



## Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues

**Brock, Andreas Libonati; Kästner, M.; Trapp, Stefan**

*Published in:*  
Sar and Qsar in Environmental Research

*Link to article, DOI:*  
[10.1080/1062936X.2017.1365762](https://doi.org/10.1080/1062936X.2017.1365762)

*Publication date:*  
2017

*Document Version*  
Peer reviewed version

[Link back to DTU Orbit](#)

*Citation (APA):*  
Brock, A. L., Kästner, M., & Trapp, S. (2017). Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues. *Sar and Qsar in Environmental Research*, 28(8), 629–650. DOI: 10.1080/1062936X.2017.1365762

---

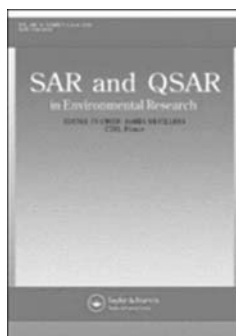
### General rights

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

- Users may download and print one copy of any publication from the public portal for the purpose of private study or research.
- You may not further distribute the material or use it for any profit-making activity or commercial gain
- You may freely distribute the URL identifying the publication in the public portal

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

## SAR and QSAR in Environmental Research



### **Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues**

Journal:	<i>SAR and QSAR in Environmental Research</i>
Manuscript ID	SQER-2017-0051.R2
Manuscript Type:	Special Issue Articles
Date Submitted by the Author:	n/a
Complete List of Authors:	Brock, Andreas; Danmarks Tekniske Universitet Institut for Vand og Miljoteknologi Kästner, Matthias; Helmholtz-Zentrum fur Umweltforschung UFZ, Department of Environmental Biotechnology Trapp, Stefan; Danmarks Tekniske Universitet Institut for Vand og Miljoteknologi
Keywords:	xenobiotics, biodegradation, microbial biomass, turnover modelling, bound residues, organic chemicals of environmental concern

SCHOLARONE™  
Manuscripts

URL: <http://mc.manuscriptcentral.com/sqer>

1  
2  
3 **1 Microbial growth yield estimates from thermodynamics and its**  
4  
5 **2 importance for degradation of pesticides and formation of biogenic**  
6  
7 **3 non-extractable residues**  
8

9  
10 4 Andreas Libonati Brock<sup>a,\*</sup>, Matthias Kästner<sup>b</sup>, Stefan Trapp<sup>a</sup>,  
11

12  
13 5 *<sup>a</sup>Department of Environmental Engineering, Technical University of Denmark, Kgs.*  
14 6 *Lyngby, Denmark*  
15

16  
17 7 *<sup>b</sup>UFZ—Helmholtz Centre for Environmental Research, Department of Environmental*  
18 8 *Biotechnology, Leipzig, Germany*  
19

20  
21 9 \*[alib@env.dtu.dk](mailto:alib@env.dtu.dk), Phone: +45 4525 1408  
22

23  
24 10  
25

26 11  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12 **Microbial growth yield estimates from thermodynamics and its**  
13 **importance for degradation of pesticides and formation of biogenic**  
14 **non-extractable residues**

15 In biodegradation studies with isotope-labelled pesticides fractions of non-  
16 extractable residues (NER) remains, but their nature and composition is  
17 rarely known, leading to uncertainty about their risk. Microbial growth  
18 leads to incorporation of carbon into the microbial mass, resulting in  
19 biogenic NER. Formation of microbial mass can be estimated from the  
20 microbial growth yield but experimental data is rare. Instead, we suggest  
21 using prediction methods for the theoretical yield based on  
22 thermodynamics. Recently, we suggested the Microbial Turnover to  
23 Biomass (MTB) method that needs minimum input data. We have  
24 estimated the growth yield on 40 organic chemicals (31 pesticides) using  
25 the MTB and two existing methods. The results were compared to  
26 experimental values, and the sensitivity of the methods was assessed. The  
27 MTB method performed best for pesticides. Having the theoretical yield  
28 and using the released CO<sub>2</sub> as a measure for microbial activity, we  
29 predicted a range for the formation of biogenic NER. For the majority of  
30 the pesticides, a considerable fraction of the NER was estimated to be  
31 biogenic. This novel approach provides a theoretical foundation applicable  
32 to evaluate and predict biogenic NER formation during pesticide  
33 degradation experiments and may also be employed to interpret NER data  
34 from regulatory studies.

35  
36 Keywords: xenobiotics, biodegradation, microbial biomass, turnover modelling,  
37 bound residues, organic chemicals of environmental concern

38

## 1. Introduction

The evaluation of biodegradation of organic chemicals of environmental concern is a big challenge for risk assessment and is subject to legislation and regulation. In the European Union (EU) chemicals that are traded have to be approved by the REACH legislation if produced and sold in amounts greater than one ton per year [1-3]. The assessment of biodegradability under environmental conditions is standardised by OECD testing guidelines, such as OECD Tests Nos. 306-309 used for the assessment of biodegradation in sea water, soil, fresh water, and fresh-water-sediment systems [4-7].

Transformation and biodegradation is mostly tested with  $^{14}\text{C}$  or  $^{13}\text{C}$  labelled parent compounds. Isotopes are particularly needed for assessment of non-extractable residues (NER; also called “bound residues”) and tracing of unknown metabolites [8]. Assessment of biodegradation is well established but may still have some pitfalls for various compounds [9-12]. Although there are several approaches for the reliable prospective assessment of chemical properties and behaviour from quantitative-structure-activity-relationship (QSAR) modelling available [13], much less is available for the assessment of biodegradation [14]. The assessment of residue formation is still in its infancy and is not yet predictable.

Recently, a novel approach was suggested for modelling the formation of biogenic residues [15] which can also elucidate the black box of NER. NER may be formed by sequestration or entrapment of parent compounds or metabolites in soils and sediments (*type I NER*), and also by covalent bonding to soil organic matter (*type II NER*). Another type of residues is formed after incorporation of carbon into microbial biomass after microbial productive degradation of the parent compound (*type III = biogenic NER*) [8]. Apparently, NER are mostly comprised of all types of residues and

1  
2  
3 63 thus the assessment of the biogenic NER formation will also provide information about  
4  
5 64 the amounts of the other types of NER formed.  
6

7 65 Compounds that are poor growth substrates and do not provide sufficient  
8  
9 66 energy, carbon or nutrients to the degrader microbes give no incentive for degradation  
10  
11 67 and can be expected to be more stable under environmental conditions. Thus, the  
12  
13 68 usability of the molecule, its energy content and suitability for anabolic processes has a  
14  
15 69 profound impact on the evolutionary pressure to develop degradation pathways.  
16  
17

18 70 Thermodynamic analysis can be applied to determine the feasibility and  
19  
20 71 direction of chemical reactions under a given set of conditions (see, e.g., [16]). In  
21  
22 72 addition, thermodynamics can also be used to describe the potential growth of bacteria  
23  
24 73 [17]. Essentially, bacterial growth is simplified and split into anabolic processes (energy  
25  
26 74 demanding) and catabolic processes (energy producing) [18]. The catabolic processes  
27  
28 75 describe the energy released from the oxidation of a chemical or a substrate. In aerobic  
29  
30 76 metabolism the oxidation product is usually  $\text{CO}_2$  and  $\text{H}_2\text{O}$ . The electrons derived from  
31  
32 77 the oxidation are partly transferred to the synthesis of microbial biomass (anabolism) or  
33  
34 78 transferred to the terminal electron acceptors via the respiration chains (catabolism) in  
35  
36 79 the membranes of the organisms resulting in the release of energy, predominately from  
37  
38 80 the formation of  $\text{CO}_2$  and  $\text{H}_2\text{O}$  [19].  
39  
40  
41  
42

43 81 The anabolic processes describe the substrate and energy use for the synthesis of  
44  
45 82 new cell biomass. Element, electron and energy balances are used to describe the  
46  
47 83 processes related to the oxidation and reduction half-reactions of the catabolism and to  
48  
49 84 the energy and electron gain used for cell synthesis (see, e.g., [17, 20-22]).  
50  
51

52 85 The bacterial growth yield is defined as the mass of bacteria formed per mass of  
53  
54 86 substrate consumed (g cells per g substrate, often also g C per g C) [19].  
55

56 87 Thermodynamic growth-yield estimation methods were developed in order to model  
57  
58  
59  
60

1  
2  
3 88 metabolic processes and to estimate the amounts of biomass that can be derived from  
4  
5 89 metabolism (e.g. [15, 17, 20-25]). These estimates have previously been used for  
6  
7 90 biotechnological purposes and for the estimation of activated sludge formation in waste  
8  
9 91 water treatment processes, e.g. [26]. The different growth yield estimation methods are  
10  
11 92 based on a similar set of considerations [26, 27].  
12  
13

14 93 The Thermodynamic Electron Equivalent Model (TEEM2) developed by Perry  
15  
16 94 L. McCarty [17, 23] and Expanded Thermodynamic True Yield Prediction Model  
17  
18 95 (ETTYM) of Xiao and VanBriesen [22, 24] or their variations have been applied [28,  
19  
20 96 29] for the estimation of bacterial growth yield on xenobiotics. Both models have  
21  
22 97 evolved towards an increased need of knowledge regarding the transformation  
23  
24 98 pathways, metabolic processes and the electron and energy losses associated hereto in  
25  
26 99 order to model specific growth of various organisms. Recently, the Microbial Turnover  
27  
28 100 to Biomass (MTB) method was developed by Trapp *et al.* [15] in order to provide a less  
29  
30 101 complex method for the estimation of biomass formation during metabolism of any  
31  
32 102 organic compound. In MTB we proposed a simple method to predict just the minimum  
33  
34 103 bacterial growth yield potential without the need for information on the pathways, as  
35  
36 104 this is rarely known for the majority of chemicals of environmental concern.  
37  
38 105 Furthermore, microbial growth and decline are coupled to the formation of soil organic  
39  
40 106 matter (SOM) [30-33]. Therefore, yield estimation provides a tool for the assessment  
41  
42 107 and prediction of biogenic NER formation in the degradation assessment of chemicals  
43  
44 108 for regulatory purposes.  
45  
46  
47  
48

49 109 The objectives of the present study are i) to thoroughly compare the recently  
50  
51 110 introduced MTB method with other yield estimation methods; ii) to extend the MTB  
52  
53 111 approach for electron acceptors other than oxygen (e.g. nitrate and sulphate); iii) to  
54  
55 112 predict the biomass yields during pesticide degradation and eventually the potential for  
56  
57  
58  
59  
60

1  
2  
3 113 biogenic NER formation; and iv) to determine the sensitive parameters in the yield  
4  
5 114 prediction methods. The estimates were compared to experimentally determined growth  
6  
7 115 yields. Finally, we contrast measured NER formation of pesticides and xenobiotics with  
8  
9 116 the predicted formation of SOM due to bacterial growth and decline.  
10  
11

## 117 2. Materials and Methods

118 The bacterial growth yield prediction methods chosen for this study have a common  
119 basic approach: a stoichiometrically balanced redox reaction and the associated change  
120 of Gibbs free energy. This means that one can set up half-reactions describing the  
121 reduction of the targeted compound (be it xenobiotic or not), calculate the associated  
122 Gibbs free energy [34], and combine it with half-reactions of an appropriate electron  
123 acceptor (e.g.  $O_2$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ,  $Fe^{3+}$ ,  $Mn^{4+}$ , or even  $CO_2$  etc.) allowing for assessment of  
124 the bacterial growth yield under a multitude of redox conditions. Here we only look at  
125  $O_2$ ,  $NO_3^-$ , and  $SO_4^{2-}$ . This approach has been shown for ETTYM and TEEM2 and here  
126 we will show that the procedure of half-reactions is applicable also for the MTB  
127 method.  
128

129 A detailed summary of the methods can be found in the Supplementary  
130 Information (SI) and in the original references. The element, energy and electron  
131 balances differ between the methods, thus a brief outline of the methods will be given in  
132 Table 1, in which the final equations used to calculate the growth yield are shown.  
133

### 132 2.1 Microbial Turnover to Biomass (MTB)

133 The Microbial Turnover to Biomass (MTB) method is presented in detail in  
134 Trapp *et al.* [15]. The method is based on the work of Diekert [18]. The maximum  
135 bacterial yield is determined from the nutritional value of the substrate (N) combined  
136 with the determination of bio-available electrons from the reaction. The nutritional  
137



1  
2  
3 137 value is the inverse of the yield and describes how much substrate is needed for the  
4  
5 138 growth of bacteria [g substrate (g biomass)<sup>-1</sup>]. This is subdivided into a biomass  
6  
7 139 yielding (anabolic) and energy yielding (catabolic) part. The catabolic yield is  
8  
9 140 determined from calculation of the Gibbs free energy released from the oxidation of the  
10  
11 141 compound, the storage of this energy in ATP, and the bacterial growth yield on ATP.  
12  
13  
14 142 Microbes cannot use all electrons to generate energy and thus the  
15  
16 143 concept of bio-available electrons was introduced. [15]. Thus, energy and electron  
17  
18 144 balances are implicitly considered. The anabolic yield is calculated from the carbon  
19  
20 145 content in the compound (the carbon source) and in the bacterial cell [18], i.e. how  
21  
22 146 many grams of cell can be produced from the carbon in the compound (only carbon  
23  
24 147 availability is assumed to limit growth). Further details and examples can be found in SI  
25  
26 148 1.1 and Trapp *et al.* [15].  
27  
28  
29  
30  
31

## 32 150 **2.2 Thermodynamic Electron Equivalent Model 2 (TEEM2)**

33  
34 151 In 1965 P. L. McCarty presented a thermodynamic model to estimate the maximal  
35  
36 152 bacterial yield from a single substrate [17]. The method determines the yield on a given  
37  
38 153 substrate from the Gibbs free energy released in the redox process. Since its inception it  
39  
40 154 has been modified and expanded [34]. It was recently modified to better capture the  
41  
42 155 observed lower yields associated with C1 compounds (i.e. methanol) and reactions  
43  
44 156 involving oxygenases [22, 23]. It is based on electron and energy balances. The electron  
45  
46 157 balance considers that the electrons provided by the substrate are used either in the  
47  
48 158 synthesis of cell material (anabolism) or in energy generation (catabolism), and the  
49  
50 159 energy balance states that the energy *captured* with a specific efficiency ( $\epsilon$ ) by the  
51  
52 160 organism is used for bacterial growth. The energy capture efficiency,  $\epsilon$ , is a key  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3 161 parameter and is estimated from experimental data. Further details and examples are  
4  
5 162 found in SI 1.2 and McCarty [23] and Rittmann and McCarty [34].  
6

7 163

### 164 *2.3 Expanded Thermodynamic True Yield Prediction Model (ETTYM)*

165 The Expanded Thermodynamic True Yield Model is based on the work by McCarty and  
166 was presented in [22] and expanded in [24]. To increase the accuracy for the yield  
167 prediction on C1 compounds and substrates with low degrees of reduction, the authors  
168 proposed to include a carbon and a nitrogen balance and as a result reformulated the  
169 electron and energy balance originally proposed by McCarty. The carbon balance  
170 describes the carbon as either invested in cell synthesis or into other carbonaceous  
171 products. The nitrogen balance can be ignored if nitrogen is not limiting [24], hence, the  
172 yield can be calculated from an energy balance, carbon balance, and electron balance.

173 For further details, see SI 1.3 and Xiao and VanBriesen [22, 24].

174

175 < Table 1 >

176

### 177 *2.4 Conditions for comparison*

178 In all calculations the chemical was assumed to be the sole source of both energy and  
179 carbon. Ammonia was taken as the sole nitrogen source so electrons for the assimilatory  
180 reduction of  $\text{NO}_3^-$  was not considered, and the carbon was assumed to be used only for  
181 cell synthesis or oxidised to  $\text{CO}_2$ , hence, no other carbonaceous products are formed.

182 Gibbs energy of reaction is calculated using activities of the reactants and  
183 products assumed to be 1 M, except for  $\text{H}^+$  which is assumed to be  $10^{-7}$  M (pH 7). The  
184 Gibbs free energy of reaction for non-standard conditions can be calculated as

$$185 \Delta G_r' = \Delta G_r^{o'} + R T \ln Q = \Delta G_r^{o'} + R T \ln \left( \frac{\prod_{i=1}^n [\text{product}]_i^p}{\prod_{i=1}^n [\text{reactant}]_i^r} \right) \quad (1)$$

186

187 where  $R$  is the ideal gas constant [ $8.314 \text{ J (K mol)}^{-1}$ ],  $T$  is the absolute temperature [K],

188  $Q$  is the reaction quotient,  $[product]$  and  $[reactant]$  are the activities of products and

189 reactants, and  $p$  and  $r$  are their respective stoichiometric coefficients. From the equation

190 itself it can be seen that when  $Q < 1$  the term is negative and when  $Q > 1$  the term is

191 positive.

192 For aerobic growth,  $\text{O}_2$  was taken as the terminal electron acceptor. For

193 anaerobic growth both nitrate-reducing conditions ( $\text{NO}_3^-$  as the terminal electron

194 acceptor) and sulphate-reducing conditions ( $\text{SO}_4^{2-}$  as the terminal electron acceptor)

195 were investigated. The balanced half-reactions as reductions and their associated Gibbs

196 free energy of reaction can be found in Table 2 (refer to Table S1, Supplementary

197 Information for simple carbon substrates). In this table, the thermodynamic frame of

198 reference can also be seen:  $\text{SO}_4^{2-}$  is reduced to  $\text{H}_2\text{S}$  and  $\text{HS}^-$ ,  $\text{NO}_3^-$  to  $\text{N}_2$ ,  $\text{O}_2$  to  $\text{H}_2\text{O}$ , C

199 to  $\text{CO}_2$ . Cl is released as HCl, and P as  $\text{PO}_4^{2-}$ . The pH was assumed to be 7 and any

200 change of Gibbs free energy due to speciation was disregarded.

201 For TEEM2, the number of (putative) oxygenase reactions was explicitly taken

202 into account. However, for ETTYM this was not done. This means that the results

203 presented in this study differ from those presented in [24].

204 To assess the accuracy and precision of the model predictions, the relative error

205 ( $E(\%)$ ) and the mean average absolute error (MAE) were calculated.

206

207

&lt; Table 2 &gt;

208

209 *Sensitivity analysis*

1  
2  
3 210 To assess the sensitivity of the different parameters on the predicted growth yield, key  
4  
5 211 input parameters were varied including  $Y_{ATP}$ , the assumption of biological standard state  
6  
7 212 conditions (i.e., chemical activities), Gibbs free energy of formation of the compound,  
8  
9 213 and the default bacterial cell formula (and thus the degree of reduction and cellular  
10  
11 214 carbon content). The sensitivity analysis was done changing one-factor-at-a-time, and  
12  
13 215 the parameter sensitivity was evaluated based on the ratio of change in output to the  
14  
15 216 change in input

$$17 \quad S_i = \frac{\Delta Y}{\Delta X_i} \quad (2)$$

18  
19  
20  
21 218

22  
23 219 where  $\Delta Y$  is the change in the observed output,  $\Delta X_i$  is the change in input  $i$ , and  $S_i$  is the  
24  
25 220 sensitivity of  $i$ .

26  
27  
28 221

### 29 222 *2.5 Chemicals of environmental concern; data*

30  
31  
32 223 Thirty-one pesticides were selected for the present study because they are commercially  
33  
34 224 available and widely applied [35] and have fate data (mineralisation and formation of  
35  
36 225 non-extractable residues NER) available in the EU Pesticide Database [36, 37].

37  
38 226 Bacterial growth yields have been experimentally assessed only for very few of  
39  
40 227 the selected compounds. Where such data was available, it was used for comparing the  
41  
42 228 performance of the prediction methods. Moreover, ibuprofen and some polycyclic  
43  
44 229 aromatic hydrocarbons (PAH) were also included as the bacterial growth yield has been  
45  
46 230 experimentally determined for these. In total 40 chemicals were selected.

47  
48  
49 231 Due to the scarcity of experimentally determined bacterial growth yields on  
50  
51 232 xenobiotics, the methods were also evaluated using the growth yield determined for  
52  
53 233 simple carbon substrates used in biotechnology. The compounds selected for  
54  
55 234 comparison are based on the review and evaluation made by Xiao and VanBriesen [24].  
56  
57  
58  
59  
60

235 Information regarding the Gibbs free energies of formation, number of carbon-  
 236 hydrogen bonds,  $Y_{ATP}$ , number of (putative) oxygenase reactions, and degree of  
 237 reductance are shown along with the name of the compound and reference of Gibbs  
 238 energy of formation in Table 3 (for simple carbon substrates refer to Table S2).

239

240

&lt; Table 3 &gt;

241

## 242 2.6 Calculation of biogenic non-extractable residues

243 Chemicals labelled with carbon isotopes ( $^{14}\text{C}$  or  $^{13}\text{C}$ ) allowed the flow of carbon  
 244 to be tracked in the experimental system [38-41]. If the compound provides carbon to  
 245 anabolism and cell synthesis, the labelled carbon will end up in microbial biomass and  
 246 finally in the biogenic NER. Biogenic NER is not posing a risk to neither the  
 247 environment nor human health [8]. When a substrate  $S$  is mineralized, the amount of  
 248 biomass formed is yield times substrate,  $Y \times S$ , and the evolved  $\text{CO}_2$  is  $(1 - Y) \times S$   
 249 [15]. After the growth phase has stopped, the maximum ratio between biomass and  $\text{CO}_2$   
 250 and is thus

$$251 \frac{[X_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{YS}{(1-Y)S} \text{ or } [X_{\text{biogenic NER}}] = \frac{Y}{1-Y} [\text{CO}_2] \quad (3)$$

252 where  $X_{\text{biogenic NER}}$  is the biomass making up the living biogenic NER. After the cessation  
 253 of the growth phase, the microorganisms start to decay. The dead microorganisms are  
 254 turned over in the microbial food chain and form new biomass,  $\text{CO}_2$  and soil organic  
 255 matter (SOM) [30-33]. Then, the ratio between biogenic NER and  $^{13/14}\text{CO}_2$  becomes  
 256 [15]

$$257 \frac{[\text{SOM}_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{f \times Y}{(1-Y) + (1-f) \times Y} \text{ or } [\text{SOM}_{\text{biogenic NER}}] = \frac{f \times Y}{(1-Y) + (1-f) \times Y} [\text{CO}_2] \quad (4)$$

258

259 where  $SOM_{biogenic\ NER}$  is the non-living biogenic NER,  $f$  is the fraction of decaying  
260 biomass turned over into both living biomass and non-living SOM (0.5, [33]), and  $1-f$  is  
261 the fraction of label released as CO<sub>2</sub>. Eq. (3) can be used to estimate NER formation  
262 during short-term experiments, whereas Eq. (4) holds for long-term experiments. It can  
263 be seen that a high mineralization and CO<sub>2</sub> formation together with a high bacterial  
264 growth yield leads to a high formation of biogenic NER.

### 265 3. Results

#### 266 3.1 Comparison of predicted bacterial growth yields

##### 267 Pesticides and chemicals of environmental concern

268 In Table 4, the predicted bacterial growth yields under aerobic conditions are shown and  
269 compared to experimentally determined growth yields. In Table 5, the predicted growth  
270 yields under anaerobic conditions are shown.

271

272 *Aerobic conditions.* With oxygen as the terminal electron acceptor, only chlorothalonil  
273 was predicted to have a bacterial growth yield of zero and only by the MTB method.  
274 The reason being the absence of carbon-hydrogen bonds. All other compounds except  
275 pyrene were predicted to have a bacterial growth yield of >0.3 g cell carbon (g substrate  
276 carbon)<sup>-1</sup> by all methods. The yield predictions from ETTYM were higher than for  
277 TEEM2 which in turn were higher than the predictions by MTB (except for NTA,  
278 where TEEM2 adjusts for energy lost due to one oxygenase reaction). Experimental  
279 yields were found for 13 of the 40 compounds selected. The mean absolute error was  
280 found to be 49% for MTB, 82% for TEEM2, and 111% for ETTYM (Table 4).

281 Moreover, a strong positive and highly significant ( $p < 0.01$ ) linear correlation between  
282 degree of reductance and predicted yield was found for ETTYM ( $Y = 0.093 \gamma_s + 0.14$ ,

1  
2  
3 283  $R^2 = 0.70$ ). A weaker but still highly significant ( $p < 0.01$ ) correlation was found for  
4  
5 284 TEEM2 ( $Y = 0.09 \gamma_s + 0.15$ ,  $R^2 = 0.53$ ), and a significant ( $p < 0.05$ ) but rather weak  
6  
7 285 correlation was found for MTB ( $R^2 = 0.14$ ).  
8

9  
10 286 For 2,4-D and carbofuran, MTB predicted the yields with an absolute error of  
11  
12 287 4.7% and 1.4%, respectively, whereas TEEM2 overestimated them by 51% and 16%,  
13  
14 288 respectively, and ETTYM overestimated by 56% and 25%, respectively. For glyphosate  
15  
16 289 and anthracene, the predicted growth yields by all three methods deviate greatly from  
17  
18 290 the observed value (more than 180% overestimation).  
19

20 291  
21  
22 292 *Anaerobic conditions.* With nitrate as the electron acceptor, the predicted yields  
23  
24 293 decrease 2-5% for all methods (Table 5). All 40 chemicals are predicted to be  
25  
26 294 degradable under nitrate-reducing conditions when considering their energy and carbon  
27  
28 295 content (the only exception still being chlorothalonil for MTB). The linear correlation  
29  
30 296 between yield and degree of reduction was similar to the aerobic case. Again, ETTYM  
31  
32 297 predictions were higher than TEEM2 predictions, which in turn are higher than the  
33  
34 298 MTB predictions.  
35  
36  
37

38 299 With sulphate as the terminal electron acceptor, the predicted yields decrease  
39  
40 300 54-93% for MTB, 50-118% for TEEM2, and 47-91% for ETTYM, compared to the  
41  
42 301 yields found under aerobic conditions (Table 5). For both aerobic and nitrate-reducing  
43  
44 302 conditions, the ranking of the estimated yields was close. Under sulphate-reducing  
45  
46 303 conditions this was not the case. In fact, TEEM2 predicted negative yields for benzene,  
47  
48 304 benzoate, and EDTA (highlighted in bold), while they were positive (albeit low) for  
49  
50 305 ETTYM and MTB. MTB predicted 20 chemicals to have a yield of  $<0.1$  g cell carbon  
51  
52 306 (g substrate carbon)<sup>-1</sup>, while ten were predicted using TEEM2, and only four using  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3 307 ETTYM. No linear correlation between the predicted yield and the degree of reductance  
4  
5 308 could be found for any of the methods.

