



Future global mortality from changes in air pollution attributable to climate change

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1 **FUTURE GLOBAL MORTALITY FROM CHANGES IN AIR POLLUTION**
2 **ATTRIBUTABLE TO CLIMATE CHANGE**

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40 Ground-level ozone and fine particulate matter (PM_{2.5}) are associated with premature human
41 mortality¹⁻⁴; their future concentrations depend on changes in emissions, which dominate the
42 near-term⁵, and on climate change^{6,7}. Previous global studies of the air quality-related health
43 effects of future climate change^{8,9} used single atmospheric models. However, in related studies,
44 mortality results differ among models¹⁰⁻¹². Here we use an ensemble of global chemistry-climate
45 models¹³ to show that premature mortality from changes in air pollution attributable to climate
46 change, under the high greenhouse gas scenario RCP8.5¹⁴, is likely positive. We estimate 3,340
47 (-30,300 to 47,100) ozone-related deaths in 2030, relative to 2000 climate, and 43,600 (-195,000
48 to 237,000) in 2100 (14% of the increase in global ozone-related mortality). For PM_{2.5}, we
49 estimate 55,600 (-34,300 to 164,000) deaths in 2030 and 215,000 (-76,100 to 595,000) in 2100
50 (countering by 16% the global decrease in PM_{2.5}-related mortality). Premature mortality
51 attributable to climate change is estimated to be positive in all regions except Africa, and is
52 greatest in India and East Asia. Most individual models yield increased mortality from climate
53 change, but some yield decreases, suggesting caution in interpreting results from a single model.
54 Climate change mitigation will likely reduce air pollution-related mortality.

55 Climate change can affect air quality through several pathways, including changes in the
56 ventilation and dilution of air pollutants, photochemical reaction rates, removal processes,
57 stratosphere–troposphere exchange of ozone, wildfires, and natural biogenic and lightning
58 emissions^{6,7}. Overall, changes in these processes are expected to increase ozone in polluted
59 regions during the warm season, especially in urban areas and during pollution episodes, but
60 decrease ozone in remote regions due to greater water vapour concentrations leading to greater
61 ozone destruction. These effects are exacerbated by the greater decomposition of reservoir
62 species such as PAN⁷. PM_{2.5} will also be affected by climate change, but impacts vary in sign

63 among models and show regional variation related to differences in precipitation, wildfires,
64 biogenic emissions, PM_{2.5} composition, and other factors.

65 Previous studies have examined the impact of future climate change on human health via air
66 quality globally^{8-9,15}, in the US^{10, 16-20}, and in Europe²¹. However, only two studies have
67 previously used an ensemble of models to assess air pollution-related mortality attributable to
68 climate change: one for the US¹⁰, and our previous global work with the same ensemble used
69 here, but evaluating the effects of historical climate change prior to 2000¹¹. Both studies found a
70 large spread of mortality outcomes depending on the atmospheric model used. Silva et al.¹¹
71 found that the multi-model average suggested a small detrimental effect of climate change on
72 global present-day air pollution-related mortality, but individual models yielded estimates of
73 opposing sign.

74 The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP) ensemble
75 (Supplementary Table 1) simulated air quality in 2000, and in 2030, 2050 and 2100 for the four
76 global Representative Concentration Pathway scenarios (RCPs)²². We previously estimated
77 future air pollution premature mortality under all four RCP scenarios, estimating the net effect of
78 both emissions changes and climate change¹². Under RCP8.5, ozone concentrations increase in
79 most locations in 2100 relative to 2000, due to increases in methane emissions and the effect of
80 climate change^{7,23}, but PM_{2.5} decreases in 2100 due to a projected decrease in particulate and
81 precursor emissions²⁴. These changes in pollutant concentrations lead to 316,000 (95% C.I.: -
82 187,000 to 1.38 million) ozone-related excess deaths yr⁻¹ and -1.31 (-2.04 to -0.17) million
83 PM_{2.5}-related (avoided) deaths yr⁻¹ in 2100¹². Here we present results from additional ACCMIP
84 simulations that were designed to isolate the influences of future climate change under RCP8.5,
85 by simulating the projected climates of 2030 and 2100 (imposed by prescribing sea-surface

86 temperatures, sea ice cover, and greenhouse gas concentrations for radiation) together with air
87 pollutant emissions from 2000. The effects of climate change are then isolated by a difference
88 with historical 2000 simulations. Premature mortality attributable to RCP8.5 climate change is
89 estimated following the methods of Silva et al.¹², including projected population and baseline
90 mortality rates (see Methods), such that mortality estimates here can be compared directly with
91 overall changes in air pollution-related mortality in RCP8.5.

