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4 **Assessing environmental pollution in birds: a new methodological approach for**
5 **interpreting bioaccumulation of trace elements in feather shafts using**
6 **geochemical sediment data**

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30

31 **Abstract**

- 32 1. Environmental trace element composition can have an important impact on ecosystem and
33 population health as well individual fitness. Therefore carefully assessing bioaccumulation
34 of trace elements is central to studies investigating the ecological impact of pollution.
35 Colonial birds are important bioindicators since non-invasive sampling can easily be
36 achieved through sampling of chick feathers, which controls for some confounding factors of
37 variability (age and environmental heterogeneity). However an additional confounding
38 factor, external contamination (ExCo), which remains even after washing feathers, has
39 frequently been overlooked in the literature.
- 40 2. We developed a new method to reliably interpret bioaccumulation of 10 trace elements (As,
41 Cd, Cr, Cu, Hg, Ni, Pb, Se, Sn, and Zn) in feathers using chicks of a colonial species: the
42 Greater Flamingo, *Phoenicopterus roseus*. First, only shafts were used to remove ExCo
43 retained in vanes. Second, we applied a thorough washing procedure. Third, we applied a
44 new analytical method to control for ExCo, which assumes that ExCo is mainly due to
45 adhered sediment particles and that the relative concentration of each trace element will be
46 similar to sediment geochemical composition of sampling sites. We validated this new
47 methodology by comparing trace element composition and particle composition (by
48 scanning electron microscopy and mass-spectrometry) of washed and unwashed feathers.
- 49 3. The washing procedure removed < 99% of K indicating that most of the ExCo from salt was
50 removed. Scanning electron microscopy and mass-spectrometry revealed that some sediment
51 particles remained after washing, especially clays which are likely to severely bias
52 bioaccumulation interpretation. We successfully controlled for ExCo by calculating the ratio
53 of ExCo due to sediment using the geochemical fingerprint of sediment samples. Our
54 methodology leads to conservative estimates of bioaccumulation for As, Cd, Cr, Cu, Hg, Ni,
55 Pb, Se, Sn, and Zn.

56 4. We have validated a new more reliable method of analysing trace element concentrations in
57 feathers, which effectively controls for ExCo, if geochemical sediment data can be
58 meaningfully compared to ExCo of feathers. We have demonstrated that overlooking ExCo
59 leads to potentially erroneous conclusions and we urge that the method applied in this study
60 be considered in future studies.

61

62 **Introduction**

63 Most metals and trace elements are omnipresent in the environment as a consequence of natural
64 processes and anthropogenic activities. Some of them play an essential role in biological processes
65 (e.g. metabolism, neuronal functions). However, other elements (e.g. mercury, lead, cadmium,
66 arsenic, etc.; Kabata-Pendias & Pendias 2001) may also exert detrimental, toxic effects on species if
67 they accumulate in the food chain (Amaral *et al.* 2006) which will negatively affect fitness and life
68 history traits of plants and animals, as well as cause diseases in wildlife and humans (Nriagu 1989;
69 Järup 2003).

70 During the last centuries, the anthropogenic exposure level of trace elements has hugely
71 increased after the industrialization era, especially in wetlands, which, in many cases, act as
72 geochemical endpoints and tend to accumulate pollution (Reddy & DeLaune 2008). The total
73 concentration of metals in soil and sediments persists for a long time because they do not undergo
74 microbial degradation (Kirpichtchikova *et al.* 2006). It has been demonstrated that metals from
75 anthropogenic inputs are often weakly associated to the finest fraction of the top layers of sediment
76 and organic matter (e.g. Salomons & Förstner 1984; Palanques *et al.* 1995; Migani *et al.* 2015) and
77 consequently tend to be much more bioavailable and bioaccumulable than the same elements of
78 natural origin (Bryan *et al.* 1979; Di Giuseppe *et al.* 2014). Monitoring environmental metal
79 contamination and investigating how organisms are affected by the excess of trace element intake
80 or, more generally, the alteration of the natural geochemical profile is of central importance in

81 evolutionary ecology and human and wildlife health. A prerequisite for such monitoring is to
82 develop reliable methods to correctly measure metal exposure, intake, and bioaccumulation.

83 For several decades, birds have proven to be valuable biomonitors for various types of
84 pollutants, including metals (Furness & Greenwood 1993). Ecotoxicological studies in the last three
85 decades have frequently used feathers in order to assess metal accumulation in birds and feather
86 analysis has proven to be a very informative tool to unravel various physiological, ecological and
87 toxicological processes inherent to individuals and populations (Burger 1993; Smith *et al.* 2003;
88 Tsipoura *et al.* 2008). An important advantage of feathers with respect to blood metal concentration
89 is that feathers are relatively easy to collect, preserve, and transport and sampling is virtually
90 harmless to birds (Burger 1993). Moreover, metal accumulation in feathers generally represents a
91 longer-term contamination process, while levels in blood represent a recent contamination directly
92 associated with feeding (Carvalho *et al.* 2013). Since concentration levels in feathers reflects the
93 body accumulation during the entire time of feather development, potential age biases can be
94 circumvented by restricting the analyses to chick feathers. However, external contamination (ExCo)
95 has always challenged researchers and has often been overlooked (but see: Hahn *et al.* 1993; Fasola
96 *et al.* 1998; Ek *et al.* 2004; Hollamby *et al.* 2006; Valladares *et al.* 2010; Borghesi *et al.* 2016).
97 ExCo is defined as the part of the concentration that is not attributable to bioaccumulation in the
98 keratin structure (i.e. metals stored during feather growth as an effect of internal bioaccumulation
99 and metabolic processes, hereafter referred to as bioaccumulation for brevity). ExCo is normally
100 attributed to atmospheric dust, water, or deposition of contaminants on feathers during preening
101 (Dmowski 1999; Dauwe *et al.* 2002; Jaspers *et al.* 2004). However, a recent study on the Greater
102 Flamingo, *Phoenicopterus roseus*, pointed out the major importance of sediment particles in
103 complicating the interpretation of analytical results (Borghesi *et al.* 2016). Most of the previous
104 field studies have tried to remove ExCo through washing, however to date no washing procedure is
105 completely effective in ensuring the total removal of ExCo from feathers (Cardiel *et al.* 2011; Espín
106 *et al.* 2014). Furthermore, so far no studies have tried to quantify the magnitude of ExCo and to

107 consequently validate the bioaccumulation data of trace elements. To continue to use feathers as
108 indicators of bioaccumulation of trace elements, it is important to improve the methodology by
109 reducing the relevance of ExCo, and at the same time, find new methods for estimating more
110 accurate data about bioaccumulated concentrations.

111 In our study, we adopted five measures for that purpose: 1) we used only shafts, because
112 feathers deprived of vanes capture dirt less efficiently (Cardiel *et al.* 2011); 2) we sampled chicks,
113 which avoids variability due to age; furthermore chicks have sediment particles of proven origin
114 entangled in their plumage; 3) we used local geochemical information from sediments collected
115 around nesting islets, in order to compare the local geochemical fingerprint to the element ratios in
116 feathers (Borghesi *et al.* 2016); 4) we chose an extensive set of elements (14), including some of
117 which are supposed to have little or no bioaccumulation and are useful to check for ExCo in the
118 investigated sites as they are indicators of clays (i.e. Al and La) and other fine fractions of the
119 sediment such as oxides and hydroxides (i.e. Fe), and salt (i.e. K). The comparison between
120 sediment and feather concentrations has been performed by adopting a new method capable of
121 estimating the relative importance of ExCo for each element and to correct the analytical result for
122 ExCo. The aim of this study is to validate this new method.

123 In order to achieve this goal, we used the Greater Flamingo as a model species. The ecology
124 and biology of this species are well known due to long term studies (Johnson & Cézilly 2007), a
125 major advantage for an ecotoxicological study. The Greater Flamingo has a large breeding range
126 including many important Mediterranean wetlands (Balkız *et al.* 2007), feeds mainly on small
127 benthonic invertebrates by filtering sediments of brackish wetlands and saltpans. During feeding it
128 can ingest a considerable quantity of sediments from which the organic matter contained therein is
129 digested as a component of diet (Jenkin 1957). Their particular feeding behaviour leads flamingos to
130 be directly exposed to polluted sediments. In addition, flamingos feed their chicks with a liquid
131 secreted from the upper digestive tract, rich in proteins, fat, carotenoids, blood cells, and, as a
132 consequence, with part of the pollutants previously bioaccumulated and metabolized (Lang 1963;

133 Fisher 1972). All of these reasons make greater flamingo chicks a good choice among birds as an
134 environmental indicator of the effect of trace element accumulation in Mediterranean wetlands
135 (Borghesi *et al.* 2011, 2016). However, from the age of 3 weeks old, chicks form a large crèche in
136 the muddy and brackish wetland near the vicinity of the breeding islet (Johnson & Cézilly 2007),
137 leading to high exposure to local environmental elements. Therefore, as highlighted by Borghesi *et*
138 *al.* (2016), ExCo can dominate trace element concentration of Greater Flamingo chick feathers.

