to be high within 50 m of the 13 source of drill spoil and in areas where spoil was clearly visible on the seabed, with depressed abundances (Foinaven 1900 individuals ha⁻¹; Schiehallion 2178 individuals 14 ha⁻¹) and diversity (H' = 1.75 Foinaven; 1.12 Schiehallion) as a result of smothering 15 16 effects. These effects extended to around 100 m from the source of disturbance, 17 although this was variable, particularly with current regime and nature of drilling 18 activity. Further from the source of disturbance, megafaunal assemblages became 19 more typical of the background area with increased diversity (H' = 2.02 Foinaven; 1.77 Schiehallion) and abundance (Foinaven 16484 individuals ha⁻¹; Schiehallion 20 21 5477 individuals ha⁻¹). Visible effects on megafaunal assemblages as a result of 22 seabed drilling were limited in extent although assemblage responses were complex, 23 being controlled by differing effects to individual species often based on their 24 motility.

1 INTRODUCTION

2

3 Disturbance is an important source of temporal and spatial heterogeneity in 4 natural communities (e.g. Sousa 1984). The importance of disturbance has been 5 highlighted (e.g. Connell 1978) in maintaining species diversity by preventing 6 competitive exclusion by dominant species in an assemblage. Physical disturbance is 7 a key factor in controlling spatial and temporal composition of shallow-water benthic 8 communities. Like shallow waters, deeper waters are now increasingly subject to a 9 range of anthropogenic perturbations that include commercial trawling (Kaiser 1998), 10 mining (Radziejewska and Stoyanova 2000) and increasingly oil exploration (Jones et 11 al. 2006). Oil exploration activities are becoming more important in the Faroe-12 Shetland Channel with the majority of fields located on the upper slope (Figure 1). 13 This area supports a high diversity of deep-water fauna primarily controlled by its 14 unusual temperature regime (Turrell et al. 1999; Bett 2001). This study will compare 15 the effects of disturbance from drilling on megafaunal assemblages in the three major 16 thermal regimes encountered in the Faroe-Shetland Channel between 200 and 1000 m 17 water depth. This study extends that of Jones et al. (2006) to include new megafaunal 18 data from contrasting thermal regimes.

Human induced disturbance in the deep sea typically has a large impact on benthic communities (Bluhm 2001). Deep-sea environments are typically stable in comparison with often more dynamic shallow-water habitats (Gage and Tyler 1991). Physical habitat characteristics are important in controlling benthic community structure (Levin et al. 2001). Anthropogenic disturbance from a variety of sources can alter these characteristics very rapidly by smothering the existing seabed with sediments from elsewhere (Stronkhorst et al. 2003; Jones et al. 2006). In addition,

1 large-scale disturbance can also occur naturally in deeper waters from turbidity 2 currents, debris flows and benthic storms (Gage and Tyler 1991). In communities with 3 limited food supply and with invertebrate assemblages depauperate in both abundance 4 and biomass, disturbance effects are likely to be greater and recovery times longer 5 (Bluhm 2001). The greatest change in communities may be expected to occur in areas 6 where disturbance alters habitat type most radically. The upper slope of the Faroe-7 Shetland Channel harbours an unusual deep-sea habitat consisting of a heterogeneous 8 mix of sediments with a preponderance of hard substratum (cobble, boulder). 9 Introduction of drill spoil acts to reduce hard substratum availability and homogenise 10 the habitat, which can directly influence the abundance, diversity, species composition 11 and distribution of the local benthic fauna (Jones et al. 2006).

12 The effects of anthropogenic disturbance, such as oil drilling activity, on the 13 benthic environment is conventionally assessed by sampling (typically by grab) a 14 range of chemical parameters and occasionally macrofauna from the source of effect 15 at geometrically increasing distances along four radiating transects (Gray et al. 1990; 16 Kingston 1992). The effects on macrofauna are usually recorded as changes in 17 diversity indices (Davies et al. 1989; Kingston 1992), although multivariate 18 approaches may be more effective (Olsgard and Gray 1995). The effects of physical 19 disturbance on whole assemblages is less well known, with shallow-water studies 20 suggesting a range of responses depending on the severity of disturbance and nature 21 of the assemblage (Airoldi 2003; Dernie et al. 2003).

Typical environmental assessments do not specifically address the larger epibenthic megafauna, yet these organisms play an important role in benthic processes (Piepenburg and Schmid 1997). Monitoring of megafauna has been shown to be effective in evaluating the impacts of disturbance on the seafloor (Bluhm 2001).

Photographic studies allow fine scale survey of megabenthic abundance, diversity and
 distribution (Piepenburg and Schmid 1997) and can have a much larger spatial extent
 than is usually possible with conventional macrofaunal sampling techniques.
 Remotely Operated Vehicles (ROV) are ideal tools for such surveys, capable of high
 resolution, systematic video and photographic investigation of epibenthic megafaunal
 assemblages (e.g. Jones et al. 2006).

The main objectives of this study are to: 1) describe the composition, diversity and distribution of megafaunal assemblages on the Faroe-Shetland Channel slope, 2) determine the effect and extent of physical disturbance from drilling operations on benthic megafauna, 3) determine the differences between benthic megafaunal assemblages and their responses to disturbance at two contrasting study sites and 4) compare results from these with existing data collected in an identical manner at a deeper Faroe-Shetland Channel site (Jones et al. 2006).

