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1 Microbial electrolytic capture, separation and
2 regeneration of CO₂ for biogas upgrading

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12 Abstract

13 Biogas upgrading to natural gas quality is essential for the efficient use of biogas in various
14 applications. Carbon dioxide (CO₂) which constitutes a major part of the biogas is generally removed
15 by physicochemical methods. However, most of the methods are expensive and often present
16 environmental challenges. In this study, an innovative microbial electrolytic system was developed to
17 capture, separate and regenerate CO₂ for biogas upgrading without external supply of chemicals, and
18 potentially to treat wastewater. The new system was operated at varied biogas flow rates and external
19 applied voltages. CO₂ was effectively separated from the raw biogas and the CH₄ content in the outlet
20 reached as high as 97.0±0.2% at the external voltage of 1.2 V and gas flow rate of 19.6 mL/h.
21 Regeneration of CO₂ was also achieved in the regeneration chamber with low pH (1.34±0.04). The
22 relatively low electric energy consumption (≤ 0.15 kWh/m³) along with the H₂ production which can
23 contribute to the energy input makes the overall energy need of the system low, and thereby makes the
24 technology promising. This work provides the first attempt for development of a sustainable biogas
25 upgrading technology and potentially expands the application of microbial electrochemical
26 technologies.

27 Keywords: Biogas upgrading; Microbial electrolytic system; Bipolar membrane; Biomethane; CO₂
28 capture and regeneration; Wastewater treatment

29

30 Introduction

31

32 Biogas as an alternative energy source could contribute to the reduction of fossil fuel utilization and
33 greenhouse gas emissions, and thus, is getting increased attention worldwide (especially in Denmark).¹

34 Biogas is primarily composed of 60-70% v/v methane (CH₄), 30-40% v/v carbon dioxide (CO₂), and
35 some trace gases such as hydrogen sulfide (H₂S).² Raw biogas with CO₂ presence exhibits a
36 significantly low Wobbe index, heating value and may adversely affect the engine performance.³ Thus,
37 those unwanted impurities especially CO₂ need to be removed from biogas before used as vehicle fuel
38 or injected to natural gas grids. The conventional technologies for CO₂ removal are water scrubbing⁴,
39 pressure swing adsorption, chemical absorption^{5,6}, amine scrubbing, membrane separation⁷, etc. Some
40 of them have several limitations such as high energy requirement, difficulty in recovery of CO₂,
41 corrosion or clogging problems. Among them, alkaline absorption is characterized of high efficiency,
42 low working pressure, high selectivity and high reaction capacity.⁶ However, the technology still
43 encounters several challenges. In addition to alkali, acids may also be required for further regeneration
44 of CO₂ to reuse/recover the captured CO₂ for subsequent sequestration or synthesis of biofuels and
45 commodity chemicals. Currently, alkali and acids are mainly produced through electrolysis of salt
46 solutions termed chlor-alkali process, which consumes a lot of electrical energy and the transportation
47 to the biogas upgrading plants increases the cost.⁸ Therefore, the high cost of chemicals and safety
48 issues associated with the transportation and storage of chemicals may affect the application of the
49 technology.

50 In the past years, microbial electrochemical technologies (METs) have gained increasing attention
51 as potentially sustainable approaches for the production of electricity, biofuels or chemicals, and water
52 treatment with low energy investment.^{9, 10} Recently, one of METs, namely microbial electrolysis

53 desalination and chemical production cell (MEDCC), has been demonstrated as an efficient and cost-
54 effective method for water desalination, acid and alkali production.¹¹ With an external applied voltage
55 up to 1.2 V, pH of 0.68 and 12.9 was achieved in the acid-production chamber and cathode chamber,
56 respectively. Notably, the production cost of alkali in MEDCC was much lower than the market price.
57 Latterly, the feasibility of using the acid and alkali solutions produced by a MEDCC for mineral
58 carbonation and CO₂ sequestration in an external reactor was successfully investigated.¹² However,
59 similar systems, which produce alkali and acids for simultaneous biogas upgrading and CO₂ recovery
60 through absorption and regeneration process, have never been reported. Moreover, a more simplified
61 microbial electrochemical system which could take the advantages of the MEDCC for alkali and acid
62 production should be pursued to reduce the construction costs. Furthermore, the energy intensive
63 aeration associated with the alkali synthesis in cathode (e.g., in MEDCC-like reactor) could be further
64 replaced with energy-gained process such as hydrogen evolution.

65 In this study, an innovative microbial electrolytic capture, separation and regeneration cell (MESC)
66 was developed as an alternative, efficient and cost-effective approach to realize simultaneous biogas
67 upgrading, CO₂ recovery and wastewater treatment. Such a system consists of an anode chamber, a
68 regeneration chamber and a cathode chamber, separated by a bipolar membrane (BPM) and an anion
69 exchange membrane (AEM), respectively. With an external applied voltage, electrons derived from
70 bacterial oxidation of organic matter transfer from the anode to cathode where H₂ evolution and OH⁻
71 accumulation occur. Meanwhile, H⁺ and OH⁻ are produced in the BPM through water dissociation,
72 resulting in an acidic environment in the regeneration chamber and neutralization of OH⁻ with H⁺
73 released from substrate oxidation in the anode chamber. Beside prevention of dramatic pH drop in the
74 anode for better bacterial growth, BPM has the proven potential of more efficient and cost-effective
75 production of acid and alkali in (bio)electrolytic process compared to conventional anion and cation

76 exchange membranes.^{12, 13} In our system, inlet CO₂ from the biogas is initially captured as HCO₃⁻ and
77 CO₃²⁻ by the alkali solution produced in the cathode (Equation 1 and 2). Because of the potential
78 difference between the middle and cathode chambers, HCO₃⁻ and CO₃²⁻ subsequently transport through
79 AEM to the regeneration chamber where they combine with H⁺ to form CO₂ again (Equation 3 and 4).
80 Therefore, through such process biogas upgrading, CO₂ recovery, wastewater treatment and H₂
81 production could be achieved simultaneously. The aim of the present study is to demonstrate the
82 feasibility of such novel technology by varying the biogas flow rates as well as the external voltages.
83 The system performance was evaluated in terms of the current output, chemical oxygen demand (COD)
84 removal efficiency, quality of the outlet gas and energy input. This work is among the first attempts to
85 provide proof of concept for a new and value-added microbial electrochemical platform technology for
86 biogas upgrading.



