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Investigation of a *Microcystis aeruginosa* cyanobacterial freshwater harmful algal bloom associated with acute microcystin toxicosis in a dog

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1   **TITLE PAGE**

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15   Running title: Cyanobacterial bloom investigation and microcystin toxicosis in a dog

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1   **Abstract.** Microcystin poisoning was diagnosed in a dog exposed to a *Microcystis aeruginosa*-

2   dominated freshwater harmful algal bloom at Milford Lake, Kansas, which occurred during the

3   summer of 2011. Lake water microcystin concentrations were determined at intervals during the

4   summer, using competitive enzyme-linked immunosorbent assays, and indicated extremely high,

5   localized microcystin concentrations of up to 126,000 ng/ml. Multiple extraction and analysis

6   techniques were utilized in the determination of free and total microcystins in vomitus and liver

7   samples from the poisoned dog. Vomitus and liver contained microcystins, as determined by

8   enzyme-linked immunosorbent assays, and the presence of microcystin LR was confirmed in

9   vomitus and liver samples using liquid chromatography coupled with tandem mass spectrometry.

10   Major toxic effects in a dog presented for treatment on the day following exposure included

11   fulminant liver failure and coagulopathy. The patient deteriorated rapidly in spite of aggressive

12   treatment, and was euthanized. Postmortem lesions included diffuse, acute, massive hepatic

13   necrosis and hemorrhage, and acute necrosis of the renal tubular epithelium. A diagnosis of

14   microcystin poisoning was based on the demonstration of *M. aeruginosa* and microcystin-LR in

15   the lake water, as well as in vomitus produced early in the course of the poisoning, the presence

16   of microcystin-LR in liver tissue, and on a typical clinical course.

17   **Key words:** *Microcystis aeruginosa*; microcystin; liver failure; renal failure.

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Introduction

We describe a case of acute microcystin poisoning in a dog following ingestion of lake water during an algal bloom containing *Microcystis aeruginosa* cyanobacteria. Ingestion of an acutely toxic dose of microcystins by dogs typically results in fulminant liver failure.[7,21]

Cyanobacteria are among the most ancient organisms, and have adapted to a wide range of marine and terrestrial habitats.[1,10] Aquatic cyanobacteria, also known as blue-green algae, are important primary producers in aquatic ecosystems.[17] They assimilate organic nutrients, inorganic nutrients, carbon dioxide, and light to produce biomass through oxygenic photosynthesis.[23] Cyanobacteria are present in most aquatic ecosystems, but algal blooms occur when physical-chemical conditions allow rapid growth and reproduction, and populations are not effectively constrained by predation.[19]

Hazardous blooms are often referred to as HABs (harmful algal blooms), or FHABs (freshwater harmful algal blooms) when they occur in freshwaters.[14] Several species of cyanobacteria produce toxins. Microcystins, produced by *Planktothrix* spp*./Oscillatoria* spp., *Microcystis* spp., *Coelosphaerium* spp., and *Anabaena* spp., are the most commonly found cyanobacterial toxins in FHAB's in the Midwestern United States.[11] Microcystins are a family of cyclic heptapeptides, which are potent inhibitors of the serine threonine family of protein phosphatases.[25] Microcystin production, and consequently poisoning risks during an FHAB, are dependent on the bacterial strains involved in the FHAB.[6]

FHABs are increasing across the world, and it raises growing concerns about impacts on water use for recreation and commerce, potable water availability, and human and animal disease.[14] The costs incurred due to directs losses and control measures are high. In the United States, for example, the annual costs range between 2.2 and 4.4 billion US$.[8] The general public is often ignorant of the occurrence of FHABs, or is unaware of the severity of the associated health risks. Continued use of public waters for recreation during FHABs potentially leads to poisoning, such as in the case described here.

Clinical history

A 6 year old, intact Briard bitch, weighing 35.5 kg, was presented to the Veterinary Medical Teaching Hospital (VMTH) at Kansas State University at 12:30 pm on 09/25/2011. The dog had visited Milford Lake, a publicly accessible, 6,357 hectare lake straddling Clay, Geary, and Dickenson Counties, Kansas, the previous day. The dog had walked along the lake shore at a swimming beach close to the town of Milford, and drank water from the lake. In the evening, the dog lost its appetite, vomited green-colored fluid, and developed diarrhea. About 50 ml vomitus from the second vomiting event was collected by the owner, and was submitted to the Kansas State Veterinary Diagnostic Laboratory (KSVDL) at Kansas State University, on 09/26/2011. Intermittent vomiting and diarrhea continued during the morning of the following day, and the dog became recumbent and unresponsive at about 11:00 am. At this time, approximately 24 hours after exposure, the dog was transported to the VMTH, and a possible seizure during transport was reported by the owner. On arrival at the VMTH the dog was laterally recumbent. It had a temperature of 37.2 ^{0}C, a pulse rate of 120 beats per minute, and a respiratory rate of 20

breaths per minute. The mucous membranes were pink, and the capillary refill time was less than

two sec. There were no abnormal lung or heart sounds. A rectal examination indicated the

presence of melena and hemorrhagic diarrhea, and the dog displayed signs of discomfort on

abdominal palpation.