14

15 MATERIALS AND METHODS

16

17 Study background:

Investigations for this study were carried out aboard the semi-submersible oil drilling platforms Paul B Loyd Junior (PBLJ, 14-28/5/2003) and Transocean Leader (TOL, 27/9-3/10/2003). The PBLJ was operating in the Foinaven field (507-509 m depth, 60°18.68′ N 4°20.33′ W) and TOL in the Schiehallion field (420-421 m depth, 60°22.95′ N 04°05.95′ W). The Foinaven and Schiehallion oil reservoirs were discovered in 1992 and 1993 respectively. Drilling activities began in the Foinaven region in 1997 and in the Schiehallion region in 1998.

1 Drill spoil was deposited on the seabed during the initial phases of drilling 2 (tophole drilling) as sediment is displaced directly. Once this phase is completed all 3 subsequent rock cuttings were recirculated with drilling mud back to the rig where 4 they were cleaned and depending on potential contamination, discharged at the 5 surface or shipped back to land. With the high current regime in the Faroe-Shetland 6 Channel the rock cuttings disposed at the surface were spread widely and very little 7 was deposited on the seabed close to the rig (Aurora 2004). In this environment, the 8 majority of spoil deposited on the seabed close to the drill site was produced for less 9 than a day per well during tophole drilling. In the area investigated at Foinaven, 14 10 wells were present at the time of investigation within 50 m of each other. These wells 11 had been drilled over a 5 year period with the most recent still being drilled during 12 this investigation. Seabed spoil resulting from this drilling operation was produced 14 13 days prior to the start of the investigation (1/5/2003), over a period of approximately 14 24 hours. In the area investigated at Schiehallion there was only one well, which was 15 also being drilled during this investigation. Seabed disturbance from tophole drilling 16 occurred twice owing to operational problems, 12 and 14 days before this 17 investigation began (on 13 and 15/9/2003; both for approximately 24 hours).

18 For survey purposes the seabed around the drilling activity was divided into 50 19 m zones radiating from the outer limit of all seabed installations (Figure 2). The area 20 to the southeast of the Foinaven well could not be surveyed owing to operational 21 constraints. The distance of the ROV from the rig was calculated from transect 22 duration and the length of the ROV tether released from the Tether Management 23 System with an approximate error of ± 1 m. Data were collected using an industry-24 operated work-class Pioneer HD ROV, following the methodology of Jones et al. 25 (2006).

2 Data analysis

3

Abundances were standardised to numbers per hectare. Each transect was
partitioned into 50 m zones and analysis was carried out on data from each zone.
There were 5 zones at Foinaven (0 – 250 m) and 4 at Schiehallion (0 – 200 m, owing
to limited data in the 250 m zone at Schiehallion). Counts for each individual transect
50 m zone formed the sampling unit.

9 A range of univariate diversity indices were calculated to assess both the 10 dominance and species richness aspects of diversity (Magurran 2003). Confidence 11 intervals (95%) for abundance and diversity were calculated using a bootstrapping 12 technique (Manly 1998). As implemented here, 1000 bootstrap samples were 13 calculated and a 95% confidence interval determined from the resultant data.

14 Variations in taxon composition were assessed by multivariate analysis 15 (hierarchical group-average clustering and non-metric multi-dimensional scaling, 16 MDS) following a square root transformation and calculation of Bray-Curtis 17 similarity coefficients (Clarke and Warwick 2001). The difference in assemblage 18 composition between distance zones was assessed using analysis of similarities 19 (ANOSIM). Multivariate dispersion (MVDISP) was used to measure within-zone 20 multivariate assemblage dispersion (Clarke & Warwick 2001). Data analysis was 21 performed using the computer programmes PRIMER (Clarke and Warwick 2001), 22 Biodiversity Pro (Natural History Museum, London and Scottish Association for 23 Marine Sciences, Oban), MATLAB (MathWorks Inc.) and MINITAB (Minitab Inc.).

24

3 Foinaven

4 General observations

5

6 At Foinaven 1075 megabenthic organisms from 33 nominal taxa were recorded in a total area of 1519 m² (Table 1; Figure 3 & 4). Crustaceans were the 7 dominant megafaunal group (47% megafauna; 5253 ha⁻¹), predominantly represented 8 9 by the ubiquitous squat lobster Munida sarsii, but also included hermit crabs, natant 10 decapods and Siphonocetes tube dwelling amphipods. Porifera (27% megafauna, 3382 ha⁻¹) were abundant. Echinoderms (22% megafauna, 2046 ha⁻¹) were dominated by 11 12 Echinus acutus and the holothurian Stichopus tremulus. Asteroids (Porania pulvillus pulvillus, Ceramaster granularis and Henricia pertusa), comatulid crinoids and 13 14 ophiuroids were also present. The remainder (4%) of the megabenthos was made up 15 of molluscs, polychaetes, cnidarians and demersal fish. Seabed structures at Foinaven 16 attracted large numbers of fish (predominantly Sebastes viviparus, Brosme brosme 17 and *Pollachius virens*) but in disturbed areas benthic megafauna were relatively 18 sparse, being largely represented by motile deposit feeders principally echinothurid 19 urchins and Munida sarsii.

Despite the large number of drilling sites in the Foinaven study area, drill spoil was constrained to a ~ 50 m zone around the drill sites. Outside the disturbed area the seabed consisted of a heterogeneous mix of sand, gravel and occasionally larger cobbles and boulders characteristic of the "iceberg ploughmark zone" (Bett 2001; Masson 2001).

