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Evaluating the productivity of four main tree species in Germany under climate change with static reduced models

Martin Gutsch¹ · Petra Lasch-Born¹ · Felicitas Suckow¹ · Christopher P.O. Reyer¹

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Abstract

• **Key message** We present simple models of forest net primary production (NPP) in Germany that show increasing productivity, especially in mountainous areas, under warming unless water becomes a limiting factor. They can be used for spatially explicit, rapid climate impact assessment.

• **Context** Climate impact studies largely rely on process-based forest models generally requiring detailed input data which are not everywhere available.

• **Aims** This study aims to derive simple models with low data requirements which allow calculation of NPP and analysis of

climate impacts using many climate scenarios at a large amount of sites.

• **Methods** We fitted regression functions to the output of simulation experiments conducted with the process-based forest model 4C at 2342 climate stations in Germany for four main tree species on four different soil types and two time periods, 1951–2006 and 2031–2060.

• **Results** The regression functions showed a reasonable fit to measured NPP datasets. Temperature increase of up to 3 K leads to positive effects on NPP. In water-limited regions, this positive effect is dependent on the length of drought periods. The highest NPP increase occurs in mountainous regions.

• **Conclusion** Rapid analyses, using reduced models as presented here, can complement more detailed analyses with process-based models. Especially for dry sites, we recommend further study of climate impacts with process-based models or detailed measurements.

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Contribution of the co-authors Martin Gutsch: design of study, model simulations and analysis, interpretation of results, writing the paper. Petra Lasch-Born: design of study, interpretation of results, writing the paper. Felicitas Suckow: design of study, interpretation of results, writing the paper. Christopher Reyer: interpretation of results, writing the paper

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✉ Christopher P.O. Reyer
reyer@pik-potsdam.de
Martin Gutsch
gutsch@pik-potsdam.de
Petra Lasch-Born
lasch@pik-potsdam.de
Felicitas Suckow
suckow@pik-potsdam.de

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1 Introduction

Climate change influences forest ecosystems and the goods and services they provide in many ways (Alcamo et al. 2007; Lindner et al. 2010; Maroschek et al. 2009). Thus, future forests have to satisfy different and often conflicting social, ecological and economic needs under uncertain future conditions.

During the last 50 years the net primary production (NPP) of forests has increased in Europe (Ciais et al. 2008; Nemani et al. 2003; Spiecker 2002). Besides changing management, age structure, forest area and nutrient balances, climate change also partly explains this trend. Firstly, photosynthesis is not

¹ Potsdam Institute for Climate Impact Research (PIK), P.O. Box 60 12 03, 14412 Potsdam, Germany

saturated at the current carbon dioxide concentration (CO₂) and thus tree growth may benefit from increasing atmospheric CO₂ concentrations (Körner 2006; Norby et al. 2005). However, interactions between the drivers of tree growth complicate the relationship between the CO₂ response of photosynthesis and growth (Körner et al. 2005; Körner 2006; Norby et al. 2010). Secondly, changes in climatic trends influence regional temperature and water balances and most likely enhance the productivity of temperate forest ecosystems at sites which are not nutrient or water limited (Boisvenue and Running 2006; Ciais et al. 2008).

On the other hand, drought stress decreases productivity and increases the predisposition to biotic (insect calamities, fungi infections) and abiotic disturbances (forest fire, storm events). For example, Norway spruce (*Picea abies* L. Karst.) forests impaired by drought stress are predisposed to bark beetle attacks (Wermelinger 2004).

Forest tree species are sensitive to climatic changes, and as a result, species distribution and composition may be altered (Thuiller et al. 2005). This will broaden the silvicultural portfolio in mountainous and other temperature-limited forest ecosystems. Besides these potential positive effects at the leading edge, negative effects at the trailing edge of species distributions are also reported: Scots pine (*Pinus sylvestris* L.) in its southernmost range may experience reduced growth and survival with modest warming (Reich and Oleksyn 2008), while beech (*Fagus sylvatica* L.) forests show a decrease in basal area growth due to long-term drought in the Apennines (Piovesan et al. 2008).

Consequently, numerous simulation studies strive to ascertain how sensitive global and regional forest productivity will be to climate change in forthcoming decades (e.g. Lasch et al. 2002a, b, 2005; Piao et al. 2009; Sitch et al. 2008; Reyer et al. 2014). However, these climate impact studies rely on process-based forest models which require detailed information such as soil, stand and meteorological data, which are usually not available for most sites in a region. Or, they result from global scale studies, which are of limited value for a specific region due to the coarse resolution of the models. Therefore, regional, spatially explicit studies of climate impacts on forests would benefit from models with low data requirements applicable for a large set of stands. Furthermore, to account for uncertainties associated with climate change projections, it is necessary to use a broad variety of climate change scenarios to derive probabilities of the expected impacts. In process-based models, this comes at the cost of large amounts of computing time. These limitations can be circumvented by applying simple regression models to the results of process-based models (e.g. Sallaba et al. 2015). We call these simple models here *static reduced models* (SRM) to distinguish this model type from time-sensitive dynamic models and complex process-based models with high data demand.