91

92 **Material and methods**

93

94 **The Setup and Operation of the MESC.** A three-chamber reactor made of nonconductive
95 polycarbonate plates was consisted of the anode chamber, regeneration chamber and cathode chamber
96 which were separated by a BPM (fumasep FBM, FuMA-Tech GmbH, Germany) and an AEM (AMI
97 7001, Membrane international, NJ), respectively. The working area of both membranes was 25 cm²

98 (5×5 cm). The working volume of the three chambers was 50, 50 and 100 mL, for the anode,
99 regeneration and cathode chamber, respectively. The reactor was tightened using rubber gaskets and
100 screws to avoid leakage. The anodic electrode is made of carbon brush (5.0 cm in diameter, 5.0 cm in
101 length, Mill-Rose, USA). The cathode electrode is a titanium woven wire mesh (4×5 cm, 0.15 mm
102 aperture, William Gregor Limited, London) coated with 0.5 mg/cm² Pt. A power supply (HQ PS3003,
103 Helmholtz Elektronik A/S, Denmark) was used to provide an additional voltage to the circuit with a 10
104 Ω external resistance.

105 The anode electrode was originally pre-acclimated in a two-chamber microbial fuel cell and
106 inoculated with a mixed bacterial culture from domestic wastewater collected from a primary clarifier
107 (Lyngby Wastewater Treatment Plant, Copenhagen, Denmark). When stable voltage was obtained
108 around 0.5 V with a 1000 Ω resistance for several cycles, the electrode was transferred into the
109 microbial electrolytic reactor. The anolyte was circulated over a 250 mL recirculation bottle with a
110 peristaltic pump at a flow rate of 25 mL/min. The total anolyte volume was 250 mL. The system was
111 then operated in a batch mode and a continuous mode. The batch experiment lasted for 7 days. During
112 the batch experiment, the recirculation bottle was filled with 250 mL wastewater from the same
113 primary clarifier mentioned before and 5 mL (1 M) sodium acetate was dosed to achieve a COD
114 concentration approximately 2000 mg/L. The regeneration and cathode chambers were filled with 0.05
115 M sodium chloride solution. Prior to each batch run, all chambers and the recirculation bottle were
116 purged with N₂ for 15 min to ensure anaerobic conditions. The synthetic biogas (60% CH₄ and 40%
117 CO₂) was introduced to the bottom of the cathode chamber by a column diffuser (pore diameters 0.5-
118 1.0 μm, Maltbazaren ApS, Denmark) via a peristaltic pump at varying gas flow rates. The gas retention
119 time in the chamber was between 3.6-7.5 h according to different gas flow rates. The upgraded biogas
120 was collected from the cathode with a gas bag. The collected gas was recirculated through the cathode

121 at the same gas flow rate of the raw biogas to enhance the mixing and increase CO₂ absorption. A
122 control experiment without external applied voltage was also operated in the batch mode at 19.6 mL/h
123 gas flow rate. In the continuous mode, the medium in the anodic recirculation bottle was renewed when
124 the current density decreased below 0.8 A/m² while the solution in the regeneration and cathode was
125 not refilled throughout the whole period. The reactor was operated for 40 days with 19.6 mL/h gas flow
126 rate. The external voltage was 1.2 V at the beginning and then increased to 1.4 V and further to 1.5 V at
127 the end. The experiments were carried out at least in triplicate and at room temperature (22±2°C). All
128 chemicals utilized were of reagent grade without purification.

129 **Electrochemical Analyses and Calculations.** Voltage drop across the 10 Ω resistance was
130 measured and voltage readings were taken every 30 mins by a digital multimeter (Model 2700,
131 Keithley Instruments, Inc.; Cleveland, OH, USA). Current was calculated according to ohm's law.
132 Current density was calculated as $i=I/A$, where I (A) is the current and A (m²) is the project surface area
133 of the cathode electrode. Conductivity and pH were measured using a CDM 83 conductivity meter
134 (Radiometer) and a PHM 210 pH meter (Radiometer), respectively. Methane and carbon dioxide were
135 determined by a GC-TCD fitted with paralleled column of 1.1 m × 3/16 'Molsieve 137 and 0.7 m × 1/4'
136 with hydrogen as the carrier gas (MGC 82-12, Micolab A/S, Denmark). Hydrogen was analyzed by a
137 GC-TCD fitted with a 4.5 m × 3 mms-m stainless column packed with Molsieve SA (10/80). Cl⁻ was
138 detected via Flame atomic absorptions spectrophotometer (FAAS, Shimadzu). Total inorganic carbon
139 (TIC) was determined by a TC-2000 (Shimadzu, Kyoto, Japan). COD was analyzed according to
140 APHA. Coulombic efficiency (CE), current efficiency (CTE), electric energy consumption by the
141 power source (W_E) and the energy efficiency (η_E) relative to the ratio of produced H₂ energy to the
142 electricity input were calculated as previously described¹⁴ and equations are shown in the supporting

143 information. Methane loss was calculated as well and the formula is shown in the supporting
144 information.