3 Megafaunal abundance (Figure 3A) was significantly different between zones 4 (Kruskal-Wallis H = 22.81, df = 4, p < 0.001) increasing with distance from drilling 5 activity at Foinaven (Spearman's rank correlation r' = 0.9, p < 0.05). Large changes in 6 abundance with distance from disturbance were observed particularly for dominant 7 phyla, with Porifera and Crustacea displaying changes in abundance by almost an 8 order of magnitude. A marked increase in abundance between 0 and 100 m from 9 drilling activity was noted for major faunal groups with post-hoc non-parametric 10 multiple comparisons of total faunal abundance (after Miller 1981) revealing 11 significant differences (p < 0.05) between the 0-50 m zone and all other zones, but no 12 significant differences between zones > 50 m from the drilling disturbance. Both 13 motile and sessile taxa abundances were significantly different between zones 14 (Kruskal-Wallis: motile: H = 20.80, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 20.80, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87, df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sessile: H = 23.87; df = 4, p < 0.001; sess 15 (0.05). Sessile taxa increased continuously in abundance from very low values close to 16 the source of disturbance. Motile taxa had low abundances close to disturbance but 17 increased beyond 50 m. Beyond 50 m from the source of disturbance there was no 18 significant differences in motile megafaunal abundance (p > 0.5 in post-hoc non-19 parametric multiple comparisons).

20

21 **Diversity**

22

Univariate diversity measures (Figure 4A) revealed significant changes in diversity with distance from the disturbance. Species richness was lowest close to the source of disturbance, increased to peak values at intermediate distances and dropped slightly in the least disturbed 250 m zone. Heterogeneity diversity (H[']) changed

significantly between zones (Kruskal-Wallis H = 22.31, df = 4, p < 0.001), revealing lowest diversity in the area of drill spoil disturbance, particularly that within structures. This was primarily driven by lack of rarer species, revealed in type I indices (those that emphasis the rarer component of the assemblage). Heterogeneity diversity followed similar trends to species richness, increasing at intermediate distances and dropping slightly further away.

7

8 Composition

9

10 Multivariate analyses showed significant differences in megafaunal 11 assemblages with distance from disturbance (ANOSIM R = 0.56, p < 0.01). Transects 12 between structures were highly scattered in the MDS ordination and cluster diagram 13 (Figure 5&6, respectively) with very low similarity to other zones (37% similarity). 14 Fauna in the 50 m zone on transects away from structures had higher, but still 15 generally low similarity compared to other zones (55%). Areas close to the 16 disturbance displayed high dispersion of sample similarities (PRIMER MVDISP = 17 1.07). Outside the zone of drill spoil there was a faunal transition zone, which 18 occurred between 50 and 100 m from the nearest disturbed area. Beyond this, 19 multivariate similarities were much more similar between samples (MVDISP 20 decreases from 0.425 in 100 m zone to 0.179 for 250 m zone).

- 21 22
- 23 Schiehallion
- 24
- 25 General observations
- 26

At the Schiehallion site a total of 1133 megabenthic organisms from 17 nominal taxa were recorded from a total area of 2715 m² (Table 1). Porifera were the

dominant group (62% megafauna, 2819 ha⁻¹) followed by Echinodermata (26% 1 megafauna, 1170 ha⁻¹) which were dominated by *Cidaris cidaris* urchins, the 2 3 holothurian Stichopus tremulus and various asteroids (Porania pulvillus pulvillus, 4 Ceramaster granularis, Asterias rubens and Henricia pertusa). Crustaceans (11% total megafauna, 495 ha⁻¹) were predominantly represented by Munida sarsii but also 5 6 included hermit crabs, Geryon sp. and Cancer pagarus. The remainder (1%) of the 7 megabenthos comprised molluscs and polychaetes. In the area of drill spoil fish were 8 present (predominantly Gadus morhua, Helicolenus dactylopterus dactylopterus and 9 Molva molva), and were most abundant around drilling structures although in lesser 10 numbers than at Foinaven.

At Schiehallion, despite there being only one drill site, the extent of spoil was greater than at Foinaven, extending to over 155 m in places. Outside the disturbed area the seabed consisted of a heterogeneous mix of sand, gravel and occasionally larger cobbles and boulders.

15

16 Abundance

17

18 Large changes in abundance with distance from drilling activity were 19 observed, particularly for dominant phyla. Megafaunal abundance (Figure 3B) was 20 significantly different between zones (Kruskal-Wallis H = 15.45, df = 3, p < 0.001); 21 increasing with distance from drilling activity (r' = 0.982, p < 0.001). A large increase 22 in abundance between 0-50 and 50-100 m from drilling activity was noted for major 23 faunal groups with post-hoc non-parametric multiple comparisons of total faunal 24 abundance (after Miller 1981) revealing significant differences (p < 0.05) between the 25 0-50 m zone and all other zones but no significant differences between zones > 50 m 26 from the drilling disturbance. Both motile and sessile taxa abundances were

1	significantly different between zones (Kruskal-Wallis: motile: $H = 26.96$, $df = 3$, p <
2	0.001; sessile: $H = 8.11$, $df = 3$, $p < 0.05$). Sessile taxa abundance increased
3	continuously with distance from disturbance while motile faunal abundances peaked
4	at intermediate distances before declining again between 150-200 m.

6 **Diversity**

7

8 Species richness was low close to the source of disturbance, increased to a 9 maximum in the 150 m zone and dropped slightly in the zone furthest from 10 disturbance (Figure 4B). Significant differences in heterogeneity diversity were 11 observed between zones in H^{\prime} (Kruskal-Wallis H = 17.63, df = 3, p < 0.001; Figure 12 4B), owing to depressed megafaunal diversity in the 50 m zone (particularly in indices 13 weighted towards rarer species). Outside this zone there was no significant difference 14 in H^{\prime} (post-hoc non-parametric multiple comparisons p > 0.05).

