1 Assessment of the Water Chemical Quality Improvement based on Human Health

- 2 Risk Indexes: application to in Drinking Water Treatment Plants incorporating
- 3 membrane technologies
- 4

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#### 17 ABSTRACT

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19 A methodology has been developed in order to evaluate the potential risk of drinking 20 water for the health of the consumers. The methodology used for the assessment 21 considers systemic and carcinogenic effects caused by oral ingestion of water based on 22 the reference data developed by the World Health Organization (WHO) and the Risk 23 Assessment Information System (RAIS) for chemical contaminants. The exposure 24 includes a hypothetical dose received by drinking this water according to the analysed 25 contaminants. The assessment of the chemical quality improvement of produced water 26 in the Drinking Water Treatment Plant (DWTP) after integration of membrane 27 technologies using human health risk indexes has been performed.

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29 Series of concentration values covering up to 261 chemical parameters during 5 years 30 (2008-2012) of raw and treated water in the Sant Joan Despi DWTP, at the low part of 31 the Llobregat River basin (NE Spain), have been used. After the application of the 32 methodology, resulting global indexes are located below the thresholds except for 33 carcinogenic risk in the output of DWTP, where the index is slightly above the 34 threshold during 2008 and 2009 before the upgrade of the treatment works with 35 membrane technologies was executed. Annual evolution of global indexes shows a 36 decrease on the global values for all situations: Ho systemic index based on RAIS 37 descends from 0.64 to 0.42 for surface water and from 0.61 to 0.31 for drinking water; 38 R carcinogenic index based on RAIS is negligible for input water and varies from  $4.2 \times 10^{-05}$  to  $7.4 \times 10^{-06}$  for drinking water; W systemic index based on WHO moves from 39 40 0.41 to 0.16 for surface water and from 0.61 to 0.31 for drinking water. A specific 41 analysis for the indexes associated to trihalomethanes (THMs) shows the same pattern. 42 Those indexes have been presented as a tool to show the improvement of the produced 43 water, especially after the year 2009 where the ultrafiltration (UF) and reverse osmosis 44 (RO) membrane technologies where installed.

45

#### 46 **KEYWORDS**

- 47
- 48 Human health risk assessment; Global risk indexes; Membrane technologies; Llobregat
- 49 River; Trihalomethanes

### 50

#### 51 52

- Indexes based on health risk assessment for the treatment of surface water
   (Llobregat river) to drinking water have been designed
  - Annual evolution of indexes shows a decrease on the global index for all situations
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- An upgrade in the water treatment is specially remarkable to improve the carcinogenic risk index for drinking water

#### 60 1. INTRODUCTION

HIGHLIGHTS

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62 In developed countries, a wide implementation of water treating technologies and a 63 proper management has led to a remarkable reduction of the risks associated to water 64 ingestion. Good practices have led to a decrease of the pollution in origin and to a better removal of the contaminants. In the European Union (EU) the Drinking Water Directive 65 (98/83/EC) concerns the quality of water intended for human consumption. According 66 67 to this legislation, a total of 48 microbiological, chemical and indicator parameters must 68 be monitored and tested regularly. Nevertheless, the list of contaminants that should be taken into account is continuously growing as the studies to define the effects on health 69 70 are progressing.

71

72 Water safety plans are considered by the World Health Organization (WHO) as the 73 most effective means of maintaining a safe supply of drinking water to the public. 74 Hazards and risks should be identified, and appropriate steps towards minimizing these 75 risks are then investigated (WHO, 2005). Additionally, the incidence of global driving forces, including climate change, increasing water scarcity, population growth, 76 demographic changes and urbanization are expected to affect the resilience of water 77 78 supply and sanitation systems and services, forcing also managers to adapt their 79 infrastructures to these driving forces (Guha-Sapir et al., 2011).

80

81 Membrane technologies have been identified as the most robust and flexible 82 technologies used to improve water quality and taste by removing undesirable compounds and pathogens (Rahardianto et al., 2007; Reverberi and Gorenflo, 2007). 83 84 Reverse osmosis (RO), nanofiltration (NF) and reverse electrodialysis (RED) are being 85 applied worldwide to meet these needs (Birnhack and Lahav, 2007; Greenlee et al., 86 2009; Wang et al., 2006). The selection of a suitable membrane technology is based on 87 technical criteria (removal of contaminants) and economic aspects (capital operation 88 and maintenance). Implementation of new technologies in drinking water treatment 89 plants (DWTP), as membrane technologies, improves quality of potable water as 90 remove toxic contaminants and reduces human health risk associated to its 91 consumption.

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However, it should be stressed that in order to ensure the minimization of pathogens the required treatment generates disinfection by-products (DBPs), one of the main drawbacks of the drinking water production. Those compounds are produced by the reaction between chemical disinfectants and naturally occurring organic material in the source water (Boorman et al., 1999; Krasner, 2009). The trihalomethanes (THMs), the most abundant DBPs, are human carcinogens according to the WHO (2005). From January 1<sup>st</sup> 2009, a limit of total THMs of 100  $\mu$ gL<sup>-1</sup> is established in the EU (98/83/EC). Although values have been established for a number of DBPs, risks
associated with an inadequate disinfection are far greater than potential risks from longterm exposure to DBPs (WHO, 2014).

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104 It is widely accepted that all stakeholders responsible for water safety should place 105 efforts in improving risk management and risk communication to the consumers, that is, 106 the provision of information and health-based assessments on the various microbial, 107 chemical, radiological and physical human health hazards that may be present in the 108 water cycle. Evaluation of existing and emerging hazards in water should include a 109 proper monitoring at source, after treated and along the distribution network for 110 reducing the risks and a proper approach to manage those associated risks.

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112 Assessing exposure and the health consequences of chemicals in drinking water is 113 challenging: exposures are typically at low concentrations, measurements in water are 114 frequently insufficient, chemicals are present in mixtures, exposure periods are usually 115 long, multiple exposure routes may be involved, and valid biomarkers reflecting the 116 relevant exposure period are scarce. In addition, the magnitude of the relative risks 117 tends to be small (Villanueva et al., 2013). Studies to assess the exposure of 118 contaminants due to drinking water ingestion detected values of arsenic and THMs 119 above the threshold in Turkey (Caylak, 2012) and perfluorooctane sulfonate (PFOS) in Taiwan (Chimeddulam and Wu, 2013). Industrial contamination led to high risk indexes 120 121 due to metals in India (Krishna and Mohan, 2014) and Pakistan (Muhammad et al., 122 2011). Studies in developed countries are more oriented to emerging compounds but 123 they are limited to the availability of reference data. Risk of adverse health effects of 124 pharmaceuticals appeared to be negligibly low in the Netherlands (Houtman et al., 125 2014). Schriks (2010) concluded that the majority of the compounds evaluated pose individually no appreciable concern to human health in the Rhine and Meuse Rivers. 126 127 Ribera (2014) used a combination of Life Cycle Assessment (LCA) and human health 128 risk assessment in order to select the percentage of water in DWTPs that should be 129 nanofiltered. Results show a reduction of one order of magnitude for the carcinogenic risk index when NF produces 100% of drinking water when is compared to the direct 130 131 consumption without treatment.

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133 In this work, a methodology to determine the evolution of the chemical hazard of water 134 has been developed. Additionally, an assessment is included on how this risk has been 135 impacted after the implementation of the new treatment processes. The methodology is 136 supported on toxic effects assessment, exposure assessment and risk indexes 137 characterization (Durham and Swenberg, 2013). The exposure assessment in the present 138 work only considers ingestion of drinking water containing pollutants through oral route 139 as unique pathway and two typologies of effects on the human health have been 140 considered: a) systemic toxicity that refers to adverse effects on any organ system 141 following absorption and distribution of a chemical throughout the body; and b) 142 carcinogenic effects.