Whole blood and serum samples were submitted to the KSVDL for routine blood cell counts and

blood chemistry panels (Table 1). Blood cell counts revealed a mild lymphopenia. The

hematocrit, erythrocyte concentration, and hemoglobin concentration were high. The blood

plasma was icteric. The prothrombin time and activated partial thromboplastin time were

markedly increased. The blood chemistry panel revealed a moderately low blood glucose

concentration, mildly elevated creatinine, elevated phosphorus, low bicarbonate, an increased

calculated anion gap, markedly elevated alanine transaminase, elevated alkaline phosphatase,

elevated creatine kinase, and elevated total bilirubin.

An intravenous catheter was placed, and approximately 400 ml of sodium chloride 0.9% were

given as a bolus. Also, given the hypoglycemia, 30 ml of 50% dextrose mixed with 30 ml of

saline were bolused upon presentation. Serum glucose increased to 268 mg/dl, hematocrit

decreased to 55% and total solids decreased to 5 g/dl. Following the bolus, intravenous fluid

therapy was continued, at twice the maintenance rate (5ml/kg/h), using sodium chloride 0.9%

supplemented with 2.5 % dextrose. Vitamin K1 was administered subcutaneously, at a dose of 5

mg/kg, and 2 ml of vitamin B complex[a] were added to the first bag of saline. Two hours later,

the rate was decreased to maintenance, and a fresh frozen plasma (FFP) transfusion was started

after placing a second intravenous catheter. A total of 520 ml of FFP (approximately 15 ml/kg)

were given over 4 hours, by slowly increasing the rate of administration. At the end of the

transfusion, crystalloids were switched from saline to Normosol-R (supplemented with 15 mEq

KCl and 2.5% dextrose) and the rate was increased back to two times maintenance. N-acetyl

cysteine was administered intravenously at a dose of 140 mg/kg. Intermittent, hemorrhagic

diarrhea, and intermittent vomiting producing blood-tinged vomitus, continued. Dolasetron, a

serotonin 5-HT$_3$ receptor antagonist with both central and peripheral antiemetic properties [18],

was administered intravenously at a dose of 0.6 mg/kg to control vomiting.

After six hours of fluid therapy, less than 10 ml of urine had been produced. A urinary catheter

was placed at 7 hours after presentation, and 120 ml of dark urine was drawn off. A small sample

was submitted for complete urinalysis. Urine was dark-yellow and cloudy. Urinalysis showed a

specific gravity of 1.027, urine pH of 8.5, 2+ bilirubin with a positive Ictotest, 3+ heme and 3+

protein (>100 mg/dl according to sulfasalicylic acid test). Microscopic examination of the urine

sediment revealed 10 to 50 erythrocytes per high power field, a few large round epithelial cells

and many amorphous crystals. Intermittent diarrhea and vomiting continued, and another

antiemetic, maropitant, was administered subcutaneously at a dose of 1 mg/kg. At 9 hours after

presentation the temperature was 38.3 ^{0}C, the pulse rate was 124 beats per minute, and the

respiratory rate was 40 breaths per minute. The mucous membranes were slightly icteric, and the

capillary refill time was less than 2 sec. The packed cell volume had decreased to 43 %, and the

blood glucose concentration was at 128 mg/dl. A second dose of N-acetyl cysteine was

administered at 70 mg/kg intravenously.

At 11 hours after presentation, and four hours after placing the urinary catheter, no additional

urine had been produced. A brief ultrasound of the caudal abdomen showed a small urinary

bladder and a patent urinary catheter, indicating that the patient had been anuric. The rate of

fluid administration had been decreased to maintenance, and a second transfusion of FFP was

started. Despite the intensive care, the patient's depressive state continued to deepen. The patient

was euthanized at 2:00 am, using an overdose of pentobarbital sodium, and placed in a

refrigerator until necropsied the following day at 1:30 pm at the Kansas State Veterinary

Diagnostic Laboratory.

Materials and methods

Tissue sampling

A necropsy was performed, and tissue samples, including liver, kidney, spleen, lung,

myocardium, the mesentery, uterus, stomach and intestines, bladder, skeletal muscle, tonsils,

salivary glands, and brain, were fixed in 10% neutral buffered formalin. Tissue sections were

processed routinely, embedded in paraffin, and stained with hematoxylin and eosin and the

kidney sections were stained by the periodic acid-Schiff (PASH) method, using a haematoxylin

counterstain. Samples for toxin analysis, including a fresh liver sample for toxin analysis, and an

early vomitus sample collected by the dog's owner, were frozen at -20°C.