The first aim of this study was to derive SRMs for four tree species (European beech, Norway spruce, Scots pine and oaks (*Quercus robur* L. and *Quercus petraea* Liebl.)), which allow calculation of the annual average NPP. The derivation was based on annual NPP values simulated with the process-based model 4C (Lasch et al. 2005). Independent variables of the SRMs for NPP prediction were climate and soil variables across a spectrum of mono-species forest stands with different site conditions. Secondly, this study aimed to analyze the impacts of a broad variety of climate scenarios with the SRMs for the four selected tree species at a large number of sites all over Germany. Hence, we were answering two main questions: What regional trends in NPP change are projected up to 2060 in Germany and how much does the SRM-based NPP estimation vary under a broad set of climate change scenarios? This variability served as an indication of the uncertainty of different climate change scenarios.

2 Material and methods

2.1 Climate data and climate scenarios

We used observed and homogenized meteorological time series for the period 1951–2006 with a daily time resolution for 2342 meteorological stations (Österle et al. 2006) in Germany. Based on these observed daily meteorological data, climate change scenarios from 2007 until 2060 with daily resolution were generated with the statistical regional climate model STAR 2.0 (Orlowsky et al. 2007). The observed meteorological data and 50 realizations of a climate change scenario based on the A1B SRES scenario (Nakicenovic et al. 2000) with a mean temperature increase in Germany of about 2 K by 2060 provide the climate dataset for deriving the SRMs. The realizations are replications of the same climate change scenario with the same temperature trend but reflecting different precipitation trends.

Once the SRMs were established, we drove them with seven climate scenarios generated with STAR 2.0 assuming a 0.5 K stepwise increase of temperature up to 3 K. For example, the 2 K climate scenario contains a temperature increase of 2 K in 2060 against the starting point of the scenario in 2007. For each of the seven climate scenarios (0 K, 0.5 K, ..., 3 K), 50 realizations spanning a wide range of precipitation trends were available. The climate scenarios show a rather linear increase in temperature, radiation and drought index while a slight trend to lower annual precipitation sums was calculated across the climate stations between the period 1951–2006 and 2031–2060 (see Online Electronic Resources (OER) Fig. 1). It is important to note that recently, the STARS algorithm has been shown to overestimate radiation trends in future scenarios (Wechsung and Wechsung 2015).

2.2 Derivation of the SRM

2.2.1 Model runs with 4C

We use the process-based forest growth model 4C, which is described in more detail in OER Text 1, Lasch et al. (2005) and Reyer et al. (2014). The average annual NPP data to be used for the regression analysis were calculated separately for mono-specific stands of the four tree species with 4C at the 2342 climate stations in Germany. At each station, four different soil types were assumed. Hence, we simulated the average annual NPP for every tree species and soil type with the observed climate data of a climate station (1951–2006) and the future climate scenario provided by STAR 2.0 based on the A1B emission trend (2007–2060). At a lower level of detail, the soil types were parameterized for 4C based on four reference profiles, representing typical forest soil conditions in Germany, taken from the soil database for Germany (BÜK 1 000, BGR 1998). They differ in the amount of plant available soil water and the carbon/nitrogen ratio (Table 1) as follows:

- poor soil with low water availability (PL)
- rich soil with low water availability (RL)
- poor soil with high water availability (PH)
- rich soil with high water availability (RH)

The stands were simulated for a time period of 56 years (1951–2006) in the past and 54 years (2007–2060) in the future. The stand data for the initialization of 4C originate from yield tables (Table 2). We only considered young stands with an initial age of 15 years to simulate the period with the highest annual productivity of an even-aged forest stand. The simulated annual NPP was averaged over the simulation period.

To eliminate sites with extreme growing conditions such as high mountain areas, which may not be captured well by 4C, only sites with a simulated average annual NPP value $Y_{NPP,4C}$ greater than 1 ton carbon $ha^{-1} year^{-1}$ were considered for deriving the species-specific regression functions. That resulted in 2340 (climate stations) times four (soil types) times two (base and future climate scenario), hence, 18,720 simulated average annual NPP values ($Y_{NPP,4C}$) for each of the tree species.