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A set of water quality data registered during five years from the DWTP monitoring program has been used to implement the risk assessment methodology. The results obtained will be used to quantify numerically the improvement of the water quality by the use of risk indexes. This study should contribute to develop new managing practices based, not only on the occurrence, but also on the potential hazard of the chemical contaminants.

#### 150 2. MATERIALS AND METHODS

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### 152 2.1. Case study description: Llobregat River and Sant Joan Despí DWTP 153

In recent decades, the drinking water supply network of the Barcelona Metropolitan Area (BMA), 635 km<sup>2</sup> and a population of 4.5 million inhabitants, has been primarily based on surface water resources from the Llobregat and Ter Rivers. Those resources are suffering the effects of mining and industrial discharges, as well as a reduction in quantity, decreasing the quality of the raw water. Additionally, due to the Mediterranean climate, the natural water resource availability is periodically lower than the water demand in the area (López-Roldán et al., 2013).

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162 To improve the water quality of the Llobregat River and its tributaries, more than 30 163 waste water treatment plants (WWTPs) treating a mixture of urban and industrial wastewaters have been set up along the river. The main industries sited along the 164 165 Llobregat River are tannery, food products, textile, pulp and paper industries, discharging a broad spectrum of organic chemicals into the river. Therefore the river 166 167 receives effluents from these WWTPs and surface runoff from agricultural areas. The 168 removal of contaminants by WWTPs is not complete; consequently they can enter into 169 environment via sewage effluents and thus become a potential risk to the receiving 170 bodies and in addition, to the production of drinking water (González et al., 2012; 171 Köck-Schulmeyer et al., 2011; Valero and Arbós, 2010).

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Sant Joan Despí DWTP treats water from the Llobregat River following the process 173 174 flow-sheet described in Figure 1. The plant has a maximum treatment capacity of 5.5 175 m<sup>3</sup>s<sup>-1</sup>, and provides almost 50% of the annual drinking water in the BMA. In 2009, an 176 improved treatment line began its operation. The new process uses membrane 177 technology and treats 50% of the water flow with a pre-treatment via micro-coagulation 178 and ultrafiltration (UF) as protection for the RO step. Water is remineralised before 179 being blended with water from the conventional treatment and sent to the post-180 chlorination stages. This process, the membrane treatment line according to Figure 1, is 181 placed after the sand bed filtration where the flow is split and 50% is treated with the 182 new process; the remaining 50% will undergo ozonisation and granular activated carbon 183 (GAC) filtration as before.

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#### 185 **2.2. Chemical data quality collection and management**

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187 The Llobregat River has been object of several studies dealing with the presence of contaminants in surface water and related compartments (e.g. sediments, fishes). In this 188 189 article only compounds detected in the water matrix are taken into account. Most of 190 these studies focus on the lower and medium part of the river basin, where most of 191 WWTPs, DWTPs and population are located, and therefore, it is the area with higher 192 pressures. Pesticides, surfactants, estrogens, pharmaceuticals and personal care products 193 (PPCPs) and even abuse drugs are the main groups detected in different studies, 194 reporting alterations in species composition, abundance or biomass and endocrine 195 disruption measured by alterations in enzymatic activity or specific protein production 196 (González et al., 2012). Nevertheless, a long list of these compounds is not monitored 197 routinely. The lack of data on their presence and their toxicity makes difficult to include 198 them in risk assessment studies on a large time basis. Table 1 provides a list of the 199 chemical compounds routinely monitored by the Sant Joan Despí DWTP in the year

200 2012. The selection of the parameters to be analysed is done because of the legislation
201 requirements, local characteristics, occurrence according to historical data and
202 assessment of the efficiency of treatment technologies, among other causes.
203

Series of data covering monthly averages of 261 chemical parameters during 5 years (2008-2012) of raw and treated water in Sant Joan Despí DWTP have been used. Not all parameters were measured along the five years, as monitoring programmes have been periodically adapted. Moreover, some compounds are only measured in surface water while other compounds are only measured in drinking water.

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#### 210 **2.3. Fundamentals of the risk assessment methodology**

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Chemicals that display environmental and biological persistence, bioaccumulation, toxicity and long range transport have been previously assessed quantitatively by national and international health agencies (Szabo and Loccisano, 2012). Among the databases that offer information on the toxicity of the compounds that can be found in water, two of the most widely used are the Risk Assessment Information System (RAIS) and the WHO guidelines (WHO, 2011).

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219 RAIS uses the Reference Dose (RfD), expressed as an oral dose per kilogram of body weight (given in units of mgKg<sup>-1</sup>day<sup>-1</sup>), as an estimate of the lowest daily human 220 221 exposure that is likely to occur without appreciable risk of deleterious, non-cancerous 222 effects during a lifetime. WHO proposes a very similar reference value called the 223 Tolerable Daily Intake (TDI) as an estimate of the amount of a substance in food or 224 drinking-water, also expressed on a body weight basis that can be ingested daily over a 225 lifetime without appreciable health risk (WHO, 1991). The TDI values take into account 226 both systemic and carcinogenic effects but risk index is calculated as systemic.

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The exposure assessment of this work only considers ingestion of drinking water containing pollutants through oral route as unique pathway. The oral dose for each contaminant present in water have been calculated by eq 1:

1)

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232 
$$D_i = \frac{C_W \times EF \times ED \times IR_W}{BW \times AT \times 365_{days/year}}$$
 (eq

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where Di represents the dose of contaminant by water ingestion (mg Kg<sup>-1</sup>day<sup>-1</sup>), C<sub>w</sub> is the annual average concentration of the contaminant in water annual average (mgL<sup>-1</sup>), EF is the exposure frequency to the contaminated media (days year<sup>-1</sup>), ED is the exposure duration (year), IR<sub>w</sub> is the rate of water intake (L day<sup>-1</sup>), BW is the body weight of the receptor (Kg), and AT is the average time (year).

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Table 2 shows the exposure values for the pathway of oral ingestion of water according to RAIS and WHO for the calculation of doses. For systemic risk  $D_i$  is calculated by using AT=ED. Then, three different indexes (systemic and carcinogenic for RAIS and an index for WHO) have been calculated:

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a) the systemic effect index according to RAIS  $(H_{Qi})$  is calculated on the dose basis according to RAIS reference values as a ratio between the dose (D) and the dose reference level (RfD) by eq 2:

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249 
$$H_{Qi} = \frac{D_i}{R_{fDi}}$$
 (eq 2)  
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where the ratio of the average daily dose to a RfD below 1 implies that adverse effects are very unlikely to occur. The guideline values were calculated separately considering the risk for individual substances, without specific consideration of additivity. Although it may result in risk underestimations, unless there is evidence to the contrary, it is appropriate to assume that the toxic effects of these compounds are additive (Backhaus and Faust, 2012). Thus, a global systemic effect is obtained as contribution of the individual index values by eq 3:

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$$H_Q = \sum H_{Qi}$$
 (eq 3)

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If H<sub>Q</sub> is below 1 it implies that adverse effects are very unlikely to occur.

b) The individual carcinogenic effects are only considered in RAIS approach and the
 individual carcinogenic effect index (R<sub>i</sub>) is calculated by eq 4:

266 
$$R_i = D'_i \times SF_i \pmod{4}$$

where SF is the Slope Factor (Kg day mg-1) that express a linear relationship of  $D_i$ versus the risk Ri at low doses. Cancer risk is calculated by multiplying the estimated dose or exposure level by the appropriate measure of carcinogenic potency. A guideline value of  $10^{-5}$  means one additional cancer case per 100 000 of the population ingesting drinking-water containing the substance at the guideline value for 70 years (Cothern et al., 1986). Following the principle of additivity of compounds, the global risk index for all compounds is calculated as an addition of individual risk indexes by eq 5:

$$R = \sum R_i \pmod{5}$$

c) the individual WHO (W) index is developed by using eq 6.