6  
7 309 The lack of experimental observations under a multitude of redox conditions  
8  
9 310 makes assessment of the prediction accuracy unachievable.

11 311

12  
13  
14 312 *Simple carbon substrates*

15  
16 313 More experimental data are available for simple carbon substrates, and the result of the  
17  
18 314 growth yield predictions are shown in Table 6. The lowest MAE between experimental  
19  
20 315 and predicted growth yield was observed for TEEM2 (MAE = 15%). MTB showed the  
21  
22 316 highest MAE of 23%. It should be noted, that while TEEM2 took energy losses due to  
23  
24 317 oxygenase reactions into account, the implemented ETTYM method did not.

25  
26  
27 318 Subsequently, the work presented in [24] resulted in an even lower MAE when both the  
28  
29 319 pH and related speciation, and oxygenase reactions were taken into account.

30  
31  
32 320 All three methods predict oxalate to have the lowest yield. Additionally, the  
33  
34 321 TEEM2 and ETTYM predictions showed similar ranking of the chemicals.

35  
36 322 Generally, the highest deviations from the experimental yields were similar for  
37  
38 323 all methods. For MTB, these were formate, oxalate, and glycine; for TEEM2, these  
39  
40 324 were oxalate, glycine, and formate; and for ETTYM, these were oxalate, phenol, and  
41  
42 325 formate.

43  
44  
45 326 Excluding oxalate, formate, and glycine from the results, all three methods  
46  
47 327 predicted the experimental growth yield with a MAE below 15%, with TEEM2 still  
48  
49 328 giving the lowest MAE (9%). These three compounds all have a degree of reductance  
50  
51 329 below that of the assumed cell formula for either method.

52  
53  
54 330

55  
56 331 < Table 4 >



332

333

&lt; Table 5 &gt;

334

335

&lt; Table 6 &gt;

336

### 337 **3.2 Sensitivity Analysis**

338 The calculated average sensitivity  $\bar{S}_i$  towards the varied parameters can be found in

339 Table 7.

340

341  $Y_{ATP}$

342 All methods are relatively sensitive to changes in  $Y_{ATP}$ , especially MTB.  $Y_{ATP}$  is an  
343 uncertain parameter because no such values have ever been determined with pesticides  
344 or other xenobiotics as substrate. The chosen default value for xenobiotics of 5 g cell  
345 dw (mol ATP)<sup>-1</sup> ([18], for methanol) used in the MTB method does not lead to large  
346 errors (cf. Table 4).

347

348 *Gibbs free energy of formation ( $\Delta G_f^{\circ'}$ )*

349 If  $\Delta G_f^{\circ'}$  is positive (e.g. benzene) then an increase in this value would lead to an  
350 increase in the predicted yield. Conversely, a negative  $\Delta G_f^{\circ'}$  which is made more  
351 negative leads to a decrease in the predicted yield (e.g., EDTA). All three methods show  
352 a low sensitivity towards changes of the Gibbs energy of the substrate ( $\bar{S}_i < 0.1$ ). While  
353 TEEM2 and ETTYM also both have low  $\bar{S}_i$  values for changes in  $\Delta G_f^{\circ'}$ , the standard  
354 deviation is high. Yield estimates for chemicals with few carbon atoms (and thus low  
355 formation of CO<sub>2</sub>) and high negative  $\Delta G_f^{\circ'}$  (e.g. glyphosate) are especially sensitive to  
356 changes.

357

358 *Standard state conditions*

359 The deviation from standard state conditions was found to have only a very small effect  
360 on the predicted yields. The relation between  $RT\ln Q$  and  $\Delta G_r'$  in Eq. (1) is logarithmic.  
361 Setting  $\Delta G_r^{\circ}$  to zero shows that varying  $Q$  from  $10^{-25}$  to  $10^{25}$  leads to a change in Gibbs  
362 free energy of only  $\pm 143 \text{ kJ mol}^{-1}$  (at standard pressure and temperature).

363

364 *Cell formula*

365 While all methods are sensitive to the cell formula used (Table 7), MTB is the method  
366 least affected. The effect of the cell formula in TEEM2 and ETTYM is not only on the  
367 energy costs related to synthesis [24], but also on the conversion to g cell carbon (g  
368 substrate carbon)<sup>-1</sup> in TEEM2. This is due to the degree of reduction of the cell ( $\gamma_s$ ) used  
369 in converting the units. For MTB, the predicted yield is only affected by changes in the  
370 assumed cellular carbon content ( $\sigma_C$ ). For MTB, a higher carbon content per mass of  
371 cell leads to a higher yield in g cell carbon (g substrate carbon)<sup>-1</sup> but lower yield in gram  
372 bacteria (gram substrate)<sup>-1</sup> (Figure S1).

373

374 **3.3 Prediction of biogenic NER formation based on the predicted growth yields**

375 The growth yields predicted with the MTB method were used to estimate the biogenic  
376 NER formation from the CO<sub>2</sub> produced during degradation experiments (Table 8). The  
377 formation of biogenic NER was predicted to make up a considerable fraction of the  
378 experimentally determined NER for most of the chemicals. Except for one compound  
379 (glyphosate, caused by the production of the metabolite aminomethylphosphonic acid  
380 (AMPA)), the predicted biogenic NER was smaller than the measured total NER. For  
381 daminozide, almost all of the formed NER is suggested to be made up of biogenic NER

1  
2  
3 382 (94%). For MCPA and MCPB, approximately 50% of the NER is biogenic. For  
4  
5 383 bifentate, iprodione, pendimethalin, phenmedipham, and pymetrozine the biogenic  
6  
7 384 NER is suggested to make up less than 10% of the formed NER. This then suggests that  
8  
9 385 for these chemicals, type I and II NER make up the majority of the formed NER.

10  
11 386 The experimental period for ibuprofen and 2,4-D (64 days) [38, 39], and  
12  
13 387 glyphosate and metamitron (80 days) [42, 43] was shorter than the experiments  
14  
15 388 reported in [36]. Eq. (3), which calculates living biomass  $X$  as biogenic NER, was  
16  
17 389 additionally used to interpret these experiments. In these four studies, the carbon label  
18  
19 390 found in amino acids was reported. For living microbes, about half of the carbon is in  
20  
21 391 proteins. This fraction increases during decay and turnover of microbial biomass  
22  
23 392 because proteins are the most stable fraction of the cells [30, 33, 15]. Except for  
24  
25 393 glyphosate, the measured label in amino acids is within the range of biogenic NER  
26  
27 394 predicted by Eqs. (3) and (4), and the measured total NER is greater.

28  
29  
30  
31  
32 395 < Table 8 >

33  
34 396

## 35 36 397 **4. Discussion**

### 37 38 398 ***4.1 Comparison of predicted bacterial growth yields***

39  
40 399 For the pesticides and chemicals of environmental concern, the lack of experimental  
41  
42 400 data for the bacterial growth yield under different redox conditions made it difficult to  
43  
44 401 assess the error associated with the predicted growth yields. Only 14 yield values could  
45  
46 402 be used in the comparison, with some of them used earlier in [15]. Although MTB  
47  
48 403 performs better than TEEM2 and ETTYM, the MAE was still found to be ~50%.

49  
50 404 The bacterial growth yield estimation methods are all developed to predict the  
51  
52 405 true yield at optimal growth conditions for microorganisms. The observed value is  
53  
54 406 typically a net yield accounting only for the formation of new cell mass and removal of

1  
2  
3 407 the parent compound [15, 23, 24, 44]. The difference between the two is that for the  
4  
5 408 observed yield energy and carbon expenditure, due to non-growth purposes, are not  
6  
7 409 captured (e.g. energy spent on maintenance, formation of metabolites or soluble  
8  
9 410 microbial products and extracellular polymeric substances), unless a dynamic model  
10  
11 411 was used for fit. Hence, the observed yield is typically lower than the true yield.  
12  
13 412 Additionally, the prediction methods assume a complete degradation of the compound.  
14  
15 413 If hardly degradable or insoluble metabolites are formed and rendered not bioavailable  
16  
17 414 (as NER I or II), the observed yield will be lower than the predicted true yield.  
18  
19 415 Compounds which have a known toxic effect (e.g. phenolic compounds [45]) can also  
20  
21 416 result in a higher amount of energy being spent on maintenance leading to an observed  
22  
23 417 yield lower than the predicted true yield.  
24  
25

26  
27 418 The bacterial yields for anthracene and glyphosate are by all methods  
28  
29 419 overestimated by >100%. Anthracene is readily adsorbed and is scarcely soluble in  
30  
31 420 water [46]. In the experiments with glyphosate [42], the intermediate AMPA  
32  
33 421 accumulated, resulting in an observed yield much that was lower than the predicted  
34  
35 422 yield (Table 4). If these two are removed from the calculations, the MAE is reduced to  
36  
37 423 20% for MTB, 40% for TEEM2, and 58% for ETTYM.  
38  
39

40  
41 424 The presence of other sources of carbon or energy (mixed substrate use) also  
42  
43 425 adds uncertainty to the observed value. Interestingly, in the experiments with 2,4-D and  
44  
45 426 carbofuran [29, 47], great care was taken in the experimental setup to minimise  
46  
47 427 confounding factors due to other carbon sources, and here the MTB predicted yields are  
48  
49 428 very close to the experimentally determined values.  
50

51  
52 429 The observed differences might also be attributed to their high hydrophobicity  
53  
54 430 and limited bioavailability [8, 48-50], which means that truly dissolved concentrations  
55  
56 431 are low. Under these conditions, microbes use most of the growth substrate just for  
57  
58  
59  
60

1  
2  
3 432 maintenance [51]. Despite the explicit consideration of energy losses related to  
4  
5 433 (putative) oxygenase reactions for PAHs in TEEM2, its errors were higher than for  
6  
7 434 MTB. Helbling *et al.* [29] successfully matched the predicted bacterial yield with the  
8  
9 435 measured bacterial yield on carbofuran by taking oxygenase reactions into account as  
10  
11 436 suggested in [28].

12  
13  
14 437 Under sulphate-reducing conditions the predicted bacterial yields were much  
15  
16 438 lower than the predicted yields under aerobic conditions, which can be expected  
17  
18 439 considering the lower energy associated with the reduction of sulphate in comparison to  
19  
20 440 oxygen reduction. An interesting observation was that the decrease in yield is similar  
21  
22 441 across all three methods. This shows that the half-reaction approach using various  
23  
24 442 electron acceptors used in ETTYM and TEEM2 can also be used with MTB, which has  
25  
26 443 not been shown before.

27  
28  
29 444 Under anaerobic conditions, the energy released from the majority of redox  
30  
31 445 reactions might not be sufficient to fuel bacterial growth. This suggests that there would  
32  
33 446 not be an evolutionary incentive to develop metabolic pathways for anaerobic  
34  
35 447 degradation where the chemical is the electron donor (unless the chemical can provide  
36  
37 448 other macro- or micronutrients, e.g. nitrogen and carbon). For all methods, the predicted  
38  
39 449 growth yield is so small that there is no relationship between degree of reductance and  
40  
41 450 bacterial growth yield.

42  
43  
44 451 For the simple carbon substrates, both TEEM2 and ETTYM perform better than  
45  
46 452 MTB (lower MAE). However, one has to consider the fact that the efficiency parameter  
47  
48 453 in the TEEM2 method was calibrated to the data in order to produce yield estimates  
49  
50 454 close to experimentally determined growth yields [23]; and experimental yields were  
51  
52 455 converted to g cell carbon (g substrate carbon)<sup>-1</sup> using the cell formula (CH<sub>2</sub>O<sub>0.6</sub>N<sub>0.2</sub>)  
53  
54 456 proposed for the ETTYM method [24].  
55  
56  
57  
58  
59  
60

1  
2  
3 457 While ETTYM and TEEM2 both overestimated the yield for oxalate, MTB  
4  
5 458 estimated it as zero due to the absence of C-H bonds (which points to the need for a  
6  
7 459 model modification as bacteria are able to grow on this substrate).  
8  
9  
10 460

#### 11 461 **4.2 Sensitivity analysis**

12 462 All the methods were shown to be sensitive to the choice of cell formula but exhibited  
13  
14 463 low sensitivity to variations of the formation energy,  $\Delta G_f^\circ$ , of the chemical of interest.  
15  
16 464 All methods are based on the Gibbs energy of reaction and knowledge of the Gibbs  
17  
18 465 energy of formation of the chemical of interest is needed. If the value has not been  
19  
20 466 determined experimentally (e.g. [52]), it can be estimated using group contribution  
21  
22 467 methods [53-58] (method [53] is implemented in the freely available ChemProp [59]),  
23  
24 468 or by component contribution methods [60] (implemented in the free accessible  
25  
26 469 database eQuilibrator [61]), or calculated using quantum mechanics [62]. For  
27  
28 470 xenobiotics, the applicability of these estimation methods may be limited.  
29  
30 471 Consequently, we also tested the sensitivity of Gibbs energy of formation of the  
31  
32 472 xenobiotic substrate by setting this value to 0 kJ mol<sup>-1</sup> (Table S3). The MTB method has  
33  
34 473 surprisingly low sensitivity. Compounds having a large negative Gibbs energy of  
35  
36 474 formation (e.g. NTA, EDTA, and glyphosate) and few carbon-hydrogen bonds (6, 8,  
37  
38 475 and 4, respectively) show a maximum deviation of around 20% from the predictions  
39  
40 476 done with correct Gibbs energy of formation. Overall, the average deviation is only 6%.  
41  
42 477 In comparison, TEEM2 and ETTYM have considerably higher average deviation (14%  
43  
44 478 and 11%, respectively).  
45  
46  
47  
48  
49  
50

51 479 While deviation from the standard state conditions might be needed to render a  
52  
53 480 reaction step thermodynamically feasible [16], the effect on the overall yield prediction  
54  
55 481 would only be seen for reactions where the Gibbs free energy of reaction is low, either  
56  
57  
58  
59  
60

1  
2  
3 482 due to the low energy associated with the oxidation of substrate (e.g. formate or  
4  
5 483 formaldehyde), or the low energy associated with the reduction of the electron acceptor  
6  
7 484 (e.g.  $\text{SO}_4^{2-}$ ). This means that the true concentrations or activities during the reaction can  
8  
9 485 be neglected without significant error.

10  
11 486 The effect of pH on Gibbs energy of reaction was not investigated in this study  
12  
13 487 as this was recently done in [24]. At pH 7, taking the distribution of the inorganic  
14  
15 488 carbon species into account only changed the predicted yield approximately 1%.  
16  
17 489 However, speciation of the substrate also has an effect on its Gibbs energy of formation.  
18  
19 490 Similarly, the sensitivity of the energy capture efficiency parameter  $\varepsilon$  (or variations  
20  
21 491 thereof) has been assessed elsewhere [22, 23].  
22  
23  
24  
25

#### 26 492 27 493 ***4.3 Prediction of biogenic NER formation and implication for degradation in the*** 28 29 494 ***environment***

30  
31 495 The experiments cited from [36] in Table 8 were run for more than 100 days. The peak  
32  
33 496 in living biomass is usually after a few days to weeks [38, 39], and therefore we expect  
34  
35 497 that Eq. (4) ( $SOM_{biogenic\ NER}$ ) is more appropriate for these experiments than Eq. (3)  
36  
37 498 ( $X_{biogenic\ NER}$ ) as the majority of the living biomass has decayed and been incorporated  
38  
39 499 into SOM after 100 days. Results obtained by Eq. (4) are smaller than the measured  
40  
41 500 NER, which confirms the results of this equation. The only exception is glyphosate. For  
42  
43 501 daminozide, the chemical with the highest predicted yield, the calculated biogenic NER  
44  
45 502 and measured NER are almost equal.  
46  
47  
48

49 503 The examination of Table 8 gives no significant correlation between measured  
50  
51 504 total NER and predicted  $X_{biogenic\ NER}$  or  $SOM_{biogenic\ NER}$ . Such a correlation should not be  
52  
53 505 expected since the processes leading to NER I, II and III are competing. If a pesticide is  
54  
55 506 not degraded it can undergo aging and irreversible sorption (type I NER) and covalent  
56  
57  
58  
59  
60

1  
2  
3 507 binding of the parent compound or its metabolites (type II NER) [8]. The estimation of  
4  
5 508 the various fractions of NER can be rebuilt in a dynamic simulation model. We  
6  
7 509 suggested such a model in [8] and used it successfully for the prediction of the NER  
8  
9 510 formation from 2,4-D and ibuprofen with pre-estimated yield data [15].  
10

11         The data compiled by [36] give no hints into which form the NER are present,  
12 and thus cannot serve to validate the estimation equation. However, it is likely that high  
13  
14 512 true yields are connected to high experimental yields, which stimulate bacterial growth  
15  
16 513 and thus microbial degradation. Of course, the enzymatic pathways to facilitate the  
17  
18 514 degradation and energy exploitation of the molecule also need to be present.  
19  
20 515  
21

22         In degradation experiments with metamitron [43], glyphosate [42], ibuprofen  
23 [39, 63] and 2,4-D [38, 63], the formation of biogenic NER was investigated by  
24  
25 517 tracking the distribution of stable carbon or nitrogen isotope ( $^{13}\text{C}$  or  $^{15}\text{N}$ ) in  $\text{CO}_2$ , amino  
26  
27 518 acids, fatty acids, metabolites, and parent compounds. Experiments of this kind are very  
28  
29 519 helpful to discriminate between the various types of NER and to validate our biogenic  
30  
31 520 NER estimation approach. Shrestha *et al.* [11] observed that the formation of NER  
32  
33 521 occurred simultaneously with the degradation and release of  $\text{CO}_2$ . This shows the  
34  
35 522 coupling of the formation of NER to microbial activity, and to the growth and decay of  
36  
37 523 biomass. Mamy *et al.* [13] observed a lack of QSAR approaches to predict the  
38  
39 524 formation of NER. The method applied in this study provides process-based theoretical  
40  
41 525 background that may be used to interpret NER data derived in degradation experiments.  
42  
43 526 Before routine application though, further confirmation by targeted experiments is still  
44  
45 527 needed.  
46  
47 528  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60



1  
2  
3 530 **5. Conclusions**  
4

5 531 The MTB method was compared with two widely used bacterial growth yield  
6  
7 532 estimation methods, TEEM2 and ETTYM. The results showed that TEEM2 and  
8  
9 533 ETTYM methods performed better than MTB in estimating the yield on simple  
10  
11 534 substrates, while MTB performed better when estimating the yield on organic chemicals  
12  
13 535 of environmental concern in general and in particular on pesticides. It was also shown  
14  
15 536 that the MTB approach can be expanded to electron acceptors other than oxygen, like  
16  
17 537 sulphate and nitrate.  
18

19  
20  
21 538 The sensitivity analysis revealed that all three methods are relatively sensitive to  
22  
23 539 changes in  $Y_{ATP}$ , an uncertain parameter. TEEM2 and ETTYM are also sensitive to  
24  
25 540 changes in the cell formula due to the change in the degree of reductance. All methods  
26  
27 541 showed low sensitivity to variations in the Gibbs energy of formation of the organic  
28  
29 542 chemicals because most of the Gibbs energy of reaction stems from the formation of the  
30  
31 543 oxidation products carbon dioxide and water. The growth yield estimates were then  
32  
33 544 successfully used to estimate the formation of biogenic non-extractable residues. The  
34  
35 545 approach applied in this study provides a theoretical foundation that can be used to  
36  
37 546 predict biogenic NER formation during pesticide degradation experiments. It can also  
38  
39 547 be employed to interpret NER data derived during regulatory studies.  
40  
41  
42  
43  
44  
45  
46

47 549 **Acknowledgements**

48 550 This research project was financially supported by the Technical University of  
49  
50 551 Denmark and the Helmholtz Centre for Environmental Research UFZ. We thank  
51  
52 552 Fabio Polesel, Carson Odell Lee, and Ulrich Bay Gosewinkel for valuable  
53  
54 553 suggestions and discussions. We also wish to acknowledge the comments and  
55  
56  
57  
58  
59  
60

1  
2  
3 554 suggestions provided by the anonymous reviewers which helped to improve the  
4  
5 555 manuscript.

6  
7 556 The MTB theoretical yield tool is available both in spreadsheet and Python code  
8  
9 557 on request from the first author.

10  
11  
12  
13 558 **Disclosure statement**

14 559 The authors declare no financial interest.

15  
16 560

For Peer Review Only

561 **References**

- 562 [1] EU. Regulation (EC) No 1907/2006 of the European Parliament and the Council of  
563 18 December 2006 concerning Registration, Evaluation, Authorisation and  
564 Restriction of Chemicals (REACH), Official Journal of the European Union  
565 (2006), L 136.
- 566
- 567 [2] EU. Regulation (EC) No 1107/2009 of the European Parliament and the council of  
568 21 October 2009 concerning the placing of plant protection products on the  
569 market and repealing Council Directives 79/117/EEC and 91/414/EEC, Official  
570 Journal of the European Union (2009), L 309/1.
- 571
- 572 [3] EU. Commission regulation (EU) No 283/2013 setting out the data requirements for  
573 active substances, in accordance with Regulation (EC) No 1107/2009 of the  
574 European Parliament and the Council concerning the placing of plant protection  
575 products on the market. Official Journal of the European Union (2013), L 93/1.
- 576 [4] OECD. *Test No. 306: Biodegradability in Seawater*. OECD Publishing, Paris, 1992.  
577 DOI 10.1787/9789264070486-en
- 578 [5] OECD. *Test No. 307: Aerobic and Anaerobic Transformation in Soil*. OECD  
579 Publishing, Paris, 2002a. DOI 10.1787/9789264070509-en
- 580 [6] OECD. *Test No. 308: Aerobic and Anaerobic Transformation in Aquatic Sediment*  
581 *Systems*. OECD Publishing, Paris, 2002b. DOI 10.1787/9789264070523-en
- 582 [7] OECD. *Test No. 309: Aerobic Mineralisation in Surface Water – Simulation*  
583 *Biodegradation Test*. OECD Publishing, Paris, 2004. DOI  
584 10.1787/9789264070547-en
- 585 [8] M. Kästner, K. M. Nowak, A. Miltne, S. Trapp, and A. Schäffer, *Classification and*  
586 *Modelling of Nonextractable Residue (NER) Formation of Xenobiotics in Soil – A*  
587 *Synthesis*, Crit. Rev. Env. Sci. Technol. 44 (2014), pp. 2107–2171.
- 588 [9] K. Fenner, S. Canonica, L.P. Wackett, and M. Elsner, M., *Evaluating Pesticide*  
589 *Degradation in the Environment: Blind Spots and Emerging Opportunities*,  
590 Science 341 (2013), pp. 752–758.
- 591 [10] M. Honti, and K. Fenner, *Deriving persistence indicators from regulatory water-*  
592 *sediment studies - Opportunities and limitations in OECD 308 data*, Environ. Sci.  
593 Technol. 49 (2015), pp. 5879–5886
- 594 [11] P. Shrestha, T. Junker, K. Fenner, S. Hahn, M. Honti, R. Bakkour, C. Diaz, and D.  
595 Hennecke, *Simulation Studies to Explore Biodegradation in Water–Sediment*  
596 *Systems: From OECD 308 to OECD 309*, Environ. Sci. Technol. 50 (2016), pp.  
597 6856–6864.
- 598 [12] M. Honti, S. Hahn, D. Hennecke, T. Junker, P. Shrestha, and K. Fenner, *Bridging*  
599 *across OECD 308 and 309 Data in Search of a Robust Biotransformation*  
600 *Indicator*, Environ. Sci. Technol. 50 (2016), pp. 6865–6872.