92 We estimate that global ozone mortality attributable to RCP8.5 climate change will be 3,340 (-
93 30,300 to 47,100) deaths yr⁻¹ in 2030 and 43,600 (-195,000 to 237,000) deaths yr⁻¹ in 2100
94 (Figures 1a and 2a). In 2100, ozone mortality increases in most regions, especially in highly
95 populated and highly polluted areas, with marked spatial differences within regions that include
96 both positive and negative mortality changes (Figure 3a, Supplementary Table 2, Supplementary
97 Figures 1 and 2a). The effect on ozone mortality in 2100 is greatest in East Asia (45,600 deaths
98 yr⁻¹, 41 deaths yr⁻¹ per million people), India (16,000 deaths yr⁻¹, 8 deaths yr⁻¹ per million people)
99 and North America (9,830 deaths yr⁻¹, 13 deaths yr⁻¹ per million people), but some areas within
100 these and other regions show decreases in mortality. East Asia has high mortality effects per
101 person in part because of its higher projected mortality rate from respiratory diseases. Climate
102 change contributes 14% of the overall increase in ozone mortality estimated for RCP8.5 in 2100
103 relative to 2000¹². However, three of 8 models in 2030 and three of 9 in 2100 show global
104 decreases in ozone mortality due to climate change. For each model, the uncertainty range does
105 not include zero; only the spread of models causes the overall uncertainty to span zero.
106 Uncertainty in modeled ozone concentrations contributes over 97% to the overall uncertainty in
107 both 2030 and 2100, with the remainder from uncertainties in relative risk (RR). Results from a
108 sensitivity analysis using present-day population and baseline mortality rates (Table 1) show

109 32% and 67% lower mortality estimates in 2030 and 2100, respectively, largely because the
110 projected baseline mortality rates of chronic respiratory diseases increase through 2100. The
111 models agree that ozone will increase due to climate change in some polluted regions, notably
112 the northeast US as found in other studies⁶ and decrease in the tropics over the oceans
113 (Supplementary Figures 3 and 4a). These changes are consistent with those analysed by Schnell
114 et al.²⁵ for 2100, using four of these same models, and were attributed to a greater efficiency of
115 precursor emissions to generate surface ozone in polluted regions, along with reductions in the
116 export of precursors to downwind regions.

117 The impact of climate change on PM_{2.5} mortality is estimated to result in 55,600 (-34,300 to
118 164,000) deaths yr⁻¹ in 2030 and 215,000 (-76,100 to 595,000) deaths yr⁻¹ in 2100 (Figures 1b
119 and 2b). Mean estimates of PM_{2.5} mortality increase in 2100 in all regions except Africa (-25,200
120 deaths yr⁻¹) (Figure 3b, Supplementary Table 3, Supplementary Figure 2b). The greatest
121 increases in mortality in 2100 occur in India (80,200 deaths yr⁻¹, 40 deaths yr⁻¹ per million
122 people), Middle East (50,400 deaths yr⁻¹, 45 deaths yr⁻¹ per million people) and East Asia
123 (47,200 deaths yr⁻¹, 43 deaths yr⁻¹ per million people), although the Former Soviet Union shows
124 greater mortality per million people in 2100 (11,800 deaths yr⁻¹, 57 deaths yr⁻¹ per million
125 people). Similar to ozone mortality, there are substantial spatial differences within each region,
126 including both increases and decreases in mortality. For PM_{2.5}, a large decrease in mortality is
127 projected in RCP8.5 relative to 2000 (when accounting for changes in both emissions and
128 climate)¹², but climate change alone increases mortality, partially counteracting the decrease
129 associated with declining emissions in RCP8.5. Without climate change, the decrease in PM_{2.5}-
130 related mortality would be roughly 16% greater in 2100 relative to 2000. Propagating
131 uncertainty in RR to the mortality estimates leads to coefficients of variation (CVs) of 8-31%

132 (2030) and 11-46% (2100) for the different models, but the spread of model results increases
133 overall CVs to 123% in 2030 and 106% in 2100. In both years, one model (GISS-E2-R) yields a
134 decrease in global mortality from climate change while the other three (2030) or four (2100)
135 show an increase. Uncertainty in modeled $PM_{2.5}$ concentrations in 2000 makes a similar
136 contribution to the overall uncertainty (50% in 2030 and 52% in 2100) compared with
137 uncertainty in modeled $PM_{2.5}$ concentrations in future years (50% in 2030, 48% in 2100).
138 Uncertainty in RR makes a negligible contribution in both periods (<1%), as the multi-model
139 mean is small and different models disagree on the sign of the influence. Considering present-
140 day population and baseline mortality rates (Table 1), we estimate 23% and 33% lower mortality
141 in 2030 and 2100, respectively, mostly associated with the increase in projected baseline
142 mortality rates through 2100.

143 $PM_{2.5}$ -related mortality was estimated above for the sum of $PM_{2.5}$ species reported by five
144 models, using a common formula (see Methods), to increase the number of models considered
145 and to increase consistency among $PM_{2.5}$ estimates. Additionally, we present a sensitivity
146 analysis considering the $PM_{2.5}$ concentrations reported by four models using their own $PM_{2.5}$
147 formulas, for which multi-model average mortality results are modestly higher: 15% greater in
148 2030 and 12% in 2100 (Supplementary Figure 5). The degree of agreement between the two
149 estimates varies among the four models, and for one model (GISS-E2-R) the two sources of
150 $PM_{2.5}$ estimates yield impacts of different sign in 2030.