Table 1 Soil characteristics for the four soil types that were selected to represent German forest soils in this study

Type	Symbol	Plant available water (mm)	C_{tot} (g m^{-2})	N_{tot} (g m^{-2})	C_{tot}/N_{tot}
Cambisol	PL	143	19,650	640	31
Cambisol	RL	92	4250	262	16
Cambisol	PH	231	19,650	640	31
Cambisol	RH	179	4250	262	16

PL poor soil with low water availability, RL rich soil with low water availability, PH poor soil with high water availability, RH rich soil with high water availability, C_{tot} total carbon content, N_{tot} total nitrogen content

2.2.2 Fitting of the static reduced models

The functions for NPP were fitted to $Y_{NPP,4C}$ dependent on climate and soil variables, which are easily available from either soil databases or climate stations. These were plant available soil water and the C/N ratio of the soil of four reference soil profiles and the mean, minimum and maximum temperature, precipitation, radiation, relative air humidity, cloudiness and water vapour pressure, as well as aggregated climate variables such as the climatic water balance and a drought index (see below) that could be derived from these measurements. To balance the amount of independent soil and climate variables with model complexity, we decreased the amount of climate variables by forward selection. Linear models with log-transformed response and predictor variables were tested to account for nonconstant error variance and trends in the distribution. Also, polynomial terms in the drought index variable were used resulting in multiple regression functions with the same structure for all tree species:

$$\log Y_{NPP,SRM} = \alpha + \beta_1(\log X_W) + \beta_2(\log X_{CN}) + \beta_3(\log X_T) + \beta_4(\log X_R) + \beta_5(\log X_D) + \beta_6(\log X_D)^2 + \beta_7(\log X_W \cdot \log X_{CN}) + \varepsilon \quad (1)$$

$Y_{NPP,SRM}$	annual average net primary production (t C $ha^{-1} year^{-1}$)
X_W	plant available water (mm)
X_{CN}	carbon/nitrogen ratio of the soil
X_T	mean annual temperature ($^{\circ}C$)
X_R	mean annual radiation (J cm^{-2})
X_D	drought index (days), mean annual number of successive days without rain in the growing season
α	intercept of regression
β_i	regression coefficients
ε	residual variance

The regression function’s performance in comparison to the NPP values simulated with 4C ($Y_{NPP,4C}$) was assessed with

Table 2 Stand characteristics for the simulated mono-specific forest stands that were selected to represent the most productive stage of German forests (from age 15 to 70)

	Age (year)	d_{bh} (cm)	h_{dh} (m)	Stem number (ha^{-1})
Oak	15	2.2	11.0	5734
Pine	15	4.6	7.7	6664
Spruce	15	3.3	6.0	7000
Beech	15	2.7	6.5	6463

d_{bh} mean diameter at breast height of tree, h_{dh} dominant height of tree stand

the root mean squared error (RMSE), the relative root mean squared error (RMSE%), the absolute bias (BIAS), the relative bias (BIAS%) and the coefficient of determination (R^2) (see OER Table 1). The distributions of the residuals were also examined. The derivation of the regression functions was conducted with the statistical software R (R Core Team 2015).

2.3 Validation and plausibility of SRM results for NPP

To evaluate the SRMs for each tree species, we split the $Y_{NPP,4C}$ values into two equally-sized groups (model construction group, model validation group; Vanclay and Skovsgaard 1997). The climate stations were sorted by latitude, and every second station was reserved for the model validation group. Thus, both datasets cover the entire study area (i.e. Germany) homogeneously. The $Y_{NPP,4C}$ values and their corresponding independent variables of the model construction group were used to fit the regression function (Eq. 1). The resulting functions were then run with the independent variables of the model validation group to predict the $Y_{NPP,SRM}$ values, and these were compared with simulated $Y_{NPP,4C}$ values of the validation group. It is important to note that the SRMs used for the climate change analysis were fitted to the full dataset as described above (see section 2.3) because validation is not possible for the future data.

To check the plausibility of the SRMs, we compared the $Y_{NPP,SRM}$ values for the 1951–2006 period with two different observed NPP datasets by Luysaert et al. (2007) and Pretzsch (2009). These two datasets (vegetation measurements from flux tower sites and forest inventory data, respectively) provide independent estimates of the annual NPP of the four forest species used in this study.

2.4 Experiments and analyses with the SRMs

The SRMs were then run with the seven climate scenarios and the corresponding realizations covering a temperature increase from 0 to 3 K until 2060. For each climate scenario consisting of 50 realizations, the average annual NPP of the four tree species, every climate station and the four different soil types

was calculated with the SRMs. The variables required to drive the SRMs (see Eq. 1) were averaged for a 30-year time period from 2031 to 2060.

For the spatial comparison of NPP, the results have been aggregated by calculating the mean average annual NPP of the climate stations belonging to the same forest eco-region (Wolff 2002; OER Fig. 2). To test the sensitivity of the SRMs to each climate factor (temperature, radiation, drought index), we varied each factor from -40 to $+100$ % around the total average value across all 2342 climate stations. This range covered the expected range of factors encountered in the studied area well. Only one factor was altered while the other factors were kept constant at their respective averages.