139

140 **Methods**

141 **Sample collection**

142 All of the feathers from flamingo chicks were collected between July and August 2014, during the
143 ringing operations in three breeding colonies of the western Mediterranean: Aigues-Mortes (AIG),
144 southern France (N 43° 33', E 4° 11'); Fuente de Piedra (FDP), southern Spain (37° 06'N, 04°
145 45'W) and the heavily polluted Odiel marshes (ODI) (Guillén *et al.* 2011), southern Spain (37° 17'
146 N, 06° 55' W) (Figure 1). All of the sampled birds were between 5 and 8 weeks old (Johnson &
147 Cézilly 2007). Ten feathers were obtained by cutting the distal part from random individuals using
148 stainless steel scissors. We selected the longest internal scapulars that were protected from aerial
149 deposition (Borghesi *et al.* 2011, 2016). Feathers were kept in envelopes at room temperature until
150 analysis. In addition for each sampling site we collected seven sediment samples of 200-500g within
151 and on the reeve of the water body where the breeding islet was situated. Each of the 21 sediment
152 samples was kept in plastic containers in dry room temperature conditions prior to analysis.

153

154 **Sample preparation and analysis**

155 We chose to analyse 14 elements in both sediments and feathers. Ten elements were chosen because
156 of environmental concern: As, Cd, Cr, Cu, Hg, Ni, Pb, Se, Sn, and Zn (ATSDR 1994; Hamasaki *et*
157 *al.* 1995; Hamilton 2004; Cempel & Nikel 2006; Stern 2010; Tchounwou *et al.* 2012; Walters *et al.*

158 2014; Herrmann *et al.* 2016). Aluminium, Fe, K and La were chosen as indicators of clay and the
159 finest fraction of sediment (Leeder 1982).

160 **a. Sediments**

161 Digestion and trace analysis of sediment samples was carried out by ACME Labs, Vancouver
162 (Canada). Samples were digested with a modified *aqua regia* solution of equal parts concentrated in
163 HCl, HNO₃ and DI-H₂O for one hour in a heating block within a hot water bath. Digestion of
164 sediments was done using a modified *aqua regia* solution from ACME labs in order to compare
165 sediment element concentration to element concentration obtained by nitric-chloridric acid digestion
166 of organic material. Indeed the acidic solution of both methods should have a similar dissolving
167 effect on samples (whether sediment or biological). The modified *aqua regia* solution was chosen
168 since it is even more similar to the solution used for feather dissolution than the stronger original
169 *aqua regia* solution (3HCl:1HNO₃). This sediment digestion method has previously been
170 successfully used when analysing feathers in previous studies (Borghesi *et al.* 2016). Each sample
171 volume was equalised with diluted HCl. The concentrations of 64 chemical elements (Ag, Al, As,
172 Au, B, Ba, Be, Bi, Ca, Cd, Ce, Co, Cr, Cs, Cu, Dy, Er, Eu, Fe, Ga, Gd, Ge, Hf, Hg, Ho, In, K, La,
173 Li, Lu, Mg, Mn, Mo, Na, Nb, Nd, Ni, P, Pd, Pr, Pt, Rb, Re, S, Sb, Sc, Se, Sm, Sn, Sr, Ta, Tb, Te,
174 Th, Ti, Tl, Tm, U, V, W, Y, Yb, Zn, Zr) were determined by Inductively coupled plasma mass
175 spectrometry (ICP-MS). To evaluate the analysis quality, an Internal Reference Material (IRM),
176 named DS10, with a composition similar to our sediment samples, was used. Only the 14 trace
177 elements analysed in feathers are considered in this study (Al, As, Cd, Cr, Cu, Fe, Hg, K, La, Ni, Pb,
178 Se, Sn, and Zn). As for feathers, concentrations in sediments are expressed in mg/Kg.

179 **b. Feathers**

180 Vanes were manually separated from shafts by keeping fingers of one hand on the feather tip and
181 then detaching each vane by pulling from the top to bottom with the other hand. Subsequently, the 2
182 mm distal portion (which still had some tiny barbs) was cut off. This method allowed us to obtain
183 rachides completely deprived of barbs and the cuticle connecting barbs to the shaft. From each

184 specimen, five rachides were prepared, in order to reduce variability between feathers and obtain
185 enough feather weight per sample.

186 Thirty-nine samples in AIG and 40 samples in FDP and ODI (in total 119 samples,
187 corresponding to 595 rachides) were thoroughly washed through three steps by sequentially using
188 acetone, Triton X^(TM) detergent, and deionized water. During each step, washing sonication was
189 performed for 20 minutes. After washing, feathers were dried in a dry box at room temperature.
190 From this point forward we now refer to the latter feather samples as “washed feathers”.

191 In order to test the effect of washing on feather trace element composition, we duplicated 10
192 individuals from each site (30 of the 119 individuals in total). For this treatment, five rachides from
193 each individual (a total of 150 rachides) of similar weight were directly sent to the digestion process
194 described below (i.e. no washing procedure was performed prior to digestion). From this point
195 forward we now refer to the latter feather samples as “unwashed feathers”.

196 All samples (approximately 0.100 g from each sample) were digested and analysed at the
197 Trace Element Analysis Core Laboratory of Dartmouth College, Hanover, NH, USA. Digestion was
198 carried out in 0.5 ml acid mixture (9:1 HNO₃:HCl) and then diluted to a final volume of 10 ml with
199 deionized water in polypropylene tubes. Digestion was performed with open polypropylene vessels
200 in a microwave digester reaching a temperature of 105°C (Beck et al. 2013). Similar digestion
201 methods have previously been used to analyse feathers (Latta et al. 2015) and toenails (Amaral et al.
202 2012; Davis et al. 2014; Punshon et al. 2015; Freeman et al. 2015). Total concentration of 14 trace
203 elements (Al, As, Cd, Cr, Cu, Fe, Hg, K, La, Ni, Pb, Se, Sn, and Zn) were measured by Agilent
204 8800 ICP-MS. QA/QC was evaluated by adding to the batches: blanks (frequency: one every 25
205 samples); six samples of oyster, tomato, and hair Standard Certified Materials (2 of each type).
206 Matrix duplicates and matrix duplicate spikes were also digested and analyzed (frequency: 1 every
207 21 and 19 samples, respectively), and fortified blanks added in batches (frequency: 1 every 50
208 samples). Additional QC consisted of reporting calibration checks and blanks (see Supplementary
209 Materials Tables S1-S8). The average recovery of the separate digestions from the National Institute

210 of Standards and Technology (NIST 1566b, 1573a, NIES #13), for As, Cd, Cu, Fe, Pb, Se, and Zn
211 was around 100%, for K was 117.5% (SD 11.5%), and, for Hg, La and Ni was around 80% (see
212 Supplementary Materials Table S4). For Al and Cr recovery averaged around 30% presumably
213 because these metals were in a form that is not solubilized by the open vessel acid digestion used
214 here (Beck et al. 2013). Tin analyzed in the feather samples was not certified in the NIST standards.
215 Since QA/QC of all trace elements were very good and hair was the most comparable certified
216 reference material (CRM) with feathers, Al and Hg concentrations were corrected using the hair
217 CRM whereas Cr, K, La, and Ni for the other available CRMs (tomato and oyster), since no
218 reference values for these metals were available for hair. Concentrations in feathers are expressed in
219 mg/Kg dry weight (dw). Method detection limits (MDLs) were calculated as three times the
220 standard deviation of the average value of the 6 calibration blanks, and based on a sample weight of
221 100 mg. Limits of quantification (LOQs) have also been calculated as 3 times the MDL (see
222 Supplementary Materials Table S8).

223

224 **Examination of feather with Scanning Electron Microscopes (SEM)**

225 In order to make morphological observations of external particles and possibly infer the nature of
226 external contaminants, shaft segments 1 cm long have been scanned with a Jeol JSM-5400 Multi-
227 Purpose Digital SEM equipped with WDS and EDS Systems at University of Bologna, Department
228 of Biological, Geological, and Environmental Sciences (BiGeA). Six feather shafts were selected
229 from each site (18 feather shafts in total), and prepared for SEM without any washing treatment (i.e.
230 the same treatment as unwashed feathers). In addition, 3 feather segments from ODI were scanned
231 after a thorough cleaning procedure with tap water, a commercial detergent, and acetone (i.e.
232 substantially the same treatment as washed feathers but without sonication).

233

234 **Correction of element concentrations in feathers for environmental contamination**

235 Feathers, even washed, retain a certain quantity of sediment (see results). For the sake of argument,
236 if we assume that all bioaccumulation is masked by ExCo, analytical results from chick feathers
237 should tend to represent the geochemical characteristics of local sediments instead of the actual
238 assimilation and accumulation in keratin structure of trace elements. If so, the relative abundances
239 of elements in sediment and feathers should be similar. In contrast, if elements are mostly
240 bioaccumulated then they should be in a higher concentration than expected if the chemical
241 fingerprint of feathers is only determined by ExCo. Using the 14 elements analysed in this study,
242 and investigating the ratios between concentration in feathers and sediment, we can check which
243 elements in feathers are clearly enriched with respect to expected ExCo concentrations.