145

146 **Results and Discussion**

147

148 **The MESC reactor performance.** The microbial electrolytic reactor was first operated with 19.6
149 mL/h gas flow rate and 1.2 V external voltage. In the control experiment, the reactor was operated in
150 the open circuit at the same gas flow rate. Reactor performance in terms of the current generation, COD
151 removal, pH and conductivity profiles and gas contents was shown in Figure 2.

152 Figure 2a shows the current output and COD tendency along with the operation time. With 1.2 V
153 external voltage, the current density started at 1.7 A/m^2 and gradually decreased to 1.3 A/m^2 during the
154 first four days since COD in the anode decreased from 2341 ± 320 to 420 ± 79 mg/L. On day 5, when
155 COD concentration decreased to the level below 500 mg/L, the current density dropped significantly
156 below 0.8 A/m^2 . Therefore, the concentration of organic matter can affect the circuit current by
157 affecting exoelectrogens.¹⁵ Acetate amended domestic wastewater was used in this study as the anodic
158 substrate to proof the concept of the technology. Since METs can use nearly all kinds of organic wastes
159 as substrates, wastewater with high COD levels (e.g., industrial wastewaters and anaerobic
160 fermentation effluents) could be used for practical application of the technology. In the control
161 experiment, a similar COD removal efficiency of $94.4 \pm 1.3\%$ was obtained indicating competitors to
162 exoelectrogens such as methanogen might prevail in the domestic wastewater. Indeed, CH_4 (17%) was
163 detected in the anode chamber for the control reactor. Lower amount of CH_4 (3%) was detected in the
164 closed circuit since exoelectrogens were dominant due to the anode potential¹⁹. Results show the
165 feasibility of the microbial electrolytic system for wastewater treatment. Figure 2b shows the pH

166 profile of the whole reactor. Initial pH of the wastewater and NaCl solution was 7.30 ± 0.28 and
167 7.61 ± 0.30 , respectively. In the control experiment, pH in the anode and regeneration chambers changed
168 barely while that in the cathode chamber decreased a little from 7.61 ± 0.30 to 6.32 ± 0.21 . This could be
169 due to CO_2 dissolving in water and the ionization of it to HCO_3^{-16} . When 1.2 V external voltage was
170 supplied to the reactor, pH in the regeneration chamber decreased below 2 which was contributed by
171 water dissociation between the BPM. Anodic pH was nearly stable during the experiment which was
172 due to the neutralization of H^+ from bacterial respiration with OH^- from water dissociation. Catholyte
173 pH first increased to maximum at 9.19 ± 0.12 on the first day and then decreased gradually to 7.42 ± 0.22
174 on the last day. The increase of the catholyte pH was probably due to the splitting of water for the
175 production of alkaline and H_2 at cathode, which was supported by the observed hydrogen production
176 (37 ± 3 mL) at the end of the batch run. Catholyte pH decreased gradually since dissolved CO_2 reacted
177 with alkaline and generated HCO_3^- and CO_3^{2-} . On day 5 a sharp decrease was observed in catholyte pH
178 which was consistent with the remarkable decrease in the current output. The electrolysis process might
179 cease as organic matter was completely consumed.

180 The initial conductivity of wastewater and NaCl solution were 2.74 ± 0.23 and 5.72 ± 0.26 mS/cm,
181 respectively. As shown in Figure 2c, in the control experiment conductivity of the reactor fluctuated
182 insignificantly. With 1.2 V external voltage, conductivity in the regeneration chamber first increased
183 from 5.72 ± 0.26 to 22.79 ± 0.75 mS/cm on the second day and then decreased to 11.07 ± 0.95 mS/cm in
184 the end. The increase of conductivity in the regeneration chamber could be explained by H^+ production
185 via the BPM. When pH in the regeneration chamber was 1.34, it could contribute 15.98 mS/cm in
186 conductivity (calculation was stated in Supporting Information). However, along with the decreasing of
187 current density from 1.7 to 0.8 A/m^2 , the water dissociation efficiency decreased even to zero. In
188 Bauer's work²⁰, the water dissociation efficiency decreased from 70% to 0% as the current density drop

189 from 10 to 0 mA/cm². Another reason for the decrease of conductivity was that carbonate and
190 bicarbonate ions reacted with H⁺ and CO₂ emission occurred. At the end of the experiment, 22±3 mL
191 CO₂ was collected from the regeneration chamber.

192 Alkaline absorption is an efficient way for biogas upgrading because of the large difference in the
193 solubility of CO₂ and CH₄.¹⁴ As shown in Figure 2d, when no alkaline was produced in the control
194 experiment, the CH₄ concentration was kept around 85% and CO₂ concentration decrease from the
195 original 40% to approximately 15% due to dissolution in water. When 1.2 V voltage was supplied, CO₂
196 reacted with alkaline leading to a higher CO₂ removal from biogas. For example, the CO₂ concentration
197 was blow 8% during the whole period and even to 0.5±0.2% while the highest CH₄ content could be
198 97.0±1.3%. Therefore, the microbial electrolytic reactor was an efficient approach for biogas upgrading.