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280 
$$W_i = \frac{D_i}{TDI_i}$$
 (eq 6)  
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And then, the global risk index for all compounds is calculated as an addition of individual risk indexes by eq 7:

$$W = \sum W_i \pmod{7}$$

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### 288 2.4. Contaminant concentrations data treatment and filtering of raw data289

A tool was created and validated using Microsoft Excel® programme for the calculation of risk indexes according to RAIS and WHO toxicity values and doses (eq 1) for the list of compounds by using eq 2-7. The tool was programmed to determine annual average concentration of the compounds. Values representing the limit of quantitation (LOQ) for each compound were also introduced so the tool was able to discriminate between measured values and values below LOQ. Blank cells are automatically recognised as not measured parameters in the data analysis.

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The scheme in Figure 2 has been followed in order to assess the risk related to the compounds present in water and incorporates filtering steps to obtain reliable risk indexes mentioned previously. When dealing with raw data concentrations in the calculation of indexes, three main issues where identified and, consequently, filtering steps were applied:

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a) The lack of existence of oral toxicity data for each contaminant. The methodology is
based on risk approach, so the contaminants without toxicity values given by RAIS or
WHO were excluded from index calculations. The comparison of measured
contaminants with levels present in the Directive 98/83/EC is a first step that could
determine the risk when toxicity is not available.

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b) Annual average concentrations were calculated with a mixture of values below the LOQ and quantified values. The election of LOQ/2 is usually applied and solves the uncertainty of a concentration that could be between zero and LOQ but, at the same time, introduces an uncertainty that has to be considered, as could lead to an overestimation of the risk (James et al., 2009). In order to have an idea of this uncertainty for the annual average values, an uncertainty index "U" has been calculated by using eq 8.

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318 
$$U = 1 - \frac{Avg(0)}{Avg(LOQ/2)}$$
 (eq 8)  
319

where Avg(0) is the average concentration when all the values below LOQ are considered as zero and Avg(LOQ/2) is the average concentration when all the values below LOQ are considered as LOQ/2. This U index is 1 if all the values are below LOQ (maximum uncertainty) and 0 if all the values are higher than LOQ (minimum uncertainty). U index will be useful to evaluate the uncertainty of final global indexes.

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c) Goodness of the analytical techniques for the sensitive measurement of risk indexes. It could happen that some analytical techniques are focused on the detection of contaminants just below the legislative values and are not sensitive enough to calculate the contribution of the contaminants to risk assessment when present at very low concentrations. Thus, an important role of the analytical techniques applied to risk indexes calculation would be to provide LOQ values able to quantify small amounts of risk.

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334 In order to decide which analytical techniques are sensitive enough to measure the risk 335 properly, the calculation of the risk indexes by using LOQ levels was performed. 336 Parameters giving values of individual indexes, based on LOQ, below a threshold (0.02 337 for systemic risk according to RAIS, 0.01 for systemic risk according to WHO and  $5 \times 10^{-10}$ 338 <sup>7</sup> for carcinogenic risk) will be included in the index. For the excluded parameters, risk 339 assessment should be performed by comparing the annual average concentration with 340 the limits recognised by legislation, as those thresholds have been also calculated on the 341 basis of risk to human health studies.

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343 Some extra calculations have been programmed so aggregate indexes are easily 344 calculated taking into account some variables e.g. compounds measured at the inlet 345 during the five years and compounds measured at the outlet during the five years. The 346 figures showing the evolution of the final indexes have been programmed and they are automatically updated. The Excel tool has facilitated the index calculations based on a
significant number of data and can be easily adapted to new input data.

- 350 **3. RESULTS**
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#### 3.1. Analysis of water quality improvement in the DWTP

Analysis of the annual evolution, from 2008 to 2012, of the average concentrations of the contaminants is collected in Table 3 and Table 4. Only parameters at the inlet and the outlet that have routinely measured during the five years have been included. As it could be seen, the surface water quality of the inlet water works (Table 3) have improved over the years for most of the parameters, except for arsenic, barium, cyanides, chromium, selenium, tetrachloroethene and tungsten. Other compounds like boron, calcium, strontium, magnesium, nitrates and sulphates remain constant.

362 In the case of the outlet concentrations (Table 4), a reduction of the contaminants levels could be seen for all the compounds except for chlorates and chromium. This reduction 363 364 can be explained mainly by the introduction of the RO step where at least 50% of the 365 total waterworks capacity is treated. When the evolution of the DBPs concentrations is 366 analysed, a reduction of 89% has been achieved for total THMs. It should also be taken 367 into account that the DWTP is applying disinfection by using chlorine up to 2010 and 368 by using chlorine dioxide from 2010, which is the main responsible for the formation of 369 chlorinated DBPs. However, the presence of bromide and iodide acts as precursors for 370 the formation of brominated and iodinated DBPs. Additionally, and due to the use of an 371 ozonisation step, the formation of bromates from bromide occurred. The introduction of 372 the membrane treatment unit has improved the quality and in terms of the reduction of the high salinity (ca. 0.9 g TDSL<sup>-1</sup>), the DBPs precursors and the DBPs themselves so 373 the total content of THMs below  $100 \,\mu g L^{-1}$  is easily accomplished from 2009. 374

375

The increase on the chlorate content should be attributed to the substitution of the initial chlorination steps of the treatment by using chlorine dioxide instead of chlorine. Chlorate and chlorites are disinfection by products of the use of chlorine dioxide. In the case of chromium (VI) the increase is due to the contribution of a groundwater pollution plume generated by electroplating industries, for more than 30 years, at industrial areas of the Llobregat Delta. The seasonal recharge of this plume onto surface water provides the detected peaks, always below the limits fixed by the regulation (10  $\mu$ gL<sup>-1</sup>).

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#### **3.2.** Risk indexes comparison of raw and treated water

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Global indexes for systemic risk according to RAIS and WHO reference values and carcinogenic risk based on RAIS reference data are shown at Table 5. The global risks indexes have been calculated by addition of the individual ones. For the calculation of the global indexes, only the compounds that have been measured during the five years at the surface water (inlet) on one side, and the compounds measured for the five years at the treated water (outlet) on the other side, have been included in order to obtain comparable global indexes.

393

A list of compounds is not included in the global index due to the filtering steps performed, due to the unavailability of reference data or to the low sensitivity of the analytical technique. In case they are excluded, annual average concentrations are 397 compared to the thresholds established at Directive 98/83/EC. Table 6, Table 7 and 398 Table 8 shows the concentrations of those parameters for the outlet water in the 399 calculation of H<sub>Q</sub>, R and W. Only free chlorine (2008-2009), chlorides (2008-2009) and 400 sodium (2008) show levels above Directive reference values.

401

402 Figure 3 shows the annual evolution of the global indexes. Thresholds for the three 403 types of indexes have also been included ( $H_0 < 1$ ;  $R < 10^{-5}$ ; W < 1). It should be highlighted that, although the thresholds have been designed for individual parameters, they are 404 405 being applied in this methodology to the global risk values. Annual evolution of indexes 406 shows a decrease on the global risk for all situations. The biggest reduction can be seen 407 after the first year, 2008, where a severe drought took place. The low average river flow 408 in 2008 (8.12 m<sup>3</sup>s<sup>-1</sup> compared to 12.83 m<sup>3</sup>s<sup>-1</sup>in 2009) may be associated to higher 409 average concentrations of pollutants and, therefore, an increase of the risk indexes.

410

From the methodology developed it is also possible to identify the main contaminants contributing to risk. Lists for the top 10 compounds contributing to every risk index for the year 2012 are shown in Table 9. U indexes shows the uncertainty related to the calculation of the individual risks. The closer U is to 1, the higher the uncertainty of the value of the annual concentration used for the risk calculation. The compounds posing major risk shows U close to 0, except for carcinogenic risk at the inlet that is based on compounds not found (U=1), but global index shows acceptable risk.