244 By investigating sediment element concentration, we are able to infer what the predicted
245 concentration of feather elements would be if ExCo was 100% (predicted external contamination;
246 PExCo) for each element. Here a reference element which indicates ExCo needs to be carefully
247 chosen. The reference should be an element that: 1) is analytically reliable, 2) that is dominant in the
248 source of ExCo (in our case soil and sediment) and 3) that is either negligibly or not bioaccumulated
249 (i.e. concentrations are dominated by ExCo). A previous study (Cardiel *et al.* 2011) has suggested
250 that Al is a good indicator of ExCo because it is known to be scarcely metabolized by birds (Beyer
251 *et al.* 1999) and it is a main component of clays and hydroxides (Moore & Reynolds 1989).
252 However, we found that Al is extracted in smaller concentrations by the acid digestion step than
253 most of the other elements (see above). As a consequence a certain amount of ExCo of elements
254 will be overlooked when using Al as the geochemical reference even if corrected using the CRM
255 and, analytically, Al is not a sufficiently reliable element to be used as a reference element. In
256 contrast Fe is well recovered by the methods applied in this study (see Supplementary Materials
257 Tables S1-S8 for QA/QC results), it is reported to be only negligibly bioaccumulated in shafts of
258 seabirds (Howell *et al.* 2012) and it represents a wider gamma of compounds in sediments than Al
259 and La (Reddy & DeLaune 2008). Finally, it is important to note that a small amount of Fe maybe
260 bioaccumulated, which means that we are actually using a conservative approach and may be

261 slightly overestimating ExCo. For sound biological interpretation the latter is highly preferable than
262 ignoring ExCo and reporting highly inflated bioaccumulated values. However we found a strong
263 correlation between Al, Fe and La concentration in washed feathers, further suggesting that ExCo
264 dominates bioaccumulation for these elements (see Supplementary Materials Figure S1-S4)
265 (Borghesi *et al.* 2016). For all of the aforementioned reasons we chose to infer feather PExCo using
266 Fe. We calculated the PExCo of feathers as:

$$PExCo_i = \frac{x_i y_j}{z_j}$$

267 where x_i is the concentration of Fe in the feather sample i and y_j is the concentration of the element
268 studied in sediment at the breeding colony j and z_j is concentration of Fe in sediment at breeding
269 colony j . From PExCo, we can deduce the proportion of element concentration found in feather that
270 is due to ExCo (external contamination factor; ExCoF):

$$ExCoF_i = \frac{PExCo_i}{w_i}$$

271 where w_i is the element concentration of interest of the feather sample i . Using these two simple
272 equations we estimated, for each feather, the proportion of ExCo for each element within each
273 breeding colony site. For pedagogical reasons, we also applied the above equations to median
274 feather concentrations for each breeding site and intervals which encompass 95% of the data (i.e. x_i
275 and w_i are median values or 95% intervals of each element for each breeding site instead of for each
276 individual feather). For each feather, we were then able to correct element concentration for ExCo
277 by using the following formula:

$$corrected\ w_i = w_i - (ExCoF_i * w_i)$$

278

279 **Statistics**

280 All statistics were carried out in R version 3.2.4 (R Core Team 2016). To investigate the effect of
281 washing of feathers on element concentrations we applied a paired Wilcoxon-Pratt signed-rank test

282 (Pratt 1959) between element concentrations for feathers that were not washed and for feathers that
283 were washed ($n = 30$). We calculated r as a measure of effect size which is the z-value divided by
284 the square root of the sample size (in our case 30; Pallant 2007). An r value between 0.1-0.3 is
285 considered as small, a value between 0.3-0.5 to be medium and finally any value above 0.5 is
286 considered as large. Median differences between washed feathers and unwashed feathers as well as
287 associated 95% confidence intervals were also reported.

288 To investigate the effect of correcting ExCo on element concentration of feather shaft we
289 calculated the mean difference in feather concentration between raw element concentration of
290 feather shaft and element concentration of feather shaft corrected for ExCo ($n = 119$) and the
291 associated Cohen's D (Cohen 1988) (note that applying a paired Wilcoxon-Pratt signed-rank test
292 here always yielded a significant result since ExCo correction always reduces concentration of
293 elements, however this does not allow us to assess whether ExCo correction had a negligible or
294 strong effect). Since many element concentrations were not normally distributed we calculated 95%
295 confidence intervals by bootstrapping (1000 bootstraps) as recommend by Nakagawa and Cuthill
296 (2007) using the boot package implemented in R (Canty & Ripley 2015). A Cohen's D of below 0.2
297 is considered as negligible, between 0.2-0.5 small, between 0.5-0.8 medium and larger than 0.8 as
298 large (Nakagawa & Cuthill 2007). We therefore considered that ExCo correction to have an
299 appreciable effect on element concentration when Cohen's D was equal to or greater than 0.2.

300

301 **Results**

302 **The effect of washing feathers**

303 The washing procedure significantly reduced trace element concentration for 12 of the 14 elements
304 analysed in feathers: Al, As, Cd, Cr, Cu, Fe, K, La, Ni, Pb, Se, and Zn (Figure 2). The effect was
305 strong ($r > 0.500$) for Al, As, Cu, Cd, Cr, Fe, K, La, Ni, and Zn. A medium effect ($r > 0.300$) was

306 observed for Se and Pb. For Sn and Hg washing did not significantly reduce trace element
307 composition (Figure 2).

308

309 **Examination of feathers with SEM**

310 SEM examination of 18 unwashed shaft segments of 1 cm revealed a large diversity of particles
311 which densely covered the feathers. A quantitative count of external particles was not possible due
312 to their abundance and complexity. A large number of particles (>200) were found, most of them
313 predominantly composed of sulphur (S) associated with other elements. We concluded that these
314 particles were probably mostly from organic matter derived from feathers, which were discarded
315 from further analysis. Of the remaining particles, one to six putative external contaminants per
316 segment were thoroughly examined for their dimension, shape and chemical composition (for the
317 most abundant elements only according to instrumental limitations). This resulted in a total of 66
318 lithic particles analysed for their element composition by SEM.

319 The analysed particles tended to range from less than 1 to 30 μm in all segments, although
320 on rare occasions, they measured up to 100 μm . Particles appeared as amorphous terrigenous
321 aggregations (Figure 3a), definite solid crystals (Figure 3b), piles of stacked sheets (Figure 3c),
322 electrostatically adhered soft objects, or a combination of the aforementioned.

323 By observing the spectrum, a classification of the geological nature of each x-rayed particle
324 has been provided. As shown in Table 1, a variety of Na and Mg salts emerged as the most abundant
325 components of particles in all sites. In salts, K was detectable only in AIG samples. Occasionally,
326 Ca was appreciably present in FDP salts. In all sites, clay particles were often associated with salt
327 particles.

328 Aluminium was a common element in clays in all sites, but the composition of other
329 elements changed according to sites. Potassium was detected in clay particles investigated in AIG,
330 whereas clays from FDP and ODI showed heterogeneous composition, being either calcic, sodic, or
331 potassic. Noticeably, Mg was detectable in clays only in ODI samples (7 out of 8), which were very

332 variable in their overall composition and in some cases particularly rich in Fe, Ti, Cr and potentially
333 many other metals.

334 Hydroxides were present in particles from all sites, but were not very frequent. They
335 appeared as Al-hydroxides, Mn was detectable in one particle from FDP. Minerals such as quartz,
336 mica, chlorite, muscovite and gypsum were occasionally found in ODI samples, while Ca-
337 carbonates were found in FDP. Four particles (2 in FDP and 2 in ODI samples) were apparently
338 composed uniquely of Al. This may be due to the use of metallic tools, such as scissors and
339 tweezers. A few lithic particles containing Cl or Ca remained undetermined.

340 In addition to the 18 unwashed shafts, three different shaft segments from ODI (the site
341 where external contaminants are more likely to be rich in trace elements) were analysed by SEM,
342 which were submitted to a washing procedure with water and detergents and then rinsed under
343 running water. Much less lithic particles were visibly found, but some scattered particles were still
344 present. In general, they were less frequent, smaller and seemed less complex in shape. At least 5
345 lithic objects were found and have been classified as sodium chloride crystals (2), carbonatic
346 mineral (2), metallic aluminum (1).