- 1  
2  
3 601 [13] L. Mamy, D. Patureau, E. Barriuso, C. Bedos, F. Bessac, X. Louchart, F. Martin-  
4 602 Laurent, C. Miege, and P. Benoit, *Prediction of the Fate of Organic Compounds*  
5 603 *in the Environment From Their Molecular Properties: A Review*, Crit. Rev. Env.  
6 604 Sci. Tec. 45 (2015), pp. 1277–1377.
- 7  
8 605 [14] A. Sabljic and Y. Nakagawa, Biodegradation and quantitative structure-activity  
9 606 relationship (QSAR), in *Non-First Order Degradation and Time-Dependent*  
10 607 *Sorption of Organic Chemicals in Soil*, W. Chen, A. Sabljic, S.A. Cryer and R.S.  
11 608 Kookana, eds, Book Series: ACS Symposium Series, American Chemical Society,  
12 609 Washington (DC), Volume 1174, 2014, pp. 57-84.
- 13  
14 610 [15] S. Trapp, A. Libonati Brock, K. Nowak and M. Kästner, *Prediction of the*  
15 611 *formation of biogenic non-extractable residues during degradation of*  
16 612 *environmental chemicals from biomass yields*, Environ. Sci. Technol. (2017),  
17 613 *Submitted for publication*
- 18  
19 614 [16] T. Maskow, and U. von Stockar, *How reliable are thermodynamic feasibility*  
20 615 *statements of biochemical pathways?*, Biotechnol. Bioeng. 92 (2005), pp. 223–230.
- 21  
22 616 [17] P.L. McCarty, *Thermodynamics of biological synthesis and growth*, Intl. J. Air  
23 617 Water Poll. 9 (1965), pp. 621-639.
- 24  
25 618 [18] G. Diekert, Grundmechanismen des Stoffwechsels und der Energiegewinnung, in  
26 619 *Umweltbiotechnologie*, J.C.G. Ottow and W. Bidlingmaier, eds., Fischer Verlag,  
27 620 Stuttgart, Germany, 1997; pp. 1-38.
- 28  
29 621 [19] M.T. Madigan, J. Martinko, and J. Parker, *Biology of Microorganisms*. International  
30 622 Student Edition, Pearson Inc, 2011.
- 31  
32 623 [20] J.J. Heijnen, *A new thermodynamically based correlation of chemotrophic biomass*  
33 624 *yields*, Anton. Leeuw. Int. J. G. 60 (1991), pp. 235-256.
- 34  
35 625 [21] J.J. Heijnen, and J.P. Dijken, *In search of a thermodynamic description of biomass*  
36 626 *yields for the chemotrophic growth of microorganisms*, Biotechnol Bioeng 39  
37 627 (1992), pp. 833-852
- 38  
39 628 [22] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic model for microbial true*  
40 629 *yield prediction*, Biotechnol. Bioeng 93 (2006), pp. 110–121.
- 41  
42 630 [23] P.L. McCarty, *Thermodynamic Electron Equivalents Model for bacterial yield*  
43 631 *prediction: Modifications and comparative evaluations*, Biotechnol. Bioeng 97  
44 632 (2007), pp. 377–388.
- 45  
46 633 [24] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic true yield prediction*  
47 634 *model: adjustments and limitations*, Biodegradation 19 (2008), pp. 99–127.
- 48  
49 635 [25] R. Kleerebezem, and M.C.M. Van Loosdrecht, *A Generalized Method for*  
50 636 *Thermodynamic State Analysis of Environmental Systems*, Crit. Rev. Env. Sci. Tec.  
51 637 40 (2010), pp. 1–54.
- 52  
53 638
- 54  
55  
56  
57  
58  
59  
60

- 1  
2  
3 639 [26] G. Tchobanoglous, F.L. Burton, and H.D. Stensel, *Wastewater Engineering,*  
4 640 *Treatment and Reuse*, 4th ed. McGraw-Hill: Boston, USA, 2004.
- 5  
6 641 [27] J.M. VanBriesen, *Evaluation of methods to predict bacterial yield using*  
7 642 *thermodynamics*, *Biodegradation* 13 (2002), pp. 171–190.
- 8  
9 643 [28] J.M. VanBriesen, *Thermodynamic yield predictions for biodegradation through*  
10 644 *oxygenase activation reactions*, *Biodegradation* 12 (2001), pp. 265–281.
- 11  
12  
13 645 [29] D.E. Helbling, F. Hammes, T. Egli, and H.-P.E. Kohler, *Kinetics and yields of*  
14 646 *pesticide biodegradation at low substrate concentrations and under conditions*  
15 647 *restricting assimilable organic carbon*, *Appl. Environ. Microbiol.* 80 (2014), pp.  
16 648 1306–1313.
- 17  
18 649 [30] R. Kindler, A. Miltner, H.H. Richnow, and M. Kästner, *Fate of gram-negative*  
19 650 *bacterial biomass in soil - Mineralization and contribution to SOM*, *Soil Biol.*  
20 651 *Biochem.* 38 (2006), pp. 2860–2870.
- 21  
22  
23 652 [31] R. Kindler, A. Miltner, M. Thullner, H.H. Richnow, and M. Kästner, *Fate of*  
24 653 *bacterial biomass derived fatty acids in soil and their contribution to soil organic*  
25 654 *matter*, *Org. Geochem.* 40 (2009), pp. 29–37.
- 26  
27 655 [32] A. Miltner, R. Kindler, H. Knicker, H.H. Richnow, and M. Kästner, *Fate of*  
28 656 *microbial biomass-derived amino acids in soil and their contribution to soil*  
29 657 *organic matter*. *Org. Geochem.* 40 (2009), pp. 978–985.
- 30  
31  
32 658 [33] A. Miltner, P. Bombach, B. Schmidt-Brücken, and M. Kästner, *SOM genesis:*  
33 659 *Microbial biomass as a significant source*, *Biogeochemistry* 111 (2012), pp. 41–  
34 660 55.
- 35  
36 661 [34] B.E. Rittmann, P.L. McCarty, *Environmental biotechnology: principles and*  
37 662 *applications*. McGraw-Hill: New York, NY, 2001.
- 38  
39  
40 663 [35] J. E. Ørum, and H. Hossy, *Bekæmpelsesmiddel-statistik 2014.*  
41 664 *Behandlingshyppighed og pesticidbelastning, baseret på salgsstatistik og*  
42 665 *sprøjtejournaldata*. Orientering fra Miljøstyrelsen nr. 13, 2015. Miljøstyrelsen  
43 666 (Danish EPA), 2015.
- 44  
45 667 [36] E. Barriuso, P. Benoit, and I. G. Dubus, *Formation of pesticide nonextractable*  
46 668 *(bound) residues in soil: Magnitude, controlling factors and reversibility*,  
47 669 *Environ. Sci. Technol.* 42 (2008), pp. 1845–1854.
- 48  
49  
50 670 [37] EU. *EU Pesticides Database*. Available at  
51 671 <http://ec.europa.eu/food/plant/pesticides/eu-pesticides-database/public/> (visited  
52 672 2017-04-07).  
53 673
- 54  
55 674 [38] K. M. Nowak, A. Miltner, M. Gehre, A. Schäffer, and M. Kästner, *Formation and*  
56 675 *fate of bound residues from microbial biomass during 2, 4-D degradation in soil*,  
57 676 *Environ. Sci. Technol.* 45 (2011), pp. 999–1006.
- 58  
59  
60

- 1  
2  
3 677 [39] K. M. Nowak, C. Girardi, A. Miltner, M. Gehre, A. Schäffer, and M. Kästner,  
4 678 *Contribution of microorganisms to non-extractable residue formation during*  
5 679 *biodegradation of ibuprofen in soil*, *Sci. Total Environ.* 445–446 (2013), pp. 377–  
6 680 384.
- 7  
8  
9 681 [40] C. Poßberg, B. Schmidt, K.M. Nowak, M. Telscher, A. Lagojda, and A. Schaeffer,  
10 682 *Quantitative Identification of Biogenic Nonextractable Pesticide Residues in Soil*  
11 683 *by 14C-Analysis* *Environ. Sci. Technol.* 50 (2016), pp. 6415–6422.
- 12  
13 684 [41] M. Kästner, K.M. Nowak, A. Miltner, and A. Schäffer, *(Multiple) Isotope probing*  
14 685 *approaches to trace the fate of environmental chemicals and the formation of non-*  
15 686 *extractable “bound” residues*. *Curr. Opin. Biotech* 41 (2016), pp. 73–82
- 16  
17 687 [42] S. Wang, B. Seiwert, M. Kästner, A. Miltner, A. Schäffer, T. Reemtsma, Q. Yang,  
18 688 and K.M. Nowak, *(Bio)degradation of glyphosate in water-sediment microcosms*  
19 689 *- A stable isotope co-labeling approach*, *Water Res* 99 (2016), pp. 91–100.
- 20  
21  
22 690 [43] S. Wang, A. Miltner, and K.M. Nowak, *Identification of degradation routes of*  
23 691 *metamitron in soil microcosms using 13C-isotope labeling*, *Environ. Pollut.* 220  
24 692 (2017), pp. 927–935.
- 25  
26  
27 693 [44] D. W. Tempest and O.M. Neijssel, *The Status of  $Y_{app}$  and Maintenance Energy as*  
28 694 *Biologically Interpretable Phenomena*. *Ann. Rev. Microbiol.*, 38 (1984), pp. 459–  
29 695 486.
- 30  
31 696 [45] B.I. Escher, R.W. Hunziker, and R.P. Schwarzenbach, *Interaction of phenolic*  
32 697 *uncouplers in binary mixtures: concentration-additive and synergistic effects*.  
33 698 *Environ Sci Technol* 35 (2001), pp. 3905–14.
- 34  
35  
36 699 [46] R.P. Schwarzenbach, P.M Gschwend, and D.M. Imboden, *Environmental Organic*  
37 700 *Chemistry*, 1st ed., John Wiley & Sons, Inc., New York, NY, 1993.  
38 701
- 39  
40 702 [47] N.M. Chong, S.C. Tsai, and T.N. Le, *The biomass yielding process of xenobiotic*  
41 703 *degradation*, *Bioresource Technol.* 101 (2010), pp. 4337–4342.
- 42  
43 704 [48] T.N. Bosma, P.J.M. Middeldorp, G. Schraa, and A.J.B. Zehnder, *Mass transfer*  
44 705 *limitation of biotransformation: Quantifying bioavailability*, *Environ. Sci.*  
45 706 *Technol.* 31 (1997), pp. 248-252.
- 46  
47  
48 707 [49] L.Y. Wick, T. Colangelo, and H. Harms, *Kinetics of mass transfer-limited bacterial*  
49 708 *growth on solid PAHs*, *Environ. Sci. Technol.* 35 (2001), pp. 354–361.
- 50  
51 709 [50] A. Katayama, R. Bhula, G.R. Burns, E. Carazo, A. Felsot, D. Hamilton, C. Harris,  
52 710 Y.H. Kim, G. Kleter, W. Koerdel, J. Linders, J.G.M.W. Peijnenburg, A. Sabljic,  
53 711 R.G. Stephenson, D.K. Racke, B. Rubin, K. Tanaka, J. Unsworth and R.D.  
54 712 Wauchope, *Bioavailability of xenobiotics in the soil environment*, *Rev. Environ.*  
55 713 *Contam. Toxicol* 203 (2010), pp. 1-86.
- 56  
57  
58  
59  
60

- 1  
2  
3 714 [51] A. Rein, I.K.U. Adam, A. Miltner, K. Brumme, M. Kästner, and S. Trapp, *Impact*  
4 715 *of bacterial activity on turnover of insoluble hydrophobic substrates (phenanthrene*  
5 716 *and pyrene) — Model simulations for prediction of bioremediation success*, J.  
6 717 *Hazard. Mater.* 306 (2016), pp. 105–114.
- 7  
8 718 [52] R.K. Thauer, K. Jungermann, and K. Decker, *Energy conservation in chemotrophic*  
9 719 *anaerobic bacteria*, *Bacteriol. Rev.* 41 (1977), pp. 100–180.
- 10  
11 720 [53] K. G. Joback, and R.C. Reid, *Estimation of pure-component properties from*  
12 721 *group-contributions*, *Chem. Eng. Comm.* 57 (1987), pp. 233–243.
- 13  
14 722 [54] M.L. Mavrovouniotis, *Group contributions for estimating standard Gibbs energies*  
15 723 *of formation of biochemical compounds in aqueous solution*, *Biotech. Bioeng.* 36  
16 724 (1990), pp. 1070–1082.
- 17  
18 725 [55] M. L. Mavrovouniotis, *Estimation of standard Gibbs energy changes of*  
19 726 *biotransformations*, *J. Biol. Chem.* 266 (1991), pp. 14440–14445.
- 20  
21 727 [56] M.L. Mavrovouniotis, S. Prickett, and L. Constantinou, *Object-oriented estimation*  
22 728 *of properties from molecular structure*, *Comput. Chem. Eng.* 16 (1992), pp. 353–  
23 729 360.
- 24  
25 730 [57] J. Marrero, and R. Gani, *Group-contribution based estimation of pure component*  
26 731 *properties*, *Fluid Phase Equilibr.* 183–184 (2001), pp. 183–208.
- 27  
28 732 [58] A. S. Hukkerikar, B. Sarup, A. Ten Kate, J. Abildskov, G. Sin, and R. Gani,  
29 733 *Group-contribution+ (GC+) based estimation of properties of pure components:*  
30 734 *Improved property estimation and uncertainty analysis*, *Fluid Phase Equilibr.* 321  
31 735 (2012), pp. 25–43.
- 32  
33 736 [59] G. Schüürmann, R. Kühne, F. Kleint, R.-U. Ebert, C. Rothenbacher, and P. Herth,  
34 737 *A software system for automatic chemical property estimation from molecular*  
35 738 *structure. in Quantitative Structure-Activity Relationships in Environmental*  
36 739 *Sciences – VII*, F. Chen and G. Schüürmann G, eds, SETAC Press, Pensacola  
37 740 (FL), USA, 1996, pp. 93-114.  
38 741
- 39  
40 742 [60] E. Noor, H.S. Haraldsdóttir, R. Milo, R.M.T. Fleming, *Consistent Estimation of*  
41 743 *Gibbs Energy Using Component Contributions*, *PLoS Comput. Biol.* 9 (2013).
- 42  
43 744 [61] A. Flamholz, E. Noor, A. Bar-Even, and R. Milo, *eQuilibrator - the biochemical*  
44 745 *thermodynamics calculator*, *Nucleic Acids Res.* 40 (2012), pp. 770-775.
- 45  
46 746 [62] *Mol-Instincts, A Fundamental Chemical Database based on Quantum Mechanics*  
47 747 *& QSPR*, Copyright by ChemEssen, Inc. 2017; software available at  
48 748 <http://www.mol-in.com>  
49 749
- 50  
51 750 [63] C. Girardi, K.M. Nowak, O. Carranza-Diaz, B. Lewkow, A. Miltner, M. Gehre, A.  
52 751 Schäffer, M. Kästner, *Microbial degradation of the pharmaceutical ibuprofen and*  
53 752 *the herbicide 2,4-D in water and soil - Use and limits of data obtained from*

- 1  
2  
3 753            *aqueous systems for predicting their fate in soil*, Sci. Total Environ. 444 (2013),  
4 754            pp. 32–42.
- 5  
6 755    [64] S.L. Trabue, X. Feng, A.V. Ogram, and L.T. Ou, *Carbofuran degradation in soil*  
7 756            *profiles*, J. Environ. Sci. Heal. B 32 (19976), pp. 861–878.
- 8  
9 757    [65] Z. Yuan and J. M. VanBriesen, *Yield prediction and stoichiometry of multi-step*  
10 758            *biodegradation reactions involving oxygenation*, Biotechnol. Bioeng. 80 (2002),  
11 759            pp. 100-113
- 12  
13  
14 760    [66] C.N Sawyer, P.L. McCarty, and G.F. Parkin, *Chemistry for environmental*  
15 761            *engineering and science*, 5th Edn., McGraw-Hill: New York, USA, 752 p., 2003.
- 16  
17 762    [67] P.W. Atkins, and J. de Paula, *Physical Chemistry for the Life Sciences*. 1st ed.  
18  
19 763            Oxford University Press: Oxford, UK. 2006.
- 20  
21 764    [68] L. Torång, N. Nyholm, H.-J. Albrechtsen, *Shifts in biodegradation kinetics of the*  
22 765            *herbicides MCPP and 2,4-D at low concentrations in aerobic aquifer materials*,  
23 766            Environ. Sci. Technol 37 (2003), pp. 3095-3103.
- 24  
25 767    [69] Z. Yuan, *The role of oxygenation and intermediates in biodegradation of chelating*  
26 768            *agents*, Ph.D. diss., Carnegie Mellon University, US, 2004.
- 27  
28 769    [70] J.D. Linton and R.J. Stephenson, *A preliminary study on growth yields in relation*  
29 770            *to the carbon and energy content of various organic growth substrates*, FEMS  
30 771            Microbiol. Lett. 3 (1978), pp. 95–98  
31  
32  
33  
34 772



773 **Tables:**

774 Table 1. Comparison of equations used to estimate the bacterial yield on a given substrate  
 775 serving as both electron donor and carbon source in g cell carbon (g substrate carbon)<sup>-1</sup>. The  
 776 nitrogen source is assumed to be NH<sub>4</sub><sup>+</sup>. Carbon is assumed to be incorporated into biomass or  
 777 evolved as CO<sub>2</sub>.

Method	Equation
Minimum Turnover to Biomass (MTB)	$Y_{\frac{C}{C}} = \left( \frac{\frac{n_{bio}}{n} \frac{\Delta G_r^{0'}}{\Delta G_{ATP}} \times Y_{ATP}}{\frac{M_c}{\sigma_c} \times n_c + \frac{n_{bio}}{n} \frac{\Delta G_r^{0'}}{\Delta G_{ATP}} \times Y_{ATP}} \right)$
Thermodynamic Electron Equivalent Model 2 (TEEM2)	$Y_{\frac{C}{C}} = \left( \frac{\gamma_s}{\gamma_c} \right) \left( \frac{\Delta G_d^{0'} - \Delta G_d^{0'} - \frac{q}{p} \Delta G_{xy}^{0'}}{\Delta G_d^{0'} - \Delta G_d^{0'} - \frac{q}{p} \Delta G_{xy}^{0'} - \frac{\frac{\Delta G_{fa}^{0'} - \Delta G_d^{0'}}{\epsilon^m} + \frac{\Delta G_{in}^{0'} - \Delta G_{fa}^{0'}}{\epsilon^n} + \frac{\frac{\Delta G_{ATP}}{Y_{ATP} \times 0.9} \times \frac{M_c}{\gamma_c \sigma_c}}{\epsilon}} \right)$
Expanded thermodynamic true yield model (ETTYM)	$Y_{\frac{C}{C}} = f_{cell} = \frac{-K (\gamma_s \Delta G_{e-O_2} - \Delta G_d)}{\frac{\Delta G_{acetate} - \Delta G_d}{K^m} + \frac{\Delta G_{ATP} \times M_{cell}}{Y_{ATP} \times 0.9} - K (\gamma_c \Delta G_{e-O_2} - \Delta G_d)}$

778 The parameters are: MTB:  $n_{bio}$ : bio-available electrons;  $n$ : electrons transferred in the redox reaction;  $\Delta G_r^{0'}$ : Gibbs free energy of  
 779 the redox reaction;  $\Delta G_{ATP}$ : Gibbs free energy of hydrolysis with ~40% efficiency taken into account [80 kJ mol<sup>-1</sup>];  $Y_{ATP}$ : bacterial  
 780 yield on ATP [g cell dw (mol ATP)<sup>-1</sup>], which is assumed to be dependent on the chemical structure;  $M_c$ : molar mass of carbon  
 781 [12.01 g mol<sup>-1</sup>];  $\sigma_c$ : fraction of carbon in dry cell [g C (g cell dw)<sup>-1</sup>];  $n_c$ : number of carbon atoms in the substrate.  
 782 TEEM2:  $\gamma_s$ : degree of reductance of the substrate;  $\gamma_c$ : degree of reductance of the cell;  $\Delta G_a^{0'}$ : Gibbs free energy of reduction of the  
 783 electron acceptor [kJ eqq<sup>-1</sup>];  $\Delta G_d^{0'}$ : Gibbs free energy of reduction of the electron donor [kJ eqq<sup>-1</sup>];  $q$ : number of oxygenase  
 784 reactions [oxygenase reactions mol<sup>-1</sup>];  $p$ : number of electron equivalents per mole substrate [eeq mol<sup>-1</sup>];  $\Delta G_{xy}$ : reduction potential of  
 785 NADH/NAD<sup>+</sup> oxidation [= -219.2 kJ mol<sup>-1</sup>];  $\Delta G_{fa}^{0'}$ : Gibbs free energy of reduction of formaldehyde [= 46.53 kJ eqq<sup>-1</sup>];  $\Delta G_{in}^{0'}$ :  
 786 Gibbs free energy of reduction of acetyl-CoA [= 30.9 kJ eqq<sup>-1</sup>];  $\epsilon$ : energy capture efficiency [=0.37];  $\Delta G_{ATP}$ : hydrolysis of ATP at  
 787 standard biological conditions [= 30.53 kJ mol<sup>-1</sup>];  $Y_{ATP}$ : the bacterial yield on ATP [=10.5 g cell dw (mol ATP)<sup>-1</sup>];  $\sigma_c$ : fraction of  
 788 carbon in the cell;  $M_c$ : molar mass of carbon [12.01 g mol<sup>-1</sup>];  $m$ : +1 if  $\Delta G_{fa}^{0'} > 0$ , else =  $n$ ;  $n$ : +1 if  $\Delta G_{in}^{0'} - \Delta G_d^{0'} > 0$ , else = -1.

1  
2  
3 789 ETTYM:  $K$ : efficiency parameter [=0.41];  $\gamma_s$ : degree of reductance of the substrate;  $\Delta G_{e-O_2}$ : Gibbs free energy of reduction of  
4 790 oxygen [ $\text{kJ eq}^{-1}$ ];  $\gamma_c$ : degree of reductance of the cell;  $\Delta G_d$ : Gibbs free energy of the carbon source [ $\text{kJ (mol C)}^{-1}$ ];  $\Delta G_{acetate}$ : Gibbs  
5 791 free energy of acetate reduction (=  $106.3 \text{ kJ (mol C)}^{-1}$ );  $\Delta G_{ATP}$ : hydrolysis of ATP at standard biological conditions [=  $30.53 \text{ kJ mol}^{-1}$ ];  
6 792  $Y_{ATP}$ : bacterial yield on ATP [=  $10.5 \text{ g cell dw (mol ATP)}^{-1}$ ];  $M_{cell}$ : cell mass per mol carbon (=  $26.4 \text{ g (mol C)}^{-1}$ ) with cell formula  
7 793  $\text{C}_5\text{H}_{10}\text{O}_3\text{N}$ ;  $m$ :  $m$  is +1 if  $(\Delta G_{acetate} - \Delta G_{CS}) > 0$  else  $m$  is = -1.  
8  
9 794  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

For Peer Review Only

795 Table 2. Balanced half-reactions as reductions of the three different terminal electron acceptors  
 796 and pesticides and xenobiotics and their associated Gibbs free energy of the half-reaction  
 797 ( $\Delta G_r^{\circ}$ ) in  $\text{kJ mol}^{-1}$  and  $\text{kJ (electron equivalent (eeq))}^{-1}$  at standard state conditions, except for  
 798  $\text{H}^+$  ( $=10^{-7} \text{ M}$ ).

	Half-reaction	$\Delta G_r^{\circ}$	
		[ $\text{kJ mol}^{-1}$ ]	[ $\text{kJ eeq}^{-1}$ ]
<b>Terminal electron acceptor</b>			
Oxygen, $\text{O}_2$	$\text{O}_2 + 4\text{H}^+ + 4\text{e}^- \rightleftharpoons 2 \text{H}_2\text{O}$	-314.88	-78.72
Nitrate, $\text{NO}_3^-$	$\text{NO}_3^- + 6\text{H}^+ + 5\text{e}^- \rightleftharpoons \frac{1}{2}\text{N}_2 + 3 \text{H}_2\text{O}$	-358.8	-71.76
Sulphate, $\text{SO}_4^{2-}$	$\text{SO}_4^{2-} + 9.5\text{H}^+ + 8\text{e}^-$ $\rightleftharpoons \frac{1}{2}\text{HS}^- + \frac{1}{2}\text{H}_2\text{S}$ $+ 4 \text{H}_2\text{O}$	170.16	21.27
<b>Electron donor, Pesticides and xenobiotics</b>			
2,4-D	$8 \text{CO}_2 + 2 \text{HCl} + 30 \text{H}^+ + 30 \text{e}^-$ $\rightleftharpoons \text{C}_8\text{H}_6\text{Cl}_2\text{O}_3$ $+ 13 \text{H}_2\text{O}$	1286	42.9
2,4-DB	$10 \text{CO}_2 + 2\text{HCl} + 42 \text{H}^+ + 42 \text{e}^-$ $\rightleftharpoons \text{C}_{10}\text{H}_{10}\text{Cl}_2\text{O}_3$ $+ 17 \text{H}_2\text{O}$	1778	42.3
Acetamiprid	$10 \text{CO}_2 + \text{HCl} + 4\text{NH}_3 + 38\text{H}^+ + 38\text{e}^-$ $\rightleftharpoons \text{C}_{10}\text{H}_{11}\text{ClN}_4$ $+ 20\text{H}_2\text{O}$	1695	44.6
Acetochlor	$14 \text{CO}_2 + \text{HCl} + \text{NH}_3 + 68\text{H}^+ + 68\text{e}^-$ $\rightleftharpoons \text{C}_{14}\text{H}_{20}\text{ClNO}_2$ $+ 26\text{H}_2\text{O}$	2091	30.8
Alachlor	$14 \text{CO}_2 + 68 \text{H}^+ + 68 \text{e}^- + \text{HCl} +$	2890	42.5

	$\text{NH}_3 \rightleftharpoons \text{C}_{14}\text{H}_{20}\text{ClNO}_2 + 26 \text{H}_2\text{O}$		
Anthracene	$14 \text{CO}_2 + 66\text{H}^+ + 66\text{e}^-$ $\rightleftharpoons \text{C}_{14}\text{H}_{10} + 28\text{H}_2\text{O}$	2204	33.4
Atrazine	$8 \text{CO}_2 + \text{HCl} + 5\text{NH}_3 + 30\text{H}^+ + 30\text{e}^-$ $\rightleftharpoons \text{C}_8\text{H}_{14}\text{ClN}_5$ $+ 16\text{H}_2\text{O}$	1629	54.3
Azoxystrobin	$22 \text{CO}_2 + 3\text{NH}_3 + 86\text{H}^+ + 86\text{e}^-$ $\rightleftharpoons \text{C}_{22}\text{H}_{17}\text{N}_3\text{O}_5$ $+ 39\text{H}_2\text{O}$	3406	39.6
Benalyxyl	$20 \text{CO}_2 + 94 \text{H}^+ + 94 \text{e}^- + \text{NH}_3$ $\rightleftharpoons \text{C}_{20}\text{H}_{23}\text{NO}_3$ $+ 37 \text{H}_2\text{O}$	3652	38.9
Benzene	$6 \text{CO}_2 + 30 \text{H}^+ + 30 \text{e}^-$ $\rightleftharpoons \text{C}_6\text{H}_6 + 12 \text{H}_2\text{O}$	848	28.3
Benzoate	$7 \text{CO}_2 + 31 \text{H}^+ + 32 \text{e}^-$ $\rightleftharpoons \text{C}_7\text{H}_7\text{O}_2^- + 12 \text{H}_2\text{O}$	1043	32.6
Bifenazate	$17 \text{CO}_2 + 76 \text{H}^+ + 76 \text{e}^- + 2 \text{NH}_3$ $\rightleftharpoons \text{C}_{17}\text{H}_{20}\text{N}_2\text{O}_3$ $+ 31 \text{H}_2\text{O}$	3395	44.7
Carbofuran	$12 \text{CO}_2 + \text{NH}_3 + 54\text{H}^+ + 54\text{e}^-$ $\rightleftharpoons \text{C}_{12}\text{H}_{15}\text{NO}_3$ $+ 21\text{H}_2\text{O}$	1676	31.0
Chlorothalonil	$8 \text{CO}_2 + 22 \text{H}^+ + 22 \text{e}^- + 4 \text{HCl}$ $+ 2 \text{NH}_3$ $\rightleftharpoons \text{C}_8\text{Cl}_4\text{N}_2 + 16 \text{H}_2\text{O}$	977	44.4