Vomitus microscopy

Microscopy was done on an unstained, early vomitus sample, using a stereo microscope[b], by

placing drops of vomitus on a glass slide, and covering them with a cover slip. A magnification

range from 23-184 x was used to locate and identify colonies. Illuminating the sample from

above assisted in locating colonies under lower magnifications because of their greenish

coloration. Higher magnifications were then used for identification.

Bloom characterization

Weekly water samples, from 07/18/2011 to 10/10/2011, were collected from Milford Lake and

analyzed for cyanobacteria content according to Standard Method 10200F [4] at five constant

locations, including Curtis Creek dock ($39^0 05'23"$N; $96^0 57'30"$W), Milford Lake State Park

beach ($39^0 05'40"$N; $96^0 54'16"$W), Milford swimming beach ($39^0 10'09"$N; $96^0 54'52"$W),

Timber Creek ramp ($39^0 12'39"$N; $96^0 58'21"$W), and Wakefield beach ($39^0 12'41"$N;

$97^0 00'19"$W). The water samples were also analyzed for total microcystins utilizing a

competitive enzyme-linked immunosorbent assay kit (ELISA)[c] in duplicate with appropriate

dilutions to bring the concentration within the working range of the standard curve. To induce

cell lyses and release of toxins into the water prior to analysis, samples were cooled to -30^0C

overnight, and thawed the next day. Microscopy was used to verify that cells were successfully

lysed. The two water samples collected at the northern end of the lake, on 09/12/2011, and

09/20/2011, in the area of the major inflow into the lake from the Republican River were

prepared utilizing 3 freeze-thaw cycles and analyzed by ELISA[e] with appropriate dilutions to fall

within range of the standard curve.

1 **Toxin characterization in vomitus and liver**

2 *Free microcystins*

3 One ml of vomitus was dispensed into a glass tube, and extracted with 10 ml 75% methanol in

4 0.1 M acetic acid. A duplicate spiked sample (1 mL) was also prepared by adding 1 μg of

5 microcystin-LR standard to 1 ml of vomitus prior to extraction. Samples were sonicated in a

6 water bath for 30 minutes to assure homogenization followed by centrifugation at 1,500 x g for

7 10 minutes. The supernatant of the sample and respective spike were collected and extraction of

8 the pellet was repeated two additional times with the acidified methanol solution. The

9 supernatants were pooled and blown to dryness with nitrogen gas at 60˚C. Samples were

10 reconstituted in deionized (DI) water (18 MΩ/cm), and solid phase extraction (SPE)[d] was

11 utilized to clean the sample matrix. The SPE cartridges were conditioned with 100% methanol,

12 equilibrated with 100% DI water, loaded, washed with 5% methanol, and eluted with 2% formic

13 acid solution in methanol. The elution fraction was blown to dryness and reconstituted in DI

14 water for analysis. The final extractant was analyzed with an indirect competitive ELISA[e] in

15 duplicate with appropriate dilutions to bring the concentration within the working range of the

16 standard curve.

17

18 The liver sample was frozen, and lyophilized to dryness at -50^0C^f. A coffee grinder was utilized

19 to homogenize the sample to a fine powdered material. One hundred milligrams (100 mg) of

20 sample was weighed out with a duplicate spiked sample also prepared (0.1 μg microcystin-LR

21 added to 0.1 g dried sample). Five milliliters of 75% methanol solution in 0.1M acetic acid was

22 utilized for the extraction. Samples were sonicated for 30 minutes in a sonicating bath and

23 centrifuged for 10 min at 1,500 x g. The supernatants were removed and extraction was repeated

1 on the pellets two additional times. The pooled supernatants were blown to dryness and

2 reconstituted in 3 ml of DI water for subsequent solid phase extraction[c] in the same manner as

3 the vomitus sample. The eluate was blown to dryness and reconstituted in DI water. The

4 reconstituted eluate was further diluted to bring the concentration within range of the ELISA[e]

5 standard curve and allow for quantitation. The spiked sample was analyzed at a concentration of

6 0.001g sample/ml solution. The assay was sensitive down to a limit of quantification (LOQ) /

7 limit of detection (LOD) of 0.15 ng/ml for the vomitus sample and 150 ng/g for the liver sample

8 as determined from spike recoveries, dilution factors, and kit sensitivity (0.15 µg/l).