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419 For systemic risk according to RAIS reference values, H<sub>0</sub>, the compounds posing a 420 major risk at the inlet are nitrates and nitrites. Nitrates are found at high concentrations 421 and the main risk is linked to their potential of becoming nitrites. Some other 422 compounds like bromates, strontium, boron, barium and nickel are also contributing 423 considerably to the global risk. At the outlet, boron and the halogenated 424 chlorodibromomethane, bromoform, trichloroethene and 1,1,2-trichlorethane appears at 425 the highest position in the list along with the former ones. Some other compounds like 426 the fluorides, free chlorine and metals like thallium, lithium and chromium (VI) and 427 non-metals as arsenic and antimony do not contribute to the global index calculation as 428 the techniques for performing the analysis have been regarded as not sensitive enough. 429 Thus indicates that LOQs should be improved so they can be included in the global risk 430 assessment.

431

432 The situation according to the systemic risk index based on WHO values, W, is not so 433 different to the index based on RAIS values,  $H_Q$ , regarding the compounds showing the 434 highest contribution to the global index. In this case, the filter has only excluded the 435 pesticide atrazine for the inlet plus free chlorine, chlorates and chlorites in drinking 436 water.

437

438 For carcinogenic risk, R, no compounds have a significant risk at the inlet, and values 439 are two order of magnitude below the threshold. Risk is higher at the outlet as DBPs can 440 only be found at the treated water. Compounds like bromates, chromium and arsenic are 441 discarded for contributing to the risk in a big extent due to their high LOQs but were 442 below the values established at the Directive 98/83/CE. At the outlet, legislated THMs 443 and 1,1,2-Trichloroethane have the maximum contribution to the general added risk. A 444 reduction in the risk is especially remarkable beyond 2009. This improvement is 445 associated to the implementation of the RO treatment step, where it expected a 446 reduction of the concentrations of the DBPs precursors (bromide, iodide and dissolved 447 organic matter) and also a reduction of the DBP concentration formed in the
448 chlorination step before coagulation (see Figure 1). This reduction is exemplified at
449 Figure 4 where the evolution of the levels of THMs is shown.

450 451

### 3.3. Contribution of disinfection by-products on risk indexes

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453 A special analysis has been done to disinfection by-products at drinking water. The four 454 (bromoform, THM's included in the legislation bromodichloromethane, 455 dibromochloromethane, chloroform) are some of the top-ranking compounds in 456 contributing to the global carcinogenic risk. Figure 4 shows the decrease of risk over 457 time due mainly to a reduction of the concentration of THMs. It should be stated that 458 for the four regulated THMs, U is equal to 0, showing no uncertainty in the risk 459 calculations as the compounds are always quantified above their LOQs.

460

461 This reduction can be explained by two main factors: the upgrade of the treatment line 462 by inclusion of the RO desalination treatment in 2009 and the substitution of chlorine 463 by chlorine dioxide with a weaker oxidation potential, and then, with a lower capacity 464 for formation of DBPs. The introduction of a desalination step treating 50% of the in-let 465 flow rate is reducing the concentration of the DBPs precursors, both inorganic species 466 as bromide and iodide, and organic species, mainly dissolved organic matter (natural 467 and non natural). The reduction of the DBP concentrations up to four times could only 468 be explained by the combination of both changes on the treatment line.

469

### 470 **3.4. Risk indexes methodology advantages and constraints.**

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The main advantage of these calculations is related to the fact that they are based on three accepted approaches on the assessment of health risks, differentiating between systemic and carcinogenic risk. Those indexes are considering all measured parameters even if the monthly average concentrations are below the LOQ. The obtained indexes can be recalculated as long as new substances are analyzed and WHO or RAIS is recognizing new toxicity values.

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479 But as those global indexes integrate individual values of specific pollutants, it is 480 noticeable that final risk values would increase as new parameters are measured, even if 481 the results of the analytics are below the quantification limit. In order to cope with this 482 problem, it is important to establish filters so substances presenting high-risk values 483 when concentrations are below quantification limits do not overestimate global health 484 risks.

485

486 Series of data show some limitations when the methodology is applied. Monthly 487 averages are calculated on the base of different frequency of measurements depending 488 on the specific parameter, so the number of analysis and the time where they were 489 performed can have an influence on the results. Additionally, not the same list of 490 parameters has been registered during the 5 years. In order to be able to perform an 491 annual comparison, global indexes only includes the parameters that have been 492 measured during the five years covered in the study.

493

Analytical techniques have their own constraints as no concentration values can be
 reported under LOQ. This LOQ is not only dependent on the technique, but on the
 specific compound, water matrix, and the methodology applied for the analytical

497 measurements. Due to some facts like the replacement of the instruments and the
498 criteria for the calculation and acceptance of these limits, an evolution of LOQs can also
499 be observed, making more difficult the interannual comparison.

500

501 Another issue to be faced is presented when trying to assess the risk of produced water 502 in comparison with raw water at the inlet of a DWTP. This analysis can be interesting 503 when evaluating the performance of the treatment technologies in removing certain substances. The difficulties arise when some analytes are only measured in one of the 504 505 water streams as their presence is not expected in the outlet, due to the optimum 506 removal efficiency, or in the inlet, due to their production as result of the treatment of 507 the water flow, e.g. DBPs. The differences in the list of compounds analysed and a 508 change in the limit of quantification, due to the analysis in different water matrixes, 509 pose an additional difficulty in the assessment.

510

#### 511 4. CONCLUSIONS

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A methodology has been developed in order to assess globally the chemical risk of drinking water and its source water. Indexes have been created including those parameters that have passed all the quality filters (existence of reference toxicological values and concentration measured with a sensitive analytical technique). The average concentration of the parameters that were excluded from the hazard indexes has been compared to the threshold established at the legislation.

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520 The annual evolution of the global indexes at the intake and the outlet of a DWTP has 521 reported a continuous decrease of the toxicity from 2008 to 2012. After the application 522 of the methodology, resulting global indexes are located below the thresholds except for 523 carcinogenic risk in the output of the DWTP, where the index is slightly above the threshold during 2008 and 2009 before the upgrade of the treatment works with 524 525 membrane technologies. Annual evolution of indexes shows a decrease on the global 526 values for all situations: H<sub>0</sub> systemic index based on RAIS descends from 0.64 to 0.42 for surface water and from 0.61 to 0.31 for drinking water; R carcinogenic index based 527 on RAIS is negligible for input water and varies from  $4.2 \times 10^{-05}$  to  $7.4 \times 10^{-06}$  for drinking 528 529 water; W systemic index based on WHO moves from 0.41 to 0.16 for surface water and 530 from 0.61 to 0.31 for drinking water. A specific analysis for the indexes associated to 531 trihalomethanes (THMs) shows the same pattern.

532

Form the second group of parameters, not included in the calculation of the indexes,
only free chlorine and chlorides at 2008 and 2009, and sodium at 2009 showed average
concentrations slightly above the threshold for drinking water.

536

Although risk indexes have been calculated in order to help the decision of the stakeholders in charge of water treatment works and administrations dealing with health issues, it is important not to forget that legislation (e.g. Directive 98/83/EC in Europe) is the main reference when assessing the compliance of water quality to health standards. Those indexes have been presented as a tool to show the improvement of the produced water, especially after the year 2009 where the UF and RO membrane technologies where installed.

544

The methodology developed in the form of risk indexes has included more parameters than the existing in the legislation to provide a tool based on risk assessment and not 547 only on the concentration of legislated parameters. Those indexes take into account 548 different effects (systemic and carcinogenic) and are based on reference values given by 549 international organizations taken into account oral ingestion doses. Indexes developed 550 provide a quantification of the quality improvement that could be integrated with Life 551 Cycle Assessment (LCA) and Life Cycles Costing (LCC) analysis.