3 Results

3.1 Model evaluation

The parameters of the four species-specific SRMs are significantly different from zero and have the same order of magnitude across the different species (OER Table 2). The SRMs fit quite well to the estimates of the average annual NPP as simulated with 4C (Table 3). The values of NPP calculated by the SRMs are nearly unbiased, with a relative bias (BIAS%) less than 1 %. The high R^2 indicates that the SRMs work equally well for all four tree species. Because the SRMs differ only slightly in their statistical characteristics for the four tree species, we show results only for Scots pine if not indicated otherwise.

Although the *Bartlett*-test of homogeneity of the residual variance detects significant differences between the soil groups, a visual analysis of the distribution of the data per soil group does not show a clear trend in the distribution of the residuals (OER Fig. 3). The two data clouds result from the rich and poor soils (OER Fig. 3 (left), see Table 1).

The validation results from the partitioned datasets also prove the suitability of the SRMs. The coefficient of

Table 3 Absolute and relative root mean square error (RMSE and RMSE%) and bias (BIAS and BIAS%) as well as coefficient of determination (R^2_{adj}) of the average annual NPP ($t\ C\ ha^{-1}\ year^{-1}$) calculated with the SRMs compared to 4C at the forest stand level

	Beech	Oak	Pine	Spruce
RMSE	0.290	0.256	0.248	0.203
RMSE%	5.054	4.662	3.433	3.964
BIAS	0.005	0.001	0.002	0.003
BIAS%	0.080	0.016	0.033	0.050
R^2_{adj}	0.978	0.984	0.987	0.988
N	18,720	18,720	18,720	18,720

Table 4 Results of the linear regression between NPP estimated with the static reduced models ($Y_{NPP,SRM}$) and the corresponding results of NPP estimated with 4C ($Y_{NPP,4C}$)

	Beech	Oak	Pine	Spruce
Equation $Y_{NPP,4C}$	$0.98Y_{NPP}-0.2$	$1.03Y_{NPP}-0.3$	$0.98Y_{NPP}-0.08$	$0.96Y_{NPP}-0.01$
R^2	0.98	0.99	0.99	0.99

determination (R^2) for the linear regression between the NPP calculated with the SRM ($Y_{NPP,SRM}$) for the model validation group and the corresponding values of NPP from 4C ($Y_{NPP,4C}$) (OER Fig. 4) is very high for all tree species (Table 4). However, for all tree species, the intercept is different from zero and the slope differs from 1 (Table 4), meaning that the relationship between the 4C-modelled and the SRM-calculated NPP differs from the perfect 1:1 line. Nevertheless, the discrepancy between the $Y_{NPP,SRM}$ and $Y_{NPP,4C}$ values is small.

The sensitivity analysis shows that the polynomial terms in the SRM describe the nonlinear relationship between the NPP and the drought index well (OER Fig. 5C). Similarly, the assumed log-linear relationship between NPP and temperature is obvious in the simulated 4C results, whereas the dependence between radiation and NPP was almost linear in the 4C results (OER Fig. 5A-B).

3.2 Comparison of SRM results with observed data

The mean annual NPP values calculated with the SRMs for the four tree species vary from 3.5 to 9.1 t C ha⁻¹ year⁻¹ (Fig. 1). For Scots pine on poor soils (PL and PH, see Table 1), there is a good correspondence between the SRM and the values of Luyssaert et al. (2007) and Pretzsch (2009).

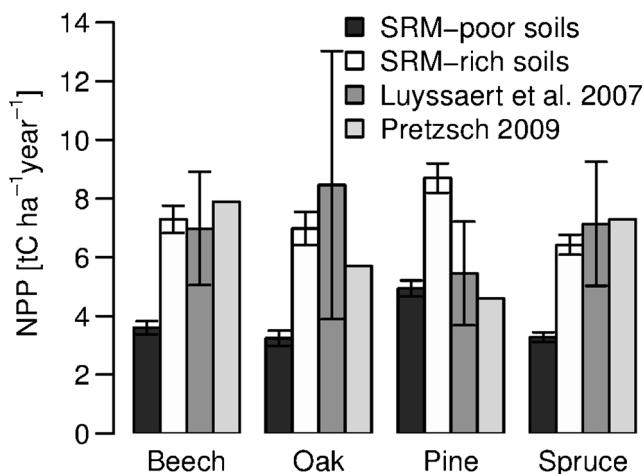


Fig. 1 Comparison of the NPP estimates for four tree species with static reduced models (SRM) on poor and rich soils and two independent observed NPP datasets by Luyssaert et al. (2007) and Pretzsch (2009). The SRM values are averaged over two types of plant available water (high and low). Error bars indicate the single standard deviation within the NPP data of one tree species

For Scots pine on rich soils (RL and RH, see Table 1), the NPP values of the SRM are the highest compared with the other tree species and larger than the observed data. Norway spruce, oak and beech show a good correspondence between SRM values and observed data for rich soil conditions. In contrast, the NPP values of the SRMs for beech, oak and Norway spruce on poor soil conditions are lower in comparison with the two measured datasets (Fig. 1).

