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1 Assessment of the status of groundwater arsenic at Singair upazilla, Manikganaj Bangladesh;
2 exploring the correlation with other metals and ions

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14

15 **Abstract**

16 Comparative study was conducted to correlate Arsenic (As), Iron (Fe), Copper (Cu), Manganese
17 (Mn), Calcium (Ca²⁺), Magnesium (Mg²⁺), Potassium (K⁺), Nitrate (NO₃⁻), Phosphate (PO₄³⁻),
18 and Ammonia (NH₃) by determining their concentration at different depth of the tube-wells in
19 the selected study area at Singair, Manikganj Bangladesh. Total 99 tube-well water samples were
20 collected from the study area. In most of the sampling points the present concentrations of **As**
21 were less than the previous concentrations. The correlation between As and Fe was positively
22 significant. It can be suggested possible adsorption/coprecipitation of As with Fe in shallow
23 aquifer. However, the relationship between As and Mn was not significantly observed. On the
24 other hand, relationship between Cu and As showed a positive significant correlation. The
25 correlation between As and PO₄³⁻ was also significant, although the correlation between As and
26 NO₃⁻ was not significant. PO₄³⁻ may be comes from phosphate fertilizers application and can be
27 a contributor of As in the shallow aquifer. The PCA biplot also indicated the significant
28 relationship between As, Cu, Fe and PO₄³⁻. Excessive withdrawal of tubewell water along with
29 aquifer dynamics along with ionic interference might be responsible for the mobilization of As in
30 the study area.

31 **Keywords:** Groundwater; Chromatography; Contamination; Arsenic geochemistry; Shallow
32 tubewell; Arsenic mobilization.

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1 Introduction

2 Geochemical cycling of arsenic (As) in the water ecosystems is increasing gradually, which
3 provokes anxiety because of its severe toxicity effects for the human. Ingestion of inorganic As
4 can cause both non-cancer health effects and external and internal cancers in the human body
5 (WHO, 1981; Smith et al., 1992; Tondel et al., 1999). Since the 1970s, the World Health
6 Organization (WHO) has drilled more than two million tube-wells in Bangladesh for the
7 population to have a safe and bacteria-free drinking water supply. Unfortunately, under certain
8 conditions, the groundwater frequently contains elevated levels of natural arsenic, greater than
9 the WHO's maximum contaminant level (MCL) of 0.01 ppm (equivalent to 0.13 mM), and often
10 greater than the Bangladesh's MCL of 0.05 ppm (0.66mM). As, a notorious bio-accumulating
11 poison, has been adversely affecting the health of millions of people in the Bengal Basin
12 (Bangladesh and the West Bengal State, India) for the last twenty years (Das et al., 1995, 1996;
13 Mandal et al., 1996; Dhar et al., 1997; Karim, 2000; Smith et al., 2000). The Bengal Basin lies
14 within in the floodplain of the Ganges and Brahmaputra Rivers. As one of largest river in the
15 world, the GBR is the single largest sediment flux and the fourth highest water discharge to the
16 oceans (Milliman and Meade, 1983). The groundwater As occurrences (higher than permissible
17 limit 0.05 mg/L) in 52 out of 64 districts (total area of 118,012 km²) of Bangladesh were
18 considered as the biggest As calamity in the world (**Karim, 2000**). Arsenic is extensively cycled
19 at the earth surface and has a complex dynamic biogeochemistry. It has multiple oxidation states
20 and it is present in aquatic environment as both in organic and inorganic compounds, which may
21 be interconverted through chemical and biological activity in the natural environment (Ferguson
22 and Gavis, 1972). One of the major sinks of As in aquatic environment is the iron rich sediments
23 (Peterson and Carpenter, 1986) but absorption of arsenic in Mn-oxides has also been reported
24 (Takamatsu *et al.*, 1985). Once it is in the sediments, as diagenesis can be largely controlled by Fe,
25 Mn and organic matter redox reactions (Branonon and Patrick, 1987). Moreover, if As aquifer
26 has a porosity of 25% and the sediments contains 1 mg/kg of labile As, the complete dissolution
27 of that As would lead to a groundwater As concentration of 7.95 mg /L far in excess of any
28 drinking water standard (BGS, 2001 and DPHE, 2000). Accumulation of As in food crops poses
29 potential health threat to humans. A number of studies have already been reported on As
30 concentrations of soil in Bangladesh. Ahsan et al., (2009) reported that the range of As in
31 floodplain soils from Faridpur district of Bangladesh was 18-65 mg/kg whereas the range of As
32 in Dhamrai and Manikganj soils was 3.1-8.9 mg/kg. Contamination of naturally occurring As has
33 recently become an alarming environmental problem in deltaic plain of Ganges-Meghna-
34 Brahmaputra River in Bangladesh and West Bengal (Nickson *et al.*, 1988). It is now well
35 documented that more than 35 million peoples in Bangladesh are under detrimental health threat
36 of As poisoning (DHC report, 1997). In addition, there have some studies reported that
37 agricultural fertilizers and organic wastes may promote As mobilization by ion-exchange with
38 phosphorus derived from fertilizers.(Acharya et al., 2000; Chowdhury et al.,1999; Anawar et al.,
39 2006 and Brömssen et al., 2008)