9

10 Confirmation of microcystin presence was conducted using liquid chromatography coupled with

11 mass spectrometry (LC/MS)[g], and tandem mass spectrometry (LC/MS/MS)[g]. A flow rate of 200

12 µl/min and two mobile phases were used, A: 100% DI with 3.6 mM formic acid and 2 mM

13 ammonium formate and solvent B: 95% (v/v) acetonitrile with 3.6 mM formic acid and 2 mM

14 ammonium formate. Separation was achieved with a Synergi™ 4 µm Hydro-RP (150mm x 2.0

15 mm) column (Phenomenex, Torrance, CA). Single ion monitoring (SIM) was conducted utilizing

16 the $[M+H]^+$ ions for microcystin LA (*m/z* 910.5), microcystin LR (*m/z* 995.5), microcystin RR

17 (*m/z* 519.5), microcystin YR (*m/z* 1045.6), and desmethyl-microcystin LR (*m/z* 981.5). The

18 LOD and LOQ for each variant were 25 ng/ml, and 50 ng/ml, respectively. Selected reaction

19 monitoring (SRM) was also utilized for the quantification of microcystin LR. The $[M+H]^+$ ion

20 for microcystin LR (*m/z* 995.5) was fragmented and the major product ions (*m/z* 599.5 and

21 553.4) monitored. The microcystin-LR retention time was 15 min. This methodology established

22 an LOD of 10 ng/ml vomitus (10 ng/g liver), and a LOQ of 20 ng/ml vomitus (20 ng/g liver), for

23 microcystin-LR.

1

2 *Total microcystins*

3 The MMPB (2-methyl-3-(methoxy)-4-phenylbutyric acid) method was utilized for the

4 determination of total microcystin by analyzing the chemically cleaved Adda (3-amino-9-

5 methoxy-2,6,8-trimethyl-10-phenyldeca-4,6-dienoic acid) group, a common chemical structure

6 in all microcystin variants.[22] The Lemieux oxidation technique for MMPB formation was

7 modified from Sano et al.[22] and an extraction method was developed and applied to the vomitus

8 and liver tissue. Theoretically, the MMPB technique can be used to measure both bound and

9 unbound fractions of microcystin. Calculations for MMPB determinations are made based on a

10 1:1 molar ratio of MMPB formation from microcystins. Since the range of molecular weights of

11 many microcystins is small (910-1070 g/mol), an estimated calculation can be made by applying

12 a 1000 g/mol qualifier for all microcystin variants. The molecular weight of MMPB is 208

13 g/mol; therefore the theoretical MMPB formation is 0.208 g for every 1 g of microcystin

14 oxidized.

15

16 One ml of vomitus was dispensed into a glass tube with a duplicate spiked sample prepared with

17 10 µg of microcystin-LR standard. Samples were oxidized at a pH of 9 with a 5 mL solution of

18 0.01 M $KMnO_4$, 0.05 M $NaIO_4$ in 0.2 M $KHCO_3$. Solutions were allowed to react at room

19 temperature for 2.5 hours in the dark. Reactions were finalized with the drop-wise addition of

20 40% $NaHSO_3$. The samples were then adjusted to a pH <2 with the addition of 10% H_2SO_4.

21 Liquid-liquid extraction with ethyl acetate was utilized to extract MMPB from excess reagents.

22 Ten milliliters of ethyl acetate was added, and samples were centrifuged at 20,000 x g for 25

23 minutes at 5˚C. The organic layer was removed and extraction on the aqueous layer was repeated

1   with 5 additional mls. The pooled organic (ethyl acetate) layer was blown to dryness with

2   nitrogen gas at 30˚C, and reconstituted in 5% methanol for SPE. The SPE cartridges were

3   conditioned with 100% methanol, equilibrated with 100% DI water, loaded, washed with 5%

4   methanol and eluted with 2% formic acid solution in methanol. The solid phase extraction eluate

5   was blown to dryness and reconstituted in 1 ml of 5% methanol. One hundred milligrams (100

6   mg) of homogenized dried liver sample was weighed out, with a duplicate spiked sample also

7   prepared (1 μg microcystin-LR added to 0.1 g dried sample). Oxidation, liquid-liquid extraction,

8   and solid phase extraction were conducted in the same manner as the vomitus sample. The final

9   solution was analyzed at a concentration of 1g/ml (1:1) with LC/MS/MS.

10

11   A polar reverse phase column[h] was utilized for chromatographic separation. The same LC/MS

12   system and mobile phases utilized in microcystin analysis were utilized for MMPB analysis.

13   The $[M-H]^-$ ion for MMPB (*m/z* 207) was fragmented and the product ion (*m/z* 131) monitored.

14   The retention time was 7 min. The established detection limit for this methodology is 10,000

15   ng/ml (microcystins) or 2,000 ng/ml (MMPB) for the vomitus sample and 10,000 ng/g

16   (microcystins) or 2,000 ng/g (MMPB) for the liver sample.