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553 To summarize, the methodology introduced is able to estimate the risk reduction win 554 when a change on the treatment line is introduced and could be used to estimate 555 potential health benefits for such investment.

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563 564

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- 657

### 659 660 Table 1. Chemical parameters routinely monitored at the Sant Joan Despí DWTP (surface water and produced drinking water) in 2012

Deremeter	Frequency inlet	Frequency outlet	Deremeter	Frequency inlet	Frequency outlet
	measurements	measurements		measurements	measurements
1,1-dichloroethane	every 2 weeks	every week	Fluorides		every month
1,1-dichioloethene	every year	every week	Collium	N/A every day	every nour
1 1 1 2-tetrachloroethane		every week	Geosmin	every 2 weeks	every 2 weeks
1 1 2-trichlorethane	every year	every week	Heptachlor	every week	every month
1.1.2.2-tetrachloroethane	every year	every week	Heptachlor epoxide	every week	every month
1.2-dibromoethane	every year	everv week	Indene(1.2.3-c.d)pyrene	every week	every month
1,2-dichloroethane	every year	every week	Indium	every day	every day
1,2-dichloropropane	every year	every week	Iron	every day	every 8 hours
2-methylisoborneol	every 2 weeks	every 2 weeks	Lanthanum	every day	every day
4,4'-DDD	every week	every month	Lead	every day	every day
4,4'-DDE	every week	every month	Lindane	every week	every month
4,4'-DDT	every week	every month	Lithium	every day	every day
Acenaphthene	every week	every month	m+p-Xylene	every 2 weeks	every month
Acenaphthylene	every week	every month	Magnesium	every day	every day
Alachlor	every week	every month	Malathion	every week	every month
Aldrin	every week	every month	Manganese	every day	every day
alpha-Endosultan	every week	every month	Mercury	every week	every day
Aluminium	every week	every month	Metalashlar	every week	every month
Amotrupo	every day	every 8 nours	Melinete	every week	every month
Ameriyne	every week	every monun	Mohimate	every week	every monun
Anthrocopo		every 12 nours	Naphthalana	every day	every day
Antimony	every day	every honun	Nickel	every 4 hours	every month every day
Arsenic	every day	every day	Nitrates		every day
Atrazine	every week	every month	Nitrites	every week	every month
Barium	every day	every day	Non-ionic tensioactives	every 2 weeks	N/A
Benzene	every 2 weeks	every month	o-Xvlene	every 2 weeks	every month
Benzo(a)anthracene	every week	every month	Palladium	every day	every day
Benzo(a)pyrene	everv week	every month	Parathion	every week	every month
Benzo(b)fluorantene	every week	every month	Pendimethalin	every week	every month
Benzo(g,h,i)perylene	every week	every month	Phenanthrene	every week	every month
Benzo(k)fluorantene	every week	every month	Phenols	every 2 weeks	N/A
Beryllium	every day	every day	Phosphorus	every day	every day
beta-Endosulfan	every week	every month	Pirimicarb	every week	every month
Bismuth	every day	every day	Potassium	every day	every day
Boron	every day	every day	Prometrine	every week	every month
Bromates	every week	every day	Propanil	every week	every month
Bromides	every day	every month	Propazine	every week	every month
Bromochloroacetonitrile	N/A	every week	Pyrene	every week	every month
Bromotorm	every year	every day	Rubidium	every day	every day
c-1,2-Dichloroethene	every year	every week	Selenium	every day	every day
C-1,3-Dichloropropene	every year	every week	Silicon	every day	every day
Cadmium	every day	every day	Silver	every day	every day
Calcium	every 2 months		Sodium	every day	every monun
Chlorates	N/A	every day	Strontium	every day	every day
Chlorfenvinnhos	every week	every month	Sulfates	every day	every day
Chlorides	every day	every day	Sum 4 PAHs Dir 98/83/CF	every week	every month
Chlorites	N/A	every day	Sum THMs Dir 98/83/CF	every year	every month
Chlorodibromomethane	every year	every day	t-1,2-Dichloroethene	every year	every week
Chloroform	every year	every day	t-1,3-Dichloropropene	every year	every week
Chlorpyrifos	every week	every month	Terbuthylazine	every week	every month
Chromium	every day	every day	Terbutryn	every week	every month
Chromium (VI)	every 4 hours	every 4 hours	Tetrachloride carbon	every year	every week
Chrysene	every week	every month	Tetrachloroethene	every 2 weeks	every week
Cobalt	every day	every day	Thallium	every day	every day
Copper	every day	every day	Tin	every day	every day
Cyanides	every 8 hours	every month	Tiobencarb	every week	every month
Diazinon	every week	every month	Titanium	every day	every day
Dibenzo(a,h)anthracene	every week	every month	Toluene	every 2 weeks	every month
Dipromoacetonitrile	N/A	every week	i otal Haloacetonitriles	IN/A	every week
Dichlopenii	every week	every month	I Utal Pesticides	every week	every month
Dichlorobromomothono		every week		N/A	every udy
Dieldrin	EVELY YEAL	every month	Trichloroethono	N/A	every week
Endrin	every week	every month	Trichloroethene + Tetrach	OVERY 2 WEEKS	every week
Ethofumesate	EVERY WEEK	every month	Trifluralin	EVERY MOOK	every month
Ethylbenzene	every 2 weeks	every month	Tunasten	every day	every day
Fenitrothion	every week	every month	Uranium	every 2 months	every vear
Fluoranthene	everv week	every month	Vanadium	every day	every day
Fluorene	every week	every month	Zinc	every day	every day
					, .,

662 663 664 Table 2. Exposure parameters for oral ingestion of water according to RAIS and WHO

Parameters	RAIS	WHO
EF(days year <sup>-1</sup> )	350	365
ED* (years)	24	-
IR (L day <sup>-1</sup> )	2	2
BW (kg)	70	60
AT* (years)	70	-