15

16 **Composition**

17

18 Multivariate analyses (Figure 5&6) showed significant differences in 19 megafaunal assemblages with distance from disturbance (ANOSIM R = 0.26, p < 20 0.05). The fauna in the 50 m zone were highly scattered in the MDS ordination 21 (65.86% similarity). While distinct from other stations, the 50 m zone samples had 22 high within zone dispersion of samples (MVDISP = 1.24). Beyond this transition zone 23 diversity was high and distance from the source of drilling did not appreciably affect 24 the assemblage, all of these outer zones formed a grouping on the MDS plot and 25 showed high similarity with cluster analysis (> 85%; Figure 5&6). There was 26 relatively low dispersion of within zone samples (mean MVDISP = 0.93).

2 Comparison between sites

3

Total megafaunal abundance was higher in Foinaven except in the areas less than 50 m from the source of disturbance. Motile faunal abundance followed similar patterns at both sites but was typically around 4 times greater in Foinaven. Sessile faunal abundances were considerably lower close to disturbance in Foinaven but were approximately equal outside this area.

9 Megafaunal species richness and heterogeneity diversity were significantly 10 lower at Schiehallion when compared to Foinaven (based on grand site totals: 11 Schiehallion S = 18, H' = 1.66; Foinaven S = 33, H' = 2.05) despite the larger survey 12 extent at Schiehallion. Notable differences in megabenthic assemblage composition 13 were observed between Schiehallion and Foinaven (ANOSIM R = 1.00, p < 0.01; 14 Figure 6). Although 15 of the observed taxa were common to both areas, there were 15 some notable differences in important taxa. For example *Cidaris cidaris* was the only 16 echinoid observed at Schiehallion, however at Foinaven no cidarids were recorded 17 and large numbers of Echinus acutus and other Echinus sp. urchins dominated 18 instead. There were also differences amongst megafaunal scavengers: at Schiehallion 19 crabs appeared to be predominant (particularly Geryon sp.), whereas at Foinaven 20 these were not present and natant decapods and whelks were more common.

21

22 **DISCUSSION**

23

24 Changes in megabenthic assemblages with disturbance

1 The extent of disturbance was shown to drive changes in megafaunal 2 abundance; low megafaunal numbers were associated with drill spoil and close to 3 sites of recent drilling impact. Physical smothering and burial of organisms was likely 4 to be the most important cause for reduction in megafaunal numbers (Stronkhorst et 5 al. 2003). Highly motile organisms responded by moving away from the disturbance, 6 as has been found in other studies (e.g. Bluhm 2001; Jones et al. 2006). For less 7 motile taxa, reduced motility led to increased mortality. Where disturbance was 8 partial, the megafaunal response to disturbance in this study was based not only on 9 motility but also on feeding mode, particle removal rate and degree of disturbance.

10 Sessile megafauna increased in abundance with a reduction in disturbance. 11 Impact of drilling disturbance on sessile forms was related directly to their ability to 12 clear particles from their feeding and respiratory surfaces as shown in many sessile 13 shallow-water organisms (Rogers 1990). Sessile megafauna were less disturbed at 14 Schiehallion where abundance was significantly greater than Foinaven (particularly in 15 the area close to disturbance) as a result of reduced overall disturbance. At the deep-16 water (600 m) Laggan site sessile fauna showed a similar response to Foinaven. The 17 Laggan site is also situated in the Faroe-Shetland Channel ($60^{\circ}57^{\circ}N$, $02^{\circ}53^{\circ}W$) in an 18 area with similar substratum but colder seabed temperatures (-1 to 2° C) than those 19 investigated here (Jones et al. 2006).

Megafaunal species diversity generally increased with distance from the point of disturbance as reduced levels of sedimentation increased survival of sessile and other less resilient organisms. Some diversity indices showed a small decrease in diversity at maximal distance and minimal disturbance. Diversity was depressed by high disturbance, but intermediate levels may have increased diversity levels through influx of vagrant scavenging animals or motile fauna taking advantage of decreased competition as a result of reduced numbers of dominant species (Connell 1978) as was also found to occur at Laggan (Jones et al. 2006). Multivariate measurements for the whole assemblage revealed an increased similarity of megabenthic assemblages with decreased disturbance. This trend has been commonly observed in community measures in many marine disturbance settings (Clarke and Warwick 2001) and also found at the Laggan site deeper in the Faroe-Shetland Channel (Jones et al. 2006).

7 Timing and extent of disturbance appears to have been an important factor in 8 this study with the least disturbed Schiehallion site having less discernable changes in 9 assemblage structure than the repeatedly disturbed Foinaven site. The Laggan site 10 (Jones et al. 2006) had two drilling events similar to Schiehallion but relatively higher 11 disturbance (greater coverage of drill spoil) and greater changes in assemblage 12 structure were observed. Frequent disturbance has been shown to have dramatic and 13 long lasting effects on shallow-water communities of the North Sea (Stronkhorst et al. 14 2003) and a similar effect would be expected in deeper water. It was apparent, 15 however, that there was some immigration of selected mobile megafaunal taxa into 16 disturbed zones, this also occurred at Laggan (Jones et al. 2006) and has been found in 17 studies of fishing disturbance (Ramsay et al. 1998). With disturbance from drilling 18 leading to reductions in suspension feeder abundance and an increase in availability of 19 fine particles of high organic matter content it is likely that deposit feeding forms such 20 as echinoids and holothurians may preferentially colonise drill spoil as individual 21 animals can select and retain fine particles without the need to sort through more 22 heterogeneous sediment complexes (Hudson et al. 2004). Although this study was 23 based on two distinct points in time the first phases of recovery were already apparent. 24 Physical disturbance observed at the study sites resulted in complete coverage 25 with sediment (presumed mortality) and potentially non-lethal effects from physical

1 smothering. These effects have also been observed in studies on disposal grounds for 2 dredged material (Stronkhorst et al. 2003), however the associated chemical changes 3 were not investigated in this study. Increase in drilling derived particulates as a result 4 of disturbance may have lead to non-lethal effects such as clogging of filter feeding 5 apparatus of some organisms (Sharma et al. 2001). Although redistribution of nutrient 6 rich subsurface layers could lead to an increase in population size over time 7 (Raghukumar et al. 2001; Sharma et al. 2001), it is likely that the initial impacts of 8 changes in seabed habitat will have had the dominant effect on the benthic 9 communities.