199 **Influence of other ionic species on HCO₃⁻/CO₃²⁻ transportation.** To achieve efficient CO₂
200 separation and capture, the key point is to drive HCO₃⁻ and CO₃²⁻ ions through the membrane to the
201 regeneration chamber. However, the cathode was filled with 0.05 M NaCl solution and alkali was
202 produced during H₂ evolution. Cl⁻ and OH⁻ might also move through the membrane and thereby are
203 competing with HCO₃⁻ and CO₃²⁻ for transportation. Therefore, pH and Cl⁻ concentration in the cathode
204 and regeneration chambers were measured along with the operation time. The profile of Cl⁻
205 concentration is shown in Figure 3. When catholyte pH increased to 9.2, the OH⁻ concentration was as
206 low as 1.58×10⁻⁵ M and therefore would have insignificant effect on electron flux balance. As shown in
207 Figure 3 and Table S2, no significant changes in Cl⁻ concentrations were observed in the regeneration
208 chamber and cathode chamber in the open circuit. On the first day of the closed circuit operation, when
209 current densities were the highest and HCO₃⁻ concentration was the minimum, the Cl⁻ concentration
210 reached the maximum of 2404±50 mg/L followed by a slightly decrease in the regeneration chamber.
211 In the cathode, the Cl⁻ concentration decreased to 1040±40 mg/L to compensate the flux of electron

212 which was consistent with the slightly lower conductivity (4.69 ± 0.17 mS/cm) on the first day.
213 Subsequently, the slight transfer of Cl^- from the regeneration chamber to cathode chamber occurred due
214 to back diffusion or ion exchange with $\text{HCO}_3^-/\text{CO}_3^{2-}$. The final Cl^- concentrations in the regeneration
215 and cathode chamber in the closed circuit were 0.053 and 0.045 M, respectively. Results indicated the
216 final flux of Cl^- transported from the cathode to the regeneration chamber was limited ($\leq 10\%$). It has
217 been reported that ion transfer can be prevented due to similar ion concentrations in both compartments
218 separated by the membrane^{17, 18}. According to the C balance shown in Figure 5, at the end of the batch
219 test 0.261 g of TIC was detected in the regeneration chamber in closed circuit which equals equivalent
220 of 0.022 mol HCO_3^- . A current efficiency of 81% could be achieved during the whole period meaning
221 that the transfer of every mol of electrons to the cathode was accompanied by 0.81 mol HCO_3^-
222 transportation. Thus, the HCO_3^- transportation would not be adversely affected by Cl^- and OH^- .

223 **Effects of the gas flow rate and applied voltage on the system performance.** The influence of
224 biogas flow rate on the system performance was investigated at the external voltage of 1.2 V. The gas
225 flow rate was control by the peristaltic pump from 13.4 to 27.7 mL/h. The summary of current outputs,
226 pH and conductivity profiles, and outlet gas composition during the experiment is shown in Figure 4.

227 Under different gas flow rates, the current density showed similar trend along with the operation
228 time (Fig. 4a). The current density fluctuated around 1.6 A/m^2 and significantly dropped below 0.8
229 A/m^2 after the fourth day due to the substrate consumption and high polarization resistance of the BPM.
230 The current output was mainly influenced by the microorganism community and the substrate
231 concentration. However, with 13.4 mL/h gas flow rate, the reactor produced relatively lower current
232 densities. Since gas introduction to the catholyte might function as mixing and enhance the mass
233 transfer. Relatively low gas flow rate and recirculation rate resulted in small disturbances and slow
234 mass transfer which was adverse to the cathodic hydrogen evolution.

235 In Fig 4b, cathodic pH was in the range of 7 to 10 while pH in the regeneration chamber was
236 between 1.5 and 4. With the gas flow rate increasing, cathodic pH decreased obviously indicating the
237 enhanced CO₂ absorption in the alkaline solution. With 13.4 mL/h gas flow rate, it was found that
238 cathodic pH reached the highest value around 10 on the second day. It is because relatively less CO₂
239 reacted and consumed OH⁻. In Fig. 4c, the maximum conductivity in the regeneration chamber
240 increased from 14.08±0.75 to 22.79±0.75 mS/cm along with the increasing of gas flow rate from 13.4
241 to 19.6 mL/h. When the gas flow rate went up further, conductivity did not increase further.
242 Correspondingly, when the gas flow rate was below 20 mL/h, CH₄ composition kept above 90% and
243 could reach to maximum at 98.1±0.2%. CO₂ was removed effectively and could be lower than 2%
244 during the first 4 days when the gas flow rate was below 20 mL/h. When the gas flow rate increased
245 from 25.3 to 27.7 mL/h, CH₄ composition dropped from 88% to 84% as average during the whole
246 period. Therefore, a proper feeding gas flow rate is crucial to improve the system performance. When
247 the gas flow rate was low, the biogas upgrading capacity might decrease while high flow rates might
248 decrease the purity of CH₄. Further strategies for the improvement of biogas flow rate are thus required
249 in future studies for accelerating the commercialization.

250 The effect of external voltage on the CO₂ removal efficiency and CH₄ enrichment was studied. The
251 system performance was compared at external voltage of 0, 1.0, 1.2 and 1.4 V which are shown in
252 Figure S1 and S2. When the gas flow rate was the same but the external voltage elevated, the maximum
253 catholyte pH was increased from 6.44±0.05 (0 V, 19.6 mL/h), 7.37±0.12 (1.0 V, 19.2 mL/h) to
254 9.19±0.11 (1.2 V, 19.6 mL/h) and from 7.38±0.08 (1.2 V, 27.7 mL/h) to 7.52±0.10 (1.4 V, 28.0 mL/h)
255 as well as the volume of generated H₂ from 0 mL (0 V, 19.6 mL/h), 22±2 mL (1.0 V, 19.2 mL/h) to
256 37±3 mL (1.2, 19.6 mL/h V) and from 37±3 mL (1.2 V, 27.7 mL/h) to 40±3 mL (1.4 V, 28.0 mL/h),
257 respectively. The applied voltage would induce fast electron transfer kinetic and compensate the

258 electrode overpotential, which could also enhance the substrate consumption by anode respiring
259 bacteria²¹ and favorite the electrolysis process. When the applied voltage was changed from 0, 1.0 to
260 1.2 V at around 19.6 mL/h gas flow rate, the maximum CH₄ content of the output gas increased from
261 85.5±0.3%, 89.2±0.4% to 97.0±0.2%. However, when the gas flow rate was higher than 25.3 mL/h, the
262 concentration of CH₄ could not be improved further with even higher applied voltage (1.4 V). The
263 biogas upgrading capacity probably has a threshold and the system performance might be determined
264 by other limiting factors.