347 **Using geochemical data to assess the importance of ExCo on shaft trace element concentration**

348 We found strong variation between elements of the importance of ExCo on trace element
349 concentration in feathers. For Cu, Hg, Se, and Zn we found a median ExCoF lower than 0.5% in all
350 the investigated sites and 95% of the data (95% interval; hereon referred to as 95%Iter) ranged
351 between 0.3% and 0.7% indicating that Cu, Hg, Se, Zn are clearly bioaccumulated in feathers and
352 dominate ExCo (Table 2). In contrast, for Al, K, and La median ExCoF were much higher than
353 100% (Table 2) suggesting that ExCo dominates any bioaccumulation for these elements. The ExCo
354 was less clear cut for the other elements (As, Cd, Cr, Ni, Pb, and Sn; Table 2). Among these
355 elements, Sn seems to be mostly bioaccumulated, with little variation between sites and median
356 ExCoF ranging between 2-5% (Table 2). For As, Cd, Cr, Ni, and Pb, the ExCoF was more variable
357 between the sampling sites. Arsenic had ExCoF of 14% (95%Iter = 3-43%) in AIG, ExCoF of 38%

358 (95%Iter = 18-87%) in FDP, and only a ExCoF of 3% (95%Iter = 2-8%) in ODI (Table 2). There
359 was a lower variation of the effect of ExCo for Pb which has a ExCoF of 13% (95%Iter = 6-20%) in
360 AIG, ExCoF of 22% (95%Iter = 14-37%) in FDP and a ExCoF of 10% (95%Iter = 5-15%) in ODI
361 (Table 2). For Cd, Cr, and Ni, there was strong variation of the effect of ExCo on trace element
362 concentrations within site, although there was little variation between sites (Table 2). For Cd, we
363 calculated a median ExCoF of 21%, 32% and 35% in ODI, AIG and FDP respectively (Table 2).
364 For Ni, median ExCoF ranged between 26-47% (AIG>FDP>ODI), with 95%Iter within site ranging
365 between 13-78%, 10-91%, 9-39% in AIG, FDP, and ODI, respectively (Table 2). For Cr, AIG and
366 ODI had a median ExCoF of 20% and 22% respectively (95%Iter=5-92% and 4-53% respectively),
367 while FDP had the highest ExCo for this element (median ExCoF=31%, 95%Iter = 7-78%; Table 2).

368

369 **Using geochemical data to correct for external contamination of feathers**

370 Correcting each individual sample mirrored median ExCoF results (Figure 4). Prior to ExCoF
371 correction Al, La and K could erroneously be interpreted as bioaccumulated (Figure 4). However,
372 ExCoF correction revealed that actually Al, La and K concentrations in feather is likely to be almost
373 entirely due to external contamination and bioaccumulation is either highly unlikely or below
374 instrumental detection limits (Figure 4). Of the remaining elements ExCoF had an appreciable effect
375 (Cohen's $D > 0.200$) on element concentration for Ni and Pb (Figure 4). However, ExCoF
376 correction had a negligible effect (Figure 4; Cohen's $D < 0.200$) on the concentration of
377 bioaccumulation for As, Cd, Cr, Cu, Hg, Se, Sn and Zn (Figure 4).

378

379 **Discussion**

380 Our results show that our novel methodological approach efficiently dealt with external
381 contamination found in feather shafts and significantly changed interpretation of feather element
382 concentration. We sampled chicks which allowed us to control for the effect of age on
383 bioaccumulation and the shorter time of exposure to external environmental agents than adults

384 (Burger 1993). Prior to analysis, we took two methodological measures to minimise ExCo and
385 unreliable biological interpretations. First, unlike most studies in feathers, in this study we removed
386 the vanes in order to limit the tendency of feathers to entangle dirt among barbs (Cardiel *et al.*
387 2011). Furthermore, vane and shaft sequester metals differently, (Bortolotti 2010; Howell *et al.*
388 2012) which may confuse biological interpretation if analysed together. Indeed, high resolution
389 images from X-ray fluorescence microscopy of shearwater chick breast feathers revealed a different
390 distribution of As, Br, Ca, Fe, and Zn among the calamus, shaft and vane (Howell *et al.* 2012). The
391 latter study was preliminary and did not give a physiological explanation of such a finding but
392 pointed out that elements can be mostly concentrated in the calamus (Ca), shaft (As, Br, and Zn), or
393 vane (Fe). In addition, most of the mass of a feather is the shaft for a given section of a feather and
394 this may consequently affect the concentration, according to Bortolotti (2010), which advocates two
395 mechanisms related to bioaccumulation in feathers depending on each trace element: mass-
396 dependent and time-dependent accumulation. Scanning electron microscopy on our samples
397 highlighted that a huge quantity of lithic particles and salt crystals are trapped in unwashed feathers,
398 even when deprived of vanes. Therefore, our second measure was to wash shafts, combining the
399 most common methods applied in the literature to date (Ansara-Ross *et al.* 2013; Costa *et al.* 2013;
400 Carvalho *et al.* 2013; Rubio *et al.* 2016) with a prolonged ultrasonic bath treatment (Weyers *et al.*
401 1988). SEM observation also revealed that some ExCo remained in washed feathers and that ExCo
402 cannot be ignored prior to data analysis. We successfully controlled for the remaining ExCo by
403 calculating the ratio of ExCo due to sediment using the geochemical fingerprint of sediment
404 samples. Our methodology allowed us to have conservative estimates of 10 bioaccumulated
405 elements (As, Cd, Cr, Cu, Hg, Ni, Pb, Se, Sn, and Zn).

406

407 **The effect of washing**

408 After washing, more than 99% of K was removed (the concentration of K went from 116-809 mg/kg
409 in unwashed shafts to 0.103-5.051 mg/kg in washed shafts), a much higher percentage than any of

410 the other elements analysed in this study. Since K is a dominant element in salt, we can conclude
411 that the washing effect was near complete in removing salt, which is likely to be a dominant residue
412 in coastal bird feathers. We note however, that despite the effectiveness of washing, some K
413 remained (0.103 - 5.051 mg/kg), suggesting that either some residual ExCo remained (rare small
414 salt crystals were observed even in washed shafts by SEM), or some bioaccumulation or both.

415 The washing treatment of feathers also significantly reduced the concentration of 11 of the
416 remaining 13 elements, with only Sn and Hg not significantly reduced. For Sn either trace element
417 concentration from environmental contaminants was negligible relative to the concentration from
418 bioaccumulation, or the efficiency of the washing procedure was lower than for the other trace
419 elements. Like most elements, very little is known about the characteristics of Hg ExCo, however
420 feather concentration of Hg is considered to be a good indicator of bioaccumulation irrespective of
421 washing procedures (Jaspers *et al.* 2004; Pedro *et al.* 2015). Previous studies have shown that Hg
422 levels in feathers are highly correlated with Hg concentration in the diet (Lewis & Furness 1991,
423 1993; Hahn *et al.* 1993; Monteiro & Furness 1995) and in internal tissues (Thompson *et al.* 1991)
424 even when potential ExCo is ignored. Furthermore, Hg concentration in feathers is stable over time
425 under various experimental environmental treatments suggesting that ExCo has little effect on this
426 element (Appelquist *et al.* 1984). Our study is therefore consistent with the literature that ExCo of
427 Hg is irrelevant regardless of the washing treatment.

428 Observing shafts by SEM demonstrated that unwashed feathers are very rich in lithic
429 particles and are likely to be the main contributors of ExCo in feathers. Most lithic particles are salt
430 crystals, clays and other fine residuals which can be removed in part by washing (Font *et al.* 2007).
431 In fact, SEM observations on some ODI washed samples revealed that since these lithic particles are
432 electrostatic and very small (typically 1-30 μm) some ExCo remain, even after the thorough
433 washing treatment. ExCo of lithic particles from salt crystals is essentially made of Na, Mg, and K
434 chlorides and Ca and Mg carbonates. However, we believe that any remaining ExCo by salt is likely
435 to have a negligible effect on metal concentration because K is hundreds of times more concentrated

436 in salt than Cu, Cr, and Zn (1,800-3,900 mg/Kg of K, 0-1.2 mg/Kg of Cu, 12-14 mg/Kg of Cr, and
437 7.4-7.5 mg/Kg of Zn in two collected and analysed samples of salt from Aigues-Mortes water; see
438 dryad data: “Dryad hyperlink if accepted”). In contrast, terrigenous particles, such as clays, hydroxides
439 and organic matter contain higher concentrations of metals, and the presence of a few of these lithic
440 particles in a feather sample is sufficient to mask bioaccumulation for some elements (Borghesi *et*
441 *al.* 2016). Therefore, further analytical methods are necessary to soundly interpret feather data.

442

443 **Assessing the importance of ExCo on shaft trace element concentration**

444 We found strong variation between elements on the importance of ExCo on trace element
445 concentration in feathers. On the one hand, ExCo had a negligible effect on trace element
446 concentrations for some elements (median ExCoF less than 0.5% for Cu, Hg, Se, and Zn; and
447 around 5% for Sn), while on the other hand, ExCo dominated element concentrations for Al, K, and
448 La (median ExCoF more than 100%). The latter is consistent with the hypothesis that residual ExCo
449 after washing is essentially made of clays. Indeed, K is incorporated in the structure of certain clay
450 minerals such as illite, and commonly adsorbed on the surfaces of many others (Salminen 2005),
451 and clays have the capability of adsorbing rare earth elements (REEs) released/dissolved during
452 weathering, with La being one of the most abundant REEs (Moldoveanu & Papangelakis 2012).
453 Aluminum, K, and La concentrations are therefore good signals of residual ExCo and should only
454 be used as controls of ExCo in trace element studies in feathers. Regarding the remaining five
455 elements (As, Cd, Cr, Ni, and Pb), ExCo had a more nuanced effect on trace element concentration
456 (depending on the element and the site), and the use of these elements in feathers to infer
457 bioaccumulation needs some ExCoF corrections in order to avoid inflated interpretation of
458 bioaccumulated concentrations.