\*Sistemic risk: AT=ED

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		2008		2009	2009		2010			2012	2008/2012	
Parameter (inlet)	Units	Avg Conc	U	Reduction*								
Aluminium	µg/l	110,285	0,000	108,792	0,000	76,006	0,000	70,283	0,000	65,689	0,000	40%
Antimony	µg/l	1,386	0,030	0,864	0,289	0,500	1,000	0,750	1,000	0,750	1,000	100%
Arsenic	µg/l	0,838	0,298	0,816	0,358	0,927	0,225	1,024	0,488	1,851	0,068	-121%
Barium	µg/l	129,628	0,000	135,570	0,000	149,878	0,000	146,448	0,000	166,529	0,000	-28%
Boron	µg/l	198,792	0,000	154,549	0,000	175,617	0,000	196,130	0,000	211,309	0,000	-6%
Bromates	µg/l	16,584	0,014	6,432	0,034	2,173	0,072	2,500	1,000	2,500	1,000	100%
Bromides	mg/l	0,894	0,000	0,627	0,000	0,580	0,000	0,644	0,000	0,632	0,000	29%
Calcium	mg/l	110,442	0,000	110,509	0,000	112,958	0,000	109,647	0,000	106,272	0,000	4%
Cyanides	µg/l	0,542	0,846	0,500	1,000	0,500	1,000	0,815	0,460	3,905	0,320	-621%
Chlorides	mg/l	390,809	0,000	268,052	0,000	254,409	0,000	255,901	0,000	263,749	0,000	33%
Chromium	μg/l	2,125	0,343	1,512	0,689	1,926	0,379	1,846	0,395	2,487	0,293	-17%
Chromium (VI)	µg/l	4,067	0,461	2,924	0,712	3,374	0,556	2,712	0,845	2,723	0,842	33%
Diazinon	µg/l	0,034	0,009	0,024	0,066	0,031	0,081	0,056	0,028	0,006	0,410	82%
Strontium	mg/l	1,658	0,000	1,689	0,000	1,790	0,000	1,733	0,000	1,708	0,000	-3%
Iron	µg/l	61,401	0,000	55,864	0,000	57,235	0,000	47,908	0,000	50,766	0,000	17%
Phosphorus	µg/l	267,809	0,000	175,152	0,000	182,537	0,000	163,824	0,000	181,920	0,000	32%
Gallium	μg/l	1,370	0,836	1,250	1,000	1,250	1,000	1,250	1,000	1,250	1,000	100%
Geosmin	ng/l	9,492	0,077	1,250	1,000	2,500	1,000	3,727	0,610	8,250	0,253	13%
Lithium	μg/l	27,402	0,000	19,406	0,000	21,095	0,000	18,575	0,000	20,731	0,000	24%
Magnesium	mg/l	33,988	0,000	32,559	0,000	32,955	0,000	33,250	0,000	32,485	0,000	4%
Malathion	μg/l	0,006	0,573	0,004	1,000	0,004	1,000	0,004	1,000	0,004	1,000	100%
Manganese	μg/l	54,221	0,000	32,831	0,000	27,360	0,000	24,742	0,000	26,520	0,000	51%
Mercury	μg/l	0,059	0,000	0,018	0,487	0,014	0,805	0,014	0,837	0,014	0,812	76%
Molybdene	µg/l	2,100	0,020	1,457	0,029	1,521	0,000	1,383	0,000	1,400	0,030	33%
Nickel	μg/l	10,669	0,000	6,888	0,000	5,083	0,000	6,189	0,000	8,924	0,000	16%
Nitrates	mg/l	9,949	0,000	9,810	0,000	11,818	0,000	10,422	0,000	9,372	0,000	6%
Nitrites	mg/l	0,491	0,000	0,339	0,000	0,290	0,000	0,328	0,000	0,164	0,038	67%
Potassium	mg/l	37,768	0,000	26,167	0,000	24,248	0,000	26,619	0,000	29,519	0,000	22%
Rubidium	μg/l	30,274	0,017	13,611	0,230	21,171	0,049	9,978	0,365	12,261	0,127	59%
Selenium	μg/l	0,799	0,365	0,611	0,682	0,500	1,000	0,818	0,840	1,029	0,486	-29%
Silicon	mg/l	1,740	0,000	1,830	0,000	2,159	0,000	1,720	0,000	1,375	0,000	21%
Sodium	mg/l	199,810	0,000	142,642	0,000	133,269	0,000	144,095	0,000	147,260	0,000	26%
Sulphates	mg/l	170,747	0,000	167,688	0,000	176,347	0,000	170,975	0,000	164,838	0,000	3%
Thallium	µg/l	1,779	0,527	1,391	0,824	1,250	1,000	1,603	0,585	1,486	0,701	16%
Non-ionic tensioactives	mg/l	0,090	0,000	0,068	0,000	0,092	0,000	0,081	0,000	0,083	0,000	8%
Terbuthylazine	µg/l	0,042	0,000	0,072	0,000	0,012	0,161	0,020	0,108	0,018	0,136	56%
Terbutryn	µg/l	0,007	0,796	0,005	0,665	0,004	1,000	0,004	0,795	0,004	1,000	100%
Tetrachloroethene	µg/l	0,138	0,264	0,260	0,022	0,125	1,000	0,145	0,788	0,156	0,735	-13%
Titanium	µg/l	1,938	0,000	1,837	0,023	1,232	0,135	1,745	0,072	1,608	0,078	17%
Total Pesticides	µg/l	0,150	0,764	0,125	1,000	0,141	0,815	0,141	0,007	0,020	0,063	87%
Vanadium	μg/l	2,388	0,000	1,492	0,056	1,295	0,064	1,137	0,073	1,002	0,166	58%
Tungsten	μg/l	2,241	0,232	1,495	0,697	1,368	0,838	1,837	0,397	3,332	0,031	-49%

Table 3. Annual average concentrations of the compounds at the inlet and its percentage of reduction in the year 2012 compared to 2008

\*Percentage of reduction of the concentration of the parameter comparing 2008 and 2012. 100% of reduction means that the concentration in 2012 has fallen below LOQ