17

18   **Results**

19   **Pathology**

20   At necropsy the carcass was mildly icteric. The liver was dark red and congested, but otherwise

21   grossly normal. The stomach was filled with reddish-brown fluid. There were extensive

22   hemorrhages on the serosa of the stomach and intestines, but the mucosa of the gastrointestinal

1 tract appeared normal. There was extensive hemorrhage in the mesentery surrounding the

2 pancreas, the uterine mucosa, the mediastinum, the diaphragm, the myocardium, and the

3 pericardial sac. There was a 2 cm spherical mass in the spleen that oozed blood from the cut

4 surface. The lungs were congested, but otherwise grossly normal. There were no gross

5 abnormalities in the pancreas, adrenal glands, thyroid glands, parathyroid glands, urinary tract, or

6 brain.

7

8 Microscopic examination revealed diffuse, acute, massive hepatic necrosis and hemorrhage. In

9 most lobules there was a complete loss of viable hepatocytes, or the only remaining viable

10 hepatocytes consisted of small groups or thin layers 1-2 cells thick adjacent to portal tracts

11 (Figure 1).   Throughout the renal cortex were large coalescing areas where the tubular

12 epithelium was brightly eosinophilic, there was loss of individual cell detail, and nuclei were

13 absent or karyorrhectic, which was interpreted as acute coagulative necrosis (Figure 2.)

14 Interspersed between the necrotic tubules were glomeruli and groups of tubules in which the

15 cells were intact and appeared to be viable. There were groups of tubules lined by highly

16 vacuolated cuboidal epithelium and individual tubules lined by shorter, more eosinophilic

17 cuboidal cells that were not vacuolated (Figure 3). The vacuolated epithelium possessed a PASH

18 positive brush border, which identified the tubules as proximal convoluted tubules. The shorter,

19 more eosinophilic cells lacked a brush border consistent with distal convoluted tubules.

20 Occasional distal convoluted tubules contained casts of necrotic epithelial cells (Figure 3).

21 Similar epithelial casts were present in larger numbers in the collecting ducts in the medulla (not

22 shown). The spherical nodule in the spleen was a hematoma. There was diffuse, acute,

23 pulmonary edema. Acute, multifocal hemorrhages were present in the myocardium, the

1 mesentery, the uterine mucosa, and the serosa of the stomach and intestines. No significant

2 microscopic lesions were found in the bladder, skeletal muscle, tonsils, salivary glands, or brain.

3

4 Microscopic examination of the vomitus obtained from the second vomiting event, collected by

5 the owner during the evening following the dog's exposure to the lake water, demonstrated

6 multiple *M. aeruginosa* colonies (Figure 4).

7

8 **Cell counts and microcystin concentrations**

9 Cyanobacteria were identified to the lowest practical taxonomic level based on morphology and

10 toxicological relevance, which was the genus level because morphological characteristics at this

11 taxonomic level are relatively unambiguous, and toxin production potential below the genus

12 level is strain-dependent.[12,17] The cyanobacterial communities were dominated by *M.*

13 *aeruginosa*, while lower numbers of *Aphanizomenon* spp., and *Anabaena* spp. were present in

14 most samples. The cell counts for *M. aeruginosa* are summarized in Figure 5. Samples of chalk-

15 green algal scum, which formed continuous, floating mats along some areas of the shoreline at

16 Milford Lake State Park, consisted of 100 % *M. aeruginosa* cells by volume.

17

18 Microcystin concentrations and cell counts were variable between sites, and over time. The

19 maximum cell count (5,575,500 cells/ml), and the maximum microcystin concentration (1,600

20 ng/ml), both occurred at Wakefield beach on 08/22/2011 (Figure 5). *M. aeruginosa* cell counts at

21 the site where the dog was exposed were variable, and ranged from 61,110 to 787,500 cells/ml.

22 Microcystin concentrations, estimated by ELISA, were also variable, and ranged from 0.5 ng/ml

1   to 250 ng/ml. The last cell counts and microcystin concentrations measured before exposure

2   occurred were on 09/19/2011, and were 8,505 cells/ml, and 6 ng/ml, respectively. Microcystin

3   concentration in a water sample collected the day following exposure contained 0.5 ng/ml

4   microcystins.

5

6   Linear regression between microcystin concentrations and cell counts, including all data points,

7   resulted in an R^2-value of 0.8128 ($y = 0.0003x - 3.085$), where y is microcystin concentration

8   (ng/ml), and x is cell count (cells/ml). A Pearson product moment correlation resulted in a

9   correlation coefficient of 0.902 (P <0.0001). The degree of the correlation was, however, heavily

10   dependent on the high values at Wakefield beach on 08/22/2011. When these peak values were

11   excluded from the analysis the R^2-value of the linear regression declined to 0.4452, and the

12   Pearson product moment correlation resulted in a correlation coefficient of 0.667 (P <0.0001),

13   indicating that the correlation was weaker at the lower concentrations and cell counts, but

14   remained significant.