265 Experimental results at various operation conditions are summarized in Table 1. As shown in
266 Figure 4a, with the fixed external voltage supply (1.2 V), the current density increased along with the
267 increasing of gas inflow rate since it enhanced the mixing and mass transfer. However, cathodic pH and
268 upgrading performance showed opposite trends (Figure 4b and 4d). It could be due to a change in
269 equilibria between alkali production capacity at cathode (expressed as pH) and the dissolving gasses. In
270 this case, the gases dissolution rate was higher than the OH⁻ generation rate. As expected, the current
271 density, cathodic pH and upgrading performance increased with the increasing external voltage supply
272 (Figure S1 and S2). However, the increase in those parameters turned to be limited when the gas inflow
273 rate was over a certain level. Thus, operational parameters applied such as voltage and gas inflow rate
274 need to be better adjusted to maintain equilibrium between cathodic pH and the dissolving gasses for
275 efficient upgrading. Considering the final pH and Cl⁻ concentration in the cathode (Table S2), the
276 changes in the transportation of other anions than HCO₃⁻ and CO₃²⁻ were not significant under the
277 tested conditions. The collected CO₂ increased with the current density since more HCO₃⁻ ions were
278 transported via the membrane driven by higher current and then CO₂ was formed at the lower pH.
279 Results indicate that both the applied voltage and biogas flow rate are crucial to the system

280 performance. Eligible CH₄ content (>90%) was obtained when the gas flow rate was in the range of
281 13.4 and 19.6 mL/h with 1.2 V applied voltage.

282 **Energy analysis and mass balance.** The electrical energy consumption for biogas upgrading
283 technologies is an important issue for the practical application. Our lab-scale system consumed
284 between 0.09 and 0.15 kWh of electrical energy for upgrading one m³ raw biogas, which was quite
285 competitive compared with commercial technologies. For instance, the electricity demand for biogas
286 upgrading is about 0.2-0.32 kWh/Nm³ raw gas with water scrubbing, 0.1-0.15 kWh/Nm³ with chemical
287 absorption and 0.15-0.22 kWh/Nm³ with membrane technology^{3, 22}. It should be noted that the
288 electricity consumption is the sole “consumable cost” of microbial electrolytic system for biogas
289 upgrading while additional of chemicals (e.g., acids and alkali), high pressure or temperature are
290 required in most of the conventional methods. Since the microbial facilitated release of electrons stored
291 in organic matter from wastewater, the applied voltage was much lower than that used in the water
292 electrolysis process.¹⁴ Moreover, an energy carrier (H₂) is produced in the cathode which could
293 substitute 21-27% of the electric energy input. Since the substrate was degraded and about 16-33% of
294 the chemical energy was converted to the electrical energy, the energy provided by the organic
295 substrate was 140-334 kJ/Nm³ which was calculated according to the energy content of acetate based
296 on the heat combustion. The energy benefits could be even higher if energy savings for wastewater
297 treatment was considered. Besides operational and maintenance costs for in real scale application,
298 investment should also be considered. However, investment costs per m³ raw biogas would decrease
299 with increasing capacity.

300 At the end of the batch run with and without 1.2 V external voltage at 19.6 mL/h gas flow rate,
301 carbon balance was performed to study the carbon capture mechanism. Rude biogas was introduced
302 into the reactor as the source of CO₂. In the cathode, CO₂ was converted to (bi)carbonate and

303 transported to the regeneration chamber. Then (bi)carbonate reacted with H^+ and CO_2 was regenerated.
304 Volume of initial CO_2 and regenerated CO_2 was measured and TIC of the solution was determined. The
305 carbon balance is shown in Fig. 5. In the control experiment, approximately 30% of the CO_2 was
306 remained and 16% of the CO_2 existed in the cathode chamber as the form of soluble $NaHCO_3$ and
307 Na_2CO_3 . No CO_2 was collected from the regeneration chamber since acid was not produced. When an
308 external voltage of 1.2 V was supplied, carbon distribution was quite different. The remained CO_2 in
309 the outlet accounted for 15% of the total initial CO_2 and 37% existed as (bi)carbonate in the
310 regeneration chamber. The regenerated CO_2 and TIC in the cathode chamber shared 2% and 12% of
311 the initial carbon, respectively. The imbalance on CO_2 species could be due to diffusion to the anode
312 and measurement error.

313 **Experiment under the continuous mode.** The microbial electrolytic system was then operated in
314 the continuous mode for 40 days. For the initial 26 days the system was operated with 1.2 V external
315 voltage and 19.6 mL/h raw biogas flow rate. The external voltage was increased to 1.4 V since day 26th
316 and enlarged further to 1.5 V from day 33rd to the end of the experiment at the same gas flow rate. CH_4 ,
317 H_2 and CO_2 contents in the outlet gas in the continuous mode are showed in Figure 6. During the first
318 six days, CH_4 content was above 90% and could reach $97.5 \pm 0.74\%$ at maximum while CO_2 content
319 was below 5% and reached $1.0 \pm 0.48\%$ at minimum. Subsequently, CH_4 and CO_2 content were kept
320 around 88.3% and 8.3%, respectively. During the whole period, H_2 was produced and the content was
321 around 3.8%. Approximately, 18.8 L rude biogas was treated and the energy consumption for per m^3
322 raw biogas was 0.18 ± 0.01 kWh. During the 40 days operation, NaCl solution has not been changed and
323 the saturation state was not reached yet, which shows a huge absorption capacity of our system with
324 less maintenance. When the external voltage was increased to 1.4 V and 1.5 V, a peak occurred at the
325 corresponding CH_4 content. However, the enhancement on the biogas upgrading was limited which

326 indicated that there were probably other limiting factors dominating the system performance other than
327 external voltage in the continuous mode. Membranes were not cleaned or replaced during the whole
328 period which exhibits the tolerance of membranes.