3.3 Overall characteristics of the net primary production calculated with the SRMs

The tree species with the highest NPP calculated by the SRMs over the 2342 climate stations is Scots pine. The NPP of all four tree species on rich soils is around twice the NPP of the poor soils (see OER Table 3). There is almost no difference in terms of NPP between the soils with low and high plant available water (OER Table 3). The sensitivity of NPP to climate factors is most pronounced on rich soils. The range between the minimum and the maximum NPP of a forest stand at a specific climate station is smaller on poor soils (OER Table 3). There is an overall trend of rising NPP with increasing temperature for the analyzed stands. However, there are differences between the tree species; oak stands benefit the most while spruce stands benefit the least from temperature increase (OER Table 3).

The median of the range within a climate scenario alternates around 0.5 t C ha⁻¹ year⁻¹ for all tree species. There is a slight increasing trend of the median of the range with rising temperature. Also, the 1.5-fold interquartile distance of the ranges increases with temperature increase, except for Norway spruce (Fig. 2).

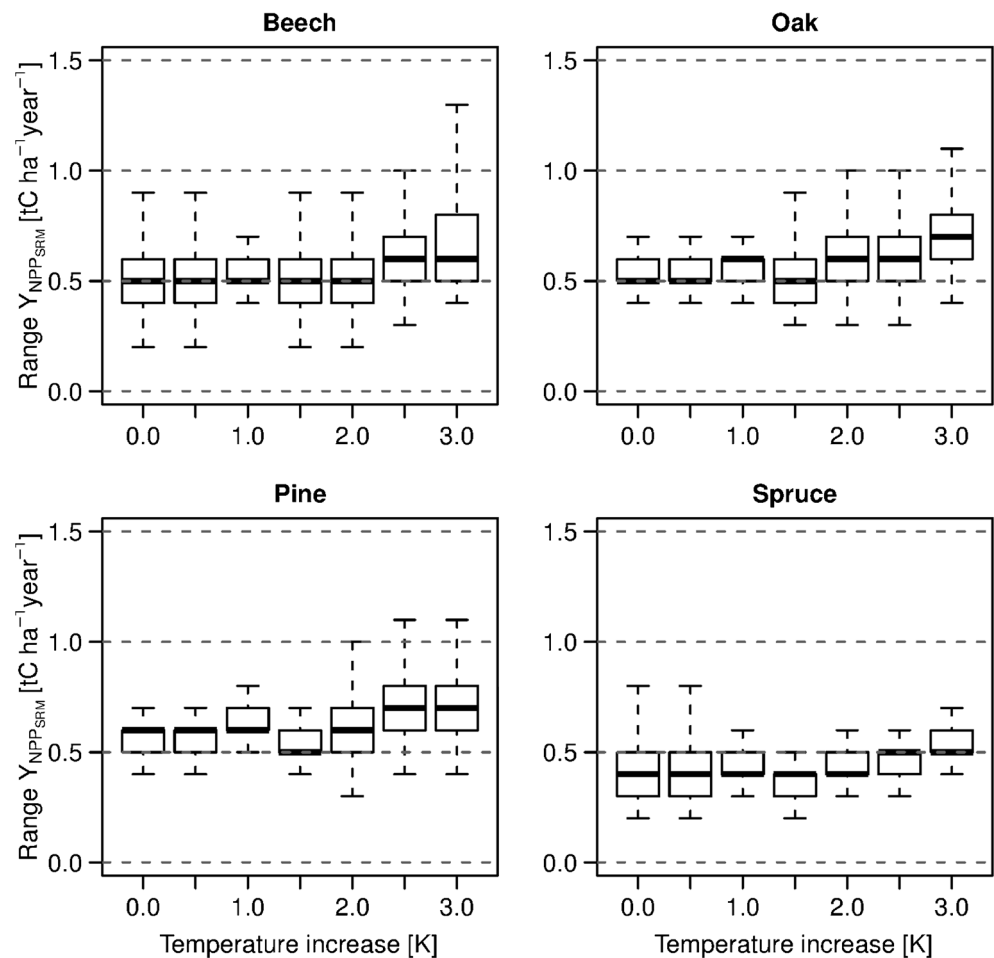
3.4 Regional specific analyses for Scots pine

The positive effects of climate change are stronger in mountainous regions (Fig. 3; OER Fig. 6). The lowest NPP increase for Scots pine occurs in the eastern low lands, where the median of 50 realizations with the 3 K scenario varies from 14–16 % (Fig. 3), and the uncertainty range across realizations is largest (Fig. 3; OER Fig. 6). The maximum spread over the 50 realizations of the 3 K climate scenario is 1.6 t C ha⁻¹ year⁻¹ in the north-eastern forest eco-regions of Germany. In the southern, western and mountainous eco-regions, different realizations lead to a spread of only 0.8 t C ha⁻¹ year⁻¹ with the least variation occurring in mountainous regions (Fig. 3).