15

16   Two water samples collected at the northern end of the lake, on 09/12/2011, and 09/20/2011, in

17   the area of the major inflow into the lake from the Republican River, contained extremely high

18   microcystin-LR concentrations of *c.* 28,000 ng/ml, and *c.* 126,000 ng/ml, respectively.

19

20   The results of microcystin analyses in vomitus and liver tissue are summarized in Table 2.

21   Chromatograms of microcystin analyses in liver and vomitus are summarized in Figure 6.

22

23   **Discussion**

1 The occurrence of FHABs remains difficult to predict, and widespread, effective, timely

2 monitoring of freshwaters is difficult to achieve with available resources. The incidence of

3 cyanobacterial blooms is related to environmental factors such as light, temperature, water

4 column stability, lake levels, and water nitrogen and phosphorus loads.[11,16,24] The relationships

5 between nutrient availability and toxin production are non-linear, and current understanding only

6 allows partial prediction of water microcystin concentrations based on nutrient availability.[9,11]

7 An inability to accurately predict toxin levels makes effective risk communication more difficult.

8 The problem is further compounded by a lack of understanding of the risks by the general public,

9 and by a politically and economically motivated reluctance to close public waters when FHABs

10 occur. This situation often leads to exposure in people and animals when people are unaware of

11 the existence of FHABs, or they are ignorant of the degree of risk associated with water contact

12 during an FHAB. The case described here is an example of a situation where pet owners made

13 use of a publicly accessible lake with an ongoing FHAB, and were apparently unaware of the

14 risks involved. Warnings had been posted at access points to the lake at the time when the

15 exposure occurred, and the FHAB had received an unusually high level of attention in local news

16 media. The poisoned dog's owners were visitors, and unfamiliar with the lake, which may have

17 played a role in their lack of awareness of the risk. Continued public access to lakes during

18 FHABs that produce high levels of toxins inevitably results in continued risks to people and pets

19 in spite of risk communication strategies employed by state and local authorities. However,

20 closure of public waters is a politically sensitive action that results in economic losses for local

21 businesses, and a complete elimination of risk is therefore not always deemed to be in the public

22 interest.

23

1 Multiple extraction and analysis techniques were utilized in the determination of microcystins in

2 both the vomitus and liver samples. The highest microcystin values were demonstrated using the

3 ELISA assay. This could be due to the congener independent sensitivity of the assay, or matrix

4 interactions. The ELISA, although a relatively easy and sensitive tool, has limited applicability

5 for matrices other than water. Biological matrices may interfere with ELISA, and can lead to

6 erroneous results and conclusions if interferences are not recognized. It was therefore important

7 to validate positive ELISA assay values with another analysis technique, such as LC/MS. The

8 LC/MS method for microcystin is very specific, but is limited to variants of microcystin with

9 available standards. The presence of free microcystins in the vomitus sample was confirmed

10 utilizing both LC/MS and LC/MS/MS, and the variant microcystin-LR was detected and

11 quantitated. The ELISA data indicated the presence of *c*. 1,000 ng/g free microcystins in the

12 liver, and microcystin-LR presence was confirmed using LC/MS/MS.

13

14 The MMPB method can reportedly detect all the variants of microcystins equally and has the

15 ability to detect both free and bound microcystins.[22] Microcystins bind reversibly, and

16 irreversibly, to protein/peptide complexes in the body for metabolic uptake and/or depuration.

17 This makes a total microcystin concentration determination difficult since the analyte is not free

18 to react with binding sites in the ELISA kit, and does not allow for a direct determination of the

19 compound. MMPB analysis of the vomitus sample resulted in microcystin levels of *c*. 16,000

20 ng/ml. This was higher than the microcystin-LR concentration determination by LC/MS (6,000

21 ng/ml), but lower than the free microcystin determination by ELISA (50,000 ng/ml). This

22 supports the occurrence of matrix effects on the ELISA assay and indicates other microcystin

23 variants may be present but not directly analyzed for with LC/MS. MMPB was not detected in

1   the liver sample although it is theorized that the bulk of microcystins present in the liver will be

2   bound. This is most likely due to high detection limits for this analyte (2,000 ng/g MMPB; and

3   10,000 ng/g total microcystins).

4

5   The demonstration of *M. aeruginosa* colonies in vomitus from the second observed vomiting

6   event suggests that collecting early vomitus, or stomach contents collected early during the

7   course of illness, may be helpful in confirming ingestion of cyanobacteria. Microcystin ingestion

8   could also be confirmed by demonstrating the presence of microcystins in vomitus by both

9   ELISA and LC/MS/MS methods. The presence of microcystins in a postmortem liver sample

10   provided evidence that microcystins had distributed to the primary target organ. These findings,

11   together with clinical signs associated with liver failure and gastroenteritis, and associated gross

12   and microscopic postmortem lesions, clearly supported a diagnosis of microcystin poisoning.