151 There is considerable agreement among models regarding the increase in $PM_{2.5}$ concentrations in
152 many locations in 2100, including most polluted regions, due to RCP8.5 climate change
153 (Supplementary Figure 4b). Allen et al.²⁶ analysed four of these same models in 2100 and found
154 that global average surface $PM_{2.5}$ concentrations increased due to climate change, reflecting

155 increases in nearly all relevant species for each model. They attributed this increase in $PM_{2.5}$
156 mainly to a decrease in wet deposition associated with less large-scale precipitation over land.
157 Our multi-model mean estimates of global population-weighted changes for $PM_{2.5}$ and individual
158 species (Supplementary Table 4; Supplementary Figure 6) are similar to those of Allen et al.²⁶.
159 Unlike Allen et al.²⁶, however, GISS-E2-R shows a net decrease in global population-weighted
160 concentrations of total $PM_{2.5}$ and of each $PM_{2.5}$ species except sea salt, in 2100, likely due to
161 projected concentration decreases over densely-populated eastern China. Models also differ
162 strongly in the sign and magnitude of changes in dust, particularly over North Africa and the
163 Middle East; HadGEM2 projects increases in $PM_{2.5}$ for all species except dust, but a strong
164 decrease in dust over the Middle East and South Asia. In Africa, the decrease in $PM_{2.5}$ near the
165 equator is likely caused by increased precipitation, whereas $PM_{2.5}$ increases are associated with
166 precipitation decreases in Southern Africa²⁶. Differences in $PM_{2.5}$ (and ozone) responses to
167 climate change among models likely result from differences in large-scale meteorological
168 changes, and different treatments of atmospheric chemistry and feedback processes among the
169 models (such as the response of dust to climate change).

170 In the US, our multi-model mean mortality estimates for the impact of RCP8.5 climate change
171 for ozone (1,130 deaths yr^{-1} in 2030; 8,810 deaths yr^{-1} in 2100) compare well with those of Fann
172 et al.²⁰, who report 420 to 1900 ozone-related deaths yr^{-1} for RCP8.5 climate change in 2030,
173 despite differences in concentration-response functions and population and baseline mortality
174 projections. These results for ozone and those for $PM_{2.5}$ (6,900 deaths yr^{-1} in 2030; 19,400 deaths
175 yr^{-1} in 2100) are also consistent with the increases in mortality and spatial heterogeneity
176 attributed to climate change in 2050 by Bell et al.¹⁶ for ozone and Tagaris et al.¹⁷ for ozone and
177 $PM_{2.5}$, although these studies used different climate change scenarios besides other

178 methodological differences. Across models, our estimates for ozone mortality in the US vary
179 between -435 and 4,750 deaths yr⁻¹ in 2030 and between -1,820 and 27,012 deaths yr⁻¹ in 2100.
180 This spread of model results, with a few models suggesting avoided mortality due to climate
181 change, is similar to that of Post et al.¹⁰ (-600 to 2,500 deaths yr⁻¹ in 2050) using SRES scenarios
182 of GHG emissions. Similarly, results show spatial heterogeneity within several regions (Figure
183 2) that is similar to Post et al.¹⁰ for the US and Orru et al.²¹ for Europe.

184 The spread of results among models highlights the uncertainty in the effect of climate change on
185 air quality. Further improvements in chemistry climate models are needed to better model the
186 interaction and feedbacks between climate and air quality, including the sensitivity of biogenic
187 emissions to climate change, the effects of meteorological changes on air quality (e.g., aerosol-
188 cloud interactions, secondary aerosol formation, wet deposition, and gas-aerosol partitioning),
189 and the impact of climate change on wildfires. Stratosphere-troposphere exchange of ozone is
190 also important, as is the impact of land use changes on regional climate and air pollution. Our
191 results are specific to climate change as projected under RCP8.5 and would differ for other
192 scenarios. We estimate the effect of climate change as the difference between simulations with
193 future climate and year 2000 climate, both with year 2000 emissions, although global emissions
194 of PM_{2.5} and its main precursors decrease under RCP8.5. Had we instead modelled future
195 emissions with present vs. future climate, we would likely have attributed smaller changes in air
196 pollution and mortality to climate change, given the projected emission reductions. Whereas the
197 net effect of missing and uncertain processes does not clearly indicate an under- or overestimate
198 for the effect of climate change on air quality, we likely underestimate the magnitude of the
199 health impact by omitting mortality for people under 25, and morbidity effects. We also neglect
200 possible synergistic effects of a warmer climate to modify air pollution-mortality relationships.

201 Although a few studies have suggested stronger relationships between ozone²⁷ and PM_{2.5}²⁸ and
202 health at higher temperatures, there is insufficient evidence to include those effects here.

203 Despite these uncertainties, this study is the first to use a multi-model ensemble to show that
204 global air pollution-related mortality attributable to climate change is likely positive. The spread
205 of results among models within the ensemble, including differences in the sign of global and
206 regional mortality estimates, suggests that results from studies using a single model and a small
207 number of model years should be interpreted cautiously. Actions to mitigate climate change,
208 such as reductions in long-lived GHG emissions, will likely benefit human health by reducing
209 the effect of climate change on air quality in many locations. These health benefits are likely to
210 be smaller than those from reducing co-emitted air pollutants²⁹, but both types of health benefits
211 via changes in air quality would add to reductions in many other influences of climate change on
212 human health³⁰.