459

460 **Correcting for the effect ExCo on shaft trace element concentration**

461 By calculating an ExCoF for each individual sample we were able to correct concentration values
462 for ExCo for each sample by subtracting from the element concentration the proportion of element
463 concentration that was estimated to be due to ExCo. Of the ten elements of environmental concern
464 analysed in this study (As, Cd, Cr, Cu, Hg, Ni, Pb, Se, Sn and Zn), ExCoF correction for Ni and Pb
465 did appreciably change mean concentrations (Figure 4). This may have important consequences
466 when investigating the relationship between element concentration and other variables (such as
467 body condition or fitness traits), including differences in bioaccumulation between sites which is
468 beyond the scope of this study.

469 Trace element concentration in sediment at ODI was higher for 13 out of 14 elements (the
470 exception was Se which was similar in ODI and FDP, and lower in AIG). These results are
471 consistent with the extensive literature which demonstrates that ODI is one of the most polluted
472 estuarine areas in the world (Guillén *et al.* 2011). However, following careful consideration of
473 ExCo, appreciably higher trace element concentrations in feathers in ODI were only found for As
474 and Pb. The latter suggests that there are important differences in how chicks metabolise each
475 element during feather development and that not all trace elements in feathers are reliable
476 environmental bioindicators. For example, Al, La and K were negligibly bioaccumulated and
477 therefore poor bioindicators, whilst the other analysed trace elements were bioaccumulated to some
478 extent and may be good bioindicators. However, there is strong indication that the bioaccumulation
479 rate in feathers is not the same for all elements (e.g. the level of Cu, Hg, Se, Sn and Zn in feather
480 shafts, while high in all samples, appears to be relatively independent of environmental levels,
481 whilst As, Cd, Cr, Ni and Pb levels seem to be more heterogeneous between individuals and sites).
482 A detailed interpretation of bioaccumulation and differences between sites is beyond the scope of
483 this article.

484 In conclusion, as pointed out by previous studies, without careful consideration of ExCo,
485 conclusions about the validity of the concentration of element bioaccumulation in feathers are
486 unreliable. We have developed a new more reliable method of analysing trace element

487 concentrations in feather shafts which effectively controls for ExCo. While Fe was used as the
488 reference element to infer ExCo in feathers in this study, a different reference may be used in other
489 studies depending on sampled species and environmental characteristics, in other words, the
490 predicted main source of external contamination and the pollutants which are the object of the
491 research. We also note that while our study focused on feathers, a similar strategy can easily be
492 applied to other non-invasive organic samples when residual soil/sediment particles may bias
493 interpretation of bioaccumulation (for example when assessing trace elements in plants,
494 invertebrates, faeces and hair samples of vertebrates). Many studies continue to overlook ExCo
495 leading to potentially erroneous conclusions and we urge that methods applied in this study be
496 considered in future studies investigating bioaccumulation of trace elements in organic samples in
497 contact with the external environment.

498

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507

508 **Data accessibility**

509 All the data used in this study is available in the public database Dryad: “Dryad hyperlink if accepted”

510

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707

708 **Figure 1: Map showing the location of the three breeding colonies sampled for greater flamingo chick**
709 **feathers and sediment. Sample sizes of washed feather shafts are in brackets.**

710

711 **Figure 2: Boxplot of paired unwashed and washed shaft feathers for the 14 elements investigated ($n =$**
712 **30). Median difference between washed and unwashed feathers, 95% confidence intervals, the z -value**
713 **and p -value of the paired Wilcoxon-Pratt signed-rank test and the effect size r are shown within the**
714 **boxplots of each element. Elements with a * were plotted on the log scale (but were not log transformed**
715 **prior to analysis).**

716

717 **Figure 3: Electron microscopy pictures of three typical examples of three types of particles found in**
718 **unwashed feather shafts: a. amorphous terrigenous aggregations; b. definite solid crystals; and, c.**
719 **piles of stacked sheets**

720

721 **Figure 4: Boxplot of paired shaft feathers not corrected for ExCo and corrected for ExCo for the 14**
722 **elements investigated ($n = 119$). Mean difference and Cohen's D between not corrected and corrected**
723 **concentrations and associated 95% confidence intervals (calculated by bootstrap, $n = 1000$) are shown**
724 **within the boxplots of each element. Elements with a * were plotted on the log scale (but were not log**
725 **transformed prior to analysis).**

726

727

728 **Table 1: Summary of element composition for each particle analysed by SEM showing the number of particles**
 729 **with a certain element composition and its geological interpretation within each site: Aigues-Mortes (AIG),**
 730 **Fuente de Piedra (FDP) and Odiel marshes (ODI).**

731

Site	Number of particles	Element composition	Geological interpretation
AIG	5	Na, Mg, Cl	Salt crystals
	2	Na, Cl	Salt crystals
	2	Mg, Cl	Salt crystals
	2	Na, K, Cl	Salt crystals
	1	N	Organic matter
	1	K, Cl	Salt crystals
	1	Al, O	Al hydroxide
	1	Al	Aluminum
	1	Al, Si, K, Ca	Clay
	1	Al, Si, K	Clay
	1	Mg, Cl, Al	Salt and Al hydroxide/oxide
	1	Si	Quartz
	FDP	3	Na, Mg, Cl
2		Cl	Chloride
1		Mg, Na, Ca, Cl	Salt crystals
1		Mg, Cl	Salt crystals
1		Na, Cl	Salt crystal
1		Mg, Cl, Al, Si, Ca, Fe	Salt and clay
1		K, Ca, Mg, Cl, Al, Si	Salt and clay
1		Mg, Cl, Al, O	Salt and Al oxide/hydroxide
1		Mg, Cl, Al	Salt crystal and aluminum
1		Na, Mg, Cl, Al	Salt crystal and aluminum
1		Na, Cl, Ca, C, O	Salt and carbonate
1		Na, Al, Si	Clay
1		Ca, C, O, Mn	Carbonate and Mn oxide hydroxide
ODI	9	Na, Cl	Salt crystals
	4	Mg, Cl	Salt crystal
	2	Na, Mg, Cl	Salt crystal
	2	Mg, Al, K, Ti, Fe, Si	Clay (mica)
	2	Si	Quartz
	2	Ca, S, O	Gypsum
	2	Al	Aluminum
	1	Mg, Al, Fe, Si	Clay (phyllosilicates)
	1	Na, Mg, Al, K, Cl, Si	Clay (phyllosilicates)
	1	Na, Mg, Al, K, Fe, Si	Clay (phyllosilicates)
	1	Al, K, Fe, Si	Clay (phyllosilicates)
	1	Al, Fe, Cr, Ca, Al, Si	Clay
	1	Mg, Al, Si	Clay
	1	Cl	Chloride
1	Ca	Calcium oxide	

732

733

734

735 **Table 2: Summary statistics for each samplings site showing median concentration of washed**
 736 **feathers prior to ExCo correction (Feather), median sediment concentration (Sediment), the**
 737 **predicted concentration if feather concentration is entirely due to external contamination (PExCo,**
 738 **see methods for calculation formula), the percentage of feather median concentration explained**
 739 **by external contamination (ExCo) and intervals which encompass 95% of the data (ExCoQ).**
 740 **Elements are ordered according to ExCo within each site. Iron (Fe) is highlighed in bold since this**
 741 **element was used as the reference for PExCo, ExCo and ExCoQ calculations and feather Fe**
 742 **concentration was assumed a priori to be 100% ExCo.**