13

14   Variations in wind speed and direction may account, in part, for variations in cell counts and

15   microcystins between locations, and at specific locations over time, observed at Milford Lake in

16   the summer of 2011 (Figure 5). Buoyant cyanobacteria, such as *Microcystis* spp., are

17   horizontally transported by wind.[15] Wind direction and velocity therefore influences the

18   distribution of *Microcystis* spp. in lakes, leading to accumulations of very high densities along

19   downwind shorelines. Taste and odor compounds in algal scums may be attractive to omnivorous

20   scavengers.[5] Mice, for example, have been observed to prefer water containing *M. aeruginosa.*[20]

21   Such behavior may play a role in the susceptibility of dogs to poisoning by cyanobacterial

22   toxins.[5] The water microcystin concentration from a water sample obtained the day after

1   exposure, from the area where the exposure occurred, was low at 0.5 ng/ml, and much lower than

2   the vomitus microcystin concentration (Table 2). However, the *M. aeruginosa* associated with

3   the FHAB were capable of producing extremely high local microcystin concentrations (up to

4   126,000 ng/ml). Attraction of the dog to pockets of accumulated algal scums along the shoreline,

5   followed by ingestion of concentrated microcystins, may offer an explanation for the relatively

6   high microcystin concentration detected in the vomitus sample.

7

8   To the authors' knowledge, there are no other reports describing necrosis of renal tubular

9   epithelium in dogs, but there is a report in rats. In an acute toxicity study, Sprague-Dawley rats

10 were given intraperitoneal doses of microcystin LR ranging from 20 μg/kg to 1200 ug/kg.[13]

11 Most rats receiving 120 to 240 μg/kg microcystin died within 2 to 3 days. Rats that received 400

12 ug/kg or more microcystin LR died 6 to 8 hours following injection. At the higher doses,

13 glomerular capillaries and proximal tubules contained eosinophilic, fibrillar material and the

14 proximal tubular epithelium was vacuolated. In rats given the highest doses there was multifocal

15 necrosis of tubular epithelium. In our canine patient, we did not find eosinophilic material in

16 glomerular capillaries, but there was acute necrosis of tubular epithelium and the epithelium

17 lining proximal convoluted tubules was vacuolated. The authors of the rat study hypothesized

18 that the tubular necrosis might have been due to direct action of the toxin, but they felt that it was

19 more likely that the acute tubular necrosis associated with microcystin toxicosis was caused by

20 ischemia resulting from shock secondary to massive hepatic necrosis and hemorrhage. That also

21 seems likely in the current case. The dog likely ingested a large amount of toxin and whether or

22 not it had a direct effect on the renal epithelium is unkown. Aggressive antemortem therapy

23 likely prolonged life allowing more time for development of degenerative and/or necrotic renal

changes secondary to ischemia. Pre-existing renal disease could not be completely excluded, but was unlikely because no renal lesions of a chronic nature were observed, and no evidence in the dog's clinical history suggested pre-existing renal disease.

Although creatinine was mildly elevated at the time of admission, it was associated with dehydration and therefore did not indicate renal failure. The urine specific gravity collected by catheterization seven hours after the start of fluid therapy also indicated that the patient retained the ability to concentrate urine during this period, which would be inconsistent with renal failure. However, the patient became anuric after this time, with no additional urine production in spite of the correction of dehydration and continued fluid therapy at twice the maintenance rate. The development of anuria during the latter part of the clinical course could have had a pre-renal component associated with decreased renal perfusion. Decreased renal perfusion in microcystin toxicosis has been described in swine[2], and may lead to secondary ischemic lesions. It may also be speculated that the exposure to microcystin-LR in this case was unusually high, leading to spill-over of intact microcystins from the liver, and subsequent distribution to the kidneys leading to direct nephrotoxic effects.[3,13] A clear distinction between the prerenal toxic effects leading to the development of kidney lesions, and direct toxin effects on kidney tissues, could not be made in this case.

1   Drs. Lisa Pohlman and Ara Gupta of the Clinical Pathology Section of the Kansas State

2   Veterinary Diagnostic Laboratory for providing interpretations of clinical pathology data.

3

4   **Sources and Manufacturers**

5   a.   Vitamin B Complex, Vedco, Inc.