213

214 **Additional information**

215 Supplementary information is available in the online version of the paper.

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316

317 **Author contributions:** JJW, JFL, DTS and RAS conceived the study. All other co-authors
318 conducted the model simulations. RAS processed model output and estimated human mortality.
319 RAS and JJW analyzed results. RAS and JJW prepared the manuscript and all co-authors
320 commented on it.

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322 financial interests.

323

324

325 **Figure Legends:**

326 Figure 1 – Impact of RCP8.5 climate change on global mortality for individual models and the
327 multi-model average. Estimates are for 2030 and 2100 for (a) ozone respiratory mortality (9
328 models) and (b) PM2.5 IHD+STROKE+COPD+LC mortality (5 models). PM2.5 is calculated as
329 a sum of species. Uncertainty for each model is the 95% CI taking into account uncertainty in
330 RR. Uncertainty for the multi-model average is the 95% CI including uncertainty in RR and
331 across models.

332
333 Figure 2 – Geographical impact of climate change on mortality. Estimates are for 2030 and 2100
334 for (a) ozone respiratory mortality and (b) PM2.5 IHD+STROKE+COPD+LC mortality,
335 showing the multi-model average in each 0.5°x0.5° grid cell. PM2.5 is calculated as a sum of
336 species.

337
338 Figure 3 – Projected mortality for ten world regions. Estimates are for 2030 and 2100 for (a)
339 ozone respiratory mortality and (b) PM2.5 IHD+STROKE+COPD+LC mortality, showing the
340 multi-model regional average. PM2.5 is calculated as a sum of species. Uncertainty for the multi-
341 model regional average is the 95% CI including uncertainty in RR and across models. World
342 regions are shown in Supplementary Figure 1.

343

344

345 Table 1 – Sensitivity analysis for changes in global air pollution-related mortality attributable to
346 climate change. Estimates are for multi-model averages (deaths yr⁻¹) for the deterministic
347 results.

348

	PM _{2.5} -related mortality		Ozone-related mortality	
	2030	2100	2030	2100
Base results	56,300	218,000	10,700	128,000
PM _{2.5} using Krewski et al. ²	66,200	318,000	--	--
Present-day (2011) population	35,500	93,800	2,970	59,400
Present-day (2010) baseline mortality rates	69,600	510,000	2,790	13,300
Present-day population and baseline mortality rates	43,300	144,000	2,300	14,500

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353 **Methods**

354 The Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP)¹³ included
355 contributions from 14 modelling groups, of which 9 completed simulations that are used here
356 (Supplementary Table 1). ACCMIP models incorporate chemistry-climate interactions, including
357 mechanisms by which climate change affects ozone and PM_{2.5}, although models do not all
358 include the same interactions, and do not always agree on their net effects⁷. Of these nine, three
359 models are not truly coupled chemistry-climate models: MOCAGE is a chemical transport model
360 driven by external meteorology, and UM-CAM and STOC-HadAM3 do not model the feedback
361 of chemistry on climate¹³. As a result, these models do not fully capture the effects of changes in
362 air pollutant concentrations on processes that affect meteorology, such as through radiative
363 transfer and clouds. Prescribed anthropogenic and biomass burning emissions were very similar
364 for the different models, but they used different natural emissions (e.g. biogenic volatile organic
365 compounds, ocean emissions, soil and lightning NO_x)^{14,23}. Modelled 2000 concentrations show
366 good agreement with observations for ozone²³ and PM_{2.5}²⁴, although models tend to overestimate
367 ozone in the Northern Hemisphere and underestimate it in the Southern Hemisphere, and to
368 underestimate PM_{2.5}, particularly in East Asia.

369 We isolate the effect of climate change on air quality as the difference in concentrations between
370 ACCMIP simulations using year 2000 emissions together with future year climate, imposed by
371 prescribing RCP8.5³¹ sea surface temperatures, sea ice cover, and GHGs (for radiation) for 2030
372 and 2100 (referred to as “Em2000Cl2030” and “Em2000Cl2100”), and simulations with 2000
373 emissions and climate (“acchist2000”)¹³. We analyse results from the nine models reporting
374 ozone from the Em2000Cl2030/2100 simulations, and the five reporting PM_{2.5} (Supplementary
375 Table 1). Ozone and PM_{2.5} species surface concentrations from each model are calculated in each

376 grid cell, after regridding output from the native horizontal resolutions of each model (1.9°x1.2°
377 to 5°x5°) to a common 0.5°x0.5° resolution. To be consistent with the epidemiological studies
378 considered^{1,4}, we use the seasonal average of daily 1-hr maximum ozone concentrations for the
379 six consecutive months with highest concentrations in each grid cell, and annual average PM_{2.5}
380 concentration.

381 Seven of the nine models with Em2000Cl2030/2100 simulations reported both hourly and
382 monthly ozone concentrations, while two reported only monthly values. We calculate the ratio
383 of the 6-month average of daily 1-hr maximum concentrations to the annual average
384 concentrations, for each grid cell and each year, for those models that reported both hourly and
385 monthly concentrations; then, we apply that ratio to the annual average ozone concentrations for
386 the other two models, following Silva *et al.*^{11,12}.