Elements	Feather (mg/kg)	Sediment (mg/kg)	PExCo (mg/kg)	ExCo (%)	ExCoQ (±95%Iter)
a. Aigues-Mortes (AIG; n = 29)					
Se	1.747	0.05	5.378E-05	0.003	0.002-0.007
Hg	0.539	0.12	1.291E-04	0.024	0.008-0.073
Cu	9.848	3.84	4.130E-03	0.042	0.032-0.051
Zn	43.312	19.7	2.119E-02	0.049	0.033-0.078
Sn	0.022	1.1	1.183E-03	5.286	1.844-12.998
Pb	0.061	7.22	7.766E-03	12.673	6.396-19.743
As	0.016	2.1	2.259E-03	14.050	2.657-42.743
Cr	0.047	8.6	9.250E-03	19.793	4.687-91.492
Cd	1.656E-04	0.05	5.378E-05	32.477	7.482-107.562
Ni	0.021	9.1	9.788E-03	47.151	13.257-78.014
Fe	7.422	6900	7.422E+00	100	NA
Al	0.938	3600	3.872E+00	412.619	179.506-1013.137
La	2.884E-04	2.9	3.119E-03	1081.612	335.137-2588.256
K	0.088	1000	1.076E+00	1227.041	116.867-1227.041
b. Fuente de Piedra (FDP; n = 30)					
Hg	0.623	0.021	1.365E-05	0.002	0.001-0.006
Se	1.597	0.6	3.901E-04	0.024	0.015-0.042
Zn	40.482	19.6	1.274E-02	0.031	0.024-0.058
Cu	7.225	13.74	8.933E-03	0.124	0.102-0.239
Sn	0.025	1.2	7.802E-04	3.106	0.113-10.958
Pb	0.041	13.76	8.946E-03	22.088	13.776-36.628
Cr	0.033	16	1.040E-02	31.348	6.599-77.545
Cd	1.289E-04	0.07	4.551E-05	35.317	6.611-91.019
Ni	0.023	13.4	8.712E-03	37.380	9.961-91.267
As	0.005	3.1	2.015E-03	37.959	17.532-87.451
Fe	6.306	9700	6.306E+00	100	NA
Al	1.410	13400	8.712E+00	618.075	214.011-1218.617
La	3.187E-04	4.6	2.991E-03	938.422	319.797-1856.099
K	0.247	4200	2.731E+00	1107.145	137.813-3114.956
c. Odiel (ODI; n = 30)					
Se	1.403	0.5	8.542E-05	0.006	0.004-0.009
Hg	0.398	0.168	2.870E-05	0.007	0.002-0.020
Zn	38.652	563.2	9.622E-02	0.249	0.157-0.382
Cu	8.509	247.1	4.222E-02	0.496	0.334-0.666
Sn	0.025	2.4	4.100E-04	1.627	0.076-4.419
As	0.499	91.2	1.558E-02	3.122	1.901-8.036
Pb	0.148	83.82	1.432E-02	9.666	5.003-15.170
Cd	1.933E-04	0.24	4.100E-05	21.216	6.641-82.005
Cr	0.035	45.3	7.739E-03	21.998	3.959-53.040
Ni	0.021	30.9	5.279E-03	25.735	9.184-38.706
Fe	7.193	42100	7.193E+00	100	NA
Al	1.460	23900	4.083E+00	279.655	65.607-599.191
La	5.760E-04	15	2.563E-03	444.940	107.003-1167.760
K	0.268	7000	1.196E+00	446.842	65.872-1364.259

743

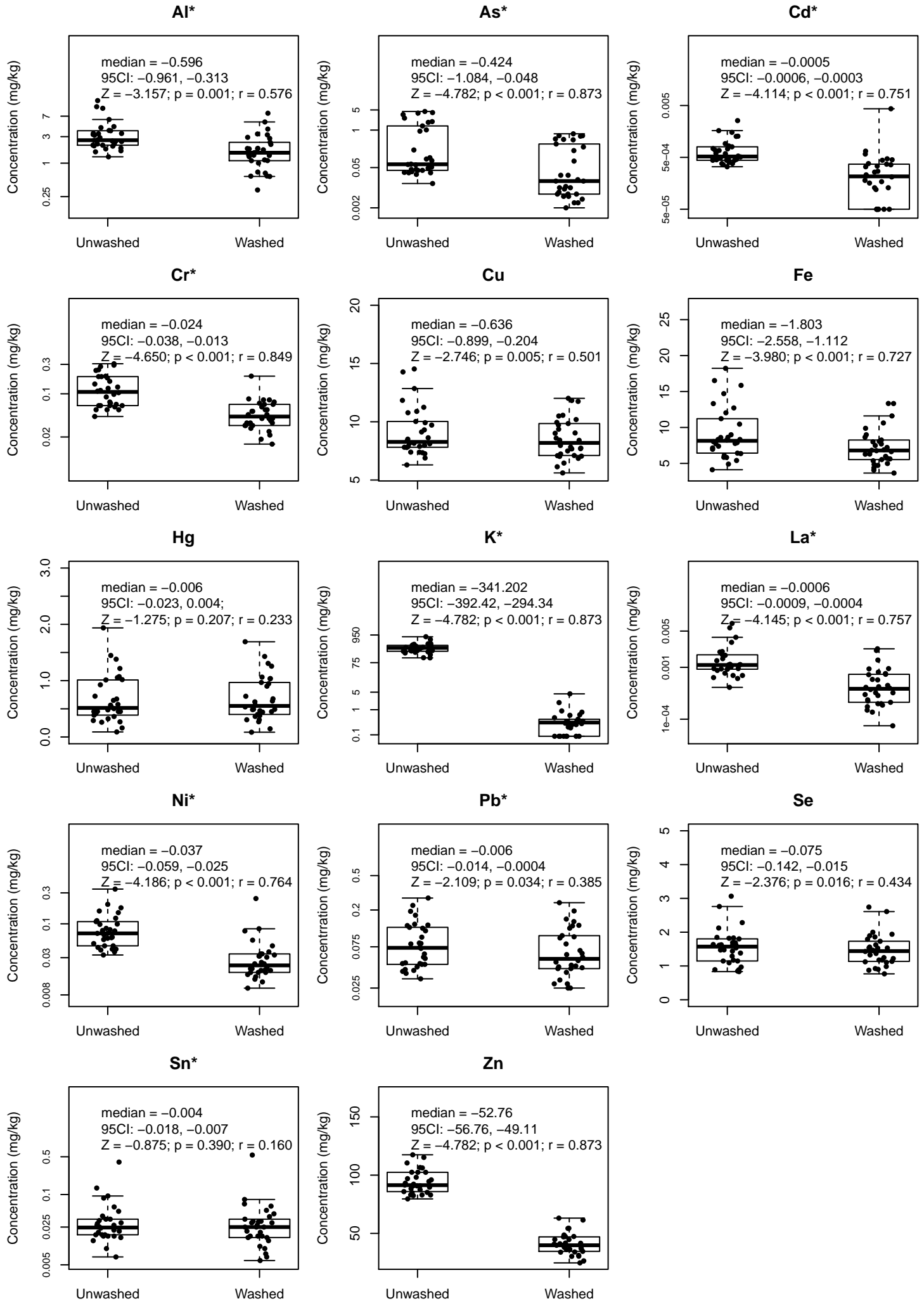


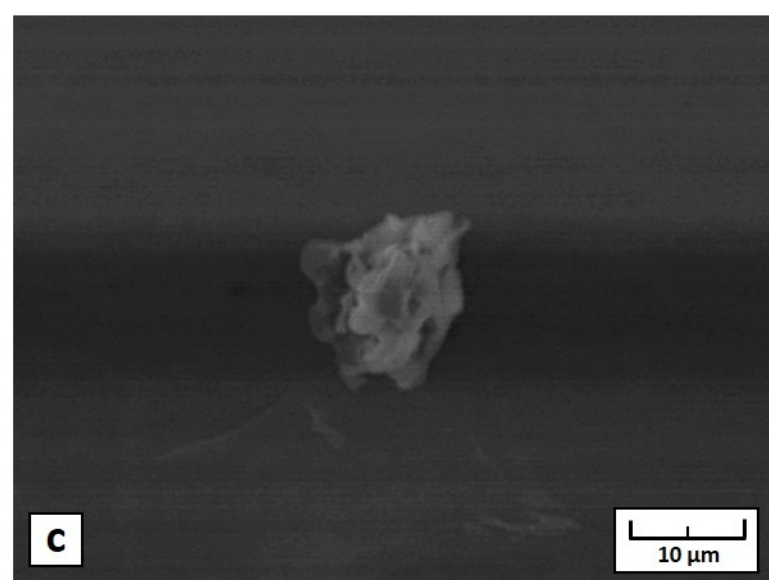
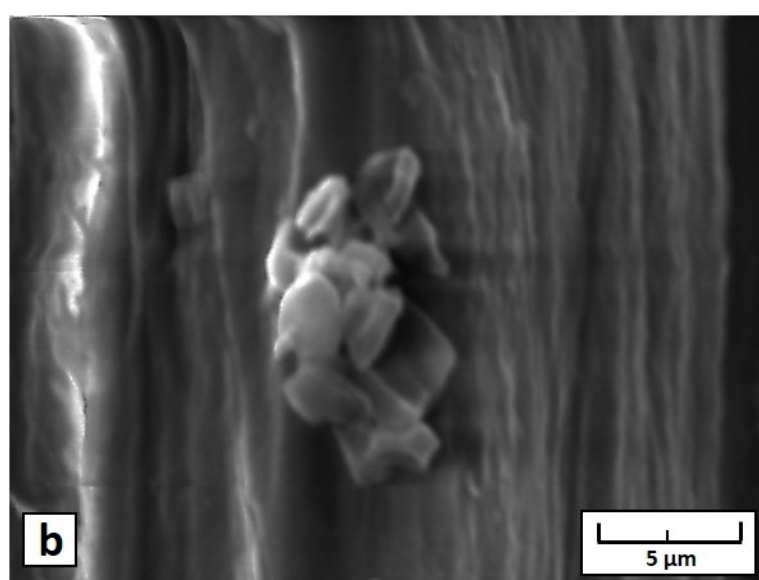
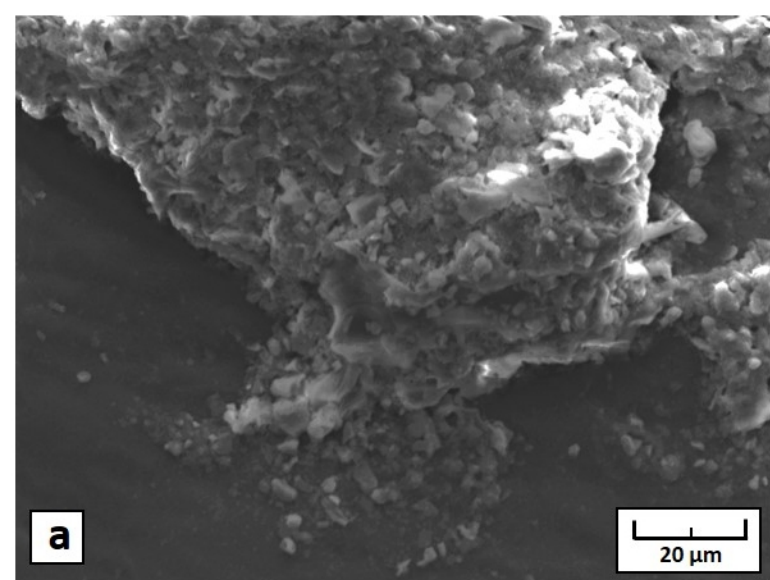
**Aigues-Mortes
(AIG; n=39)**