4 Discussion

This paper shows that SRMs, derived from a complex process-based model, provide meaningful estimates of NPP. The independent variables used in these SRMs to describe the

Fig. 2 Box plots showing the range (maximum–minimum) of NPP ($Y_{\text{NPP,SRM}}$) estimated by the static reduced models (SRM) for different levels of temperature increase. Each box plot represents the 50 realizations of every climate scenario for four tree species and for site conditions of a rich soil with low water availability (RL). Boxes show the lower quartile (25 %), the median (50 %) and the upper quartile (75 %). The whiskers represent the 1.5-fold interquartile range



site and climatic conditions can be derived easily from basic soil databases or from meteorological data. This allows for a broad application of the SRMs as opposed to complex process-based models such as 4C that usually require a much more detailed soil and climate description (Fontes et al. 2010), which are not commonly available for most sites. The application of these SRMs shows that productivity under climate change increases over a wide range of climate scenarios in Germany. Below, we discuss the model evaluation, the climate change impact analysis and the general applicability of SRMs in more detail.

4.1 Model evaluation

The goodness of fit between simulated NPP with 4C and simulated NPP with the SRMs was very high for all tree species considered in this analysis. Despite some degree of subjectivity in choosing the soil variables (but spanning a wide range of conditions from poor to rich soils), regressions for each soil type provide high goodness of fit (results not shown). However, a more balanced data selection using soil data that covers all German soil types (currently no accessible datasets available) could change this tight relationship and introduce more

variation. Regarding the problem of variance homogeneity of the residuals, we followed the suggestions of Zuur et al. (2010, see their page 6) to also add a visual analysis of significance (c.f. OER Fig. 3). In general, we note that the results of classical significance tests have to be interpreted with care when applied to large modelled datasets because they easily lead to significant differences between groups simply due to the large number of data points. For example, a randomly chosen subsample of NPP values ($n=200$) showed no significant difference ($p>0.05$) between different soil types even though the full dataset does show these differences (results not shown). Further studies, which compare different regression techniques, could enhance the robustness of SRMs and the understanding of underlying relationships.

The model bias and root mean square error were low, and there were only a few sites with deviations larger than $1 \text{ t C ha}^{-1} \text{ year}^{-1}$. The model validation did not reveal serious model errors (Table 4). However, the partitioning of the dataset into a model construction and a model validation dataset does not produce totally independent sets of data. Thus, the validation does not supply further information about model biases, which is also reported for other cross-validation methods (Vanclay and Skovsgaard 1997; Kozak and Kozak 2003).