387 We calculate PM_{2.5} concentration using the sum of PM_{2.5} species mass mixing ratios reported by
388 five models and a common formula:

$$389 \text{ PM}_{2.5} = \text{BC} + \text{OA} + \text{SO}_4 + \text{SOA} + \text{NH}_4 + 0.25 * \text{SS} + 0.1 * \text{Dust},$$

390 where BC – Black Carbon, OA – (Primary) Organic Aerosol corrected to include species other
391 than carbon, NH₄ – NH₄ in ammonium sulfate, SOA – Secondary Organic Aerosol, and SS –
392 Sea Salt, as had been done previously by Fiore *et al.*³³ and Silva *et al.*^{11,12}. The factors 0.25 and
393 0.1 are intended to approximate the fractions of sea salt and dust that are in the PM_{2.5} size range.
394 Nitrate was reported by three models, but we chose to omit nitrate from our PM_{2.5} formula to
395 avoid imposing changes inconsistent with the effect of climate change for other models,
396 following Silva *et al.*¹¹, although nitrate was included in estimates of total PM_{2.5} by Silva *et al.*¹².
397 Four of these models also reported their own estimate of PM_{2.5} (Supplementary Table 1).

398 The impact of climate change on global population-weighted differences (Em2000CI2030/2100
399 minus acchist2000) in PM_{2.5} and ozone concentrations for the different models are shown in
400 Supplementary Tables 4 and 5, respectively, while regional multi-model average differences are
401 shown in Supplementary Figures 7 and 8.

402 We estimate premature mortality by calculating the fraction of cause-specific mortality
403 attributable to long-term changes in pollutant concentrations, using methods that are identical to
404 those of Silva *et al.*¹², so that mortality attributable to climate change can be compared simply
405 with changes in mortality under the RCP scenarios. We use relative risks (RRs) from Jerrett *et*
406 *al.*¹ for ozone and respiratory diseases and Burnett *et al.*⁴ for PM_{2.5} and cardiopulmonary diseases
407 and lung cancer. Then, we apply that attributable fraction in each grid cell to future adult
408 population (age 25 and older) and baseline mortality rates based on projections from the
409 International Futures (IFs) integrated modelling system³². Using country-level projections per
410 age group, we mapped and gridded to the 0.5°x0.5° grid assuming that the present-day spatial
411 distribution of total population within each country is unchanged in the future, as well as the
412 present-day ratio of baseline mortality for the specific causes included in the epidemiological
413 studies and for three disease groups projected in IFs (chronic respiratory diseases, cardiovascular
414 diseases and malignant neoplasms). We select population projections from IFs instead of those
415 underlying RCP8.5 to ensure consistency between projections of population and baseline
416 mortality, since the latter are not available for RCP8.5, and for consistency with Silva *et al.*¹². IFs
417 projections of future total population are lower than those of RCP8.5 (-5% in 2030 and -27% in
418 2100) (Supplementary Figure 9). Had we used projections of population underlying RCP8.5, we
419 would have likely estimated greater changes in premature mortality relative to 2000. IFs
420 projections of baseline mortality rates reflect an aging population and regional demographic

421 changes, showing a steep rise in chronic respiratory diseases (roughly tripling globally by 2100),
422 particularly in East Asia and India, some regional increases in cardiovascular diseases (e.g.
423 Middle East, Africa), and global decreases in lung cancer.

424 Overall uncertainty in mortality estimates includes uncertainty from the RRs and from air
425 pollutant concentrations. First, we conduct 1000 Monte Carlo (MC) simulations separately for
426 each model-year to propagate uncertainty from the RRs to mortality estimates. For ozone, we use
427 the 95% Confidence Intervals (CIs) for RR reported by Jerrett *et al.*¹ and assume a normal
428 distribution, while for PM_{2.5} we use the parameter values of Burnett *et al.*⁴ for 1000 MC
429 simulations. Then, we calculate the average and 95% CI for the pooled results of the 1000 MC
430 simulations for each model to quantify the spread of model results. We do not include
431 uncertainties associated with population and baseline mortality rates, since these are not reported.
432 As ACCMIP models used the same anthropogenic and biomass burning emissions, we do not
433 consider uncertainty in emissions inventories, however we acknowledge that this is an important
434 source of uncertainty, especially in particular regions³⁴⁻³⁷. Our mortality estimates are affected by
435 our choices of and underlying assumptions regarding concentration-response functions,
436 population, and baseline mortality rates. Although a number of factors, such as vulnerability of
437 the exposed population and PM_{2.5} composition, vary spatially and possibly temporally, we
438 assume that the RRs estimated for the present day apply on a global scale and in future time
439 periods. Also, our assumption that the spatial distribution of population within each country is
440 constant in the future likely understates the effects of rural-to-urban migration, which is currently
441 underway and expected to continue. However, the effects of climate change on air pollutant
442 concentrations may be somewhat spatially uniform (as opposed to changes in emissions), and the

443 coarse grid resolution of global models would not resolve air pollutant concentrations well in
444 urban areas.

445

446 **Data Availability**

447 Data used in this project are archived here:

448 Air pollutant concentrations: Atmospheric Chemistry & Climate Model Intercomparison Project
449 (ACCMIP) datasets - <http://catalogue.ceda.ac.uk/uuid/b46c58786d3e5a3f985043166aeb862d> .