**Odiel marshes
(ODI; n=40)**

**Fuente de Piedra
(FDP; n=40)**

The effect of washing feathers (duplicated samples only; n = 30)





Trace elements of feather (n = 119) before corrections (Not-corrected) and after (Corrected)

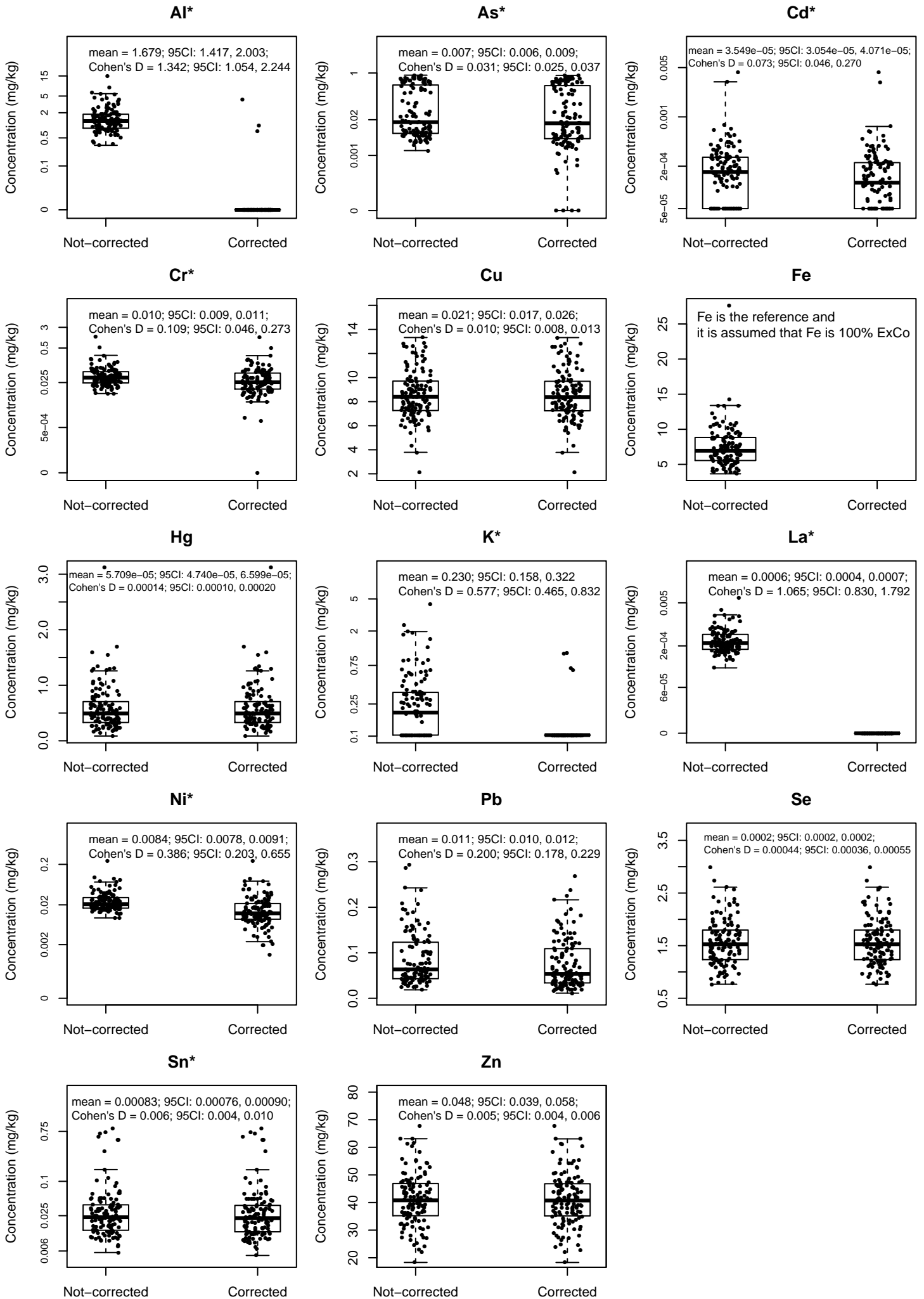


Figure S1: Pairwise Pearson correlation for each of the elements assumed to indicate ExCo (Al,Fe,K,La) analysed in washed feathers in this study prior to ExCo correction (n=119). All correlations are significant except between Fe and K

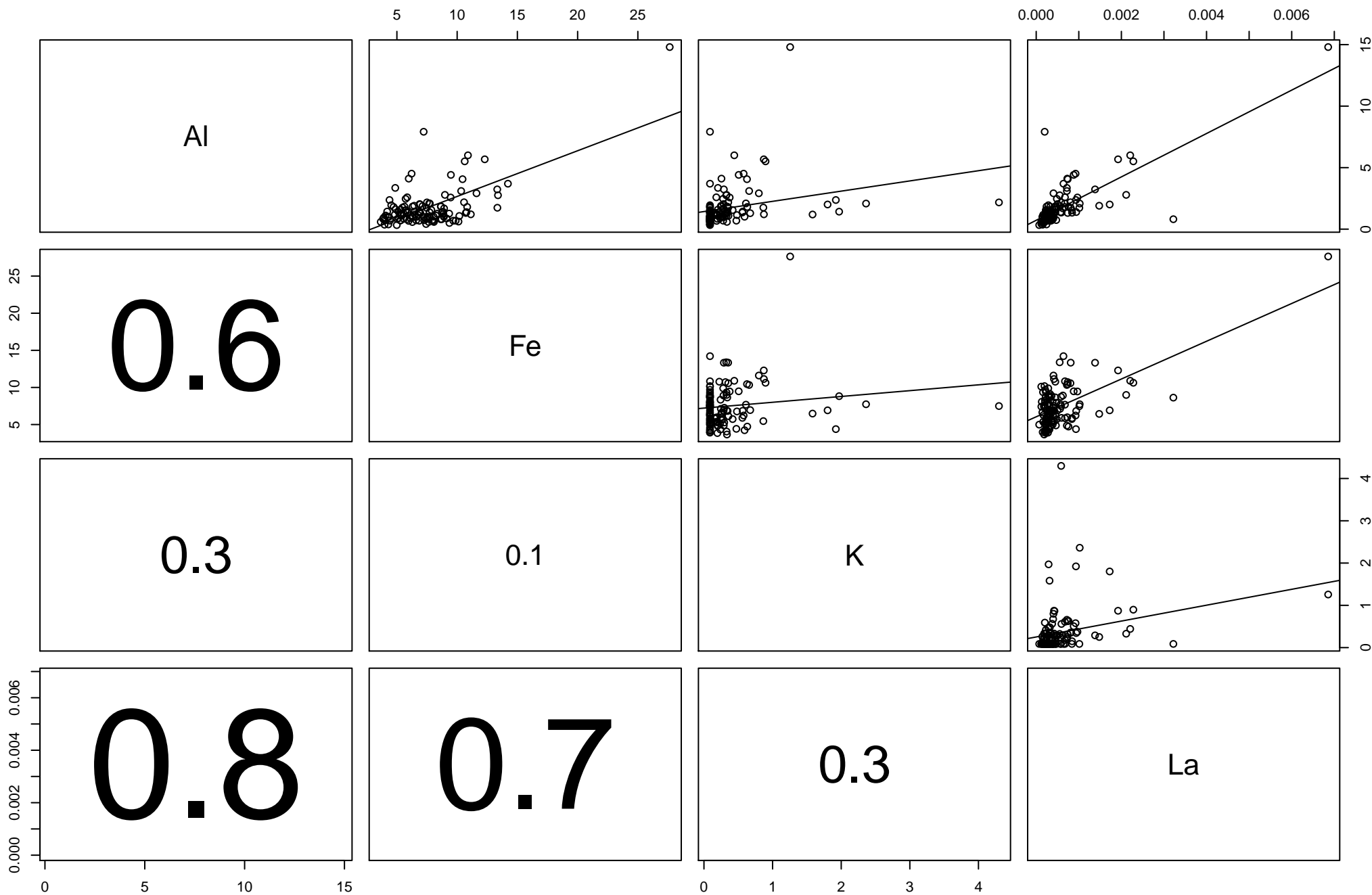


Figure S2: Pairwise Spearman correlation for each of the elements assumed to indicate ExCo (Al,Fe,K,La) analysed in washed feathers in this study prior to ExCo correction (n=119) All correlations are significant except between Fe and K