Chlorpropham	$10 \text{ CO}_2 + 44 \text{ H}^+ + 44 \text{ e}^- + \text{HCl} + \text{NH}_3$ $\rightleftharpoons \text{C}_{10}\text{H}_{12}\text{ClNO}_2$ $+ 18 \text{ H}_2\text{O}$	2223	50.5
Cypermethrin	$22 \text{ CO}_2 + 96 \text{ H}^+ + 96 \text{ e}^- + 2 \text{ HCl}$ $+ \text{NH}_3$ $\rightleftharpoons \text{C}_{22}\text{H}_{19}\text{Cl}_2\text{NO}_3$ $+ 41 \text{ H}_2\text{O}$	3762	39.2
Daminozide	$6 \text{ CO}_2 + 2 \text{ NH}_3 + 24 \text{ H}^+ + 24 \text{ e}^-$ $\rightleftharpoons \text{C}_6\text{H}_{12}\text{N}_2\text{O}_3$ $+ 9 \text{ H}_2\text{O}$	1251	52.1
DDT	$14 \text{ CO}_2 + 60 \text{ H}^+ + 60 \text{ e}^- + 5 \text{ HCl}$ $\rightleftharpoons \text{C}_{14}\text{H}_9\text{Cl}_5 + 28 \text{ H}_2\text{O}$	2508	41.8
Desmedipham	$16 \text{ CO}_2 + 2 \text{ NH}_3 + 66 \text{ H}^+ + 66 \text{ e}^-$ $\rightleftharpoons \text{C}_{16}\text{H}_{16}\text{N}_2\text{O}_4$ $+ 28 \text{ H}_2\text{O}$	3342	50.6
Dicamba	$8 \text{ CO}_2 + 30 \text{ H}^+ + 30 \text{ e}^- + 2 \text{ HCl} \rightleftharpoons$ $\text{C}_8\text{H}_6\text{Cl}_2\text{O}_3 + 13 \text{ H}_2\text{O}$	1278	42.6
Ethylenediaminetetraacetate (EDTA)	$10 \text{ CO}_2 + 2 \text{ NH}_3 + 34 \text{ H}^+ + 34 \text{ e}^-$ $\rightleftharpoons \text{C}_{10}\text{H}_{16}\text{N}_2\text{O}_8$ $+ 12 \text{ H}_2\text{O}$	1294	38.1
Famoxadone	$22 \text{ CO}_2 + 2 \text{ NH}_3 + 92 \text{ H}^+ + 92 \text{ e}^-$ $\rightleftharpoons \text{C}_{22}\text{H}_{18}\text{N}_2\text{O}_4$ $+ 40 \text{ H}_2\text{O}$	3840	41.7
Glyphosate	$3 \text{ CO}_2 + 15 \text{ H}^+ + 12 \text{ e}^- + \text{PO}_4^{3-} + \text{NH}_3$ $\rightleftharpoons \text{C}_3\text{H}_8\text{NO}_5\text{P} + 5 \text{ H}_2\text{O}$	756	63.0
Ibuprofen	$13 \text{ CO}_2 + 66 \text{ H}^+ + 66 \text{ e}^- +$ $\rightleftharpoons \text{C}_{13}\text{H}_{18}\text{O}_2$ $+ 24 \text{ H}_2\text{O}$	2566	38.9

Iprodione	$13 \text{ CO}_2 + 2\text{HCl} + 3\text{NH}_3 + 48\text{H}^+$ $+ 48\text{e}^-$ $\rightleftharpoons \text{C}_{13}\text{H}_{13}\text{Cl}_2\text{N}_3\text{O}_3$ $+ 23\text{H}_2\text{O}$	2358	49.1
MCPA	$9 \text{ CO}_2 + 38 \text{ H}^+ + 38 \text{ e}^- + \text{HCl}$ $\rightleftharpoons \text{C}_9\text{H}_9\text{ClO}_3$ $+ 15 \text{ H}_2\text{O}$	1530	40.3
MCPB	$11 \text{ CO}_2 + \text{HCl} + 50\text{H}^+ + 50\text{e}^-$ $\rightleftharpoons \text{C}_{11}\text{H}_{13}\text{ClO}_3$ $+ 19\text{H}_2\text{O}$	1964	39.3
Mecoprop (MCP)	$10 \text{ CO}_2 + 44 \text{ H}^+ + 44 \text{ e}^- + \text{HCl}$ $\rightleftharpoons \text{C}_{10}\text{H}_{11}\text{ClO}_3$ $+ 17 \text{ H}_2\text{O}$	1779	40.4
Metalaxyl-M	$15 \text{ CO}_2 + \text{NH}_3 + 70\text{H}^+ + 70\text{e}^-$ $\rightleftharpoons \text{C}_{15}\text{H}_{21}\text{NO}_4$ $+ 26\text{H}_2\text{O}$	2997	42.8
Metamitron	$10 \text{ CO}_2 + 36 \text{ H}^+ + 36 \text{ e}^- + 4 \text{ NH}_3$ $\rightleftharpoons \text{C}_{10}\text{H}_{10}\text{N}_4\text{O}$ $+ 19 \text{ H}_2\text{O}$	1391	38.6
Milbemectin	$31 \text{ CO}_2 + 154\text{H}^+ + 154\text{e}^-$ $\rightleftharpoons \text{C}_{31}\text{H}_{44}\text{O}_7 + 55\text{H}_2\text{O}$	6402	41.6
Naphthalene	$10 \text{ CO}_2 + 48\text{H}^+ + 48\text{e}^-$ $\rightleftharpoons \text{C}_{10}\text{H}_8 + 20\text{H}_2\text{O}$	1638	34.1
Nitrilotriacetate (NTA)	$6\text{CO}_2 + \text{NH}_3 + 18\text{H}^+ + 18\text{e}^-$ $\rightleftharpoons \text{C}_6\text{H}_9\text{NO}_6 + 6\text{H}_2\text{O}$	731	40.6
Paraquat	$12 \text{ CO}_2 + 2\text{NH}_3 + 56\text{H}^+ + 56\text{e}^-$ $\rightleftharpoons \text{C}_{12}\text{H}_{14}\text{N}_2 + 24\text{H}_2\text{O}$	2229	39.8

Pendimethalin	$13 \text{ CO}_2 + 54 \text{ H}^+ + 54 \text{ e}^- + 3 \text{ NH}_3$ $\rightleftharpoons \text{C}_{13}\text{H}_{19}\text{N}_3\text{O}_4$ $+ 22 \text{ H}_2\text{O}$	3082	57.1
Phenanthrene	$14 \text{ CO}_2 + 66 \text{ H}^+ + 66 \text{ e}^-$ $\rightleftharpoons \text{C}_{14}\text{H}_{10} + 28 \text{ H}_2\text{O}$	2204	33.4
Phenmedipham	$16 \text{ CO}_2 + 66 \text{ H}^+ + 66 \text{ e}^- + 2 \text{ NH}_3$ $\rightleftharpoons \text{C}_{16}\text{H}_{16}\text{N}_2\text{O}_4$ $+ 28 \text{ H}_2\text{O}$	3336	50.5
Propyzamid	$12 \text{ CO}_2 + 52 \text{ H}^+ + 52 \text{ e}^- + 2 \text{ HCl}$ $+ \text{NH}_3$ $\rightleftharpoons \text{C}_{12}\text{H}_{11}\text{Cl}_2\text{NO}$ $+ 23 \text{ H}_2\text{O}$	2036	39.2
Pymetrozine	$11 \text{ CO}_2 + 5 \text{ NH}_3 + 38 \text{ H}^+ + 38 \text{ e}^-$ $\rightleftharpoons \text{C}_{11}\text{H}_{11}\text{N}_5\text{O}$ $+ 21 \text{ H}_2\text{O}$	1880	49.5
Pyrene	$16 \text{ CO}_2 + 74 \text{ H}^+ + 74 \text{ e}^-$ $\rightleftharpoons \text{C}_{16}\text{H}_{10} + 32 \text{ H}_2\text{O}$	1968	26.6

799

800

801 Table 3. Gibbs free energy of formation in  $\text{kJ mol}^{-1}$  ( $\Delta G_f^\circ$ ), number of carbon-hydrogen bonds,  
 802  $Y_{ATP}$  in g dry weight  $(\text{mol ATP})^{-1}$ , number of oxygenase reactions ( $t_{oxy}$ ), chemical structure, and  
 803 degree of reductance ( $\gamma_s$ ) of the pesticides and xenobiotics used in the comparison.

Name	Structure	$\Delta G_f^\circ$	C-H bonds	$Y_{ATP}$	$t_{oxy}$	$\gamma_s$	Reference
2,4-D	$\text{C}_8\text{H}_6\text{Cl}_2\text{O}_3$	-241.5	5	5	0	3.75	[61]
2,4-DB	$\text{C}_{10}\text{H}_{10}\text{Cl}_2\text{O}_3$	-67.8	9	5	0	4.20	[61]
Acetamiprid	$\text{C}_{10}\text{H}_{11}\text{ClN}_4$	745.6	11	5	0	3.80	[61]
Acetochlor	$\text{C}_{14}\text{H}_{20}\text{ClNO}_2$	-128.1	20	5	0	4.86	[61]
Alachlor	$\text{C}_{14}\text{H}_{20}\text{ClNO}_2$	670.8	20	5	0	4.86	[61]
Anthracene	$\text{C}_{14}\text{H}_{10}$	695.8	10	5	1	4.71	[61]
Atrazine	$\text{C}_8\text{H}_{14}\text{ClN}_5$	811.3	12	5	0	3.75	[61]
Azoxystrobin	$\text{C}_{22}\text{H}_{17}\text{N}_3\text{O}_5$	478.1	17	5	0	3.91	[61]
Benalaxyl	$\text{C}_{20}\text{H}_{23}\text{NO}_3$	771.4	23	5	0	4.70	[61]
Benzene	$\text{C}_6\text{H}_6$	133.9	6	5	2	5.00	[66] cited in [23]
Benzoate	$\text{C}_7\text{H}_7\text{O}_2^-$	-105.4	6	5	2	4.57	[62]
Bifenazate	$\text{C}_{17}\text{H}_{20}\text{N}_2\text{O}_3$	965.2	18	5	0	4.47	[61]
Carbofuran	$\text{C}_{12}\text{H}_{15}\text{NO}_3$	-251.6	14	5	2	4.50	[61]
Chlorothalonil	$\text{C}_8\text{Cl}_4\text{N}_2$	163.8	0	5	0	2.75	[61]
Chlorpropham	$\text{C}_{10}\text{H}_{12}\text{ClNO}_2$	639.1	11	5	0	4.40	[61]
Cypermethrin	$\text{C}_{22}\text{H}_{19}\text{Cl}_2\text{NO}_3$	700	19	5	0	4.36	[61]
Daminozide	$\text{C}_6\text{H}_{12}\text{N}_2\text{O}_3$	11.2	11	5	0	4.00	[61]
DDT	$\text{C}_{14}\text{H}_9\text{Cl}_5$	583	9	5	0	4.29	[61]
Desmedipham	$\text{C}_{16}\text{H}_{16}\text{N}_2\text{O}_4$	993	14	5	0	4.13	[61]
Dicamba	$\text{C}_8\text{H}_6\text{Cl}_2\text{O}_3$	-249.8	5	5	0	3.75	[61]
Ethylenediaminetetraacetate (EDTA)	$\text{C}_{10}\text{H}_{16}\text{N}_2\text{O}_8$	-1209.2	8	5	4	3.40	[65]
Famoxadone	$\text{C}_{22}\text{H}_{18}\text{N}_2\text{O}_4$	935.9	17	5	0	4.18	[61]
Glyphosate	$\text{C}_3\text{H}_8\text{NO}_5\text{P}$	-883.5	4	5	0	4.00	[61]
Iprodione	$\text{C}_{13}\text{H}_{13}\text{Cl}_2\text{N}_3\text{O}_3$	434	12	5	0	5.08	[61]



Ibuprofen	C <sub>13</sub> H <sub>18</sub> O <sub>2</sub>	504	17	5	0	3.69	[61]
MCPA	C <sub>9</sub> H <sub>9</sub> ClO <sub>3</sub>	-105.2	8	5	0	4.22	[61]
MCPB	C <sub>11</sub> H <sub>13</sub> ClO <sub>3</sub>	11.2	12	5	0	4.55	[61]
Mecoprop (MCP)	C <sub>10</sub> H <sub>11</sub> ClO <sub>3</sub>	-15.7	10	5	0	4.40	[61]
Metalaxyl-M	C <sub>15</sub> H <sub>21</sub> NO <sub>4</sub>	434.8	21	5	0	4.67	[61]
Metamitron	C <sub>10</sub> H <sub>10</sub> N <sub>4</sub> O	414.8	8	5	0	3.60	[62]
Milbemectin	C <sub>31</sub> H <sub>44</sub> O <sub>7</sub>	1090	42	5	0	4.97	[61]
Naphthalene	C <sub>10</sub> H <sub>8</sub>	527.1	10	5	1	4.80	[61]
Nitilotriacetate (NTA)	C <sub>6</sub> H <sub>9</sub> NO <sub>6</sub>	-954.8	6	5	1	3.00	[65]
Paraquat	C <sub>12</sub> H <sub>14</sub> N <sub>2</sub>	906.4	14	5	0	4.67	[61]
Pendimethalin	C <sub>13</sub> H <sub>19</sub> N <sub>3</sub> O <sub>4</sub>	944.3	18	5	0	4.15	[61]
Phenanthrene	C <sub>14</sub> H <sub>10</sub>	695.8	10	5	1	4.71	[61]
Phenmedipham	C <sub>16</sub> H <sub>16</sub> N <sub>2</sub> O <sub>4</sub>	986.9	14	5	0	4.13	[61]
Propyzamide	C <sub>12</sub> H <sub>11</sub> Cl <sub>2</sub> NO	399.9	10	5	0	4.33	[61]
Pymetrozine	C <sub>11</sub> H <sub>11</sub> N <sub>5</sub> O	878	10	5	0	3.45	[61]
Pyrene	C <sub>16</sub> H <sub>10</sub>	301.3	10	5	1	4.63	[62]
Ammonia	NH <sub>3</sub>	-26.6					[52]
Carbon dioxide	CO <sub>2</sub>	-394.4				0	[52]
Nitrate	NO <sub>3</sub> <sup>-</sup>	-108.7					[67]
Hydrogen ion (proton) (pH 7)	H <sup>+</sup>	-39.9					[52]
Water	H <sub>2</sub> O	-237.2					[52]
Oxygen	O <sub>2</sub>	0					[67]
Hydrogen sulphide anion	HS <sup>-</sup>	12.1					[52]
Hydrogen sulphide	H <sub>2</sub> S	-27.8					[67]
Sulphate	SO <sub>4</sub> <sup>2-</sup>	-744.4					[52]
Nitrogen	N <sub>2</sub>	0					[67]
Hydrogen chloride	HCl	-131.2					[67]

804

805 Table 4. Bacterial growth yields on organic chemicals of environmental concern under aerobic  
 806 conditions in g cell carbon (g substrate carbon)<sup>-1</sup> predicted using MTB, TEEM2, and MTB.  
 807 Predictions are evaluated using available experimental data. The error and mean absolute error  
 808 are shown. The observed experimental bacterial growth yields ( $Y^{OBS}$ ) with reference are shown.  
 809 The cell formulation for MTB and TEEM2 was taken to be C<sub>3</sub>H<sub>7</sub>O<sub>2</sub>N with a degree of  
 810 reductance of 4.0. For ETTYM it was taken to be C<sub>5</sub>H<sub>10</sub>O<sub>3</sub>N with a degree of reductance of 4.2.  
 811 The predictions were made under standard state conditions (pH = 7). The entries are sorted from  
 812 low to high predicted yield of MTB. For TEEM2 a weak positive correlation and for ETTYM a  
 813 strong positive correlation exists between the degree of reductance of the compound and  
 814 predicted yield ( $R^2$  is shown).

Compound	$Y^{OBS}$	$Y^{pred}$ MTB	Error	$Y^{pred}$ TEEM2	Error	$Y^{pred}$ ETTYM	Error	Reference
Unit	g cell carbon (g substrate carbon) <sup>-1</sup>	g cell carbon (g substrate carbon) <sup>-1</sup>	[%]	g cell carbon (g substrate carbon) <sup>-1</sup>	[%]	g cell carbon (g substrate carbon) <sup>-1</sup>	[%]	
Chlorothalonil		0.00		0.35		0.36		
Pyrene	0.21- 0.31	0.27	-13 to 27	0.44	45 to 111	0.52	145	[51]
2,4-D	0.31	0.30	-4.7	0.47	51	0.48	56	[47]
Dicamba		0.30		0.47		0.48		
DDT		0.30		0.53		0.55		
Anthracene	0.11- 0.13	0.31	128 to 183	0.54	182 to 341	0.56	316 to 416	[49]
Phenanthrene	0.32	0.31	-4.1	0.54	67.3	0.56	75	[23]

Azoxystrobin		0.34		0.48		0.49		
Famoxadone		0.34		0.52		0.53		
Ethylenediaminetetraacetate (EDTA)	0.27	0.34	26	0.36	31.6	0.42	55	[69] cited in [24]
Metamitron	0.30	0.34	14	0.43	45	0.45	49	[46]
Benzoate	0.42	0.35	-18	0.49	16	0.54	28	[70] cited in [24]
Propyzamide		0.35		0.52		0.54		
Cypermethrin		0.36		0.53		0.55		
MCPA		0.37		0.52		0.53		
Benzene	0.43	0.37	-13	0.47	10	0.57	33	[23]
2,4-DB		0.38		0.52		0.54		
Naphthalene	0.47	0.38	-18	0.55	16	0.57	22	[23]
Desmedipham		0.39		0.55		0.57		
Phenmedipham		0.39		0.55		0.56		
Pymetrozine		0.39		0.45		0.47		
Iprodione		0.40		0.49		0.50		
Mecoprop (MCP)	0.3	0.40	32	0.54	80	0.56	85	[68]
Nitrioltriacetate (NTA)	0.27	0.40	46	0.35	28	0.38	39	[69] cited in [24]
Carbofuran	0.42	0.41	-1.4	0.49	16	0.52	25	[29]
MCPB		0.42		0.55		0.57		
Bifenazate		0.42		0.57		0.58		
Benalaxyl		0.43		0.57		0.58		
Acetamiprid		0.43		0.48		0.50		
Paraquat		0.43		0.57		0.59		
Chlorpropham		0.44		0.58		0.60		
Ibuprofen	0.30	0.46	52	0.54	79	0.56	85	[39]
Acetochlor		0.46		0.55		0.56		

Milbemectin		0.47		0.62		0.63		
Metalaxyl-M		0.48		0.58		0.60		
Alachlor		0.49		0.61		0.62		
Pendimethalin		0.51		0.58		0.60		
Glyphosate	0.18	0.51	183	0.58	224	0.60	234	[42]
Atrazine		0.52		0.51		0.53		
Daminozide		0.57		0.54		0.55		
Linear correlation to degree of reductance ( $R^2$ )		0.14		0.53		0.70		
Mean absolute error (MAE)			49%		82%		111%	

815

816

817 Table 5. Comparison of bacterial growth yields of organic chemicals of environmental concern  
 818 under anaerobic conditions in g cell carbon (g substrate carbon)<sup>-1</sup>. Predictions were made using  
 819 MTB, TEEM2, and ETTYM. The cell formulation for MTB and TEEM2 was taken to be C<sub>5</sub>H<sub>7</sub>O<sub>2</sub>N  
 820 with a degree of reductance of 4.0. For ETTYM it was taken to be C<sub>5</sub>H<sub>10</sub>O<sub>3</sub>N with a degree of  
 821 reductance of 4.2. The predictions were made under standard biochemical conditions (pH = 7).  
 822 The entries are sorted from low to high predicted yield of MTB under nitrate-reducing  
 823 conditions. For TEEM2 a weak positive correlation and for ETTYM and strong positive  
 824 correlation exists between the degree of reductance of the compound and predicted yield (*R*<sup>2</sup> is  
 825 shown). Values highlighted in bold are negative and thus not meaningful.

Compound	MTB	TEEM2	ETTYM	MTB	TEEM2	ETTYM
Unit	g cell carbon (g substrate carbon) <sup>-1</sup>					
Electron acceptor	NO <sub>3</sub> <sup>-</sup>			SO <sub>4</sub> <sup>2-</sup>		
Chlorothalonil	0.00	0.34	0.35	0.00	0.11	0.12
Pyrene	0.25	0.42	0.50	0.02	0.02	0.05
2,4-D	0.28	0.46	0.47	0.07	0.14	0.15
Dicamba	0.28	0.45	0.47	0.07	0.14	0.15
DDT	0.29	0.51	0.53	0.07	0.15	0.17
Anthracene	0.29	0.52	0.54	0.05	0.07	0.11
Phenanthrene	0.29	0.52	0.54	0.05	0.07	0.11
Azoxystrobin	0.32	0.46	0.47	0.07	0.12	0.14
Famoxadone	0.33	0.50	0.52	0.08	0.15	0.16
Ethylenediaminetetraacetate (EDTA)	0.33	0.34	0.41	0.07	<b>-0.06</b>	0.11
Metamitron	0.33	0.42	0.43	0.07	0.11	0.12
Benzoate	0.33	0.47	0.52	0.05	<b>-0.02</b>	0.10
Propyzamide	0.34	0.51	0.52	0.08	0.13	0.15
Cypermethrin	0.35	0.51	0.53	0.08	0.14	0.15
MCPA	0.35	0.50	0.52	0.09	0.14	0.15
Benzene	0.36	0.45	0.55	0.04	<b>-0.07</b>	0.07

2,4-DB	0.36	0.51	0.52	0.09	0.16	0.17
Naphthalene	0.37	0.53	0.56	0.07	0.07	0.12
Desmedipham	0.37	0.54	0.55	0.12	0.21	0.23
Phenmedipham	0.37	0.53	0.55	0.12	0.21	0.23
Pymetrozine	0.38	0.44	0.46	0.12	0.17	0.19
Iprodione	0.38	0.47	0.49	0.12	0.18	0.20
Mecoprop (MCP)	0.38	0.52	0.54	0.10	0.15	0.16
Nitrotriacetate (NTA)	0.38	0.34	0.37	0.10	0.04	0.11
Carbofuran	0.40	0.47	0.51	0.06	0.01	0.08
MCPB	0.40	0.53	0.55	0.10	0.14	0.16
Bifenazate	0.41	0.55	0.57	0.12	0.18	0.20
Benalaxyl	0.41	0.55	0.57	0.10	0.14	0.16
Acetamiprid	0.41	0.47	0.48	0.12	0.15	0.17
Paraquat	0.42	0.55	0.57	0.11	0.15	0.16
Chlorpropham	0.43	0.57	0.59	0.15	0.22	0.24
Ibuprofen	0.44	0.52	0.59	0.11	0.13	0.14
Acetochlor	0.45	0.52	0.54	0.07	0.08	0.09
Milbemectin	0.46	0.60	0.62	0.13	0.18	0.19
Metalaxyl-M	0.47	0.57	0.58	0.14	0.18	0.19
Alachlor	0.47	0.59	0.61	0.14	0.18	0.20
Pendimethalin	0.50	0.57	0.58	0.22	0.26	0.28
Glyphosate	0.50	0.57	0.59	0.23	0.29	0.32
Atrazine	0.51	0.50	0.52	0.21	0.22	0.24
Daminozide	0.56	0.53	0.54	0.24	0.22	0.23
Linear correlation to degree of reductance ( $R^2$ )	0.14	0.49	0.67	0.00	0.01	0.01

826

827

828 Table 6. Bacterial growth yields on simple carbon substrates under aerobic conditions predicted  
 829 using MTB, TEEM2 and ETTYM in g cell carbon (g substrate carbon)<sup>-1</sup>. Where experimental  
 830 data have been available, the predictions are evaluated with respect to these. The error and mean  
 831 absolute error are shown. The experimental bacterial growth yields ( $Y^{EXP}$ ) were taken from  
 832 Xiao and VanBriesen [24]. The predictions were done under standard biochemical conditions  
 833 (pH = 7). For ETTYM the cell formulation is CH<sub>10</sub>O<sub>3</sub>N with a degree of reductance of 4.2. For  
 834 TEEM2 and MTB it is C<sub>5</sub>H<sub>7</sub>O<sub>2</sub>N with a degree of reductance of 4.0. The entries are sorted  
 835 based on the highest absolute error calculated for MTB.