450 Data retrieved from 08/2012 to 12/2013.

451 Present-day population: Oak Ridge National Laboratory (ONRL) - LandScan 2011 Global
452 Population Dataset, <http://spruce.lib.unc.edu.libproxy.lib.unc.edu/content/gis/LandScan/> . Data
453 retrieved on 12/05/2012.

454 Present-day baseline mortality: Institute for Health Metrics and Evaluation (IHME): Global
455 Burden of Disease Study 2010 (GBD 2010) Results by Cause 1990-2010 - Country Level,
456 Seattle, United States, 2013.
457 [https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2](https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2F2010)
458 [F2010](https://cloud.ihme.washington.edu/index.php/s/d559026958b38c3f4d12029b36d783da?path=%2F2010) . Data retrieved from 12/2013 to 03/2014.

459 Future population and baseline mortality: Web-Based IFs - The International Futures (IFs)
460 modeling system, version 6.54., www.ifs.du.edu . Data retrieved on 07/2012.

461 IER model: Global Burden of Disease Study 2010. Global Burden of Disease Study 2010 (GBD
462 2010) - Ambient Air Pollution Risk Model 1990 - 2010. Seattle, United States: Institute for

463 Health Metrics and Evaluation (IHME), 2013. [http://ghdx.healthdata.org/record/global-burden-](http://ghdx.healthdata.org/record/global-burden-disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010)
464 [disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010](http://ghdx.healthdata.org/record/global-burden-disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010) . Data retrieved on
465 11/08/2013.

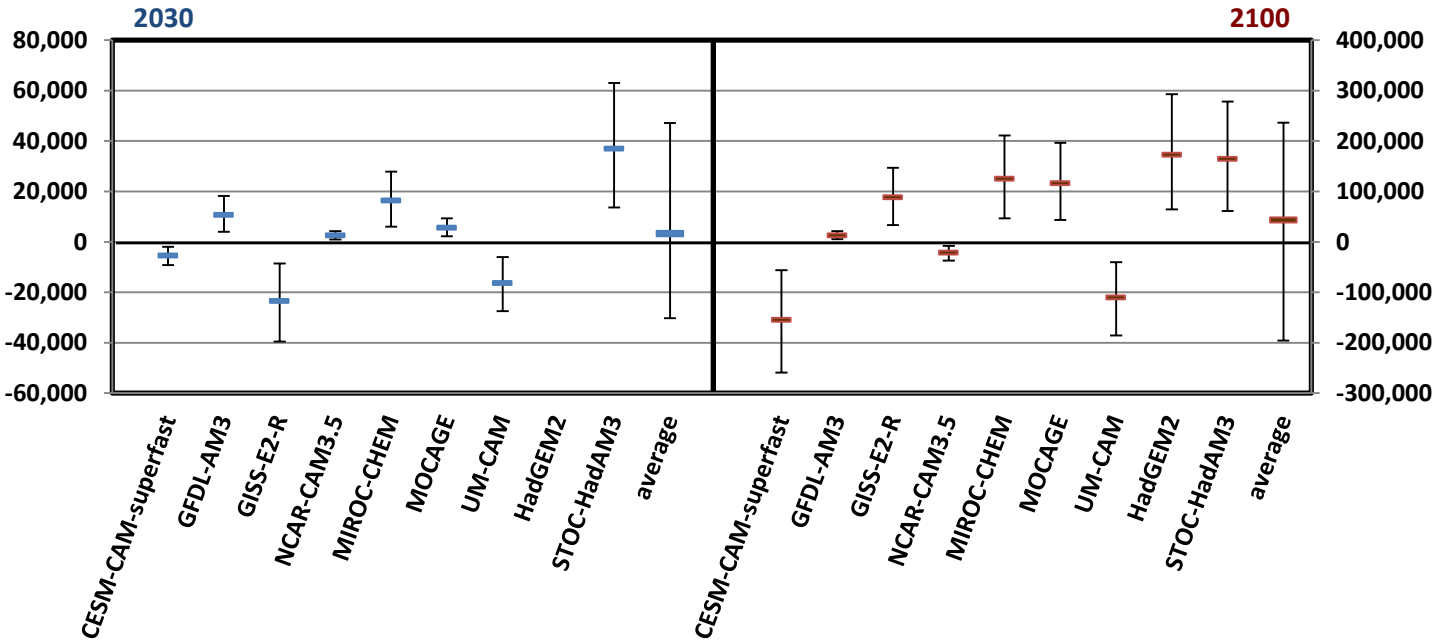
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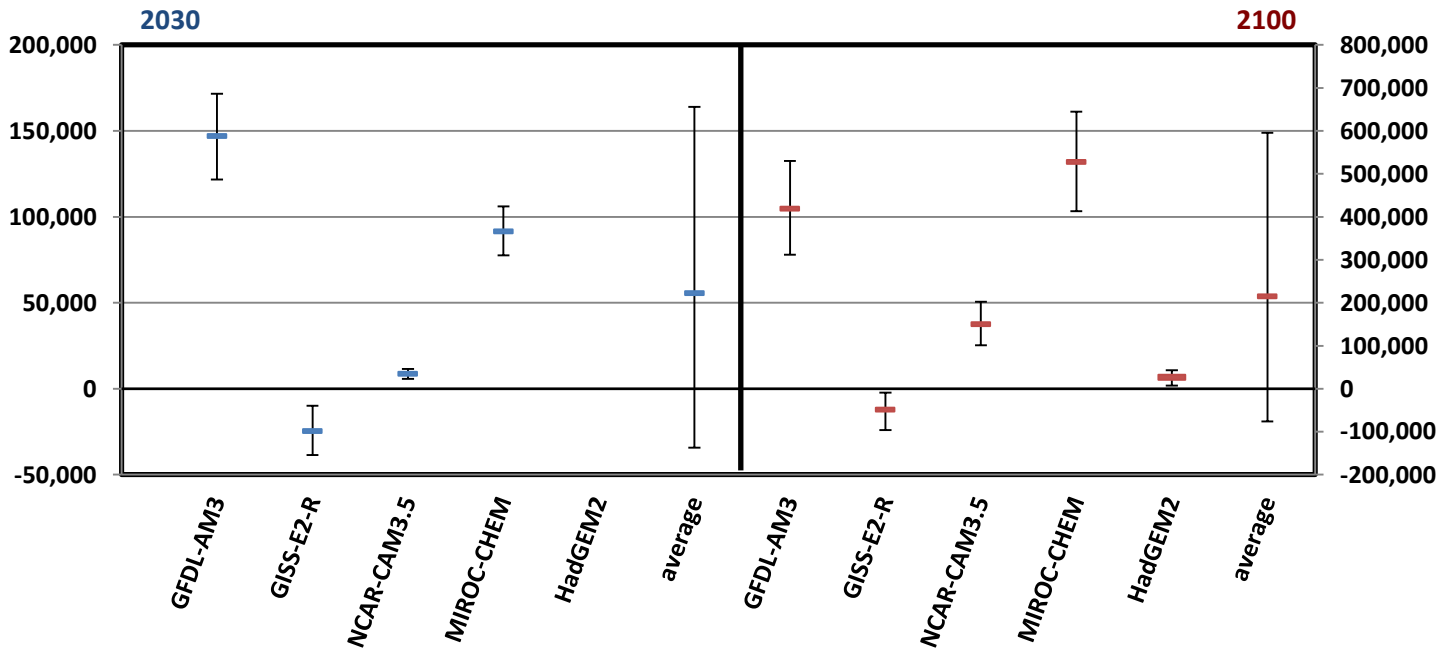
484