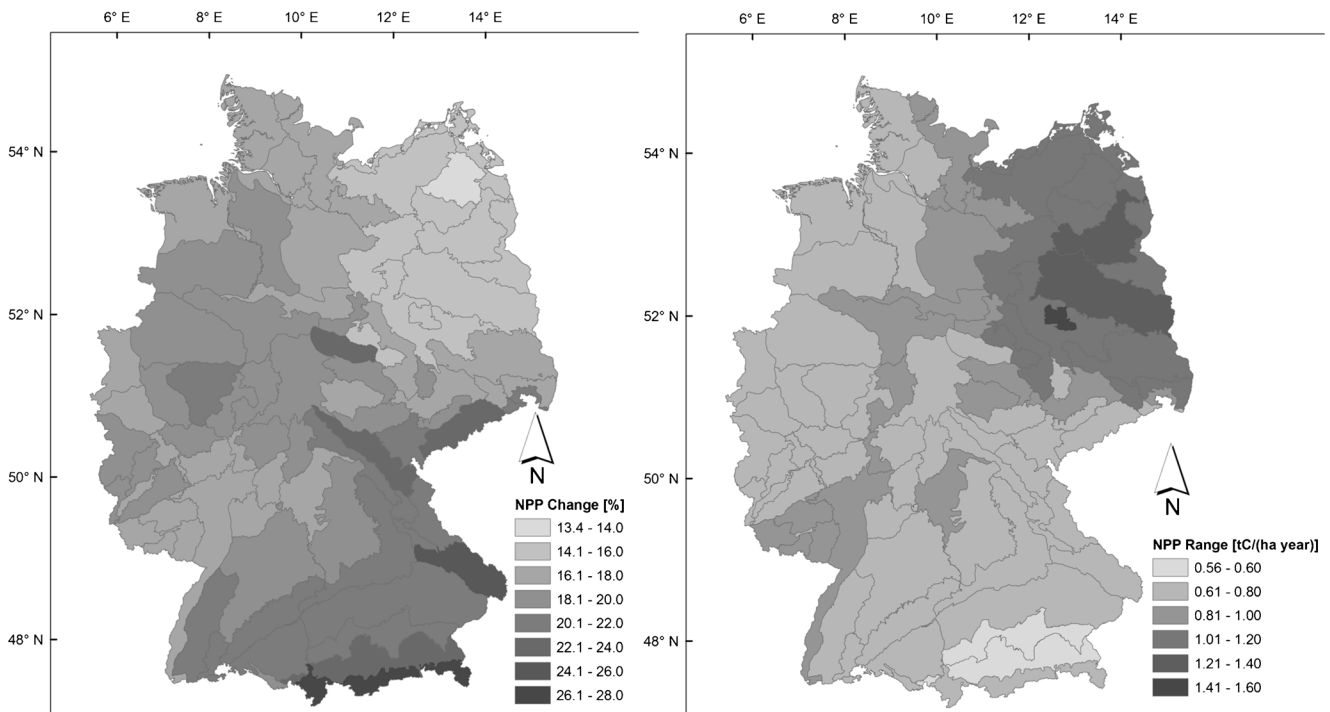


Fig. 3 Map of the interpolated, relative change in (left) and range of (right) NPP of Scots pine calculated with the static reduced model (SRM) on a typical rich soil with low water availability (RL) for 2031–2060 for the 3 K scenario. The relative change (left) refers to the

calculated NPP for 2031–2060 for the 0 K scenario in Germany. The range of NPP (right) is calculated for 2031–2060 over the 50 realizations. *Black lines* are the boundary lines for the forest eco-regions

The NPP simulated with 4C that has been used to derive the SRMs represents only young, productive stands (age 15–70). Therefore, the NPP values calculated with the SRMs are expected to be higher than NPP for middle-aged and old forest stands. The SRMs' average mean annual NPP of 3 to 10 t C ha⁻¹ year⁻¹ is in accordance with a simulation study by Oene et al. (2000). They calculated the NPP over a broad spectrum of Norway spruce and beech forest stands in Europe ranging from 3 to 12 t C ha⁻¹ year⁻¹. Jochheim et al. (2009) simulated NPP values between 5 and 7 t C ha⁻¹ year⁻¹ for Scots pine, beech and Norway spruce forests on nine level II plots in Germany. Similarly to our results, they found that the differences in NPP between sites are larger than between the different tree species.

In comparison with the datasets of Luysaert et al. (2007) and Pretzsch (2009), our NPP values for Scots pine seem too high (Fig. 1). However, Scots pine stands are very often situated on less productive sites in Germany. Therefore, the SRM values for the poor sites probably represent the actual species distribution more realistically and in fact fit much better to the observed data. In the case of beech, oak and Norway spruce, the SRM values on rich sites are within the range of the two datasets. For beech and oak, this again reflects the realistic distribution of these species on forest soils. For Norway spruce, the situation is more complicated: in comparison to Scots pine, Norway spruce shows generally lower NPP values calculated with the SRMs, which is not consistent with the

observed data. This cannot be explained by site-specific characteristics. Here, a systematic bias in the simulated NPP with 4C could be present and requires further investigation. This highlights the use of the SRMs as a tool to guide further development of process-based models.

The inclusion of the C/N ratio as a measure for nitrogen availability reflects a relationship between nitrogen availability and NPP, which was also found by Oene et al. (2000). The model 4C correctly simulated reduced tree productivity on poor soils mainly due to nitrogen limitation. In such situations, NPP is low and other factors such as the water supply have a small influence on the average annual NPP. On rich soils, no nitrogen limitation is simulated and climate factors such as temperature and precipitation have higher impacts on NPP, which explains the greater range of NPP values on these soils. In spite of its weak influence in this study, plant available water is usually considered an important site factor for tree productivity under climate change (Kellomaki and Wang 2000). One reason explaining our findings could be a low level of simulated drought stress from low soil water contents in 4C. On the other hand, a strong negative correlation between the drought index greater than four and NPP for water-limited sites could be detected according to the SRMs (OER Fig. 5C). Small changes in the precipitation sum or the drought index lead to stronger responses of the SRMs in water-limited regions such as described in other studies (Gerten et al. 2008; Sang and Su 2008). The positive

relationship between mean annual temperature and NPP in the SRMs is in good agreement with investigations in temperate forests (Boisvenue and Running 2006; Oene et al. 2000; Piao et al. 2009).