Compound	$Y^{EXP}$	MTB	E(%)	TEEM2	E(%)	ETTY M	E(%)
Unit	g cell carbon (g substrate carbon) <sup>-1</sup>						
Formate	0.16	0.40	155%	0.22	40%	0.23	45%
Oxalate	0.07	0.00	-100%	0.14	95%	0.11	59%
Glycine	0.25	0.38	55%	0.35	41%	0.36	44%
Phenylacetic acid	0.48	0.35	-27%	0.48	0%	0.52	11%
Fumaric acid	0.32	0.24	-25%	0.36	12%	0.37	15%
Formaldehyde	0.47	0.58	23%	0.51	9%	0.53	13%
Malonate	0.24	0.29	23%	0.31	30%	0.29	23%
Citrate	0.37	0.29	-22%	0.35	-7%	0.35	-7%
Glyoxylate	0.22	0.27	21%	0.27	23%	0.28	26%
a-D-Fructose	0.51	0.61	20%	0.49	-3%	0.51	0%
Gluconate	0.53	0.62	18%	0.57	7%	0.58	10%
Sorbitol	0.55	0.47	-16%	0.53	-4%	0.54	-2%
a-D-Glucose	0.53	0.61	15%	0.49	-6%	0.51	-4%
Mannitol	0.55	0.47	-15%	0.53	-4%	0.54	-1%
a-Lactose	0.51	0.44	-14%	0.50	-2%	0.51	1%
Acetate	0.42	0.47	12%	0.41	-3%	0.45	8%
Lactate	0.49	0.45	-9%	0.46	-7%	0.47	-5%
a-D-Galactose	0.56	0.61	9%	0.49	-12%	0.50	-9%
Phenol	0.36	0.33	-8%	0.44	23%	0.53	48%

Malate	0.34	0.32	-7%	0.35	2%	0.36	5%
Succinate	0.39	0.37	-4%	0.38	-3%	0.40	1%
Propionate	0.48	0.50	3%	0.47	-1%	0.53	10%
Pyruvate	0.38	0.39	3%	0.39	4%	0.40	7%
Xylose	0.49	0.48	-2%	0.58	17%	0.59	21%
Tartrate	0.28	0.27	-2%	0.35	26%	0.36	30%
Glycerol	0.62	0.62	0%	0.56	-9%	0.58	-6%
Mean absolute error (MAE)			23%		15%		16%

836

837

838

839

840

841

842

843

844



845 Table 7. Sensitivity analysis of all three methods. The average sensitivity,  $\bar{S}_i$ , and standard  
 846 deviation shown in brackets. The degree of reductance does not affect the predictions of MTB.  
 847 The standard cell formula for ETTYM is  $C_5H_{10}O_3N$  ( $\gamma_s = 4.2$ ,  $\sigma_c = 0.45$  gC (g cell dw)<sup>-1</sup>) and for  
 848 MTB and TEEM2 it is  $C_5H_7O_2N$  ( $\gamma_s = 4$ ,  $\sigma_c = 0.53$  g cell carbon (g cell dw)<sup>-1</sup>). The alternative  
 849 cell formulae used were:  $C_5H_{8.33}O_{0.8}N$  ( $\gamma_s = 4.74$ ,  $\sigma_c = 0.63$  g cell carbon (g cell dw)<sup>-1</sup>),  
 850  $C_{4.1}H_{6.8}O_{2.2}N$  ( $\gamma_s = 3.85$ ,  $\sigma_c = 0.47$  g cell carbon (g cell dw)<sup>-1</sup>).

Parameter	Method	Relative change in parameter value	Average sensitivity, $\bar{S}_i$ (standard deviation)
$Y_{ATP}$	MTB	+20%	0.56 (0.073)
	TEEM2		0.51 (0.016)
	ETTYM		0.49 (0.006)
	MTB	-20%	0.66 (0.068)
	TEEM2		0.61 (0.016)
	ETTYM		0.59 (0.0046)
$\Delta G_r^{\circ'}$	MTB	+50%	0.010 (0.090)
	TEEM2		-0.004 (0.25)
	ETTYM		0.006 (0.21)
	MTB	-50%	0.015 (0.080)
	TEEM2		0.035 (0.17)
	ETTYM		0.024 (0.15)
$C_5H_{10}O_3N$			
Carbon content	MTB	-14%	0.59 (0.12)
	TEEM2	-14%	0.71 (0.02)
Degree of reductance	TEEM2	5%	-2.05 (0.05)
$C_5H_7O_2N$			
Carbon content	ETTYM	17%	0.61 (0.03)
Degree of	ETTYM	-5%	-2.14 (0.12)

reductance			
$C_5H_{8.33}O_{0.8}N$			
Carbon content	MTB	19%	0.53 (0.12)
	TEEM2	19%	0.81 (0.40)
	ETTYM	39%	0.30 (0.01)
Degree of reductance	TEEM2	19%	0.81 (0.40)
	ETTYM	13%	0.89 (0.04)
$C_{4.1}H_{6.8}O_{2.2}N$			
Carbon content	MTB	-12%	0.60 (0.12)
	TEEM2	-12%	-0.67 (0.57)
	ETTYM	3%	-1.77 (0.15)
Degree of reductance	TEEM2	-4%	-2.17 (1.86)
	ETTYM	-8%	0.62 (0.05)

851

852

853

854

855 Table 8. Predicted yields, measured CO<sub>2</sub>, non-extractable residues (NER), amino-acids, and  
 856 predicted formation of soil organic matter (SOM) (biogenic NER) from Eq. (4) ( $SOM_{biogenic\ NER}$ )  
 857 (in brackets: result of Eq. (3), short-term experiments only). The data on measured CO<sub>2</sub> and  
 858 NER formation were taken from Barriuso *et al.* [36] except where indicated otherwise. The data  
 859 points are the maximum values reported for the experiments with the longest duration, if more  
 860 than one experiment was reported. The entries are sorted from high to low predicted biogenic  
 861 NER formation.

Compound	$Y_{MTB}^{EST}$	Measured CO <sub>2</sub>	Measured NER	Measured carbon label in amino acids	Predicted $SOM_{biogenic}$ $_{NER}$ using Eq. (4) ( $X_{biomass\ NER}$ using Eq. (3))
Unit	mol C <sub>bacteria</sub> (mol C <sub>substrate</sub> ) <sup>-1</sup>	% of applied labelled compound			
Glyphosate	0.51	80.1	8.8		28
Daminozide	0.57	59	25		24
Glyphosate <sup>1</sup>	0.51	50	30.4	11 – 12	17 (52)
MCPB	0.43	58	30		15
MCPA	0.38	67	30		15
Ibuprofen <sup>2</sup>	0.46	45	29.6	28	13 (38)
Mecoprop (MCP)	0.41	52	47		13
Desmedipham	0.39	46.4	55		11
Milbemectin	0.49	35	40		11
Metalaxyl-M	0.5	33	73		11
Cypermethrin	0.36	48	26		11

Propyzamid	0.36	48	27		11
2,4-D <sup>3</sup>	0.3	58	36	23	10 (24)
Metamitron <sup>4</sup>	0.35	49	41	15	10 (25)
2,4-DB	0.38	42.1	33.2		10
Benalaxyl	0.43	25	18.8		7
Famoxadone	0.34	32.2	51.4		7
Bifenazate	0.43	23	67.3		6
2,4-D	0.3	36	27.9		6
Carbofuran <sup>5</sup>	0.41	23.8	63		6
Phenmedipham	0.4	16.5	64.1		4
Pymetrozine	0.4	15	61		4
Azoxystrobin	0.34	14	24		3
Acetamiprid	0.43	9.6	32.3		3
Iprodione	0.41	5	40		1
Pendimethalin	0.5	2.4	10		1
Chlorothalonil	0	15	54		0

862 <sup>1</sup>[42] <sup>2</sup>[39] <sup>3</sup>[38] <sup>4</sup>[43] <sup>5</sup>[64]

863

1  
2  
3  
4 *Authors' response for*  
5  
6  
7

8  
9 Microbial growth yield estimates from thermodynamics and its importance for  
10 degradation of pesticides and formation of biogenic non-extractable residues.  
11

12 (Manuscript ID SQER-2017-0051)  
13  
14

15  
16 Andreas Libonati Brock<sup>a,\*</sup>, Matthias Kästner<sup>b</sup>, Stefan Trapp<sup>a</sup>,  
17

18  
19 *<sup>a</sup>Department of Environmental Engineering, Technical University of Denmark, Kgs. Lyngby,*  
20  
21 *Denmark*  
22

23  
24 *<sup>b</sup>UFZ—Helmholtz Centre for Environmental Research, Department of Environmental*  
25  
26 *Biotechnology, Leipzig, Germany*  
27

28 \*[alib@env.dtu.dk](mailto:alib@env.dtu.dk), Phone: +45 4525 1408  
29  
30  
31  
32  
33

34 We greatly appreciate the feedback we received from the anonymous reviewers. In this document, we  
35 have addressed all the comments made (blue font colour) and we have listed the revisions made to the  
36 manuscript in quotation marks. The revisions made to sentences are marked in blue font colour. In the  
37 manuscript, the revisions are included in the same blue font colour.  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 Referee: 1  
5

6 Comments to the Author

7 A thermodynamically based approach is presented for modelling the degradation of chemicals and the  
8 formation of non extractable residues. The manuscript extends an approach that is submitted to another  
9 journal and is still under review.

10 Although the authors complete their conclusions section on a positive note, it is difficult to see an added  
11 value of the approaches presented in this manuscript. Basically, the main conclusion is that other factors  
12 than thermodynamics are more important in modulating degradation and bound residue formation. It is  
13 especially surprising in this respect to see that at the end, setting a default value to the Gibbs free energy  
14 of formation is concluded to be a suited approach as the final results are independent of this parameter. In  
15 addition to this key issue, there is the problem of the underlying conceptual manuscript still being under  
16 review. Thus, there is currently no real basis for the key model used in this contribution.  
17

18  
19 We thank the reviewer for the comment. We do agree with the reviewer that the conceptual manuscript  
20 should be accepted (and published online) before the publication of the present manuscript. Adding to this,  
21 we think that the reviewer has misunderstood the conclusion regarding the Gibbs free energy. Our finding  
22 was not that other factors than thermodynamics are important in modulating degradation and bound  
23 residue formation (although they surely are). The finding is that the Gibbs free energy of formation of the  
24 *chemical of interest* (i.e. pesticide) can be set to 0 kJ mol<sup>-1</sup> if no reliable data is available without affecting  
25 the outcome of the MTB method more than a few percent. This is because the majority of energy in the  
26 reaction comes from the formation of CO<sub>2</sub> and H<sub>2</sub>O – especially if the pesticide contains many C and H  
27 atoms.  
28

29  
30 To make this clearer we changed the sentence in the conclusions accordingly. It now reads (lines 541-543):  
31

32 “All methods showed low sensitivity to variations in the Gibbs energy of formation of the organic chemicals  
33 because most of the Gibbs energy of reaction stems from the formation of the oxidation products carbon  
34 dioxide and water.”  
35

36 Referee: 2  
37

38 Comments to the Author

39 Authors: Andreas Libonati Brock,\* Matthias Kästner, Stefan Trapp

40 Article: Microbial growth yield estimates from thermodynamics and its importance for degradation of  
41 pesticides and formation of biogenic non-extractable residues

42 Journal: SAR and QSAR in Environmental Research

43 Manuscript: SQER-2017-0051  
44

45  
46 This is a valuable contribution on the evaluation and/or prediction of non-extractable residues formation  
47 during (bio)degradation of pesticides and I would like to see it published. However, this manuscript has  
48 several major problems, some of which is of a formal nature, and needs major revision before it will be  
49 suitable for publication. The detailed description of each critical point is given in the following section,  
50 complemented with the suggestion for possible improvement.  
51

52  
53 This is an outstanding review, helpful, detailed, and careful. We are very grateful to this reviewer for the  
54 efforts made to improve our manuscript. Thank you! The contribution has been duly noted in the  
55 acknowledgements. The critical points given in the following section are all addressed.  
56  
57  
58  
59  
60

1  
2  
3  
4 Major problems:  
5

6 1. This manuscript cannot and should not be published before reference 16, manuscript submitted by the  
7 same authors to Environmental Science and Technology (ES&T) journal, is accepted for publication and  
8 published on-line. Namely, manuscript submitted to ES&T and its content is of critical importance for the  
9 main objective of this study, comparison of MTB method with the existing yield estimation methods.  
10 Reference 16 is cited 15 times in the text of this manuscript which clearly demonstrates its critical  
11 relevance for this manuscript.  
12

13 We do agree with the reviewer that the conceptual manuscript should be accepted (and published online)  
14 before the publication of the present manuscript. Reference 16 (in the revised manuscript it is reference  
15 15) was submitted several months before this manuscript, but we had been given more time for revision.  
16 We will resubmit ref 16 earlier so that it is published soon (but of course we cannot guarantee this :-))  
17

18 2. Materials and Methods section should be significantly reduced since to a large extent it duplicates  
19 materials and results that are already published in the open literature.  
20  
21

22 Lines 118-123 now reads:

23 "The bacterial growth yield prediction methods chosen for this study have a common basic approach: a  
24 stoichiometrically balanced redox reaction and the associated change of Gibbs free energy. This means that  
25 one can set up half-reactions describing the reduction of the targeted compound (be it xenobiotic or not),  
26 calculate the associated Gibbs free energy [34], and combine it with half-reactions of an appropriate  
27 electron acceptor."  
28

29 Line 125 now reads:

30 "This approach has been shown for ETTYM and TEEM2 and here"  
31  
32

33 Line 128-131 has been changed and now reads:

34 "A detailed summary of the methods can be found in the Supplementary Information (SI) and in the original  
35 references. The element, energy and electron balances differ between the methods, thus a brief outline of  
36 the methods will be given in Table 1, in which the final equations used to calculate the growth yield are  
37 shown."  
38

39 The section has been significantly reduced (for details on each sub-section in the section see below). Table  
40 S1 showing the final equations used for the growth yield predictions has been moved from the  
41 Supplementary Information to the manuscript. The table is now listed as Table 1. Consequently, the  
42 numbering of tables has been updated throughout the manuscript. Additionally, the amount of equations  
43 has been reduced and their numbering has also been updated throughout the text.  
44  
45

46 Specifically:

47 (i) sub-section "Microbial Turnover to Biomass (MTB)" covering 3.5 pages should be either deleted or  
48 reduced to a minimum (half page). All this is already described in reference 16 and only points relevant for  
49 the main objective of this study may be briefly presented in this sub-section.  
50

51 The sub-section has been significantly reduced and now contains the following (lines 132-148):  
52

53 "The Microbial Turnover to Biomass (MTB) method is presented in detail in Trapp *et al.* [15]. The method is  
54 based on the work of Diekert [18]. The maximum bacterial yield is determined from the nutritional value of  
55 the substrate (N) combined with the determination of bio-available electrons from the reaction. The  
56 nutritional value is the inverse of the yield and describes how much substrate is needed for the growth of  
57 bacteria [g substrate (g biomass)<sup>-1</sup>]. This is subdivided into a biomass yielding (anabolic) and energy yielding  
58  
59  
60

(catabolic) part. The catabolic yield is determined from calculation of the Gibbs free energy released from the oxidation of the compound, the storage of this energy in ATP, and the bacterial growth yield on ATP.

Microbes cannot use all electrons to generate energy and thus the concept of bio-available electrons was introduced. [15]. Thus, energy and electron balances are implicitly considered. The anabolic yield is calculated from the carbon content in the compound (the carbon source) and in the bacterial cell [18], i.e. how many grams of cell can be produced from the carbon in the compound (only carbon availability is assumed to limit growth).

Further details and examples can be found in SI 1.1 and Trapp *et al.* [15]. “

The paragraph

“Gibbs energy of reaction is calculated using activities of the reactants and products assumed to be 1 M, except for H<sup>+</sup> which is assumed to be 10<sup>-7</sup> M (pH 7). The Gibbs free energy of reaction for non-standard conditions can be calculated as

$$\Delta G_r' = \Delta G_r^{o'} + R T \ln Q = \Delta G_r^{o'} + R T \ln \left( \frac{\prod_{i=1}^n [\text{product}]_i^p}{\prod_{i=1}^n [\text{reactant}]_i^r} \right) \quad (1)$$

where  $R$  is the ideal gas constant [8.314 J (K mol)<sup>-1</sup>],  $T$  is the absolute temperature [K],  $Q$  is the reaction quotient, and  $[\text{product}]$  and  $[\text{reactant}]$  are the activities of products and reactants, and  $p$  and  $r$  are their respective stoichiometric coefficients. From the equation itself it can be seen that when  $Q < 1$  the term is negative and when  $Q > 1$  the term is positive.”

has been moved to lines 182-191.

The bulk of the text has been moved to the supplementary information thus making it possible for the interested reader to easily locate a relevant summary of the method.

(ii) sub-section “Thermodynamic Electron Equivalent Model 2 (TEEM2)” covering 3 pages should be also either deleted or reduced to a minimum (half page). Content of this sub-section is covered in details in the original studies (references 18, 23, 24, 35) and there is no need for repetition here. Again, only a brief outline of points relevant for the main objective of this study may be given in this sub-section.

The sub-section has been significantly reduced and now contains the following (lines 151-162):

“In 1965 P. L. McCarty presented a thermodynamic model to estimate the maximal bacterial yield from a single substrate [17]. The method determines the yield on a given substrate from the Gibbs free energy released in the redox process. Since its inception it has been modified and expanded [34]. It was recently modified to better capture the observed lower yields associated with C1 compounds (i.e. methanol) and reactions involving oxygenases [22, 23]. It is based on electron and energy balances. The electron balance considers that the electrons provided by the substrate are used either in the synthesis of cell material (anabolism) or in energy generation (catabolism), and the energy balance states that the energy *captured* with a specific efficiency ( $\epsilon$ ) by the organism is used for bacterial growth. The energy capture efficiency,  $\epsilon$ , is a key parameter and is estimated from experimental data. Further details and examples are found in SI 1.2 and McCarty [23] and Rittmann and McCarty [34]. “



1  
2  
3  
4 The bulk of the text has been moved to the supplementary information thus making it possible for the  
5 interested reader to easily locate a relevant summary of the method.  
6  
7

8 (iii) sub-section "Expanded Thermodynamic True Yield Prediction Model (ETTYM)" covering 2.5 pages is also  
9 a waste of valuable journal space and should be reduced to a minimum. Its content is already published in  
10 references 23 and 25. Outline briefly points relevant for the main objective of this study.  
11

12 The sub-section has been significantly reduced and now contains the following (lines 165-173):  
13

14 "The Expanded Thermodynamic True Yield Model is based on the work by McCarty and was presented in  
15 [22] and expanded in [24]. To increase the accuracy for the yield prediction on C1 compounds and  
16 substrates with low degrees of reduction, the authors proposed to include a carbon and a nitrogen balance  
17 and as a result thereof reformulate the electron and energy balance originally proposed by McCarty. The  
18 carbon balance describes that the carbon is either invested in cell synthesis or into other carbonaceous  
19 products. The nitrogen balance can be ignored if nitrogen is not limiting [24], hence, the yield can be  
20 calculated from an energy balance, carbon balance, and electron balance.  
21 For further details, the reader is kindly referred to SI1.3 and [22, 24]."  
22  
23

24 The bulk of the text has been moved to the supplementary information thus making it possible for the  
25 interested reader to easily locate a relevant summary of the method.  
26  
27

28  
29 3. It seems that the authors are not familiar with the mathematical concept of logarithms. Logarithms are  
30 only defined for positive numbers, those larger than zero. Thus, there are no logarithms for negative  
31 numbers. Consequently, the statement "when  $Q < 0$  the term is negative" on page 8, line 172 is meaningless.  
32

33 Thank you. This is an obvious oversight - what is meant is that when  $Q < 1 \Rightarrow \ln(Q) < 0$ . This has been  
34 corrected. Lines 190-191 now reads:  
35

36 "when  $Q < 1$  the term is negative and when  $Q > 1$  the term is positive."  
37

38  
39 4. page 8, lines 169 and 170: There is no such unit as degree Kelvin ( $^{\circ}\text{K}$ ). The units of absolute temperatures  
40 are Kelvin (K).  
41

42 Thank you for the comment. This has been corrected on line 187.  
43

44 5. Another major drawback of this manuscript are numerous linguistic problems. A lot of sentences in this  
45 manuscript have awkward structure and are in contrast to the elementary principles of English language.  
46 This made the reading of this manuscript a demanding and time-consuming task. This manuscript should  
47 have been checked and corrected by someone proficient in English before it was submitted for a review. In  
48 a separate section "Some Linguistic Problems or Errors" I have listed some of those linguistic problems and  
49 have suggested possible corrections. However, in addition to those suggested corrections, a thorough  
50 inspection of the revised manuscript by someone proficient in English is still needed.  
51

52 We sincerely thank the reviewer for taking time to meticulously go through the manuscript. The manuscript  
53 has undergone a thorough inspection with the aid from a proficient native speaker from the USA. The  
54 changes made are listed in the following.  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 Lines 19-22: "Formation of microbial mass can be estimated from the microbial growth yield but  
5 experimental data is rare. Instead, we suggest using prediction methods for the theoretical yield based on  
6 thermodynamics."

7  
8 Lines 23-24: "We have estimated the growth yield on 40 organic chemicals"

9  
10 Lines 28-29: "Having the theoretical yield and using the released CO<sub>2</sub> as a measure for microbial activity,"

11  
12 Lines 40-41: "The evaluation of biodegradation of organic chemicals of environmental concern is a big  
13 challenge for risk assessment and is subject to legislation and regulation."

14  
15 Lines 43-45: "The assessment of biodegradability under environmental conditions is standardised by OECD  
16 testing guidelines, such as OECD Tests Nos. 306-309 used for"

17  
18 Lines 47-48: "Transformation and biodegradation is mostly tested with <sup>14</sup>C or <sup>13</sup>C labelled parent  
19 compounds."

20  
21 Lines 48-49: "Isotopes are particularly needed for assessment of non-extractable residues (NER; also called  
22 "bound residues") and tracing of unknown metabolites [8]."

23  
24 Line 51: "Although there are several approaches for the reliable prospective assessment"

25  
26 Lines 50-51: "Assessment of biodegradation is well established but may still have some pitfalls for various  
27 compounds [9-12]."

28  
29 Lines 54-55: "The assessment of residue formation is still in its infancy and is not yet predictable."

30  
31 Line 62-64: "Apparently, NER are mostly comprised of all types of residues and thus the assessment of the  
32 biogenic NER formation will also provide information about the amounts of the other types of NER  
33 formed."

34  
35 Lines 65-66: "Compounds that are poor growth substrates and do not provide sufficient energy"

36  
37 Lines 67-69: "Thus, the usability of the molecule, its energy content and suitability for anabolic processes  
38 has a profound impact on the evolutionary pressure to develop degradation pathways."

39  
40 Lines 71-73: "In addition, thermodynamics can also be used to describe the potential growth of bacteria  
41 [17]."

42  
43 Lines 73-74: "Essentially, bacterial growth is simplified and split into anabolic processes (energy  
44 demanding) and catabolic processes (energy producing) [18]."

45  
46 Line 74-77: "The catabolic processes describe the energy released from the oxidation of a chemical or a  
47 substrate. In aerobic metabolism the oxidation product is usually CO<sub>2</sub> and H<sub>2</sub>O. The electrons derived from  
48 the oxidation are partly transferred (...)"

49  
50 Lines 79-80: ", predominately from the formation of CO<sub>2</sub> and H<sub>2</sub>O [19]."

51  
52 Lines 81-82: "The anabolic processes describe the substrate and energy use for the synthesis of new cell  
53 biomass."

1  
2  
3  
4 Lines 89-91: "These estimates have previously been used for biotechnological purposes and for the  
5 estimation of activated sludge formation in waste water treatment processes,"  
6

7 Lines 91-9: "The different growth yield estimation methods are based on a similar set of considerations [26,  
8 27]."  
9

10 Lines 106-108: "Therefore, yield estimation provides a tool for the assessment and prediction of biogenic  
11 NER formation in the degradation assessment of chemicals for regulatory purposes."  
12

13 Lines 179-180: "Ammonia was taken as the sole nitrogen source so electrons for the assimilatory reduction  
14 of  $\text{NO}_3^-$  was not considered,"  
15

16 Lines 193-194: "nitrate-reducing conditions ( $\text{NO}_3^-$  as the terminal electron acceptor) and sulphate-reducing  
17 conditions ( $\text{SO}_4^{2-}$  as the terminal electron acceptor)"  
18

19 Line 202: ". However, for ETTYM this was not done"  
20

21 Lines 210-211: "key input parameters were varied including  $Y_{ATP}$ ,"  
22

23 Lines 226-227: "Bacterial growth yields have been experimentally assessed only for very few of the selected  
24 compounds."  
25

26 Lines 229-231: "Moreover, ibuprofen and some polycyclic aromatic hydrocarbons (PAH) were also included  
27 as the bacterial growth yield has been experimentally determined for these."  
28

29 Lines 232-233: "the methods were also evaluated using the growth yield determined for simple carbon  
30 substrates used in biotechnology."  
31

32 Lines 233-234: "The compounds selected for comparison are based on the review and evaluation made by  
33 Xiao and VanBriesen [24]. "  
34

35 Line 268: "In Table 4, the predicted bacterial growth yields under aerobic conditions are shown and"  
36

37 Lines 274-276: ". The reason being the absence of carbon-hydrogen bonds. All other compounds except  
38 pyrene were predicted to have a bacterial growth yield of  $>0.3 \text{ g cell carbon (g substrate carbon)}^{-1}$  by all  
39 methods."  
40

41 Lines 283-285: ". A weaker but still highly significant ( $p < 0.01$ ) correlation was found for TEEM2 ( $Y = 0.09 \gamma_s$   
42  $+ 0.15$ ,  $R^2 = 0.53$ ), and a significant ( $p < 0.05$ ) but rather weak correlation was found for MTB ( $R^2 = 0.14$ )."  
43

44 Lines 288-289: "and ETTYM overestimated by 56% and 25%, respectively. For glyphosate and anthracene,"  
45

46 Line 299: "With sulphate as the terminal electron acceptor,"  
47

48 Lines 300-301: "compared to the yields found under aerobic conditions (Table 5)"  
49

50 Lines 301-302: "For both aerobic and nitrate-reducing conditions, the ranking of the estimated yields was  
51 close."  
52

53 Lines 305-307: "MTB predicted 20 chemicals to have a yield of  $<0.1 \text{ g cell carbon (g substrate carbon)}^{-1}$ ,  
54 while ten were predicted using TEEM2, and only four using ETTYM."  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 Lines 309-310: "The lack of experimental observations under a multitude of redox conditions makes  
5 assessment of the prediction accuracy unachievable."

6  
7 Lines 318-319: "Subsequently, the work presented in [24] resulted in an even lower MAE when both the pH  
8 and related speciation, and oxygenase reactions were taken into account."

9  
10 Lines (formerly) 320-322 were deleted: "(the ranking was identical from the lowest predicted yield and 12  
11 places up; and the six substrates predicted to give the highest yield)"

12  
13 Lines 323-325: "For MTB, these were formate, oxalate, and glycine; for TEEM2, these were oxalate, glycine,  
14 and formate; and for ETTYM, these were oxalate, phenol, and formate."

15  
16 Lines 327-328: "with TEEM2 still giving the lowest MAE (9%)"

17  
18 Line 342: "All methods are relatively sensitive to changes in  $Y_{ATP}$ , especially MTB."

19  
20 Lines 344-446: "The chosen default value for xenobiotics of 5 g cell dw (mol ATP)<sup>-1</sup> ([18], for methanol) used  
21 in the MTB method does not lead to large errors (cf. Table 4)."