485

a. Ozone mortality
(deaths yr⁻¹)

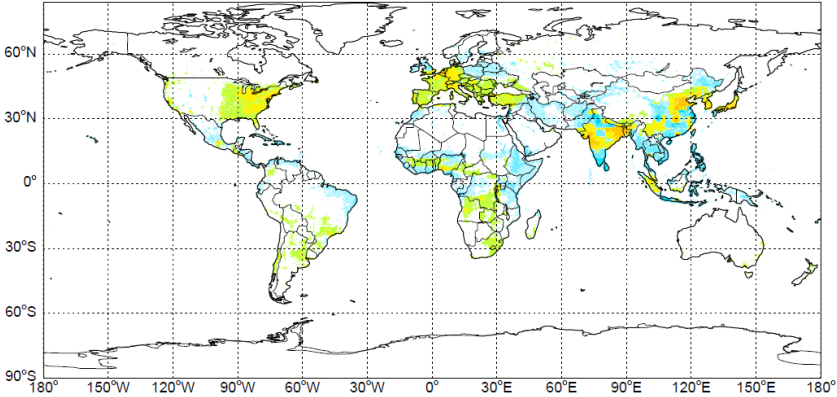


b. PM_{2.5} mortality
(deaths yr⁻¹)

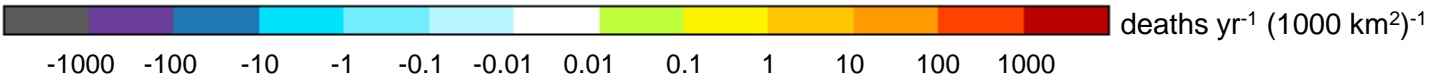
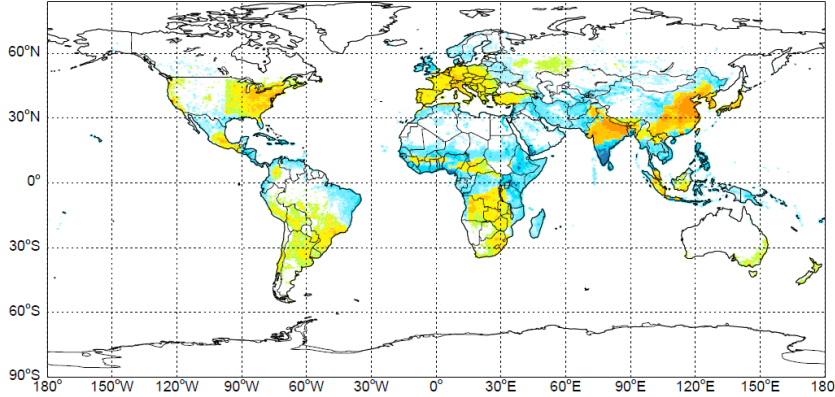