329 This study for the first time demonstrated the proof of concept of a microbial electrolytic system
330 for simultaneous CO₂ capture, separation and regeneration, biogas upgrading and wastewater treatment.
331 Compared to commercial biogas upgrading technologies, the novel system developed here has several
332 merits. First, addition of chemicals is not needed since acid and alkaline are generated and utilized in
333 situ which saves the cost/risks for chemicals transportation and storage. Secondly, methane loss (e.g.,
334 due to dissolution) during the whole period is less than 1.4% while that in traditional biogas upgrading
335 technologies is about 2%^{3,23}. Low methane loss is beneficial for economic and environmental aspects.
336 Thirdly, separated CO₂ from rude biogas and the (bi)carbonate solution can be further used in other
337 industrial applications, such as the production of algal biomass and enhanced oil recovery.³ Moreover,
338 the applied voltage to overcome the thermodynamic barrier of water splitting in our system is quite
339 lower than that used in the abiotic water electrolysis. It is because the chemical energy stored in the
340 organic matter was released with the help of bacteria to offset the difference. Organic substrates
341 provided energy to the system instead of consuming energy. However, it was not counted as energy
342 input since the wastewater has to be treated anyway. Therefore, the electrical energy consumption in
343 the system is quite low (0.09 kWh/Nm³) and H₂ produced can even substitute part of the energy costs
344 (up to 27%). In addition, the MESC as biological system could meet the needs of sustainability and
345 circular economy. The simple configuration of the MESC makes the technology more competitive
346 compared to the traditional electrolysis processes and MEDCC for the purpose of biogas upgrading and
347 CO₂ recovery. However, for practical applications, several issues should be considered. For instance,
348 H₂S and other impurities in biogas should be removed (biogas pre-upgrade) in advance to eliminate the

349 risk for adverse effect on the microbial electrolytic system. Acetate amended real domestic wastewater
350 was used here to prove the concept. Many waste streams such as industrial wastewater or fermentation
351 effluent have high COD concentrations and can potentially be used for practical application of the
352 technology. Despite long term operation of the microbial electrochemical system with domestic
353 wastewater (emended with acetate), for more than half year, biofouling of the membranes was not
354 observed. Furthermore, NaCl as electrolyte in this study may add the operational costs. However, in
355 practical application, NaCl can be added once to start up the process rather than continuous addition.
356 Wastewater, which could support the conductivity in the liquid during continuous operation, may
357 substitute NaCl. Taking this into account, the energy costs could be further reduced, and thus, further
358 improvement such as appropriate material selection and chemicals refusal might be necessary. To
359 conclude, the concept is very challenging but also full of promises and is definitely going to make a
360 significant impact on relative research field.

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366 **Supporting Information**

367 Calculation for Coulombic efficiency (CE), energy efficiency, current efficiency and methane loss.
368 Table S1, Table S2, Figure S1, and Figure S2 as noted in the text. This material is available free of
369 charge via the Internet at <http://pubs.acs.org/>

370 **Author Contributions**

371 #X.J. and Y.Z. contributed equally to the paper.

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427 **Table 1.** Experimental results at various operation conditions

Voltage V	0	1.0	1.2	1.2	1.2	1.4	1.4
Gas flow rate mL/h	19.6	19.2	13.4	19.6	25.3	27.7	28.0
pH _{max} in Cathode	6.44±0.05	7.37±0.12	10.03±0.07	9.19±0.11	8.59±0.10	7.38±0.08	7.52±0.10
pH _{min} in Regeneration	6.87±0.04	2.09±0.07	1.88±0.05	1.34±0.04	1.43±0.10	1.36±0.07	1.44±0.06
Max CH ₄ %	85.5±0.3	85.9±0.4	98.1±0.2	97.0±0.2	90.3±0.3	86.0±0.3	86.1±0.2
CO ₂ removal efficiency %	78.3±1.4	77.0±0.7	98.7±0.5	92.6±0.6	89.3±1.5	85.7±0.4	81.3±1.1
COD removal efficiency %	94.4±1.3	80.4±2.7	78.7±3.5	98.2±2.6	91.9±2.8	93.4±4.7	86.5±5.9
CE %	-	16.3±1.4	20.0±1.3	26.0±2.1	22.7±1.7	32.6±1.3	17.5±1.8
CO ₂ collected mL	-	6±3	10±3	22±3	18±2	24±2	25±2
H ₂ mL	-	22±2	24±2	37±3	32±2	31±4	40±3
Electricity consumption kWh/m ³		0.09±0.004	0.15±0.003	0.15±0.003	0.10±0.003	0.11±0.002	0.13±0.004
Energy efficiency	-	26.8±1.2	24.2±1.7	26.5±0.8	26.7±1.4	20.6±1.7	22.4±2.2
CH ₄ loss %	1.22±0.02	1.32±0.03	1.21±0.02	1.23±0.02	1.14±0.03	1.20±0.02	1.27±0.02

428 **Figure Caption**

429 **Figure 1.** The schematic diagram of MESC. BPM, bipolar membrane; AEM, anion exchange membrane.