22  
23 Lines 350-351: ". Conversely, a negative  $\Delta G_f^{\circ}$  which is made more negative leads to a decrease in the  
24 predicted yield (e.g., EDTA)."

25  
26 Line 354: "Yield estimates for chemicals with few carbon atoms"

27  
28 Line 355: "are especially sensitive to changes"

29  
30  
31 Lines 366-369: ". The effect of the cell formula in TEEM2 and ETTYM is not only on the energy costs related  
32 to synthesis [24], but also on the conversion to g cell carbon (g substrate carbon)<sup>-1</sup> in TEEM2. This is due to  
33 the degree of reduction of the cell ( $\gamma_s$ ) used in converting the units."

34  
35 Line 376: "NER formation from the CO<sub>2</sub> produced during degradation experiments"

36  
37 Lines 378-380: "Except for one compound (glyphosate, caused by the production of the metabolite  
38 aminomethylphosphonic acid (AMPA)), the predicted biogenic NER was smaller than the measured total  
39 NER."

40  
41 Lines 383-384: "and pymetrozine the biogenic NER is suggested to make up less than 10% of the formed  
42 NER"

43  
44 Line 402: "with some of them used earlier in [15]."

45  
46 Lines 404-405: "The bacterial growth yield estimation methods are all developed to predict the true yield"

47  
48 Lines 405-407: "The observed value is typically a net yield accounting only for the formation of new cell  
49 mass and removal of the parent compound"

50  
51 Lines 407-409: "The difference between the two is that for the observed yield energy and carbon  
52 expenditure, due to non-growth purposes, are not captured"

53  
54 Lines 409-410: "(e.g. energy spent on maintenance, formation of metabolites or soluble microbial products  
55 and extracellular polymeric substances)"

1  
2  
3  
4 Line 411: “, unless a dynamic model was used for fit. Hence,”  
5

6 Line 413-414: “If hardly degradable or insoluble metabolites are formed and rendered not bioavailable (as  
7 NER I or II),”  
8

9 Lines 415-417: “can also result in a higher amount of energy being spent on maintenance leading to an  
10 observed yield lower than the predicted true yield. “  
11

12 Lines 420-422: “, the intermediate AMPA accumulated, resulting in an observed yield much that was lower  
13 than the predicted yield (Table 4).”  
14

15 Lines 424-425: “The presence of other sources of carbon or energy (mixed substrate use) also adds  
16 uncertainty to the observed value.”  
17

18 Lines 431-432: “Under these conditions, microbes use most of the growth substrate just for maintenance  
19 [51].”  
20

21 Lines 441-442: “This shows that the half-reaction approach using various electron acceptors used in ETTYM  
22 and TEEM2 can also be used with MTB”  
23

24 Line 453: “the TEEM2 method was calibrated to the data in order to produce yield estimates”  
25

26 Line 456: “proposed for the ETTYM method [24]. “  
27

28 Line 458: “estimated it as zero due to the absence of C-H bonds (which points to the need for a”  
29

30 Line 489: “. However, speciation of the substrate also has an effect on its Gibbs energy of formation.”  
31

32 Line 490: “Similarly, the sensitivity of the energy capture efficiency parameter  $\epsilon$ ”  
33

34 Line 496: “The peak in living biomass is usually after a few days to weeks [38, 39],”  
35

36 Lines 508-510: “We suggested such a model in [8] and used it successfully, for the prediction of the NER  
37 formation from 2,4-D and ibuprofen with pre-estimated yield data [15].”  
38

39 Line 511: “The data compiled by [36] give no hints into which form the NER are present”  
40

41 Line 516: “done” deleted: “In degradation experiments with metatriton [43], glyphosate [42]”  
42

43 Lines 518-519: “amino acids, fatty acids, metabolites, and parent compounds.”  
44

45 Lines 525-526: “The method applied in this study provides process-based theoretical background that may  
46 be used to interpret NER data derived in degradation experiments.”  
47

48 Lines 534-535: “ while MTB performed better when estimating the yield on organic chemicals of  
49 environmental concern in general and in particular on pesticides.”  
50

51 Line 808: “Predictions are evaluated using available experimental data.”  
52

53 Line 826: “Values highlighted in bold are negative and thus not meaningful.”  
54

55 6. lines 650-666 and 733-736: This extensive discussion and related conclusions are meaningless and should  
56 be deleted. Gibbs energy of formation can be easily calculated for any chemical by quantum-mechanical  
57

1  
2  
3  
4 methods. Thus, any limitation of group contribution methods is irrelevant and it also does not make sense  
5 to set the Gibbs energy of formation to zero.  
6

7 Please allow that at this one issue we disagree with the reviewer. It may be possible to get exact values of  
8 Gibbs energy of formation, but personally we had some difficulties to find the values of several organic  
9 chemicals and we also found many conflicting data of delta G. We expect other users to have similar  
10 experiences. It is therefore a relevant message that small error or even missing delta G does not inhibit  
11 good estimates of the microbial yield of xenobiotics. Nonetheless, we shortened this section in 4.2 so that it  
12 is less extensive, and we deleted the lines in the conclusions, except one, which now reads (lines 541-543):  
13

14  
15 "All methods showed low sensitivity to variations in the Gibbs energy of formation of the organic chemicals  
16 because most of the Gibbs energy of reaction stems from the formation of the oxidation products carbon  
17 dioxide and water.  
18

19 Section 4.2 *Sensitivity analysis* now reads (lines 462-478):  
20

21 "All the methods were shown to be sensitive to the choice of cell formula but exhibited low sensitivity to  
22 variations of the formation energy,  $\Delta G_f^\circ$ , of the chemical of interest. All methods are based on the Gibbs  
23 energy of reaction and knowledge of the Gibbs energy of formation of the chemical of interest is needed. If  
24 the value has not been determined experimentally (e.g. [52]), it can be estimated using group contribution  
25 methods [53-58] (method [53] is implemented in the freely available ChemProp [59]), or by component  
26 contribution methods [60] (implemented in the free accessible database eQuilibrator [61]), or calculated  
27 using quantum mechanics [62]. For xenobiotics, the applicability of these estimation methods may be  
28 limited. Consequently, we also tested the sensitivity of Gibbs energy of formation of the xenobiotic  
29 substrate by setting this value to 0 kJ mol<sup>-1</sup> (Table S3). The MTB method has surprisingly low sensitivity.  
30 Compounds having a large negative Gibbs energy of formation (e.g. NTA, EDTA, and glyphosate) and few  
31 carbon-hydrogen bonds (6, 8, and 4, respectively) show a maximum deviation of around 20% from the  
32 predictions done with correct Gibbs energy of formation. Overall, the average deviation is only 6%. In  
33 comparison, TEEM2 and ETTYM have considerably higher average deviation (14% and 11%, respectively)."  
34  
35  
36

37 Specific Problems or Errors:  
38

39 1. line 54: References 13 and 15 are barely suitable to support the current status on modeling and  
40 assessment of biodegradation. Reference 13 is just a technical collection of published QSAR models without  
41 any critical evaluation of those models while reference 15 is only describing Biocatalysis/Biodegradation  
42 Database and the improved public access to this database. Thus, those references should be either  
43 replaced or, at least, amended by the recent critical review on the most relevant qualitative and  
44 quantitative models for estimating or evaluating biodegradation of organic chemicals and published in ACS  
45 Symposium Series:  
46  
47

48 A. Sabljic and Y. Nakagawa, Biodegradation and quantitative structure-activity relationship (QSAR), in Non-  
49 First Order Degradation and Time-Dependent Sorption of Organic Chemicals in Soil, W. Chen, A. Sabljic, S.A.  
50 Cryer and R.S. Kookana, eds, Book Series: ACS Symposium Series, American Chemical Society, Washington  
51 (DC), Volume 1174, 2014, pp. 57-84.  
52

53 We thank the reviewer for the comment. After reading the suggested literature, we have included it as a  
54 reference and removed reference 15.  
55  
56  
57  
58  
59  
60

1  
2  
3  
4 2. lines 115-117: Delete this sentence since it only repeats previous statement.  
5

6 The sentence was deleted.  
7

8 3. line 193: It seems that variable  $Y_{ATP}$  has not been defined.  
9

10 Please note, that this sentence has been deleted from the manuscript and moved to the supplementary  
11 information. (Line SI 58)  
12

13 Due to the changes made to the Material and Methods section,  $Y_{ATP}$  is now defined in Table 1.  
14

15 4. line 266: "(McCarty 2007)", wrong format for this reference  
16

17 The reference has been changed to the correct format:  
18

19 "[23]"  
20

21 Please note, that this sentence has been deleted from the manuscript and moved to the supplementary  
22 information.  
23

24 5. line 352: From here the authors have started to number sections and sub-sections. Why? Previous  
25 sections and sub-sections are not numbered.  
26

27 The un-numbered sections and subsections have now been numbered.  
28

29 Line 39: "**1. Introduction**"  
30

31 Line 117: "**2. Materials and Methods**"  
32

33 Line 132: "**2.1 Microbial Turnover to Biomass (MTB)**"  
34

35 Line 150: "**2.2 Thermodynamic Electron Equivalent Model 2 (TEEM2)**"  
36

37 Line 164: "**2.3 Expanded Thermodynamic True Yield Prediction Model (ETTYM)**"  
38

39 Line 177: "**2.4 Conditions for comparison**"  
40

41 Line 222: "**2.5 Chemicals of environmental concern; data**"  
42

43 Line 242: "**2.6 Calculation of biogenic non-extractable residues**"  
44  
45  
46

47 6. lines 363-364: "SO<sub>4</sub><sup>2-</sup> is oxidised to H<sub>2</sub>S and HS<sup>-</sup>, NO<sub>3</sub><sup>-</sup> to N<sub>2</sub>, O<sub>2</sub> to H<sub>2</sub>O,"?! Do you mean reduced? Also  
48 "oxidized" is US spelling used in this manuscript.  
49

50 Thank you for pointing this out. Yes, it is not oxidation but reduction as they are all electron acceptors.  
51

52 "Oxidised" has been changed to "reduced" on line 198: "SO<sub>4</sub><sup>2-</sup> is reduced to H<sub>2</sub>S and HS<sup>-</sup>, NO<sub>3</sub><sup>-</sup> to N<sub>2</sub>, O<sub>2</sub> to  
53 H<sub>2</sub>O, C to CO<sub>2</sub>"  
54  
55  
56  
57  
58  
59  
60

7. lines 409-442: The content of sub-section "2.3 Calculation of biogenic non-extractable residues" is described in details in reference 16 and only points relevant for the main objective of this study should be briefly presented in this sub-section.

Thank you for the comment. The sub-section has been revised and reduced one third. It now reads (Lines 243-264):

"Chemicals labelled with carbon isotopes ( $^{14}\text{C}$  or  $^{13}\text{C}$ ) allowed the flow of carbon to be tracked in the experimental system [38-41]. If the compound provides carbon to anabolism and cell synthesis, the labelled carbon will end up in microbial biomass and finally in the biogenic NER. Biogenic NER is not posing a risk to neither the environment nor human health [8]. When a substrate  $S$  is mineralized, the amount of biomass formed is yield times substrate,  $Y \times S$ , and the evolved  $\text{CO}_2$  is  $(1 - Y) \times S$  [15]. After the growth phase has stopped, the maximum ratio between biomass and  $\text{CO}_2$  and is thus

$$\frac{[X_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{YS}{(1-Y)S} \text{ or } [X_{\text{biogenic NER}}] = \frac{Y}{1-Y} [\text{CO}_2] \quad (3)$$

where  $X_{\text{biogenic NER}}$  is the biomass making up the living biogenic NER. After the cessation of the growth phase, the microorganisms start to decay. The dead microorganisms are turned over in the microbial food chain and form new biomass,  $\text{CO}_2$  and soil organic matter (SOM) [30-33]. Then, the ratio between biogenic NER and  $^{13/14}\text{CO}_2$  becomes [15]

$$\frac{[\text{SOM}_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{f \times Y}{(1-Y) + (1-f) \times Y} \text{ or } [\text{SOM}_{\text{biogenic NER}}] = \frac{f \times Y}{(1-Y) + (1-f) \times Y} [\text{CO}_2] \quad (4)$$

where  $\text{SOM}_{\text{biogenic NER}}$  is the non-living biogenic NER,  $f$  is the fraction of decaying biomass turned over into both living biomass and non-living SOM (0.5, [33]), and  $1-f$  is the fraction of label released as  $\text{CO}_2$ . Eq. (3) can be used to estimate NER formation during short-term experiments, whereas Eq. (4) holds for long-term experiments. It can be seen that a high mineralization and  $\text{CO}_2$  formation together with a high bacterial growth yield leads to a high formation of biogenic NER."

8. lines 569-570: In Table 7 the predicted and measured biogenic NERs are not similar for 2,4-D. Clarify.

This requires some explanation. In the four cases where biogenic NER was reported (Karolina Nowak and her co-workers), the authors took the measured label in amino acids and multiplied with factor 2. This factor stems from the protein content (about 50%) in living microbes. However, as seen from the NER equations (now Eq. (3) and Eq. (4)), the biogenic NER is comprised of living and dead biomass. And for dead biomass, the factor 2 is not valid because sugars and fatty acids are metabolized much faster than amino acids. It can even be seen from the author's data that the application of the factor 2 to the amino acid fraction leads to false results because for ibuprofen and 2,4-D the fraction of biogenic NER is larger than the total NER. We therefore deleted the column "Reported biogenic NER" in Table 8, refer to the measured amino acids and changed the manuscript accordingly. As can be seen, the measured label in amino acid is always within the range of biogenic NER given by Eq. 3 (living biomass) and Eq. 4 (decayed biomass).

Now line 386-394:

"The experimental period for ibuprofen and 2,4-D (64 days) [38, 39], and glyphosate and metamitron (80 days) [42, 43] was shorter than the experiments reported in [36]. Eq. (3), which calculates living biomass  $X$  as biogenic NER, was additionally used to interpret these experiments. In these four studies, the carbon label found in amino acids was reported. For living microbes, about half of the carbon is in proteins. This fraction increases during decay and turnover of microbial biomass because proteins are the most stable fraction of the cells [30, 33, 15]. Except for glyphosate, the measured label in amino acids is within the range of biogenic NER predicted by Eqs. (3) and (4), and the measured total NER is greater."



1  
2  
3  
4 9. line 613: I would suggest to cite here also the recent review on bioavailability of xenobiotica in the soils  
5 environment.  
6

7 A. Katayama, R. Bhula, G.R. Burns, E. Carazo, A. Felsot, D. Hamilton, C. Harris, Y.H. Kim, G. Kleter, W.  
8 Koerdel, J. Linders, J.G.M.W. Peijnenburg, A. Sabljic, R.G. Stephenson, D.K. Racke, B. Rubin, K. Tanaka, J.  
9 Unsworth and R.D. Wauchope, Bioavailability of xenobiotics in the soil environment, Reviews of  
10 Environmental Contamination and Toxicology 203 (2010), pp. 1-86.  
11

12 Thank you for the suggestion. After reading the publication we agree with the reviewer. It is an excellent  
13 review and is now cited as ref 50 on line 430.  
14

15  
16  
17 10. line 631: "any linearity"?! probably "there is no relationship"

18 Thank you for your comment. The sentence now reads (lines 448-450):

19 "For all methods, the predicted growth yield is so small that there is no relationship between degree of  
20 reductance and bacterial growth yield."  
21

22  
23 11. line 649: References 55-59 should be deleted since those are not references for ChemProp estimation  
24 software, only reference 14 is relevant.  
25

26 Thank you for the comment. References 55-59 are references for group contribution based methods. Lines  
27 466-469 now read:

28 "it can be estimated using group contribution methods [53-58] (method [53] is implemented in the freely  
29 available ChemProp [59]) or by component contribution methods [60] (implemented in the free accessible  
30 database eQuilibrator [61]) or calculated using quantum mechanics [62]."  
31

32 12. line 650: Quantum mechanics is used to calculate and not to estimate the Gibbs energy of formation for  
33 chemicals.  
34

35 Thank you for the comment. See above comment where 'calculated' has been added.  
36

37 13. line 684: Reference 16 should be replaced by original reference(s) on "the peak in living biomass".  
38 Besides, reference 16 is not published yet.  
39

40 Done. We cite here as example the experiments done with ibuprofen and 2,4-D (Refs 38, 39) (Line 496).  
41

42 14. line 685: expression "is more likely applying to the data" is confusing and does not make sense  
43  
44

45 Thank you for the comment. The sentence has been revised and now reads (Lines 496-499):

46 "and therefore we expect that Eq. (4) ( $SOM_{biogenic\ NER}$ ) is more appropriate for these experiments than Eq. (3)  
47 ( $X_{biogenic\ NER}$ ) as the majority of the living biomass has decayed and been incorporated into SOM after 100  
48 days."  
49

50  
51 15. lines 685-687: What is implied by "this view"? No view is described in previous sentences of this sub-  
52 section.  
53

54 Lines 499-500 have been revised and now read:

55 "Results obtained by Eq. (4) are smaller than the measured NER, which confirms the results of this  
56 equation."  
57  
58  
59  
60

1  
2  
3  
4 See also comment 14.  
5

6 16. lines 690-693: confusing long sentence  
7

8 In regards to comment 37, the sentence has been changed to (lines 504-507):

9 "Such a correlation should not be expected since the processes leading to NER I, II and III are competing. If  
10 a pesticide is not degraded it can undergo aging and irreversible sorption (type I NER) and covalent binding  
11 of the parent compound or its metabolites (type II NER) [8]."  
12  
13

14 17. lines 700-701: What is the meaning of this sentence? How do you exploit the structure and energy of  
15 molecule by enzymatic pathways?  
16

17 Thank you for the comment. The sentence has been revised and now reads (lines 514-515):

18 "Of course, the enzymatic pathways to facilitate the degradation and energy exploitation of the molecule  
19 also need to be present."  
20  
21

22 18. lines 716-718: This statement is not correct since the MTB method is already presented in reference 16.  
23

24 Thank you for the comment. Changed to "applied". It now reads (lines 525-526):

25 "The method applied in this study provides process-based theoretical background that may be used to  
26 interpret NER data derived in degradation experiments."  
27

28 19. lines 738-740: This conclusion is not correct since the MTB method is already presented in reference 16.  
29

30 Thank your pointing this out. Changed to "applied". It now reads (lines 545-546):

31 "The approach applied in this study provides a theoretical foundation that can be used to predict biogenic  
32 NER formation during pesticide degradation experiments."  
33  
34

35 Some Linguistic Problems or Errors:  
36

37 1. line 28: add comma, i.e. " for microbial activity, "  
38

39 Thank you for pointing it out. The comma is added on line 29.  
40

41 2. lines 56-57: correct sentence as "Recently, a novel approach was suggested for modelling the formation  
42 of biogenic residue [16] which can also shed light into the black box of NER."  
43  
44

45 Thank you for your suggestion. The sentence has been revised and now reads (lines 56-57):

46 "Recently, a novel approach was suggested for modelling the formation of biogenic residues [15] which can  
47 also elucidate the black box of NER."  
48

49 3. lines 59-60: "but may also be formed by covalent bonding of metabolites (type II NER)" - Covalent  
50 bonding of metabolites to what? Why only metabolites, why is covalent bonding not possible for parent  
51 compound?  
52

53 Yes, covalent bonding is possible for both parent compound and metabolites. The sentence was rewritten  
54 and now reads (lines 57-60):

55 "NER may be formed by sequestration or entrapment of parent compounds or metabolites in soils and  
56 sediments (*type I NER*), and also by covalent bonding to soil organic matter (*type II NER*)"  
57  
58  
59  
60

1  
2  
3  
4. line 63: probably “comprise of all types”

5 Thank you for your suggestion. The sentence has been revised (line 62): “Apparently, NER are mostly  
6 comprised of all types of residues and”  
7

8  
9 5. lines 94-97: correct sentence as “Thermodynamic Electron Equivalent Model (TEEM2) developed by  
10 Perry L. McCarty [18, 24] and Expanded Thermodynamic True Yield Prediction Model (ETTYM) by Xiao and  
11 VanBriesen [23, 25] or their variations have been established and applied [29,30] for the estimation of  
12 bacterial growth yield on xenobiotics.”  
13

14 Thank you for your suggestion. The sentence has been revised and the paragraph now reads (lines 93-99):  
15 “The Thermodynamic Electron Equivalent Model (TEEM2) developed by Perry L. McCarty [17, 23] and  
16 Expanded Thermodynamic True Yield Prediction Model (ETTYM) of Xiao and VanBriesen [22, 24] or their  
17 variations have been applied [28, 29] for the estimation of bacterial growth yield on xenobiotics. Both  
18 models have evolved towards an increased need of knowledge regarding the transformation pathways,  
19 metabolic processes and the electron and energy losses associated hereto in order to model specific  
20 growth of various organisms.”  
21

22 6. line 98: expression “These models have both been moving” is meaningless in this sentence, maybe “Both  
23 models have been moving”  
24

25 Thank you for your suggestion. The sentence has been revised (lines 96-97):  
26

27 “Both models have evolved towards an increased need of knowledge regarding the transformation  
28 pathways,”  
29

30 7. lines 104-105: correct text as “potential without the need for information on the pathways as this is  
31 rarely known for the majority of chemicals of environmental concern.”  
32

33 Thank you for your suggestion. The sentence has been revised as suggested (lines 102-104):  
34

35 “In MTB we proposed a simple method to predict just the minimum bacterial growth yield potential  
36 without the need for information on the pathways, as this is rarely known for the majority of chemicals of  
37 environmental concern.”  
38  
39

40 8. line 106: correct as “the microbial growth and decline are coupled”  
41

42 Thank you for your suggestion. The sentence has been revised (line 105):  
43

44 “Furthermore, microbial growth and decline are coupled to the formation of soil organic”  
45  
46

47 9. lines 110-111: correct text as “to thoroughly compare the recently introduced MTB method with other  
48 yield estimation methods”  
49

50 Thank you for your suggestion. The sentence has been revised as suggested (lines 109-110):  
51

52 “The objectives of the present study are i) to thoroughly compare the recently introduced MTB method  
53 with other yield estimation methods;”  
54

55 10. line 128: “straightforward”? probably “direct”  
56  
57  
58  
59  
60

1  
2  
3  
4 The word has been deleted and the sentence now reads (lines 122-124):

5 "and combine it with half-reactions of an appropriate electron acceptor (e.g.  $O_2$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ,  $Fe^{3+}$ ,  $Mn^{4+}$ , or  
6 even  $CO_2$  etc.) allowing for assessment of the bacterial growth yield under a multitude of redox conditions."

7  
8  
9 11. line 129: Set as new sentence, i.e. "of redox conditions. Here we only look at  $O_2$ ,  $NO_3^-$  and  $SO_4^{2-}$ ."

10 Thank you for your suggestion. The sentence has been revised as suggested (lines 124-125):

11 "of redox conditions. Here we only look at  $O_2$ ,  $NO_3^-$ , and  $SO_4^{2-}$ ."

12  
13 12. lines 160-161: correct as "the Gibbs free energy of reaction"

14 Thank you for your suggestion. The sentence has been revised as suggested. Note that the sentence has  
15 been moved from the manuscript to the supplementary information (Line SI 59).

16  
17 13. lines 237 and 240: There is a mismatch between "(catabolism)." and "where". The sentence is ending  
18 (i.e. period) on line 237 and new sentence is starting with the lower case letter on line 240. Analogous  
19 problem on lines 251 and 254.

20 The periods are removed. Note that the sentence has been moved from the manuscript to the  
21 supplementary information.

22  
23 14. lines 246 or 247: Period is missing at the end of this sentence.

24 Thank you for the comment. A period has been added. Note that the sentence has been moved from the  
25 manuscript to the supplementary information.

26  
27 15. line 250: "that all the energy"?! probably "that the whole energy"

28 Thank you for the comment. I have changed the sentence to read (correction in bold):

29 "The energy balance states that the energy *captured* by the organism from the redox reaction is used for  
30 bacterial growth."

31 Note that the sentence has been moved from the manuscript to the supplementary information (Line SI  
32 133).

33  
34 16. lines 310-312: This sentence starts with equation which does not make sense. Also "Where" should be  
35 "where" since this is not the beginning of sentence.

36 The period in line has been removed and "Where" is correctly changed to "where".

37 Note that the sentence has been moved from the manuscript to the supplementary information.

38  
39 17. line 322: "Where" should be "where" since this is not the beginning of sentence.

40 Thank you for your suggestion. The sentence has been revised as suggested.

41 Note that the sentence has been moved from the manuscript to the supplementary information.

42  
43 18. lines 342-344: This sentence also starts with equation which does not make sense or period on line 341  
44 must be deleted.

45 Thank you for pointing this out. The period has been deleted.

46 Note that the sentence has been moved from the manuscript to the supplementary information.

47  
48 19. lines 347-349: This sentence does not make sense at all. If electrons are only diverted to the reduction  
49 of the terminal electron acceptor (e.g.  $O_2$ ) and the synthesis of new cell material, and the electron donor is  
50 also the carbon source, the equation reduces to (where  $\gamma_{CO_2} = 0$ )

51 Thank you for bringing this to our attention. The sentence has been changed to:

1  
2  
3  
4 "If the electron donor is also the carbon source and the electrons are used only for reduction of the  
5 terminal electron acceptor and synthesis of new cell material, the equation reduces to (remember,  $\gamma_{CO_2} =$   
6 0)"  
7

8 I hope it is easier to make sense of it now. Note that the sentence has been moved from the manuscript to  
9 the supplementary information (Lines SI 226-228).  
10

11 20. line 363: "oxidized" is US spelling used in this manuscript

12 This has been corrected to "reduced" (line 198) – also see comment 6 under Specific Problems or Errors:  
13

14 21. line 544: correct "MTB at least."

15 Since MTB is the method least sensitive to changes in the cell formulation, the sentence has been changed  
16 to (lines 365-366):  
17

18 "While all methods are sensitive to the cell formula used (Table 7), MTB is the method least affected."  
19

20 22. line 587: delete "experimentally" since it is redundant

21 Thank you for your suggestion. Yes, it is clearly a pleonasm and it has been deleted (line 405):  
22

23 "The observed value is typically a net yield accounting only for the formation of new cell"  
24

25 23. line 594: correct as "prediction methods assume a complete degradation of compound."