2008 2009 2010 2011 2012 2008/2012 Parameter (outlet) Units Avg Conc U. Avg Conc 11 Avg Conc U. Avg Conc 11 Avg Conc U. Reduction\* 0,200 0,042 0,025 1,000 0,025 1,000 0,125 1,000 0,125 1,000 100% 1,1,1-trichlorethane µg/l 1.1.2-trichlorethane 3.857 0.001 0.397 0.026 1.117 0.000 1.939 0.000 2.265 0.009 41% µg/l 1.2-dichloroethane 0.281 0.111 0.119 0.480 0.083 0.630 0.225 1.000 0.225 1.000 100% ua/l Aluminium 44.761 0.000 86.387 0.000 57.931 0.000 53.135 0.000 34,908 0.000 22% ua/l Antimony 0,564 0,812 0,551 0,831 0,500 1,000 0,750 1,000 0,750 1,000 100% µq/l Silver 0.669 0.560 0.500 1.000 0.500 1.000 0.750 1.000 0.750 1.000 100% ua/l 0,000 0,000 Barium µq/l 54,383 52,663 36,219 0,000 31,276 0,000 31,261 0,000 43% Boron µq/l 195,303 0.000 129,547 0.000 154,589 0.000 137,830 0.000 147,601 0.000 24% Bromates µq/l 5,745 0,011 7,893 0,000 7,867 0,008 3,636 0,458 4,161 0,350 28% Bromoform µg/l 39.647 0.000 35,545 0.000 19.724 0.000 20,063 0.000 16,126 0.000 59% Bromides ma/l 0.136 0.123 0.216 0.029 0.165 0.076 0.097 0.107 0.074 0.198 46% 134,535 0,000 108,935 0,000 97,789 0,000 92,806 0,000 90,362 0,000 33% Calcium mg/l Free chlorine residual (in situ) 1.006 0.000 1.007 0.000 0.837 0.000 0.828 0.000 0.858 0.000 15% ma/l Total chlorine residual (in situ) 1,192 0,000 1,183 0,000 0,967 0,000 0,931 0,000 0,980 0,000 18% mg/l Chlorates ua/l 85,201 0.039 954,511 0.001 1012,222 0.000 857,978 0.000 822,752 0.000 -866% Chlorites µq/l 13,000 0,353 13,339 0,187 17,021 0,171 14,978 0.111 9,026 0.231 31% Chlorodibromomethane 28.842 0.000 19,148 0.000 6.204 0,000 4.891 0.000 4.277 0.000 85% µg/l Chloroform µq/l 9,373 0,000 6,023 0,000 0,483 0,000 0,745 0,000 0,619 0,000 93% 414,610 0,000 280,842 0,000 186,797 0,000 0,000 183,703 0,000 56% Chlorides mg/l 179,581 Cobalt 0.625 0.667 0.500 1.000 0.500 1.000 0,500 1.000 0.500 1.000 100% µg/l Chromium 2,171 0.384 1,956 0.479 2,631 0,277 2,239 3,921 0.027 -81% µg/l 0,326 0,713 Chromium (VI) 2,500 1,000 2,922 3,240 0,579 3.623 0.403 3,073 0.678 -23% µq/l Dibromoacetonitrile 0.281 0.119 0.050 1.000 0.102 0.287 0.050 1.000 0.089 0.469 68% ua/l Dichlorobromomethane 18,654 0,000 10,881 0,000 0,797 0,000 0,633 0,000 0,634 0,000 97% µq/l 1,103 Strontium 1,852 0,000 1,626 0,000 1,226 0,000 1,098 0,000 0,000 40% mg/l 0.038 0,000 0,045 0,000 1,000 0,125 0,250 100% Ethylbenzene µq/l 0.050 1,000 1,000 9.205 0,000 8.831 0.000 8,999 0,000 0,035 8.389 0.000 8,751 0,106 0.014 5% Iron µg/l 0,151 0.127 29% Fluorides mg/l 0.000 0.114 0.055 0.119 0.016 0.059 0,017 0,196 0,149 13,343 Phosphorus µq/l 24,928 13,242 0,063 10,617 11,174 0,094 46% 12.522 Lithium ua/l 28.483 0.000 17.996 0.000 13.654 0.000 11.287 0.000 0.000 56% 0,071 0.052 0,000 0,050 0,250 0,500 1,000 100% m+p-Xylene 0,000 1,000 1,000 µq/l Magnesium 43.063 0.000 31,344 0.000 24,860 0.000 22,281 0.000 23,147 0.000 46% ma/l 0,901 1,920 0,000 0,142 0,118 0,185 0,995 0,168 48% Manganese 1,176 1,061 µq/l Mercurv 0.046 0.045 0.014 0.794 0.013 1.000 0.013 1.000 0.013 1.000 100% ua/l Molybdene µq/l 1,104 0,189 0,975 0,214 0,500 1,000 0,846 0,246 0,636 0.589 42% Nickel 5,108 0,000 3.263 0.000 2.527 0,000 2,747 0.000 3,763 0.000 26% µg/l Nitrates mg/l 10,755 0,001 10,313 0,000 9,138 0,000 7,700 0,000 7,458 0,000 31% 0,000 23,162 0,000 19,175 0,000 45% Potassium mg/l 34,802 0,000 16,988 17,182 0,000 18,471 0.056 0.436 8,933 0,525 0,565 7,398 0.704 60% Rubidium µq/l 9.556 8,302 0,569 0.805 0.500 1.000 0.500 1.000 0.713 0.750 1.000 100% Selenium µg/l 0,876 Silicon 2,507 0,000 1,643 0.000 1,463 0,000 1,169 0,000 1,249 0.000 50% mg/l Sodium mg/l 203,257 0,000 129,339 0,000 98,539 0,000 99,359 0,000 106,460 0,000 48% 197,292 127,833 110,333 Sulphates 0,000 158,333 0.000 0,000 0,000 109,917 0.000 44% mg/l Thallium 1.446 0.793 0.815 1.250 1.000 1.474 0.707 1.474 0.706 -2% ua/l 1.406 Terbuthylazine 0,005 0,708 0,004 1,000 0,004 1,000 0,004 0,004 1,000 100% µg/l 1,000 Tetrachloroethene µq/l 0.395 0.047 0.084 0.822 0.119 0.472 0,125 1,000 0,213 0,342 46% Toluene µg/l 0.063 0,000 0.093 0.000 0.080 0,573 0,250 1.000 0.250 1.000 100% 0,242 **Total Haloacetonitriles** µq/l 0,414 0,150 1,000 0,243 0,360 0,150 1,000 0,188 0,733 55% Total Trihalomethanes 96.449 0.000 46.296 0.000 13.267 0.000 12,766 0.000 11.044 0.000 89% ua/l 71% Trichloroethene 0,562 0,000 0,173 0,000 0.356 0,000 0,249 0,293 0,163 0.638 µq/l Trichloroethene + Tetrachloroethene 0.953 0.000 0.000 0.301 0.249 0.360 0.521 0.317 0.658 67% µg/l 0.305 0,234 0,500 1,067 0,835 0,349 0.558 0,821 0,586 0,711 1,000 100% Vanadium µq/l

Table 4. Annual average concentrations of the compounds at the outlet and its percentage of reduction in the year 2012 compared to 2008

\*Percentage of reduction of the concentration of the parameter comparing 2008 and 2012. 100% of reduction means that the concentration in 2012 has fallen below LOQ

Table 5. Global risk indexes calculation for systemic risk according to RAIS reference values ( $H_Q$ ) and WHO values (W) and carcinogenic risk (R)

		Inlet		Outlet						
	Hq	R	W	Hq	R	W				
2008	0,64	4,62E-07	0,41	0,61	4,21E-05	0,32				
2009	0,50	4,51E-07	0,31	0,50	2,66E-05	0,25				
2010	0,49	4,68E-07	0,30	0,41	8,18E-06	0,19				
2011	0,47	5,11E-07	0,32	0,33	8,06E-06	0,17				
2012	0,42	5,24E-07	0,16	0,31	7,40E-06	0,17				

# Table 6. Annual averages of the compounds excluded from the $H_Q$ calculation when reference values from Directive 98/83/CE exist

Parametre (excluded from H <sub>Q</sub> )	Unit	2008	2009	2010	2011	2012
Antimony	µg/l	0,564	0,551	0,500	0,750	0,750
Arsenic	µg/l	0,558	0,500	0,500	0,750	0,907
Bromates	µg/l	-	-	-	3,636	4,161
Free chlorine residual (in situ)	mg/l	1,006*	1,007*	0,837	0,828	0,858
Chlorides	mg/l	414,610*	280,842*	186,797	179,581	183,703
Chromium	µg/l	2,171	1,956	2,631	2,239	3,921
Fluorides	mg/l	0,151	0,114	0,119	0,127	0,106
Lindane	µg/l	0,006	0,004	0,004	0,004	0,004
Pirimicarb	µg/l	0,006	0,006	0,006	0,006	0,006
Sodium	mg/l	203,257*	129,339	98,539	99,359	106,460
Sulfates	mg/l	197,292	158,333	127,833	110,333	109,917
Terbuthylazine	µg/l	0,005	0,004	0,004	0,004	0,004
Trichloroethene + Tetrachloroethene	µg/l	0,953	0,305	0,301	0,360	0,317

\*Average concentrations above the thresholds established at Directive 98/83/EC: free chlorine (1 mg/L), chlorides (250 mg/L) and sodium (200 mg/L)

## Table 7. Annual averages of the compounds excluded from the R calculation when reference values from Directive 98/83/CE exist

Parametre (excluded from R)	Unit	2008	2009	2010	2011	2012
Aldrin	µg/l	0,004	0,004	0,004	0,004	0,004
alpha-Endosulfan	µg/l	0,006	0,004	0,004	0,004	0,005
Aluminium	µg/l	44,761	86,387	57,931	53,135	34,908
Ammonium	mg/l	0,038	0,038	0,041	0,038	0,038
Antimony	µg/l	0,564	0,551	0,500	0,750	0,750
Arsenic	µg/l	0,558	0,500	0,500	0,750	0,907
beta-Endosulfan	µg/l	0,004	0,004	0,004	0,004	0,006
Boron	µg/l	195,303	129,547	154,589	137,830	147,601
Bromates	µg/l	5,745	7,893	7,867	3,636	4,161
Free chlorine residual (in situ)	mg/l	1,006*	1,007*	0,837	0,828	0,858
Chlorfenvinphos	µg/l	0,004	0,004	0,004	0,004	0,004
Chlorides	mg/l	414,610*	280,842*	186,797	179,581	183,703
Chromium	µg/l	2,171	1,956	2,631	2,239	3,921
Dieldrin	µg/l	0,004	0,004	0,004	0,004	0,004
Iron	µg/l	9,205	8,831	8,999	8,389	8,751
Fluorides	mg/l	0,151	0,114	0,119	0,127	0,106
Heptachlor	µg/l	0,006	-	-	-	-
Heptachlor epoxide	µg/l	0,004	0,004	0,004	0,004	0,004
Lindane	µg/l	0,006	0,004	0,004	0,004	0,004
Malathion	µg/l	0,004	0,004	0,004	0,004	0,004
Manganese	µg/l	1,920	1,176	1,061	0,901	0,995
Mercury	µg/l	0,046	0,014	0,013	0,013	0,013
Metolachlor	µg/l	0,004	0,004	0,004	0,004	0,004
Molinate	µg/l	0,004	0,004	0,004	0,004	0,004
Nickel	µg/l	5,108	3,263	2,527	2,747	3,763
Nitrates	mg/l	10,755	10,313	9,138	7,700	7,458
Nitrites	mg/l	0,010	0,010	0,010	0,010	0,010
Pendimethalin	µg/l	0,006	0,004	0,004	0,004	0,004
Pirimicarb	µg/l	0,006	0,006	0,006	0,006	0,006
Propazine	µg/l	0,006	0,006	0,006	0,006	0,006
Selenium	µg/l	0,569	0,500	0,500	0,876	0,750
Sodium	mg/l	203,257*	129,339	98,539	99,359	106,460
Sulfates	mg/l	197,292	158,333	127,833	110,333	109,917
Terbuthylazine	µg/l	0,005	0,004	0,004	0,004	0,004
Terbutryn	µg/l	0,006	0,004	0,004	0,004	0,004
Tiobencarb	µg/l	0,006	0,004	0,004	0,004	0,004
Trichloroethene + Tetrachloroethene	µg/l	0,953	0,305	0,301	0,360	0,317