6   b.   Olympus SZX16 Research Stereomicroscope, Olympus Corporation, Center Valley, NJ

7   c.   Quantiplate Kit for Microcystins (EP 022), Envirologix Inc., Portland, ME

8   d.   Strata X (200 mg), Phenomenex Inc., Torrance, CA

9   e.   Microcystins and nodularin (ADDA) ELISA Kit - PN 520011, Abraxis, Warminster, PA

10 f.   Thermo Savant Modulyo Freeze Dryer System, Thermo Fisher Scientific, Waltham, MA

11 g.   A Thermo Finnigan™ Surveyor HPLC system coupled with a Thermo Finnigan™ LCQ

12   Advantage MS^n, Thermo Fisher Scientific, Waltham, MA

13 h.   Synergi™ 4 μm Polar-RP (150mm x 2.0 mm) column, Phenomenex, Torrance, CA

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1 **Tables**

2

3 **Table 1.** Blood cell counts and blood chemistry from a dog that was exposed to a lethal dose of

4 microcystins, upon presentation for treatment on the day following exposure.

Analysis	Value (Reference range)
Lymphocytes	1.0 K/µl (1.5-5)
Hematocrit	70 % (37-55)
Total solids	6.6 g/dl (5.4-7.6)
Erythrocytes	9.86 M/µl (5.5-8.5)
Hemoglobin	23.0 g/dl (12-18)
Prothrombin time	>100 sec (7.1-9.1)
APTT*	>200 sec (8.2-12.7)
Glucose	59 mg/dl (73-113)
Creatinine	2.4 mg/dl (0.5-1.5)
Phosphorus	10.3 mg/dl (2.4-6.4)
Bicarbonate	7 mmol/l (18-29)
Calculated anion gap	48 mmol/l (16-26)
Alanine transaminase	39,326 units/l (28-171)
Alkaline phosphatase	206 units/l (1-142)
Creatine kinase	2,652 units/l (128-328)
Total bilirubin	2.6 mg/dl (0.1-0.3)

5 * = Activated partial thromboplastin time

26

1 **Table 2.** Estimated microcystin concentrations in vomitus and liver samples from a dog that died

2 of microcystin poisoning, analyzed by enzyme-linked immunosorbent assay (ELISA), liquid

3 chromatography coupled with mass spectrometry (LC/MS), and liquid chromatography coupled

4 with tandem mass spectrometry (LC/MS/MS).

Analysis	Vomitus ng/ml	Liver ng/g
Free microcystins (ELISA*)	*c*. 50,000	*c*. 1,000
Microcystin LR (LC/MS**)	6,514	ND
Microcystin LR (LC/MS/MS***)	6,059	10-20
Desmethyl-microcystin LR (LC/MS)**	ND	ND
Microcystin LA (LC/MS)**	ND	ND
Microcystin RR (LC/MS)**	ND	ND
Microcystin YR (LC/MS)**	ND	ND
Total microcystins (MMPB) (LC/MS/MS)****	15,588	ND

ND = Not detected above limit of detection

* Limit of detection: 0.15 ng/ml (vomitus); 150 ng/ml (liver)

** Limit of detection: 25 ng/ml (vomitus); 50 ng/ml (liver)

*** Limit of detection: 10 ng/ml (vomitus); 10 ng/ml (liver)

**** Limit of detection: 10,000 ng/ml (vomitus); 10,000 ng/ml (liver)

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1   **Figure legends**

2   **Figure 1.** Liver in which there is acute, diffuse necrosis and hemorrhage with almost complete

3   loss of hepatocytes. Hematoxylin and eosin stain. Bar = 100 μm.

4

5   **Figure 2.** Photomicrograph of kidney with multiple groups of acutely necrotic tubules in which

6   the epithelium is brightly eosinophilic and individual cells have lost their nuclei or the nuclei are

7   karyorrhectic (arrows). Note that adjacent to necrotic tubules are tubules with intact epithelium

8   with viable appearing nuclei and that glomeruli are normal and that erythrocytes are still intact

9   and individualized. H&E stain, bar = 100 μm.

10

11   **Figure 3.** Photomicrograph of kidney with acutely necrotic tubules in the upper one third of the

12   photograph. Note the intact, vacuolated proximal convoluted tubular epithelium (arrows) and the

13   intact distal convoluted tubule (arrowhead) that contains a group of sloughed, necrotic epithelial

14   cells. H&E stain, bar = 25 μm.

15

16   **Figure 4.** Spherical *Microcystis aeruginosa* colonies forming roughly circular mats of closely

17   packed individual organisms in vomitus from a dog that ingested lake water during an algal

18   bloom. Unstained. Bar = 100 μm.

19

1   **Figure 5.** Microcystis spp. cell counts and microcystin concentrations in water samples from

2   Milford Lake, KS, during the summer of 2011.

3

4   **Figure 6.** Representative chromatograms of microcystin-LR standards, and microcystin-LR in

5   vomitus and liver; as well as an MMPB (2-methyl-3-(methoxy)-4-phenylbutyric acid) standard,

6   and MMPB in vomitus.

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