26 Thank you for your suggestion. The sentence has been revised on line 412:  
27

28 "Additionally, the prediction methods assume a complete degradation of the compound."  
29

30 24. lines 610-611: correct as "other carbon sources, and here the MTB predicted yields are very close to the  
31 experimentally determined values."  
32

33 Thank you for your suggestion. The sentence has been revised as suggested (lines 426-428):  
34

35 "great care was taken in the experimental setup to minimise confounding factors due to other carbon  
36 sources, and here the MTB predicted yields are very close to the experimentally determined values."  
37

38 25. line 612: correct text as "The observed differences might also be attributed to their high hydrophobicity  
39 and"  
40

41 Thank you for your suggestion. The sentence has been revised as suggested (lines 429-430):  
42

43 "The observed differences might also be attributed to their high hydrophobicity and limited bioavailability"  
44

45 26. line 616: correct text as "reactions for PAHs in TEEM2, its errors were higher than for MTB."  
46

47 Thank you for your suggestion. The sentence has been revised as suggested (lines 433-434):  
48

49 "(putative) oxygenase reactions for PAHs in TEEM2, its errors were higher than for MTB."  
50

51 27. lines 618-619: correct text as "yield on carbofuran taking oxygenase reactions into account as suggested  
52 in [29]."  
53

54 Thank you for your suggestion. The sentence has been revised ( lines 435-436):  
55

56 "measured bacterial yield on carbofuran by taking oxygenase reactions into account as suggested in [28]."  
57

58 28. lines 620-621: correct text as "Under sulphate-reducing conditions the predicted bacterial yields were  
59 much lower than the predicted yields under aerobic conditions, which can be expected"  
60

Thank you for your suggestion. The sentence has been revised as suggested (lines 437-439):  
61

62 "Under sulphate-reducing conditions the predicted bacterial yields were much lower than the predicted  
63 yields under aerobic conditions, which can be expected considering the lower"  
64

65 29. line 623: correct "An interesting observation"

66 Thank you for your suggestion. The sentence has been revised as suggested (line 440):  
67  
68  
69  
70

1  
2  
3  
4 "An interesting observation was that the decrease in yield"  
5

6 30. lines 626-627: correct text as "the majority of redox reactions might not be sufficient to fuel bacterial  
7 growth."

8 Thank you for your suggestion. The sentence has been revised as suggested (lines 444-445):

9 "the energy released from the majority of redox reactions might not be sufficient to fuel bacterial growth"  
10

11 31. line 639: "Where" should be "While"

12 Thank you for your suggestion. The sentence has been revised as suggested (line 457):

13 "While ETTYM and TEEM2 both overestimated the yield for oxalate,"  
14

15 32. line 646: correct text as "reaction, knowledge of the Gibbs energy of formation of chemical is needed."  
16

17 Thank you for your suggestion. The sentence has been revised see comment 6 under Major Problems (lines  
18 464-464):

19 "All methods are based on the Gibbs energy of reaction and knowledge of the Gibbs energy of formation of  
20 the chemical of interest is needed"  
21

22 33. lines 670-671: correct text as "either due to the low energy associated with the oxidation of substrate  
23 (e.g. formate or formaldehyde) or the low energy associated with the reduction of electron acceptor"  
24

25 Thank you for your suggestion. The sentence has been revised (lines 481-484):

26 "either due to the low energy associated with the oxidation of substrate (e.g. formate or formaldehyde), or  
27 the low energy associated with the reduction of the electron acceptor (e.g.  $\text{SO}_4^{2-}$ ). "  
28

29 34. lines 685-687: modify this sentence as follows "Results obtained by Eq. (22) are smaller than the  
30 measured NER, which confirms this view."  
31

32 Thank you for your suggestion. The sentence has been revised (lines 499-500). See also comment 15 under  
33 Specific Problems or Errors:

34 "Results obtained by Eq. (4) are smaller than the measured NER, which confirms the results of this  
35 equation."  
36

37 35. lines 687-688: modify this sentence as follows "For daminozide, chemical with the highest predicted  
38 yield, calculated biogenic NER and measured NER are almost equal."  
39

40 Thank you for your suggestion. The sentence has been revised (lines 500-502):  
41

42 "For daminozide, the chemical with the highest predicted yield, the calculated biogenic NER and measured  
43 NER are almost equal. "  
44

45 36. line 689: modify text as "The inspection of Table 7 gives no significant correlation"  
46

47 Thank you for your suggestion. The sentence has been revised and now reads (lines 503-504):

48 "The examination of Table 8 gives no significant correlation between measured total NER and predicted  
49  $X_{\text{biogenic NER}}$  or  $SOM_{\text{biogenic NER}}$ ."  
50

51 37. lines 690-693: modify this sentence as follows "Such a correlation should not be expected since the  
52 processes leading to NER I, II and III are competing and if pesticide is not degraded it can undergo aging,  
53 irreversible sorption and covalent binding of parent compound or its metabolites [8]."  
54

55 Thank you for your suggestion. The sentence has been revised and now reads (lines 504-507):

56 "Such a correlation should not be expected since the processes leading to NER I, II and III are competing. If  
57 a pesticide is not degraded it can undergo aging and irreversible sorption (type I NER) and covalent binding  
58 of the parent compound or its metabolites (type II NER) [8]."  
59  
60

1  
2  
3  
4 38. line 702: add comma after reference 46

5 Thank you for pointing this out, see comment 39 below

6  
7 39. line 704: correct text "stable carbon or nitrogen isotope ( $^{13}\text{C}$  or  $^{15}\text{N}$ )"

8 Thank you for your suggestions. The sentence has been revised and now reads (lines 517-518)

9 "the formation of biogenic NER was investigated by tracking the distribution of stable carbon or nitrogen  
10 isotope ( $^{13}\text{C}$  or  $^{15}\text{N}$ ) in  $\text{CO}_2$  "

11  
12 40. line 707: probably "The experimentally determined values"

13 Thank you for the comment. The sentence now reads (lines 519-521):

14 "Experiments of this kind are very helpful to discriminate between the various types of NER and to validate  
15 our biogenic NER estimation approach."

16  
17 41. line 709: "although" does not make sense here since both stated results are positive

18 The sentence has been deleted as the paragraph has been revised.

19  
20 42. lines 710-714: Another example of long and confusing sentence. "together with the release of  $\text{CO}_2$ "  
21 does not make any sense here while expression "to in fact be" is meaningless and cannot be combined with  
22 "surmised".

23 Thank you for the comment. Yes, we agree, the sentence is too long and is barely readable in one breath.

24 The sentence has been tidied up and is now (lines 521-524):

25 "Shrestha *et al.* [11] observed that the formation of NER occurred simultaneously with the degradation and  
26 release of  $\text{CO}_2$ . This shows the coupling of the formation of NER to microbial activity, and to the growth and  
27 decay of biomass."

28  
29 43. lines 718-719: "dedicated experiments"?! maybe "targeted experiments"

30 Yes, targeted experiments done by dedicated researchers. This has been corrected on line 527-528:

31 "Before routine application though, further confirmation by targeted experiments is still needed"

32  
33 44. lines 722-723: correct this sentence as "The MTB method was compared with two widely used bacterial  
34 growth yield estimation methods, TEEM2 and ETTYM."

35 Thank you for your suggestion. The sentence has been revised as suggested (lines 531-532):

36 "The MTB method was compared with two widely used bacterial growth yield estimation methods, TEEM2  
37 and ETTYM."

38  
39 45. line 727: correct text as "to electron acceptors other than oxygen,"

40  
41 Thank you for your suggestion. The sentence has been revised (lines 536-537): "the MTB approach can be  
42 expanded to electron acceptors other than oxygen, like sulphate and nitrate."

43  
44 46. line 730: delete "which is" since it is redundant

45 Thank you for your suggestion. The sentence has been revised (line 539):

46 "changes in  $Y_{ATP}$ , an uncertain parameter."

47  
48 47. line 730: correct as "are also sensitive to"

49 Thank you for your suggestion. The sentence has been revised (lines 539-540):

50 "TEEM2 and ETTYM are also sensitive to changes in the cell formula"

51  
52 48. line 731-732: correct the first part of this sentence as "All methods were insensitive with respect to the  
53 imprecise data on Gibbs energy of formation"

54 Thank you for your suggestion. The sentence has been revised (lines 540-543):

1  
2  
3  
4 “All methods showed low sensitivity to variations in the Gibbs energy of formation of the organic  
5 chemicals because most of the Gibbs energy of reaction stems from the formation of the oxidation  
6 products carbon dioxide and water.  
7

8 49. line 748: probably “spreadsheet”

9 Thank you for pointing this out. This has been corrected (line 556). Apparently my MS Word automatically  
10 changes it to “spread sheet.”  
11

12 50. line 749: correct text as “on request from the first author.”

13 This has been corrected. Line 557 now reads:

14 “on request from the first author.”  
15  
16  
17  
18

### 19 Other changes

20 Other changes made to the manuscript are listed here.

21 Line 773: “and Figures” deleted.

22  
23 Lines 817-818 (deleted “anoxic”): “Comparison of bacterial growth yields of organic chemicals of  
24 environmental concern under anaerobic conditions in g cell carbon (g substrate carbon)<sup>-1</sup>.”  
25

26  
27 Lines 822-823: “The entries are sorted from low to high predicted yield of MTB under nitrate-reducing  
28 conditions.”  
29

30 In sub-section 3.3 *Prediction of biogenic NER formation based on the predicted growth yields* the  
31 following sentence was deleted as it was an unnecessary repetition:  
32

33 “High CO<sub>2</sub> formation as a measure for microbial activity typically resulted in the prediction of a large  
34 biogenic NER pool.”  
35

36 To reflect the contributions made by the anonymous reviewers, the acknowledgements now read (lines  
37 550-555): “This research project was financially supported by the Technical University of Denmark and the  
38 Helmholtz Centre for Environmental Research UFZ. We thank Fabio Polesel, Carson Odell Lee, and Ulrich  
39 Bay Gosewinkel for valuable suggestions and discussions. We also wish to acknowledge the comments and  
40 suggestions provided by the anonymous reviewers which helped to improve the manuscript.”  
41

42  
43 After having re-examined the tables, it appears that the values in column 2 ( $Y^{EXP}$ ) in Table 6 are incorrectly  
44 listed. They have somehow shifted during the transfer from the spreadsheet to the text document. The  
45 values in the other columns are correct, hence, the analysis and conclusions made based on Table 6 are all  
46 correct and unchanged.  
47

48 Moreover, wrong values for  $Y_{ATP}$  (= 5) and C-H-bonds (= 6) for gluconate have been used. Instead the values  
49 should have been  $Y_{ATP} = 10$  and C-H-bonds = 12. This has been changed in the supplementary information  
50 Table S2 and thus corrected in Table 6 and Table 7.  
51

52 Additionally, the calculations made for the preparation of Table 7 also contain a minor calculation error.  
53 Instead of using  $Y_{ATP} = 10.5$  for the calculation of the growth yield for TEEM2 and ETTYM methods, a value  
54 of  $Y_{ATP} = 10.5 \times 0.8 = 8.4$  has been used. Table 7 has been updated with correctly calculated values. The  
55 conclusions drawn on the original erroneous calculations are still valid.  
56  
57  
58  
59  
60



1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

Consequently, Figure S1 has also been changed to reflect this. The changes do not change any conclusion made.

For Peer Review Only

## Supplementary Information

### Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues

Andreas Libonati Brock<sup>a,\*</sup>, Matthias Kästner<sup>b</sup>, Stefan Trapp<sup>a</sup>,

<sup>a</sup>*Department of Environmental Engineering, Technical University of Denmark, Kgs. Lyngby, Denmark*

<sup>b</sup>*UFZ—Helmholtz Centre for Environmental Research, Department of Environmental Biotechnology, Leipzig, Germany*

\*[alib@env.dtu.dk](mailto:alib@env.dtu.dk), Phone: +45 45251408;

The supporting information contains information regarding:

- i) S1. Summary of the methods used to estimate the microbial growth yields.
- ii) Table S1. Balanced half-reactions of the simple carbon substrates and the Gibbs free energy of reaction
- iii) Table S2. Gibbs free energy of formation, C-H bonds,  $Y_{ATP}$ ,
- iv) Table S3. The relative deviation from the estimated bacterial growth yield estimated under aerobic conditions when assuming a Gibbs free energy of formation of 0 kJ mol<sup>-1</sup>.
- v) Figure S1. The effect of cell formula on the yield estimate
- vi) References

18 pages; 4 tables; 1 figure

31 **S1. Summary of the methods used to estimate the microbial growth yields.**

32 A summary of the three methods to estimate the microbial growth yields is provided in  
33 the following.

34 ***S1.1 Microbial Turnover to Biomass (MTB)***

35 The Microbial Turnover to Biomass (MTB) method presented in Trapp *et al.* [1] was  
36 based on the work by Diekert [2] where the maximum bacterial yield is determined  
37 from the nutritional value of the substrate ( $N$ ) combined with the determination of bio-  
38 available electrons from the reaction. The nutritional value describes how much  
39 substrate is needed per mass of bacteria [g substrate (g biomass)<sup>-1</sup>]. This is subdivided  
40 into a biomass yielding (anabolic) and energy yielding (catabolic) part. The nutritional  
41 value is the inverse of the yield, when the yield is defined as gram of cells per gram of  
42 substrate

$$43 \quad N = N_{ana} + N_{cata} = \frac{1}{Y} = \frac{1}{Y_{ana}} + \frac{1}{Y_{cata}} \quad (S1)$$

$$44 \quad \text{or } Y = \frac{Y_{ana} \times Y_{cata}}{Y_{ana} + Y_{cata}} \quad (S2)$$

45  
46 where  $Y_{ana}$  is the anabolic yield,  $Y_{cata}$  is the catabolic yield, and  $Y$  is the bacterial growth  
47 yield [all in g bacteria dry weight (dw) (g substrate)<sup>-1</sup>].

48  
49 ***Catabolism***

50 The catabolic yield is determined from calculation of the Gibbs free energy released  
51 from the complete oxidation of the compound, the storage of this energy in ATP, and  
52 the bacterial growth yield on ATP

$$53 \quad Y_{cata} = \frac{\frac{\Delta G^{0'}_{reaction}}{\Delta G^{0'}_{ATP} / \eta} Y_{ATP}}{M_S} \quad [\text{g cell dw (g substrate)}^{-1}] \quad (S3)$$

54

55  $\Delta G^{0'}_{ATP}$  is the Gibbs free energy of hydrolysis of adenosine triphosphate (ATP) [30.53  
 56 kJ mol<sup>-1</sup> [3]]. Bacteria have approximately 40% efficiency in the ATP energy gain ( $\eta$ )  
 57 [2, 4]. Taking this into account, the energy needed to synthesise one mole of ATP is  
 58 approximately -80 kJ (mol ATP)<sup>-1</sup>.  $Y_{ATP}$  is the bacterial yield on ATP [g cell dw (mol  
 59 ATP)<sup>-1</sup>] and its estimation is explained later.  $\Delta G^{0'}_{reaction}$  is the Gibbs free energy of  
 60 reaction,  $\Delta G^{0'}_{reaction} = \Delta G^{0'}_{products} - \Delta G^{0'}_{reactants}$ , at biological standard state conditions  
 61 [kJ mol<sup>-1</sup>].

62 The energy released from the reaction is assumed to not be fully available for  
 63 the microorganism and thus the concept of bio-available electrons was introduced [1].

64 The Gibbs free energy of reaction is related to the redox potential ( $\Delta E$ ) of the  
 65 reaction through the Nernst equation

$$66 \quad \Delta G = -n \times F \times \Delta E \quad (S4)$$

67

68 where  $n$  is the number of electrons transferred (sum of the change of oxidation status of  
 69 the carbon atoms in the substrate during oxidation), and  $F$  is the Faraday constant  
 70 [=96,485 C mol<sup>-1</sup>].

71 When C-H compounds are oxidised to CO<sub>2</sub> and H<sub>2</sub>O, the transferred electrons  
 72 are readily available for energy gain ( $2e^-$  per C-H bond). Thus, the number of bio-  
 73 available electrons is  $n_{bio} \geq \text{no. of C - H bonds} \times 2$ . Therefore; the minimum bio-available  
 74 energy can be calculated from the number of C-H bonds.

$$75 \quad \Delta G^{0'}_{bioavailable} = \frac{n_{bio}}{n} \Delta G^{0'}_{reaction} \quad (S5)$$

76

77 The “bio-available energy corrected”  $Y^*_{cata}$  can then be formulated as

$$78 \quad Y^*_{cata} = \frac{\frac{\Delta G^{0'}_{bioavailable}}{(\Delta G^{0'}_{ATP} / \eta)} Y_{ATP}}{M_S} \quad (S6)$$

79

80

81

82

83

84

85

86

87

88

89

90

79

80 Finally, as  $Y_{ATP}$  has been shown to be variable and dependent upon the substrate (and  
 81 substrate concentration), e.g. [5, 6], a simple set of rules have been proposed in order to  
 82 define this parameter. The rules are based on the  $\langle\text{CH}_2\text{O}\rangle$ -“sugar-structural-similarity”  
 83 of the compound [1]:

84

- 85 • If  $\langle\text{CH}_2\text{O}\rangle$  -  $\langle\text{C}_2\text{H}_4\text{O}_2\rangle$  is present in the molecule, or if oxygen is missing,  $Y_{ATP} = 5$   
 86 g cell dw (mol ATP)<sup>-1</sup>
- 87 • If  $\langle\text{C}_3\text{H}_6\text{O}_3\rangle$  -  $\langle\text{C}_4\text{H}_8\text{O}_4\rangle$  is present in the molecule,  $Y_{ATP} = 7.5$  g cell dw (mol  
 88 ATP)<sup>-1</sup>
- 89 • If  $\langle\text{C}_5\text{H}_{10}\text{O}_5\rangle$  -  $\langle\text{C}_6\text{H}_{12}\text{O}_6\rangle$  is present,  $Y_{ATP} = 10$  g cell dw (mol ATP)<sup>-1</sup>

90

91 Moreover, compounds not containing any “sugar-like” structure are assumed to have a  
 92  $Y_{ATP} = 5$  g cell dw (mol ATP)<sup>-1</sup>. The oxygen atoms in N-O, S-O, P-O bonds and  
 93 carboxylic groups are not counted [1].

94

### 95 *Anabolism*

96 The anabolic yield,  $Y_{ana}$ , is calculated from the carbon content in the compound (the  
 97 carbon source) and in the bacterial cell [2], i.e. how many grams of cell can be produced  
 98 from the carbon in the compound (only carbon availability is assumed to limit growth)

$$99 \quad Y_{ana} = \frac{n_C M_C}{\sigma_C M_S} \quad [\text{g cell dw (g substrate)}^{-1}] \quad (S7)$$

100

101 where  $\sigma_C$  is the fraction of carbon in dry cell (here taken as  $\sim 0.53$  g C (g cell dw)<sup>-1</sup>, with  
 102 the cell formula  $\text{C}_5\text{H}_7\text{O}_2\text{N}$  [7] but suggested to be 0.5 g C (g cell dw)<sup>-1</sup> in [2]),  $M_C$  is the

1  
2  
3 103 molar mass of carbon (12.01 g mol<sup>-1</sup>),  $M_S$  is the molar mass of the substrate [g mol<sup>-1</sup>],  
4  
5 104 and  $n_c$  is the number of carbon atoms in the substrate [mol C (mol substrate)<sup>-1</sup>].  
6

7 105 To convert the yield from g cell dw (g substrate)<sup>-1</sup> to g cell carbon (g substrate  
8  
9 106 carbon)<sup>-1</sup> the conversion can be determined from

11  
12 107 
$$f_{g/c} = \frac{1}{Y_{ana}} = \frac{\sigma_c}{M_c} \times \frac{M_s}{n_c} \quad (S8)$$

13  
14 108 where  $f_{g/c}$  has the unit (g cell carbon (g substrate carbon)<sup>-1</sup>) (g cell dw (g substrate)<sup>-1</sup>)<sup>-1</sup>.  
15  
16 109

### 17 18 110 *S1.2 Thermodynamic Electron Equivalent Model 2 (TEEM2)*

19  
20 111 In 1965 P. L. McCarty presented a thermodynamic model to estimate the maximal  
21  
22 112 bacterial yield [8]. Since its inception it has been modified and expanded [9]. As the  
23  
24 113 original model was found unable to capture the observed lower yields associated with  
25  
26 114 C1 compounds (i.e. methanol) and for reactions involving oxygenases [4], it was  
27  
28 115 recently modified [7]. The yield is calculated from an energy balance and an electron  
29  
30 116 balance.  
31  
32 117

#### 33 34 118 *Electron balance*

35  
36 119 The electron balance states that the electrons provided by the substrate are used either in  
37  
38 120 synthesis of cell material (anabolism) or in energy generation (catabolism)  
39  
40 121

41  
42 122 
$$f_s^0 + f_e^0 = 1 \quad (S10)$$

43  
44 123 where  $f_s^0$  is the fraction of electrons diverted for synthesis and  $f_e^0$  is the fraction of  
45  
46 124 electrons diverted for energy generation (used to reduce the electron acceptor). The  
47  
48 125 bacterial yield is equal to the fraction of electrons that are diverted to cell synthesis and  
49  
50 126 formation of new biomass ( $f_s^0$ ). The yield can be calculated as g cell carbon per g  
51  
52 127 substrate carbon (which is the very same as mol C per mol C) from the degree of  
53  
54  
55  
56  
57  
58  
59  
60

1  
2  
3 128 reductance of the cell carbon ( $\gamma_c$ ) and the degree of reductance of the substrate carbon  
4  
5 129 ( $\gamma_s$ ) [10].

6  
7 130 
$$Y_{C/C} = f_s^0 \frac{\gamma_s}{\gamma_c} \quad (S11)$$

8  
9 131

10 132 *Energy balance*

11  
12 133 The energy balance states that the energy *captured* by the organism from the redox  
13  
14 134 reaction is used for bacterial growth

15  
16 135 
$$-f_e^0 \epsilon \Delta G_r^{0'} = f_s^0 \Delta G_s^{0'} \quad (S12)$$

17  
18 136

19  
20 137 where  $\Delta G_s^{0'}$  is the Gibbs free energy for synthesis [ $\text{kJ mol}^{-1}$ ],  $\Delta G_r^{0'}$  is the Gibbs free  
21  
22 138 energy released from the redox reaction [ $\text{kJ mol}^{-1}$ ], and  $\epsilon$  is the energy capture  
23  
24 139 efficiency.  $\epsilon$  is a key parameter and is estimated from experimental data. In [7] a best fit  
25  
26 140 between predicted and experimental values was observed when it was set to 0.37 – i.e.  
27  
28 141 37% of the energy released from the redox reaction is captured by the bacterium to be  
29  
30 142 used in synthesis. It has generally been found to vary as a function of the growth (is it  
31  
32 143 autotrophic or heterotrophic) [4, 9].

33  
34 144 The Gibbs free energy of reaction is defined as

35  
36 145 
$$\Delta G_r^{0'} = \Delta G_a^{0'} - \Delta G_d^{0'} - \frac{q}{p} \Delta G_{xy}^{0'} \quad (S13)$$

37  
38 146

39  
40 147 where the subscript  $a$  denotes the electron acceptor while  $d$  denotes the electron donor  
41  
42 148 of the redox reaction,  $\Delta G_{xy}^{0'}$  is the reduction potential of NADH/NAD<sup>+</sup> oxidation (equal  
43  
44 149 to  $-219.2 \text{ kJ mol}^{-1}$  [7],  $q$  is the number of oxygenase reactions [oxygenase reactions mol  
45  
46 150 <sup>1</sup>],  $p$  is the number of electron equivalents (eeq) per mole substrate [eeq mol<sup>-1</sup>].

1  
2  
3 151  $\Delta G^0_s$  is given by the Gibbs free energy associated with the transformation of the  
4  
5 152 carbon source (which is often also the electron donor) to an intermediate and the  
6  
7 153 subsequent synthesis of cell material from this intermediate  
8

9  
10 154 
$$\Delta G_s^{0'} = \frac{\Delta G_{fa}^{0'} - \Delta G_d^{0'}}{\epsilon^m} + \frac{\Delta G_{in}^{0'} - \Delta G_{fa}^{0'}}{\epsilon^n} + \frac{\Delta G_{pc}^{0'}}{\epsilon} \quad (S14)$$

11  
12 155  
13  
14 156 where  $\Delta G_{fa}^{0'}$  is the Gibbs free energy of the half-reaction of formaldehyde ( $\Delta G_{fa}^{0'} =$   
15 46.53 kJ eeq<sup>-1</sup>),  $\Delta G_{in}^{0'}$  is the intermediate (assumed to be acetyl-CoA, a main  
16 157 intermediate in cell synthesis ( $\Delta G_{in}^{0'} = 30.9$  kJ eeq<sup>-1</sup>)),  $m$  is +1 when the electron donor  
17 158 is a C1 compound, else it is equal to  $n$ .  $n$  is equal to +1 if  $\Delta G_{in}^{0'} - \Delta G_d^{0'} > 0$  (energy is  
18 159 is a C1 compound, else it is equal to  $n$ .  $n$  is equal to +1 if  $\Delta G_{in}^{0'} - \Delta G_d^{0'} > 0$  (energy is  
19 160 needed); the energy needed to drive the reaction from electron donor to intermediate has  
20 161 to be divided with  $\epsilon$  to take inefficiencies into account. If  $\Delta G_{in}^{0'} - \Delta G_d^{0'} < 0$   $n$  is -1  
21 162 (energy is generated); the energy generated from the reaction is captured with efficiency  
22 163  $\epsilon$ . The efficiency term in synthesis ( $\epsilon$ ) is normally taken to be similar to the energy  
23 164 capture efficiency despite not being proven to be identical [11, 12].  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34

35 165  $\Delta G_{pc}^{0'}$  is the Gibbs free energy associated with the synthesis of cell material  
36 166 from the intermediate is given as  
37  
38  
39

40 167 
$$\Delta G_{pc}^{0'} = \frac{\Delta G_{ATP}^{0'}}{Y_{ATP} \times 0.9} \times \frac{M_C}{\gamma_c \sigma_c} \quad [\text{kJ eeq}^{-1}] \quad (S15)$$

41  
42 168  
43  
44 169 where  $\gamma_c$  is the average degree of reductance of the carbon atoms in the cell [eeq (mol  
45 170 carbon)<sup>-1</sup>],  $\sigma_c$  is the fraction of carbon in the cell [for a cell formula of C<sub>5</sub>H<sub>7</sub>O<sub>2</sub>N:  $\gamma_c = 4$   
46 171 eeq mol<sup>-1</sup>,  $\sigma_c = 0.531$  g carbon (g cell)<sup>-1</sup>],  $\Delta G_{ATP}^{0'}$  is as previously defined,  $Y_{ATP}$  is as  
47 172 previously defined and taken as a constant and equal to 10.5 g cell dw (mol ATP)<sup>-1</sup>, and  
48 173 0.9 is the percentage of organic material of a dry cell [g organic material (g cell dw)<sup>-1</sup>].  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60



175 *S1.3 Expanded Thermodynamic True Yield Prediction Model (ETTYM)*

176 The Expanded Thermodynamic True Yield Model was presented in [4] and expanded in  
177 [11] to account for oxygenase reactions and pH.