\*Average concentrations above the thresholds established at Directive 98/83/EC: free chlorine (1 mg/L), chlorides (250 mg/L) and sodium (200 mg/L)

## Table 8. Annual averages of the compounds excluded from the W calculation when reference values from Directive 98/83/CE exist

Parametre (excluded from W)	Unit	2008	2009	2010	2011	2012
1,2-dichloroethane	µg/l	0,281	0,119	0,083	0,225	0,225
Alachlor	µg/l	0,006	0,004	0,004	0,004	0,004
alpha-Endosulfan	µg/l	0,006	0,004	0,004	0,004	0,005
Aluminium	µg/l	44,761	86,387	57,931	53,135	34,908
Ammonium	mg/l	0,038	0,038	0,041	0,038	0,038
Arsenic	µg/l	0,558	0,500	0,500	0,750	0,907
Atrazine	µg/l	0,006	0,006	0,006	0,006	0,006
Benzene	µg/l	0,023	0,035	0,050	0,125	0,125
beta-Endosulfan	µg/l	0,004	0,004	0,004	0,004	0,006
Bromates	µg/l	5,745	7,893	7,867	3,636	4,161
Free chlorine residual (in situ)	mg/l	1,006*	1,007*	0,837	0,828	0,858
Chlorfenvinphos	µg/l	0,004	0,004	0,004	0,004	0,004
Chlorides	mg/l	414,610*	280,842*	186,797	179,581	183,703
Chromium	µg/l	2,171	1,956	2,631	2,239	3,921
Iron	µg/l	9,205	8,831	8,999	8,389	8,751
Fluorides	mg/l	0,151	0,114	0,119	0,127	0,106
Heptachlor	µg/l	0,006	0,004	0,004	0,004	0,004
Heptachlor epoxide	µg/l	0,004	0,004	0,004	0,004	0,004
Malathion	µg/l	0,004	0,004	0,004	0,004	0,004
Manganese	µg/l	1,920	1,176	1,061	0,901	0,995
Pirimicarb	µg/l	0,006	0,006	0,006	0,006	0,006
Propazine	µg/l	0,006	0,006	0,006	0,006	0,006
Selenium	µg/l	0,569	0,500	0,500	0,876	0,750
Sodium	mg/l	203,257*	129,339	98,539	99,359	106,460
Sulfates	mg/l	197,292	158,333	127,833	110,333	109,917
Terbutryn	µg/l	0,006	0,004	0,004	0,004	0,004
Tiobencarb	µg/l	0,006	0,004	0,004	0,004	0,004
Trichloroethene + Tetrachloroethene	µg/l	0,953	0,305	0,301	0,360	0,317

\*Average concentrations above the thresholds established at Directive 98/83/EC: free chlorine (1 mg/L), chlorides (250 mg/L) and sodium (200 mg/L)

Parameter	H <sub>Qi</sub> Parameter		Parameter	Ri Pa		Parameter Wi			Parameter H <sub>Qi</sub>		H <sub>Qi</sub> Parameter		Ri		Parameter	Wi	
T didificiei	inlet	U	1 didiffeter	inlet	U	T didificitoi	inlet	U	1 diameter	outlet	U	1 arameter	outlet	U	T didificter	outlet	U
Nitrates	0.160	0.00	Heptachlor	1.59E-07	1.00	Nitrates	0.084	0.00	Nitrates	0.128	0.00	Chlorodibromomethane	3.37E-06	0.00	Nitrates	0.067	0.00
Strontium	0.078	0.00	Benzo(a)pyrene	8.57E-08	1.00	Boron	0.041	0.00	Strontium	0.050	0.00	1,1,2-trichlorethane	1.21E-06	0.01	Bromoform	0.030	0.00
Nitrites	0.045	0.04	Dibenzo(a,h)anthracene	8.57E-08	1.00	Nickel	0.025	0.00	Bromoform	0.022	0.00	Bromoform	1.20E-06	0.00	Boron	0.029	0.00
Boron	0.029	0.00	Benzene	6.46E-08	1.00	Antimony	0.004	1.00	Boron	0.020	0.00	Dichlorobromomethane	3.70E-07	0.00	Nickel	0.010	0.00
Barium	0.023	0.00	Trichloroethene	5.40E-08	1.00	Trichloroethene	0.003	1.00	1,1,2-trichlorethane	0.016	0.01	1,1,2,2-tetrachloroethane	2.35E-07	1.00	Chlorodibromomethane	0.007	0.00
Nickel	0.012	0.00	Ethylbenzene	2.58E-08	1.00	Aldrin	0.001	1.00	Trichloroethene	0.009	0.64	1,2-dichloroethane	1.92E-07	1.00	Nitrites	0.005	1.00
Heptachlor epoxide	0.008	1.00	Atrazine	1.35E-08	1.00	Dieldrin	0.001	1.00	Heptachlor epoxide	0.008	1.00	Chloroform	1.80E-07	0.00	Antimony	0.004	1.00
Molybdene	0.008	0.03	Benzo(a)anthracene	8.57E-09	1.00	Simazine	<0.001	1.00	Beryllium	0.007	1.00	1,2-dichloropropane	1.69E-07	1.00	Trichloroethene	0.004	0.64
Beryllium	0.007	1.00	Benzo(b)fluorantene	8.57E-09	1.00	Tetrachloroethene	<0.001	0.74	Chlorodibromomethane	0.006	0.00	Heptachlor	1.59E-07	1.00	Tetrachloride carbon	0.003	1.00
Trichloroethene	0.007	1.00	Indene(1,2,3-c,d)pyrene	8.57E-09	1.00	Terbuthylazine	<0.001	0.14	Nickel	0.005	0.00	Tetrachloride carbon	8.22E-08	1.00	Chloroform	0.001	0.00

#### Table 9. Lists for the compounds having a higher contribution for every risk index and their associated uncertainty in the year 2012

Fig 1. Diagram of the DWTPs involved in the study. GAC: Granular Activated Carbon; MF: Micro Filtration; REM: Remineralization; RO: Reverse Osmosis; UF: Ultra Filtration. The box indicates the modification introduced on the treatment line including an step of Reverse osmosis



Fig 2. Chart flow representing the methodology for risk assessment and previous filtering steps



# Fig 3. Annual evolution of the global indexes for systemic and carcinogenic risk assessment according to RAIS and WHO reference data



Fig 4. Annual evolution of risk indexes for systemic and carcinogenic risk assessment for the four regulated THMs (bromoform, chlorodibromomethane, chloroform, dichlorobromomethane) at the outlet