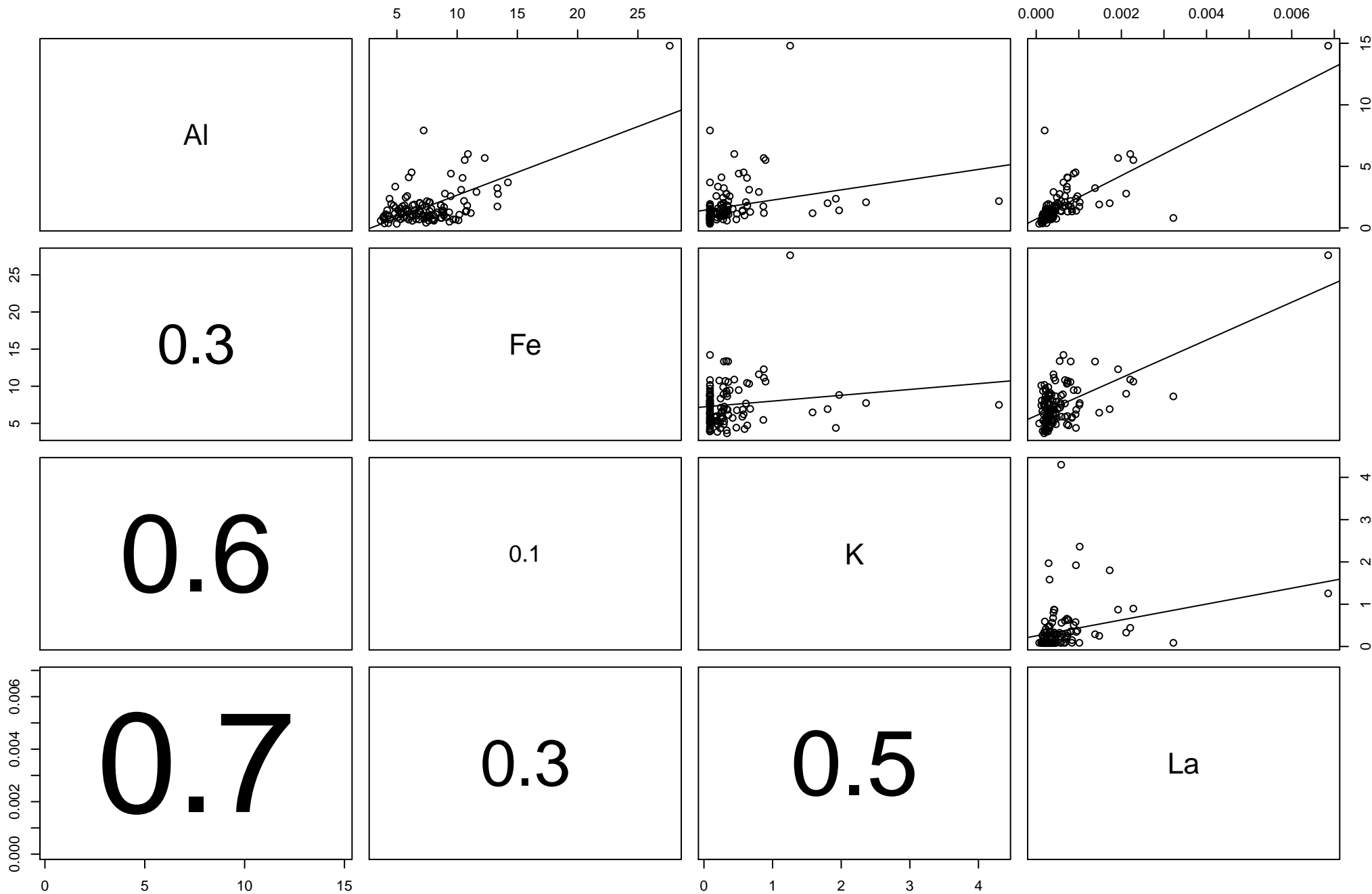


Figure S3: Pairwise Pearson correlation for each of the elements assumed to indicate ExCo (Al,Fe,K,La) analysed in washed feathers in this study prior to ExCo correction (n=118). Without outlier with Fe = 27.64; All correlations are significant except between Fe and K

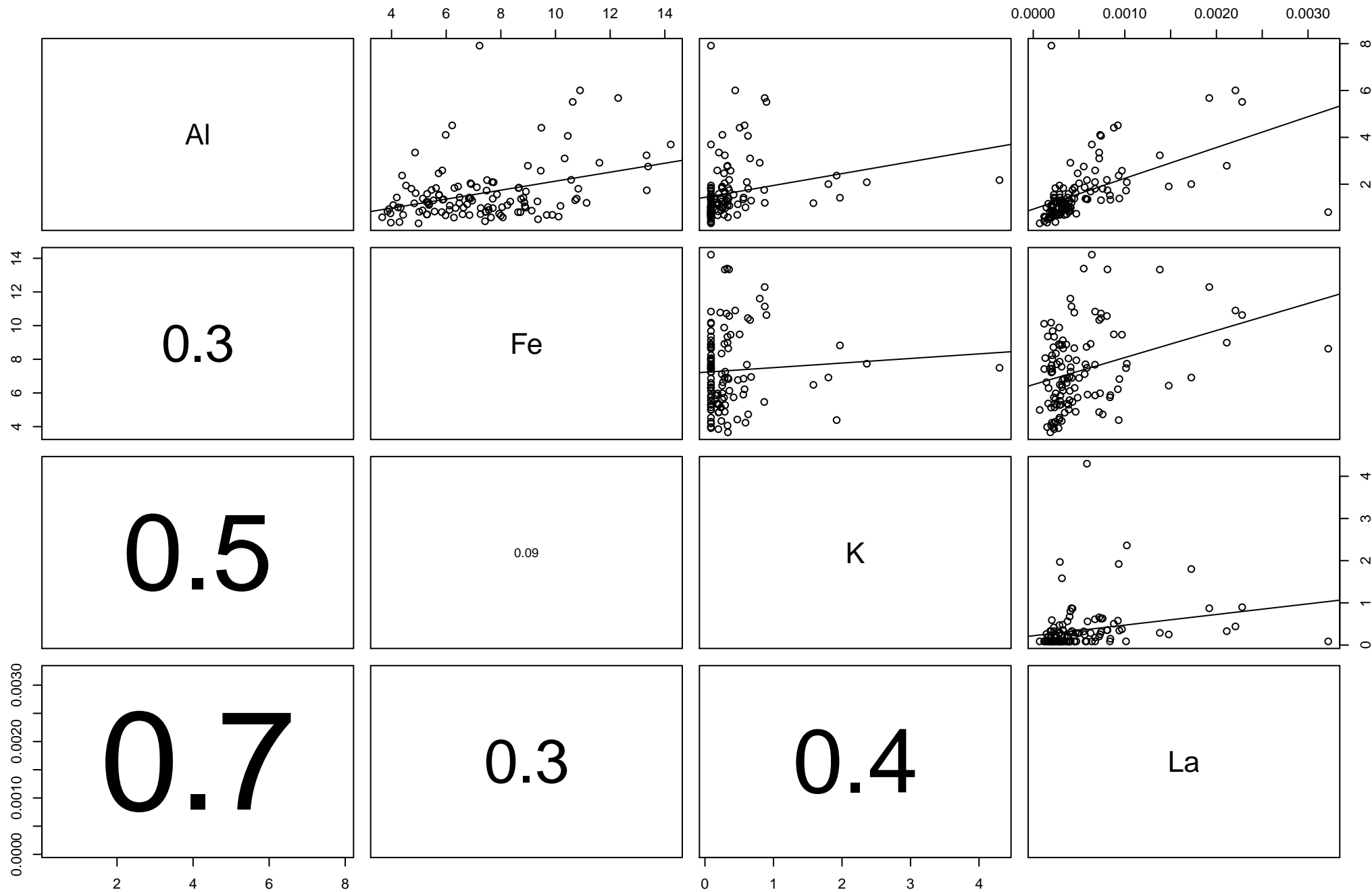


Figure S4: Pairwise Spearman correlation for each of the elements assumed to indicate ExCo (Al,Fe,K,La) analysed in washed feathers in this study prior to ExCo correction (n=118). Without outlier with Fe = 27.64; All correlations are significant except between Fe and K