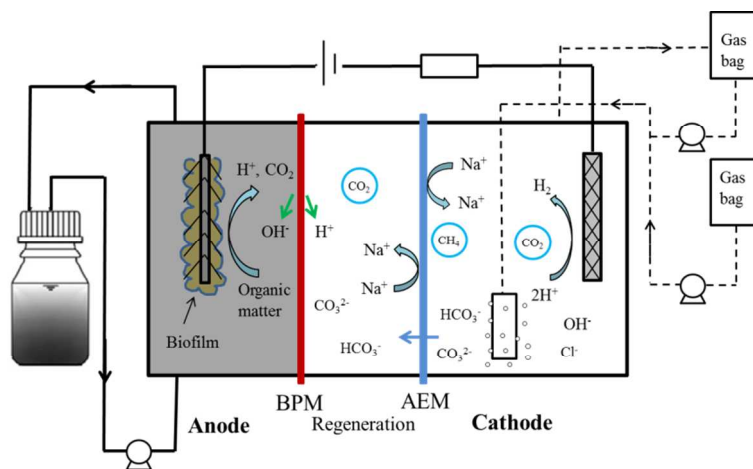
430 **Figure 2.** MESC performance at 19.6 mL/h gas flow rate with 1.2 V and control experiment. A, anode chamber;
431 R, regeneration chamber; C, cathode chamber. The current density output and COD changes (a); pH
432 profiles in three chambers (b); conductivity profiles in three chambers (c); methane and carbon dioxide
433 contents in the outlet gas (d).

434 **Figure 3.** Profiles of Cl^- concentration in the regeneration chamber and cathode chamber along with the
435 operation time.

436 **Figure 4.** MESC performance with 1.2 V and varying gas flow rates. The current density output (a); pH profiles
437 in the cathode (C) chamber and the regeneration (R) chamber (b); conductivity profiles in the
438 regeneration chamber (c); methane and carbon dioxide contents in the outlet gas (d).

439 **Figure 5.** Carbon balance in MESC system with and without 1.2 V external voltage at 19.6 mL/h gas flow rate.

440 **Figure 6.** Methane, hydrogen and carbon dioxide contents in the outlet gas in the continuous mode.

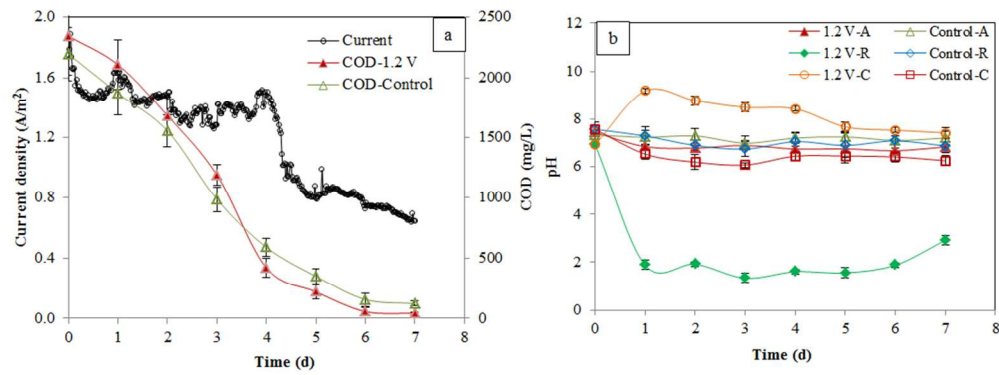


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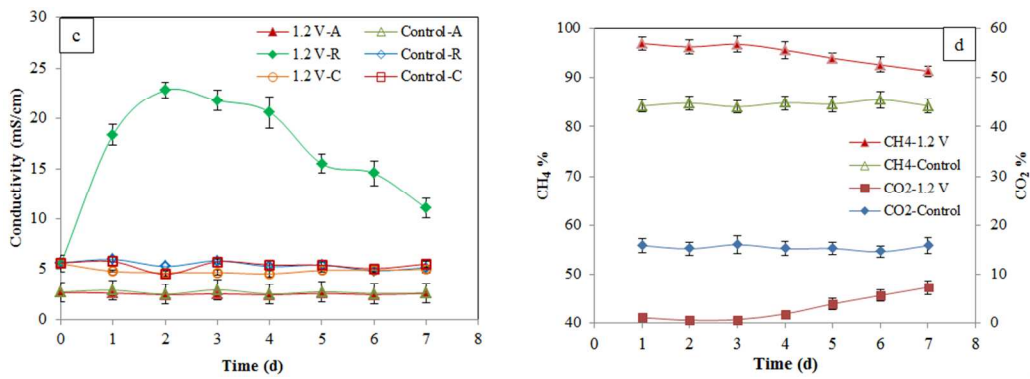
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Figure 1

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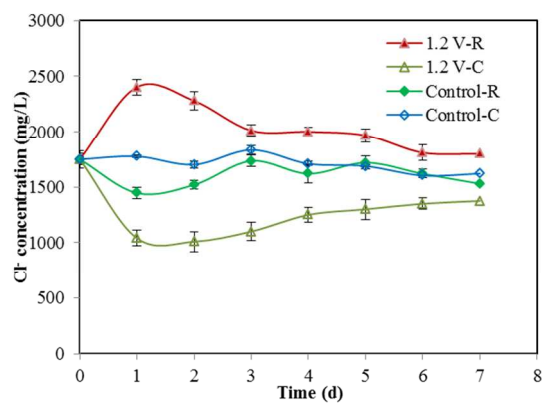