10 The composition of seabed sediments changed as a result of drilling activity, 11 from a heterogeneous substratum with extensive exposed hard surfaces to a 12 homogeneous soft substratum. This change further reduced diversity and although 13 changes are difficult to separate from those directly related to disturbance, reduction 14 in habitat heterogeneity has been shown to reduce diversity in the deep sea (Levin et 15 al. 2001). Smothering of existing sediment with that of a different composition 16 resulted in conditions unfavourable to existing communities and would therefore 17 reduce rates of re-colonisation and larval settlement, potentially prolonging recovery 18 (Snelgrove et al. 1999). Changes in substratum may also have favoured particular 19 faunal elements in the existing communities, increasing dominance and altering 20 community composition.

Outside of the area impacted by drill spoil there was a highly heterogeneous distribution of benthic megafauna. Distribution of megafauna in these 'natural' seabed areas seems primarily driven by availability of suitable microhabitats as has been found elsewhere in the Faroe-Shetland Channel (Fautin et al. 2005; Tyler et al. 2005; Jones et al. 2006). The stochastic arrangement of ice rafted larger stones may have

gone some way to structuring the distribution of megabenthos. Most sessile filter feeders lived attached to hard substrata; whereas many echinoderms preferred softer sediments, being more common on gravel and sandy areas of seabed. Many species, particularly *Munida sarsii* were cryptic, preferring to live under rocks. Strong currents (up to 0.5 ms⁻¹) observed at both sites may have had an important effect on the distribution of megafauna as have been observed by Rosenberg (1995) and Flach et al. (1998).

8

9 Comparison of the undisturbed assemblages of Schiehallion and Foinaven

10

11 Megafaunal abundance at both sites was variable (from 1,900 to 16,483 individuals ha⁻¹ at Foinaven and from 2,178 to 5,626 individuals ha⁻¹ at Schiehallion). 12 13 Megabenthic abundance has generally been found to decrease with depth (Thurston et 14 al. 1994; Piepenburg et al. 2001); however in the Faroe-Shetland Channel the 15 situation is more complex, with warm Atlantic waters overlaying cold Arctic water 16 (Turrell et al. 1999) with some indication of higher macrofaunal abundances in the 17 deeper cold water compared with the shallower warmer waters (Bett 2001). Results 18 from this study and Jones et al. (2006) suggest this may extend to megafauna. At 19 Foinaven megafaunal abundances were higher than that of Schiehallion. The fauna at 20 Foinaven were at a depth where they must experience wide temperature variations (of 21 around 5°C: from -0.5 to 4.5°C), with abundances similar to those found at deeper 22 sites (e.g. Laggan) characterised by Arctic water masses with temperatures between -1 23 and $2^{\circ}C$ (Jones et al. 2006), however few representatives of the typical Arctic faunas 24 found at greater depths in the Faroe-Shetland Channel extend into the present study 25 sites (Jones et al. 2006). Comparison with the Atlantic fauna of the Rockall Trough,

south of the Wyville-Thomson Ridge, revealed similar assemblages, particularly to
 those at Schiehallion (Gage 1986) although comparable megafaunal abundance values
 are not quoted. These comparisons suggest stronger affinities between the shallow
 Faroe-Shetland Channel and the northeast Atlantic rather than with the Norwegian
 Basin and other northern waters.

6 Megafaunal species richness values were recalculated as $ES_{(70)}$ for comparison 7 with other literature (10.8 to 15.1 for Foinaven; 9.3 to 10.7 for Schiehallion). Species 8 richness at Laggan ($ES_{(70)}$ for Laggan between 11.7 and 12.2) was comparable with 9 Foinaven but higher than Schiehallion. This supports the hypothesis that megafaunal 10 richness patterns are similar to those found in macrofauna, with increased diversity at 11 intermediate depths in the Faroe-Shetland Channel (Bett 2001; Narayanaswamy et al. 12 2005). Richness in this study was high in comparison with the Arctic stations 13 investigated by Starmans and Gutt (2002) although these were within the confidence 14 limits for the more diverse shallow Greenland station (Starmans and Gutt 2002). In 15 comparison with the Atlantic, although direct megafaunal diversity measures are 16 unavailable, from species tables and figures it appears that megafaunal diversity in the 17 two areas was similar, with a similar species complement (Gage 1986; Gage et al. 18 2000).