4.2 Climate impact analysis

For Germany, the average annual NPP calculated with the SRMs increases until 2060 for all climate scenarios and tree species. This result supports the findings of other studies that the terrestrial carbon sinks should peak around 2050 (Piao et al. 2009; Sitch et al. 2008). In Germany, two main regions could be identified where climate change impacts on NPP are most pronounced. Firstly, in mountainous regions, the temperature increase will strongly enhance the NPP. A maximum value of 28 % increase of NPP for Scots pine forest stands has been calculated by the SRM for the 3 K scenario (Fig. 3; OER Fig. 6). In comparison, Oene et al. (2000) simulated a maximum increase in NPP of 11 %, when only considering the direct effects of a temperature increase of 2 K. The different precipitation patterns of the scenario realizations only slightly affect the NPP in regions with high precipitation. Secondly, in the north-eastern lowlands of Germany, the potential temperature-induced increase of NPP is limited by the drought duration during the growing period. Here, the change in NPP under the 3 K scenario is accompanied by a higher uncertainty range of the scenario (Fig. 3; OER Fig. 6). The outcomes of temperature increase are site-specific. Considering the effect of temperature increase on plant water relations, a rising transpiration demand results in a more frequent stomata closure on water-limited sites, which reduces the NPP. Thus, future precipitation patterns in the realizations are particularly relevant in regions with low annual precipitation sums today. It is also important to note that these simulations do not consider the effects of changing disturbances and other extreme events that can potentially negate productivity increases driven by average temperature increase alone (Reyer et al. 2013, 2015; Seidl et al. 2014).

4.3 Applicability of the SRMs

Our analyses highlight the advantage of the SRMs as a tool for fast analyses of large sets of climate scenarios for a large number of sites. A simulation experiment with a set of 2342 sites, each with four soil types and four tree species, with 50 realizations of seven climate scenarios for a simulation period of 30 years with the model 4C would require 150 days of CPU time, whereas running the SRMs takes 30 min. Furthermore, as shown above, there is less demand for input data for model application compared to a process-based model like 4C. The SRMs need only two site characteristics, which can be derived from soil databases and long-term mean values of climate data. Applying the SRMs helps to detect regions that face high

climate impacts, and subsequently allow for an efficient investigation of vulnerable forest stands with more complex process-based models or targeted measurements. Thus, the SRMs do not substitute analyses with process-based models, but rather complement them.

Moreover, on a regional scale, SRMs can be used to support current studies of future forest carbon sequestration that consider stand development and management but often neglect climate change impacts (Krug et al. 2009). While in the short-term, an omission of climate effects may be acceptable because, in most cases, forest management strategies imply stronger impacts on the carbon balance of a forest than climate change, in the longer-term, this assumption of stationary conditions does not hold (Jandl et al. 2007; Gutsch et al. 2011). Therefore, for more comprehensive analyses, SRM can be coupled to forest growth models that are not climate sensitive and thus far have relied on simple scaling of growth functions to capture climate change effects (e.g. Eggers et al. 2008; Schelhaas et al. 2015).

On a global scale, our approach to derive the SRMs and their low data requirements can be used to derive simple impact functions from multi-model runs of dynamic global vegetation models. These global or continental-scale SRMs (e.g. Sallaba et al. 2015) are of increasing interest for cross-sectoral studies as part of integrated assessment model studies to assess the costs and damages from climate change at the macroeconomic level.

To increase the robustness of such analyses, integrating measurement data directly in the derivation of SRMs would be an option instead of deriving models only using simulation data. However, thus far the low variation in site conditions and species at the flux sites in Germany, the comparably short time periods covered, and limited data availability hamper such an approach. Other data sources, such as the National Forest Inventory in Germany, could provide valuable data. However, repeated measurements would be required, which are not available for all federal states in a standardized format.

There are also disadvantages to the use of SRMs. They do not allow the analysis of annual or seasonal cycles of NPP or of productivity in mixed stands. Also, questions related to reduced productivity due to biotic disturbances or extreme events in forest stands cannot be answered.

5 Conclusions

Static reduced models (SRM) can be used as simple impact functions to determine regional impacts on the productivity of mono-specific forest stands under climate change. The SRMs permit a spatial overview of general trends in productivity change under climate change, and thus help to identify regions which are likely to be most vulnerable.

Running the SRMs with a broad range of temperature increases, and different precipitation levels reveal a generally positive impact on the productivity of the four main tree species in Germany. The NPP increases for all tree species with rising temperatures, especially on temperature-limited sites such as mountainous regions or on sites with high precipitation. However, beyond a 2 K temperature increase by 2060, the uncertainty of NPP on water-limited forest sites also increases. Here, the results indicate a high risk of decreasing NPP for Norway spruce, Scots pine and beech. The NPP of oak on these sites is less sensitive. These rapid analyses can inform and thus complement more detailed analyses with process-based models. Especially, for water-limited or constrained sites, we recommend further analysis of climate impacts with process-based forest growth models or detailed measurements. Further research should focus on an improved validation of the SRMs with regard to soil factors and forest productivity.

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Compliance with ethical standards

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