178 To increase the accuracy for the yield prediction on C1 compounds and  
179 substrates with low degrees of reduction, the authors proposed to include a carbon and a  
180 nitrogen balance and as a result thereof reformulate the electron and energy balance  
181 originally proposed by McCarty [4, 11]. The nitrogen balance can be ignored if nitrogen  
182 is not limiting [11], hence, the yield can be calculated from an energy balance, carbon  
183 balance, and electron balance.

184

185 *Carbon balance*

186 The carbon balance describes that the carbon is either invested in cell synthesis or into  
187 other carbonaceous products. If the only other carbonaceous product is CO<sub>2</sub>, the  
188 equation states that either carbon is incorporated into cell mass or oxidised to CO<sub>2</sub>

189 
$$f_{cell} + \sum_i f_{CS}(i) = 1 \quad (S16)$$

190

191 where  $f_{cell}$  is the bacterial growth yield [g cell carbon (g substrate carbon)<sup>-1</sup>], and  $f_{CS}(i)$  is  
192 the yield of carbonaceous product  $i$ .

193

194 *Energy balance*

195 The energy balance describes the relationship between the Gibbs free energy of cell  
196 synthesis and the Gibbs free energy captured from oxidation and reduction of various  
197 electron donor-electron acceptor pairs. If there is only one electron donor-electron  
198 acceptor pair, the energy balance can be written as

$$f_{cell} \times \left( \frac{(\Delta G_{acetate} - \Delta G_{CS})}{K^m} + \frac{\frac{\Delta G_{ATP} \times M_{cell}}{Y_{ATP} \times 0.9}}{K} \right) = -K \times (g(1)\Delta G_a - f_{CO_2}\Delta G_d) \quad (S17)$$

200

201 where  $\Delta G_{acetate}$  is the Gibbs free energy of acetate reduction (= 106.3 kJ (mol C)<sup>-1</sup>) [11]

202  $\Delta G_{CS}$  is the Gibbs free energy of the carbon source,  $M_{cell}$  is the cell mass per mol carbon

203 (= 26.4 g (mol C)<sup>-1</sup>) with cell formula C<sub>5</sub>H<sub>10</sub>O<sub>3</sub>N,  $g(I)$  is the number of electrons gained

204 by the electron acceptor (electron equivalents (eeq) (mol carbon substrate)<sup>-1</sup>),  $f_{CO_2}$  is the

205 fraction of carbon oxidised to CO<sub>2</sub> (mol C (mol carbon substrate)<sup>-1</sup>),  $K$  is the energy

206 efficiency parameter associated with (i) energy capture and storage in ATP and (ii)

207 energy transfer from ATP to cell synthesis ( $K = 0.41$ ) [4]. While these two different

208 processes have different efficiencies, the values are close enough to lump them into one

209 parameter [4].  $m$  is +1 if  $(\Delta G_{acetate} - \Delta G_{CS}) > 0$  (energy is needed for transformation

210 of carbon source to acetate) else  $m = -1$  (energy is gained from transformation of carbon

211 source to acetate). The other parameters are as previously defined for TEEM2. All half-

212 reactions are written as reductions similarly to TEEM2 and MTB, although  $\Delta G_a$  is in kJ

213 eeq<sup>-1</sup>,  $\Delta G_d$  is in kJ (mol substrate carbon)<sup>-1</sup>.

214

### 215 *Electron balance*

216 The electron balance states that the electrons coming from the oxidation of the electron

217 donor(s) are equal to the electrons used for reducing electron acceptors: the terminal

218 electron acceptor, nitrogen source different from ammonium, electrons lost in

219 oxygenase reactions; electrons used to reduce or increase the degree of reduction of the

220 carbon source to the same level as the cell [4]

$$\sum_i f_{ED}(i) \times (\gamma_{ED} - \gamma(i)) = \sum_j g(j) \quad (S18)$$

222

1  
2  
3 223 where  $f_{ED}(i)$  is the fraction of electron donor going to oxidised product  $i$ ,  $\gamma_{ED}$  is the  
4  
5 224 degree of reduction of the electron donor,  $\gamma(i)$  is the degree of reduction of oxidised  
6  
7 225 product  $i$ ,  $g(j)$  is the number of electrons sent to electron acceptor  $j$ .  
8

9  
10 226 If the electron donor is also the carbon source and the electrons are used only for  
11  
12 227 reduction of the terminal electron acceptor and synthesis of new cell material, the  
13  
14 228 equation reduces to (remember,  $\gamma_{CO_2} = 0$ )

15  
16 229 
$$f_{CO_2} \times (\gamma_s - \gamma_{CO_2}) + f_{cell}(\gamma_s - \gamma_x) = f_{CO_2} \times \gamma_s + f_{cell}(\gamma_s - \gamma_x) = g(1) \text{ (S19)}$$

17  
18  
19 230 where  $\gamma_s$  is the degree of reduction of the electron donor (carbon source), and  $\gamma_x$  is the  
20  
21 231 degree of reduction of the cell.  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

232 Table S1. Balanced half-reactions as reductions of the simple carbon substrates and their associated Gibbs  
 233 free energy of the half-reaction ( $\Delta G_r^{\circ}$ ) in  $\text{kJ mol}^{-1}$  and  $\text{kJ (electron equivalent (eeq))}^{-1}$  at standard state  
 234 conditions, except for  $\text{H}^+$  ( $=10^{-7} \text{ M}$ ).

Electron donor	Half-reaction	$\Delta G_r^{\circ}$	
		in $\text{kJ mol}^{-1}$	$\text{kJ eeq}^{-1}$
Acetate	$2 \text{ CO}_2 + 7 \text{ H}^+ + 8 \text{ e}^- \rightleftharpoons \text{C}_2\text{H}_3\text{O}_2^- + 2 \text{ H}_2\text{O}$	223.4	27.9
Citrate	$6 \text{ CO}_2 + 15 \text{ H}^+ + 18 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_5\text{O}_7^{3-} + 5 \text{ H}_2\text{O}$	608.1	33.8
Formaldehyde	$\text{CO}_2 + 4 \text{ H}^+ + 4 \text{ e}^- \rightleftharpoons \text{CH}_2\text{O} + \text{H}_2\text{O}$	185.8	46.5
Formate	$\text{CO}_2 + \text{H}^+ + 2 \text{ e}^- \rightleftharpoons \text{CHO}_2^-$	82.8	41.4
a-D-Fructose	$6 \text{ CO}_2 + 24 \text{ H}^+ + 24 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{12}\text{O}_6 + 6 \text{ H}_2\text{O}$	982.7	40.9
Fumaric acid	$4 \text{ CO}_2 + 12 \text{ H}^+ + 12 \text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_4 + 4 \text{ H}_2\text{O}$	458.8	38.2
a-D-Galactose	$6 \text{ CO}_2 + 24 \text{ H}^+ + 24 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{12}\text{O}_6 + 6 \text{ H}_2\text{O}$	974.6	40.6
Gluconate	$12 \text{ CO}_2 + 42 \text{ H}^+ + 44 \text{ e}^- \rightleftharpoons \text{C}_{12}\text{H}_{22}\text{O}_{14}^{2-} + 10 \text{ H}_2\text{O}$	3151	71.6
a-D-Glucose	$6 \text{ CO}_2 + 24 \text{ H}^+ + 24 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{12}\text{O}_6 + 6 \text{ H}_2\text{O}$	980.9	40.9
Glycerol	$3 \text{ CO}_2 + 14 \text{ H}^+ + 14 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_8\text{O}_3 + 3 \text{ H}_2\text{O}$	540.3	38.6
Glycine	$2 \text{ CO}_2 + \text{NH}_3 + 6 \text{ H}^+ + 6 \text{ e}^- \rightleftharpoons \text{C}_2\text{H}_5\text{NO}_2 + 2 \text{ H}_2\text{O}$	208.7	34.8
Glyoxylate	$2 \text{ CO}_2 + 3 \text{ H}^+ + 4 \text{ e}^- \rightleftharpoons \text{C}_2\text{HO}_3^- + \text{H}_2\text{O}$	210.9	52.7
Lactate	$3 \text{ CO}_2 + 11 \text{ H}^+ + 12 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_5\text{O}_3^- + 3 \text{ H}_2\text{O}$	391.4	32.6
a-Lactose	$12 \text{ CO}_2 + 48 \text{ H}^+ + 48 \text{ e}^- \rightleftharpoons \text{C}_{12}\text{H}_{22}\text{O}_{11} + 13 \text{ H}_2\text{O}$	2044	42.6
Malate	$4 \text{ CO}_2 + 10 \text{ H}^+ + 12 \text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_5^{2-} + 3 \text{ H}_2\text{O}$	418.2	34.9
Malonate	$3 \text{ CO}_2 + 6 \text{ H}^+ + 8 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_2\text{O}_4^{2-} + 2 \text{ H}_2\text{O}$	269.4	33.7
Mannitol	$6 \text{ CO}_2 + 26 \text{ H}^+ + 26 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{14}\text{O}_6 + 6 \text{ H}_2\text{O}$	1035	39.8
Oxalate	$2 \text{ CO}_2 + 2 \text{ e}^- \rightleftharpoons \text{C}_2\text{O}_4^{2-}$	114.0	57.0
Phenol	$6 \text{ CO}_2 + 28 \text{ H}^+ + 28 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_6\text{O} + 11 \text{ H}_2\text{O}$	799.1	28.5
Phenylacetic acid	$8 \text{ CO}_2 + 36 \text{ H}^+ + 36 \text{ e}^- \rightleftharpoons \text{C}_8\text{H}_8\text{O}_2 + 14 \text{ H}_2\text{O}$	1131	31.4
Propionate	$3 \text{ CO}_2 + 13 \text{ H}^+ + 14 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_5\text{O}_2^- + 4 \text{ H}_2\text{O}$	390.7	27.9
Pyruvate	$3 \text{ CO}_2 + 9 \text{ H}^+ + 10 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_3\text{O}_3^- + 3 \text{ H}_2\text{O}$	354.8	35.5
Sorbitol	$6 \text{ CO}_2 + 26 \text{ H}^+ + 26 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{14}\text{O}_6 + 6 \text{ H}_2\text{O}$	1035	39.8

Succinate	$4 \text{ CO}_2 + 12 \text{ H}^+ + 14 \text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_4^- + 4 \text{ H}_2\text{O}$	415.7	29.7
Tartrate	$4 \text{ CO}_2 + 8 \text{ H}^+ + 10 \text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_6^{2-} + 2 \text{ H}_2\text{O}$	583.9	58.4
Xylose	$5 \text{ CO}_2 + 20 \text{ H}^+ + 20 \text{ e}^- \rightleftharpoons \text{C}_5\text{H}_{10}\text{O}_5 + 5 \text{ H}_2\text{O}$	1231	61.5

235

For Peer Review Only

236 Table S2. Gibbs free energy of formation in  $\text{kJ mol}^{-1}$  ( $\Delta G_f^\circ$ ), number of carbon-hydrogen bonds,  $Y_{ATP}$  in  
 237 g dry weight ( $\text{mol ATP}^{-1}$ ), number of (putative) oxygenase reactions ( $t^{oxy}$ ), chemical structure, and degree  
 238 of reductance ( $\gamma_s$ ) of the simple carbon substrates used in the comparison.

Name	Structure	$\Delta G_f^\circ$	C-H bonds	$Y_{ATP}$	$t_{oxy}$	$\gamma_s$	Reference
Acetate	$\text{C}_2\text{H}_3\text{O}_2^-$	-369.4	3	5	0	4	[13]
Citrate	$\text{C}_6\text{H}_5\text{O}_7^{3-}$	-1168.3	4	5	0	3	[13]
Formaldehyde	$\text{CH}_2\text{O}$	-130.5	2	5	0	4	[13]
Formate	$\text{CHO}_2^-$	-351.0	1	5	0	2	[13]
a-D-Fructose	$\text{C}_6\text{H}_{12}\text{O}_6$	-915.4	7	10	0	4	[13]
Fumaric acid	$\text{C}_4\text{H}_4\text{O}_4$	-647.1	2	5	0	3	[13]
a-D-Galactose	$\text{C}_6\text{H}_{12}\text{O}_6$	-923.5	7	10	0	4	[13]
Gluconate	$\text{C}_{12}\text{H}_{22}\text{O}_{14}^{2-}$	-880.2	12	10	0	3.7	[14]
a-D-Glucose	$\text{C}_6\text{H}_{12}\text{O}_6$	-917.2	7	10	0	4	[13]
Glycerol	$\text{C}_3\text{H}_8\text{O}_3$	-488.5	5	7.5	0	4.7	[13]
Glycine	$\text{C}_2\text{H}_5\text{NO}_2$	-370.8	2	5	0	3	[13]
Glyoxylate	$\text{C}_2\text{HO}_3^-$	-459.6	1	5	0	2	[13]
Lactate	$\text{C}_3\text{H}_5\text{O}_3^-$	-517.8	4	5	0	4	[13]
a-Lactose	$\text{C}_{12}\text{H}_{22}\text{O}_{11}$	-1515.2	14	5	0	4	[13]
Malate	$\text{C}_4\text{H}_4\text{O}_5^{2-}$	-845.1	3	5	0	3	[13]
Malonate	$\text{C}_3\text{H}_2\text{O}_4^{2-}$	-677.6	2	5	0	2.7	[14]
Mannitol	$\text{C}_6\text{H}_{14}\text{O}_6$	-942.6	8	5	0	4.3	[13]
Oxalate	$\text{C}_2\text{O}_4^{2-}$	-674.0	0	5	0	1	[13]
Phenol	$\text{C}_6\text{H}_6\text{O}$	-72.8	5	5	2	4.7	[14]
Phenylacetic acid	$\text{C}_8\text{H}_8\text{O}_2$	-136.9	7	5	2	4.5	[14]
Propionate	$\text{C}_3\text{H}_5\text{O}_2^-$	-361.1	5	5	0	4.7	[13]
Pyruvate	$\text{C}_3\text{H}_3\text{O}_3^-$	-474.6	3	5	0	3.3	[13]
Sorbitol	$\text{C}_6\text{H}_{14}\text{O}_6$	-942.7	8	5	0	4.3	[13]
Succinate	$\text{C}_4\text{H}_4\text{O}_4^{2-}$	-690.2	4	5	0	3.5	[13]
Tartrate	$\text{C}_4\text{H}_4\text{O}_6^{2-}$	-747.5	2	5	0	2.5	[14]

Xylose	C <sub>5</sub> H <sub>10</sub> O <sub>5</sub>	-350.9	6	5	0	4	[14] 239
--------	---	--------	---	---	---	---	----------

For Peer Review Only

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

240 Table S3. The relative deviation from the estimated bacterial growth yield estimated under aerobic  
 241 conditions when assuming a Gibbs free energy of formation of 0 kJ mol<sup>-1</sup>. The deviation is computed  
 242 relative to the yield estimated with the Gibbs free energy of formation reported in Table 3. The mean  
 243 absolute deviation is computed. MTB has the lowest mean average deviation.

	MTB	TEEM2	ETTYM
2,4-D	5%	7%	7%
2,4-DB	1%	1%	1%
Acetamiprid	-10%	-27%	-21%
Acetochlor	1%	2%	2%
Alachlor	-4%	-8%	-8%
Anthracene	-7%	-25%	-9%
Atrazine	-11%	-27%	-22%
Azoxystrobin	-3%	-5%	-5%
Benalaxyl	-4%	-8%	-7%
Benzene	-3%	-15%	-4%
Benzoate	2%	4%	3%
Bifenazate	-6%	-10%	-10%
Carbofuran	3%	5%	4%
Chlorothalonil		-6%	-8%
Chlorpropham	-7%	-11%	-11%
Cypermethrin	-4%	-6%	-6%
Daminozide	0%	0%	0%
DDT	-6%	-8%	-8%
Desmedipham	-8%	-12%	-12%
Dicamba	5%	7%	7%
Ethylenediaminetetraacetate (EDTA)	18%	39%	30%
Famoxadone	-6%	-9%	-8%
Glyphosate	20%	52%	52%
Iprodione	-5%	-7%	-7%

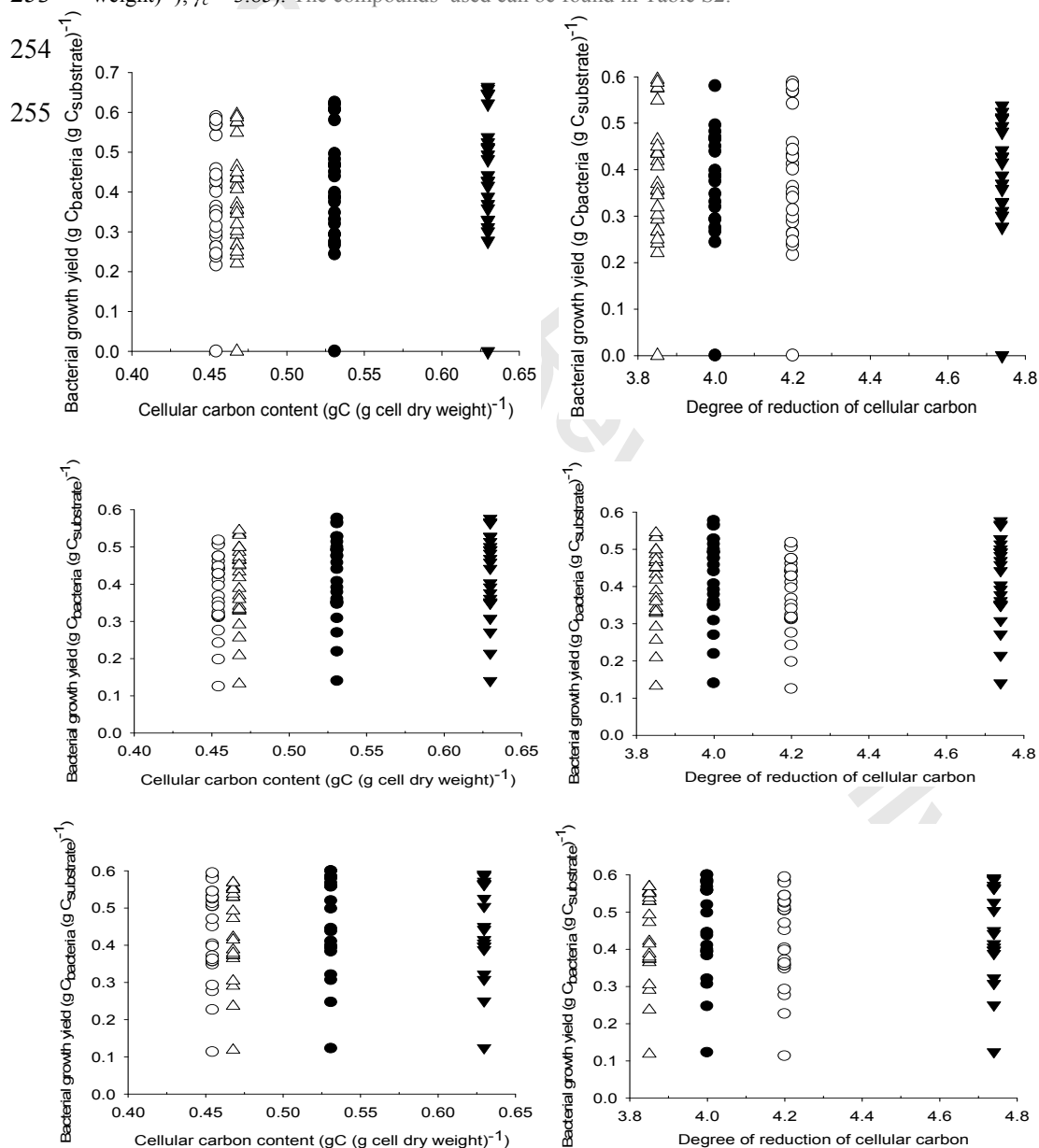


MCPA	2%	2%	2%
MCPB	0%	0%	0%
Mecoprop (MCP)	0%	0%	0%
Metalaxyl-M	-3%	-5%	-5%
Metamitron	-7%	-18%	-14%
Milbemectin	-3%	-6%	-6%
Naphthalene	-6%	-25%	-10%
Nitrilotriacetate (NTA)	23%	49%	44%
Paraquat	-8%	-27%	-14%
Pendimethalin	-7%	-13%	-13%
Phenanthrene	-7%	-25%	-9%
Phenmedipham	-8%	-11%	-12%
Propyzamide	-4%	-6%	-7%
Pymetrozine	-12%	-27%	-24%
Pyrene	-3%	-12%	-5%
Mean absolute deviation	6%±5%	14%±13	11%±11%

244

245

246 Figure S1. The effect of cell formula on the yield estimate. **Top:** Microbial Turnover to Biomass (MTB)  
 247 method; **middle:** Thermodynamic Electron Equivalent Method 2 (TEEM2); **bottom:** Expanded  
 248 Thermodynamic True Yield Prediction Method (ETTYM). **Left:** The effect of the carbon content ( $\sigma_c$ );  
 249 and **right:** the effect of degree of reductance ( $\gamma_c$ ) are shown depending on the chosen cell formula. For  
 250 ETTYM, the standard cell formula is  $C_5H_{10}O_3N$ ; for TEEM2 and MTB it is  $C_5H_7O_2N$ .  $\bullet$ :  $C_5H_7O_2N$  ( $\sigma_c =$   
 251  $0.53 \text{ gC (g cell dry weight)}^{-1}$ ,  $\gamma_c = 4$ ),  $\circ$ :  $C_5H_{10}O_3N$  ( $\sigma_c = 0.45 \text{ gC (g cell dry weight)}^{-1}$ ,  $\gamma_c = 4.2$ );  $\blacktriangledown$ :  
 252  $C_5H_{8.33}O_{0.8}N$  ( $\sigma_c = 0.63 \text{ gC (g cell dry weight)}^{-1}$ ,  $\gamma_c = 4.74$ );  $\Delta$ :  $C_{4.1}H_{6.8}O_{2.2}N$  ( $\sigma_c = 0.47 \text{ gC (g cell dry}$   
 253  $\text{weight)}^{-1}$ ,  $\gamma_c = 3.85$ ). The compounds used can be found in Table S2.



S17

256

257 **References**

- 258 [1] S. Trapp, A. Libonati Brock, K. Nowak and M. Kästner, *Prediction of the formation*  
259 *of biogenic non-extractable residues during degradation of environmental*  
260 *chemicals from biomass yields*, Environ. Sci. Technol. (2017), Submitted for  
261 *publication*
- 262 [2] G. Diekert, Grundmechanismen des Stoffwechsels und der Energiegewinnung, in  
263 *Umweltbiotechnologie*, J.C.G. Ottow and W. Bidlingmaier, eds., Fischer Verlag,  
264 Stuttgart, Germany, 1997; pp. 1-38.
- 265 [3] K. Burton, *Energy of adenosine triphosphate*. Nature 181 (1958), pp. 1594-1595.
- 266 [4] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic model for microbial true*  
267 *yield prediction*, Biotechnol. Bioeng 93 (2006), pp. 110–121.
- 268 [5] J.J. Heijnen, and J.P. Dijken, *In search of a thermodynamic description of biomass*  
269 *yields for the chemotrophic growth of microorganisms*, Biotechnol Bioeng 39  
270 (1992), pp. 833-852
- 271 [6] D. W. Tempest and O.M. Neijssel, *The Status of  $Y_{atp}$  and Maintenance Energy as*  
272 *Biologically Interpretable Phenomena*. Ann. Rev. Microbiol., 38 (1984), pp. 459–  
273 486.
- 274 [7] P.L. McCarty, *Thermodynamic Electron Equivalents Model for bacterial yield*  
275 *prediction: Modifications and comparative evaluations*, Biotechnol. Bioeng 97  
276 (2007), pp. 377–388.
- 277 [8] P.L. McCarty, *Thermodynamics of biological synthesis and growth*, Intl. J. Air  
278 Water Poll. 9 (1965), pp. 621-639.
- 279 [9] B.E. Rittmann, P.L. McCarty, *Environmental biotechnology: principles and*  
280 *applications*. McGraw-Hill: New York, NY, 2001.
- 281 [10] J.M. VanBriesen, *Thermodynamic yield predictions for biodegradation through*  
282 *oxygenase activation reactions*, Biodegradation 12 (2001), pp. 265-281.
- 283 [11] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic true yield prediction*  
284 *model: adjustments and limitations*, Biodegradation 19 (2008), pp. 99–127.
- 285 [12] J.M. VanBriesen, *Evaluation of methods to predict bacterial yield using*  
286 *thermodynamics*, Biodegradation 13 (2002), pp. 171–190.
- 287 [13] R.K. Thauer, K. Jungermann, and K. Decker, *Energy conservation in chemotrophic*  
288 *anaerobic bacteria*, Bacteriol. Rev. 41 (1977), pp. 100–180.

- 1  
2  
3 290 [14] A. Flamholz, E. Noor, A. Bar-Even, and R. Milo, *eQuilibrator - the biochemical*  
4 291 *thermodynamics calculator*, *Nucleic Acids Res.* 40 (2012), pp. 770-775.  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

For Peer Review Only