19 There was a clear difference in megafaunal species composition between 20 Foinaven and Schiehallion; this is likely to be predominantly driven by temperature. 21 The fauna at the Schiehallion and Foinaven sites are in an area of natural transition 22 between those organisms more typical of the Atlantic in the warmer shallower waters 23 of the Faroe-Shetland Channel, and those more typical of the Arctic deep Norwegian 24 Sea (Bett 2001). The boundary between warm and cold waters in the Faroe-Shetland 25 Channel oscillates between 400-600 m (Turrell et al. 1999). The fauna of Schiehallion

1 therefore predominantly live in comparatively warm Atlantic waters. The fauna of 2 Foinaven on the other hand is subject to extreme changes in temperature over very 3 short time scales. The fauna at Laggan, living in constantly cold temperatures, was 4 different again from Foinaven, with only 5 taxa common to both areas (Jones et al. 5 2006). Hydrographic regimes are important in structuring benthic communities (Gage 6 et al. 1995), and environmental temperature is a major contributing factor governing 7 the range of species found in marine communities (Gage and Tyler 1991) particularly 8 in the Faroe-Shetland Channel (Bett 2001; Narayanaswamy et al. 2005). It is also 9 likely that differences in specific taxa may be related to bathymetric gradients in 10 faunal distribution (Gage and Tyler 1982; Rex et al. 1997).

11 This study represents an important step forward in quantifying the effects of 12 anthropogenic disturbance across a number of sites in deep waters, being especially 13 relevant in the context of increasing hydrocarbon drilling at deep-water sites. The use 14 of ROVs for monitoring has been shown to be highly effective in studies of this 15 nature, which, as this technology is routinely used in these developments, may 16 increase industry and science collaboration initiatives in monitoring disturbance and 17 the subsequent recovery of benthic assemblages. Disturbance was shown to have 18 important effects on benthic assemblages particularly through smothering and 19 resultant habitat changes. These changes were difficult to predict, based on individual 20 species ecology but the study of assemblage parameters such as abundance, diversity 21 and faunal distribution reveals the ecosystem level effects of disturbance. This work 22 also provides the foundation for future studies monitoring faunal recovery in these 23 areas. It also helps to identify targets for future directed in situ ROV experimental 24 studies of individual species responses to anthropogenic disturbance.

1 ACKNOWLEDGEMENTS

2

The authors wish to acknowledge Transocean Inc. and the crew of the drilling rigs *Transocean Leader* and *Paul B Loyd Junior*, Subsea 7 and their ROV teams and the field operator BP (DBU) for supporting this project. This project was carried out as part of the SERPENT Project and DIEPS (Deepwater Industry, Environment, Policy and Science). D. Jones was funded by NERC studentship NER/S/A/2002/10397 and DIEPS grant NE/C508518/1. This work complies with the current laws of the United Kingdom.

1			

REFERENCES

4	Airoldi L (2003) The effects of sedimentation on rocky coast assemblages.
5	Oceanography and Marine Biology 41: 161-236
6	Aurora (2004) Schiehallion Environmental Statement. Aurora Environmental Ltd,
7	Orkney, UK
8	Bett BJ (2001) UK Atlantic Margin Environmental Survey: introduction and overview
9	of bathyal benthic ecology. Continental Shelf Research 21: 917-956
10	Bluhm H (2001) Re-establishment of an abyssal megabenthic community after
11	experimental physical disturbance of the seafloor. Deep-Sea Research Part II:
12	Topical Studies in Oceanography 48: 3841-3868
13	Brey T (1999) Growth performance and mortality in aquatic macrobenthic
14	invertebrates. Advances in Marine Biology 35: 153-223
15	Clarke KR, Warwick RM (2001) Changes in marine communities: An approach to
16	statistical analysis and interpretation. Plymouth Marine Laboratory, U.K.
17	Connell JH (1978) Diversity in tropical rain forests and coral reefs. Science 199:
18	1302-1310
19	Davies JM, Bedborough DR, Blackman RAA, Addy JM, Appelbee JF, Grogan WC,
20	Parker JC, Whitehead A (1989) Environmental effects of oil-based mud
21	drilling in the North Sea. In: Englehardt FR, Ray JP, Gillam AH (eds) Drilling
22	wastes. Elsevier Applied Science, London
23	Dernie KM, Kaiser MJ, Richardson EA, Warwick RM (2003) Recovery of soft
24	sediment communities and habitats following physical disturbance. Journal of
25	Experimental Marine Biology and Ecology 285-286: 415-434

1	Fautin DG, Daly M, Cappola V (2005) Sea anemones (Cnidaria: Actiniaria) of the
2	Faroe Islands: A preliminary list and biogeographic context. Annales
3	Societatis Scientiarum Faroensis Supplementum XXXXI. BIOFAR
4	Proceedings 2005, BIOFAR symposium, Torshavn 4-26 April 2003: 77-87
5	Flach E, Lavaleye M, de Stigter H, Thomsen L (1998) Feeding types of the benthic
6	community and particle transport across the slope of the NW European
7	Continental Margin (Goban Spur). Progress in Oceanography 42: 209-231
8	Gage JD (1986) The benthic fauna of the Rockall Trough: Regional distribution and
9	bathymetric zonation. Proceedings of the Royal Society of Edinburgh 88: 159-
10	174
11	Gage JD, Lamont PA, Kroeger K, Harvey R (2000) Desktop study of tranches 19-22
12	Section 2.2, Atlantic Frontier Environmental Network CD ROM. Atlantic
13	Frontier Environmental Network (AFEN)
14	Gage JD, Lamont PA, Tyler PA (1995) Deep-sea macrobenthic communities at
15	contrasting sites off Portugal. Preliminary results: 1. Introduction and diversity
16	comparisons. Internationale Revue der gesamten Hydrobiologie 80: 235-250
17	Gage JD, Tyler PA (1982) Depth-related gradients in size structure and the
18	bathymetric zonation of deep-sea brittle stars. Marine Biology 71: 299-308
19	Gage JD, Tyler PA (1991) Deep Sea Biology. Cambridge University Press,
20	Cambridge
21	Gray JS, Clarke AJ, Warwick RM, Hobbs G (1990) Detection of initial effects of
22	pollution on marine benthos: an example from the Ekofisk and Eldfisk
23	oilfields, North Sea. Marine Ecology Progress Series 66: 285-299
24	Hudson IR, Wigham BD, Tyler PA (2004) The feeding behaviour of a deep-sea
25	holothurian, Stichopus tremulus (Gunnerus) based on in situ observations and