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Figure 2

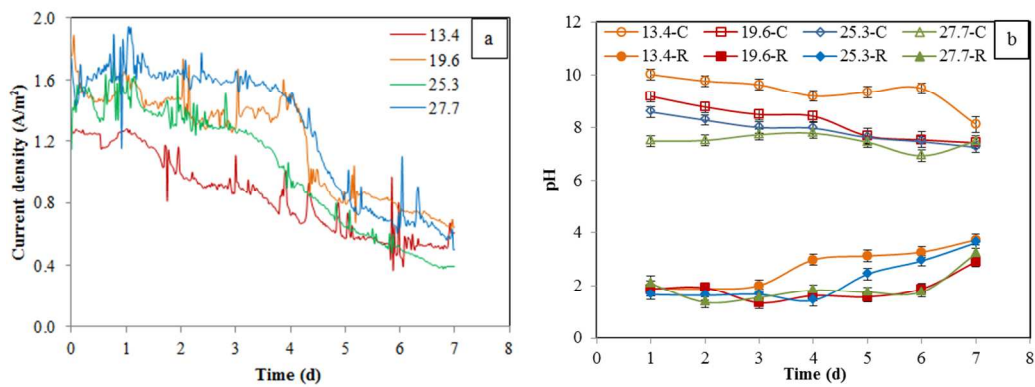


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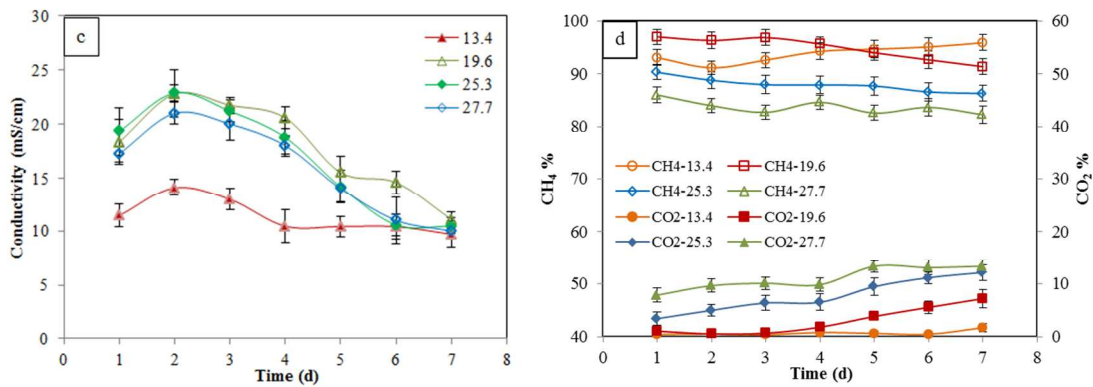
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Figure 3

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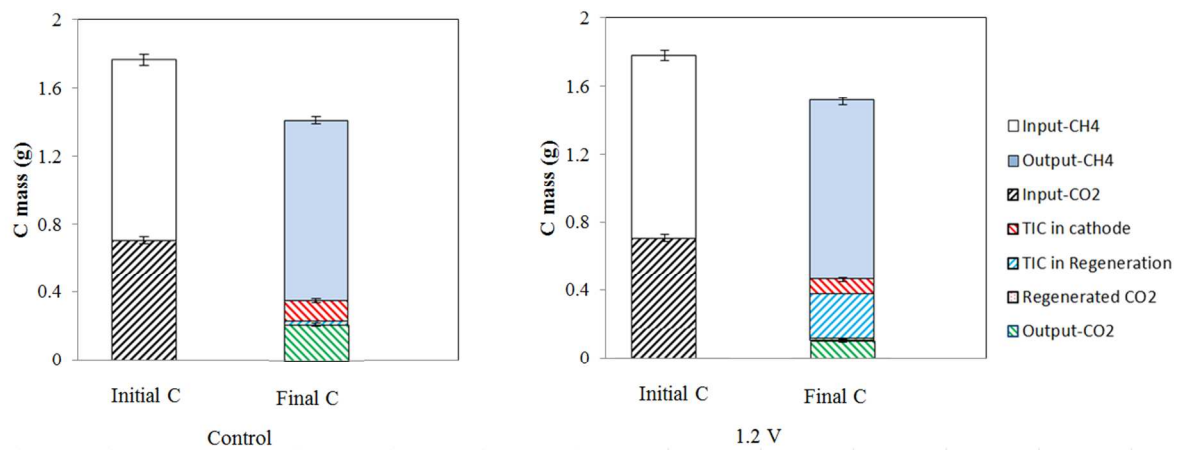


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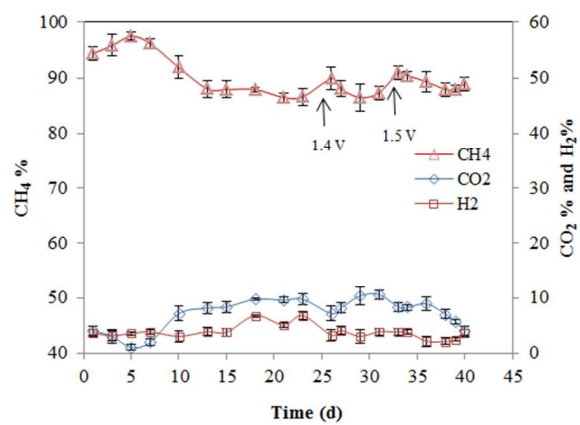
Figure 4



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Figure 5

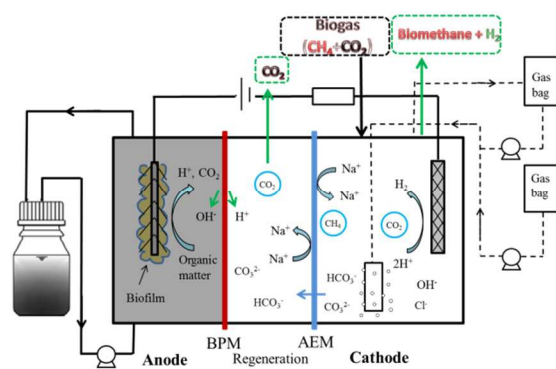


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Figure 6

455 TOC art



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