a. Ozone mortality

2030 **8 models**

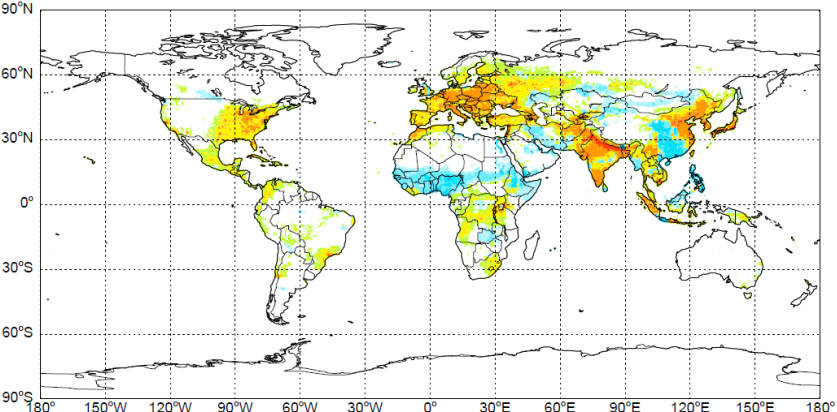


2100 **9 models**

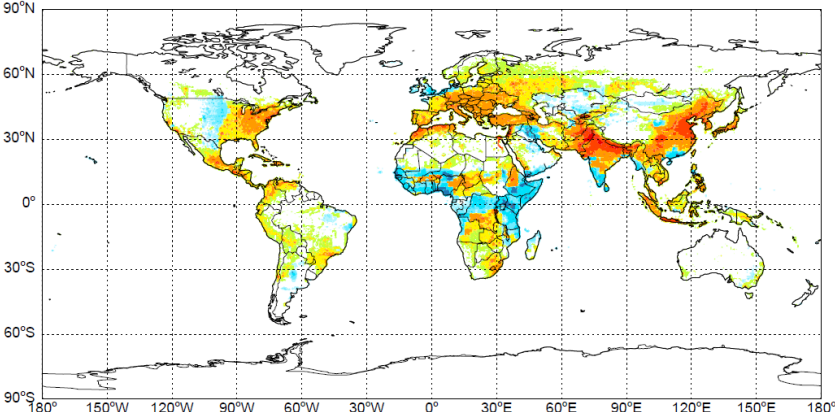


b. PM_{2.5} mortality

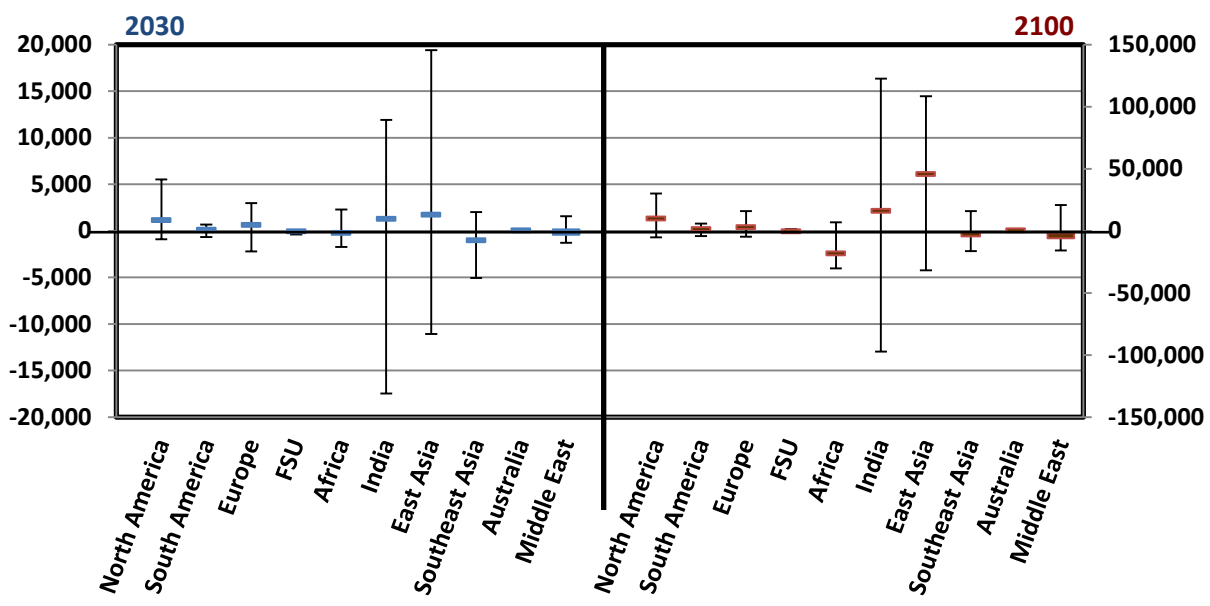
2030 **4 models**



2100 **5 models**



a. Ozone mortality (deaths yr⁻¹)



b. PM_{2.5} mortality (deaths yr⁻¹)