1	experiments using a Remotely Operated Vehicle. Journal of Experimental
2	Marine Biology and Ecology 301: 75-91
3	Jones DOB, Hudson IR, Bett BJ (2006) Effects of physical disturbance on the cold-
4	water megafaunal communities of the Faroe-Shetland Channel. Marine
5	Ecology Progress Series 319: 43-54
6	Kaiser MJ (1998) Significance of Bottom-Fishing Disturbance. Conservation Biology
7	12: 1230-1235
8	Kingston PF (1992) Impact of offshore oil production installations on the benthos of
9	the North Sea. ICES Journal of Marine Science 49
10	Levin LA, Etter RJ, Rex MA, Gooday AJ, Smith CR, Pineda J, Stuart CT, Hessler
11	RR, Pawson D (2001) Environmental influences on regional deep-sea species
12	diversity. Annual Review of Ecology and Systematics 32: 51-93
13	Magurran AE (2003) Measuring Biological Diversity. Blackwell Science, Oxford
14	Manly BFJ (1998) Randomization, bootstrap and Monte Carlo methods in biology.
15	Chapman and Hall, London
16	Masson DG (2001) Sedimentary processes shaping the eastern slope of the Faroe-
17	Shetland Channel. Continental Shelf Research 21: 825-857
18	Miller RG (1981) Simultaneous statistical inference. McGraw-Hill, New York
19	Narayanaswamy BE, Bett BJ, Gage JD (2005) Ecology of bathyal polychaete fauna at
20	an Arctic-Atlantic boundary (Faroe-Shetland Channel, North-east Atlantic).
21	Marine Biology Research 1: 20-32
22	Olsgard F, Gray JS (1995) A Comprehensive Analysis of the Effects of Offshore Oil
23	and Gas Exploration and Production on the Benthic Communities of the
24	Norwegian Continental-Shelf. Marine Ecology Progress Series 122: 277-306

1	Piepenburg D, Brandt A, von Juterzenka K, Mayer M, Schnack K, Seiler D, Witte U,
2	Spindler M (2001) Patterns and determinants of the distribution and structure
3	of benthic faunal assemblages in the northern north Atlantic. In: Schafer P,
4	Ritzrau M, Schluter M, Thiede J (eds) The Northern North Atlantic: A
5	changing environment. Springer, Berlin, pp 179-198
6	Piepenburg D, Schmid MK (1997) A photographic survey of the epibenthic
7	megafauna of the Arctic Laptev Sea shelf: distribution, abundance and
8	estimates of biomass and organic carbon demand. Marine Ecology Progress
9	Series 147: 63-75
10	Radziejewska T, Stoyanova V (2000) Abyssal epibenthic megafauna of the Clarion-
11	Clipperton area (NE Pacific): changes in time and space versus anthropogenic
12	environmental disturbance. Oceanological Studies. Gdansk 29: 83-101
13	Raghukumar C, Loca Bharathi PA, Ansari ZA, Nair S, Ingole BS, Sheelu G,
14	Mohandass C, Nath BN, Rodrigues N (2001) Bacterial standing stock,
15	meiofauna and sediment nutrient characteristics: indicators of benthic
16	disturbance in the Central Indian Basin. Deep-Sea Research Part II: Topical
17	Studies in Oceanography 48: 3381-3399
18	Ramsay K, Kaiser MJ, Hughes RN (1998) Responses of benthic scavengers to fishing
19	disturbance by towed gears in different habitats. Journal of Experimental
20	Marine Biology and Ecology 224: 73-89
21	Rex MA, Etter RJ, Stuart CT (1997) Large-scale patterns of species diversity in the
22	deep-sea benthos. Cambridge University Press, New York
23	Rogers CS (1990) Reponses of coral reefs and reef organisms to sedimentation.
24	Marine Ecology Progress Series 62: 185-202

1	Rosenberg R (1995) Benthic marine fauna structured by hydrodynamic processes and
2	food availability. Netherlands Journal of Sea Research 34: 303-317
3	Sharma R, Nath BN, Parthiban G, Sankar SJ (2001) Sediment redistribution during
4	simulated benthic disturbance and its implications on deep seabed mining.
5	Deep-Sea Research Part II: Topical Studies in Oceanography 48: 3363-3380
6	Snelgrove PVR, Grassle JP, Grassle JF, Petrecca RF, Ma HG (1999) In situ habitat
7	selection by settling larvae of marine soft- sediment invertebrates. Limnology
8	and Oceanography 44: 1341-1347
9	Sousa WP (1984) The role of disturbance in natural communities. Annual Review of
10	Ecology and Systematics 15: 353-391
11	Starmans A, Gutt J (2002) Mega-epibenthic diversity: a polar comparison. Marine
12	Ecology Progress Series 225: 45-52
13	Stronkhorst J, Ariese F, Van Hattum B, Postma JF, De Kluijver M, Besten PJD,
14	Bergman MJN, Daan R, Murk AJ, Vethaak AD (2003) Environmental impact
15	and recovery at two dumping sites for dredged material in the North Sea.
16	Environmental Pollution 124: 17-31
17	Thrush SF, Hewitt JE, Cummings VJ, Dayton PK (1995) The impact of scallop
18	dredging on marine benthic communities: what can be predicted from the
19	results of experiments? Marine Ecology Progress Series 129: 141-150
20	Thurston MH, Bett BJ, Rice AL, Jackson PAB (1994) Variations in the invertebrate
21	abyssal megafauna in the North Atlantic Ocean. Deep-Sea Research Part I:
22	Oceanographic Research Papers 41: 1321-1348
23	Tuck I, Hall SJ, Robertson M, Armstrong E, Basford D (1998) Effects of physical
24	trawling disturbance in a previously unfished sheltered Scottish sea loch.
25	Marine Ecology Progress Series 162: 227-242

1	Turrell WR, Slesser G, Adams RD, Payne R, Gillibrand PA (1999) Decadal
2	variability in the composition of Faroe Shetland Channel bottom water. Deep-
3	Sea Research Part I: Oceanographic Research Papers 46: 1-25
4	Tyler PA, Emson RH, Sumida PYG, Howell KL (2005) Ophiuroid distribution at
5	sublittoral and bathyal depths round the Faroe Islands, NE Atlantic Ocean.
6	Annales Societatis Scientiarum Faroensis Supplementum XXXXI. BIOFAR
7	Proceedings 2005, BIOFAR symposium, Torshavn 4-26 April 2003: 175-194
8	
9	

1 **FIGURES**

2

3 Figure 1: Bathymetry of the West of Shetland area, north of Scotland, UK, showing 4 the position of the sampling sites at the Schiehallion and Foinaven fields (cross 5 symbol). Laggan site also identified (star symbol) for comparison with Jones et al. 6 (2006).7 8 Figure 2: ROV video transects conducted at Foinaven and Schiehallion fields, West of 9 Shetland, showing the extent of the visible drill spoil, subsea structures and 50m 10 zones radiating from sources of disturbance. 11 12 Figure 3: Abundances of motile and sessile megafauna in Foinaven (A) and 13 Schiehallion (B) fields, West of Shetland. Error bars represent 95% confidence 14 intervals derived from bootstrapping. 15 16 Figure 4: Alpha species diversity from ROV video survey of megabenthos in 17 Foinaven (A) and Schiehallion (B) fields, West of Shetland. Shannon-Wiener Index, 18 (H' log e), Total number of taxa observed (S), plotted with distance zones from source

- 19 of disturbance.
- 20

21 Figure 5: Multidimensional scaling ordination (based on Bray-Curtis similarities

22 computed from root transformed abundances) of megafauna from ROV video footage

23 in 50m zones from drilling disturbance at Foinaven and Schiehallion fields, West of

24 Shetland.

25

1	Figure 6: Percentage similarity of ROV megafaunal video transects based on
2	Hierarchical cluster analysis (based on Bray-Curtis similarities of root transformed
3	abundances) of megafauna from ROV video footage in 50m zones radiating from
4	drilling disturbance at Foinaven and Schiehallion fields, West of Shetland.

1 **TABLES**

2

3 Table 1: Densities of megafaunal taxa within concentric 50 m zones around two oil

- 4 drilling sites, Foinaven and Schiehallion, West of Shetland. Species densities (no ha⁻¹)
- 5 tabulated by distance from source of disturbance (50 meter distance zones).
- 6







Distance from disturbance, m







Schiehallion

Foinaven

Table 1:

				Foinaven				Schiehallion			
Phylum	Class	Species	0-50	50- 100	100- 150	150- 200	200- 250	0-50	50- 100	100- 150	150- 200
Porifera	demospongia	Indet. sponge 1				12		18	44	29	18
		Indet. sponge 2		192	824	3	22				
		Indet. sponge 3				6					
		Indet. sponge 4	12	192	165	24	44	35	117	352	233
		Indet. sponge 5	169	1779	1319	212	4176	1611	2168	255	3142
		Hymedesmia paupertas?		48		12	73			44	9
		Indet. encrusting sponge 1			22	12					
		Indet. encrusting sponge 2			55						
		sponge 3		336	495	24	293	18	13	59	9
		Aplysilla sulphurea?		529	659	91	513	35	117	249	198
Cnidaria	actiniaria	Indet. actinarian 1			55						
		Indet. actinarian 2	12			6					
Annelida	polychaeta - errantia	Indet. errant polychaete	12	96	549	3	73		15		
Mollusca	gastropoda	Indet. buccinid	12	48	22	12	513				
	bivalvia	Indet. pectenid		48	55					15	18
	cephalopoda	Sepiola atlantica			55						
Arthropoda	decapoda	Pandalus borealis	61		165						
		Geryon sp.						18	15	29	72
		Pagarus sp.	36	192	385	18	147			15	
		Cancer pagarus									18
		Munida sarsii	666	69	555	65	733	89	791	63	35
	amphipoda	Siphonocetes sp.	12	144	55	12					
Echinodermata	crinoidea	Indet. comatulid				12	147				
	ophiuroidea	Indet. ophiuroid	19	144		12	147				
	asteroidea	granularis		48	55	6	73	35	44	132	18
		Asterias rubens	12						29	29	36
		Henricia pertusa Porania pulvillus			55	6	147	53	132	161	18
		pulvillus			55			18		15	
	echinoidea	echinothurid	641	1634	824	121	1538				
		Echinus acutus		48			73				
		Echinus sp.	12	144	165	48	293				
		Cidaris cidaris						177	835	1143	88
	holothuria	Stichopus tremulus	19	769	385	3	44	71	293	22	233
Chordata	chondrichthyes	Chimaera monstrosa	12				73				
	osteichthyes	Lophius piscatorius	12								
		Paraliparis sp.					73				
Total		Abundance	1900	12402	11868	13151	16484	2178	4703	5626	5477
		Number of taxa	16	18	22	22	20	12	13	16	15