40 Although serious As contamination of groundwater in Bangladesh has been frequently reported,
41 there are few articles that discussed the relationship between As, and other common metals and
42 nutrients in the groundwater. In the present study, concentration of As along with other
43 commonly occurring heavy, alkaline earth metals and nutrient ions such as, total Fe^{2+ and 3+}, Mn²⁺,
44 Cu²⁺, Na⁺, K⁺, Ca²⁺, Mg²⁺, NO₃⁻, PO₄³⁻, and NH₄⁻ were determined in the groundwater of the

1 contaminated site. The co-occurrences of these metals and nutrient ions with As were discussed
2 in terms of correlation matrices and step wise linear regression models to focus their competitive
3 effects on the release of As in the groundwater of Manikganj district, Bangladesh.
4

5 **Materials and Methods**

6 Singair Upazilla (Manikganj district) with an area of 217.38 km², is bounded by Dhamrai and
7 Manikganj sadar upazilas on the north, Nawabganj (Dhaka) upazila on the south, savar and
8 keraniganj upazilas on the east, Manikganj Sadar upazilla on the west (Fig. 1). Main rivers are
9 the Dhaleshwari, Ghazikhali and Kaliganga. The geographic location of Singair Upazilla is
10 between 23°51' north and 90°21' east (**Banglapedia, 2006**). Ground water samples were
11 collected from the same shallow tube-wells as collected by Uddin et al., 2006 in the study area
12 and the GPS locations were also recorded for GIS mapping (Fig.1). For this study, a total of 99
13 ground water samples were collected from tube-wells with different depth from the surface of
14 Singair Upazilla from August 23, 2010 to August 26, 2010. The approximate length of the
15 sampling area was 6.5 Km. Samples were collected in 250 ml polypropylene plastic bottles.
16 Acidification was carried out by 2 drops of concentrated HNO₃ to prevent precipitation of
17 dissolved iron as well as adsorption of trace metals on to the bottle surface before filtration. All
18 of the samples were filtered with 0.7 µm glass fiber filters (GF/F Whatman) and diluted before
19 analysis, and sampling bottles were carried in dark containers and stored in freezing condition
20 until analysis.

21 **Analytical Methods**

22 For the analysis of phosphate (PO₄³⁻) ion, a phosphate analyzer of ASCSII with wavelength of
23 880 nm was used. Concentrations of anions and cations were determined by Ion Chromatography
24 method (DX-120, Dionex, USA) using the columns “Ion Pac AS12A” (Dionex, Abbreviation of
25 state, USA) for anions and “Ion Pac CS3” (Dionex, Abbreviation of state, USA), for cations. The
26 inductively coupled plasma (ICP-OES) optical emission spectrometer was used (SII Nano
27 Technology inc.) for the determination of **As, Fe, Mn, and Cu**. The water quality analyses were
28 carried out in the laboratory of ecosystem studies, school of environmental science, The
29 University of Shiga prefecture, Japan. Correlation matrix and Principal component analysis (PCA)
30 was done using Origin 9.0 software (OriginLab, USA) and rest of the statistical data analyses
31 was done by MS-excel 2007.

32 **Results and discussion**

33 *Concentration of arsenic and other compounds in the study area*

34 The Singair upazilla stands on the active deltaic region on the Ganges basin and the deltaic areas
35 of Bengal basin receives large amounts of land derives and land weathered organic matter rich
36 sediments with high concentrations of trace metals (Tareq et al., 2003). The results of water
37 quality analysis of the collected samples are presented in Fig. 3(a, b). The mean concentration of
38 arsenic was 0.02 mg/L, ranging from less than detectable limit (BDL) to 0.113mg/L, as of the
39 samples collected in August 2010. Among the total sampling point (99), 51 tube-well water
40 contained different levels of As. Where as the previous mean **As** concentration of the study area

1 was 0.08 mg/L, ranging from BDL mg/L to 0.7mg/L in unpublished data (**Fig. 2**) by Uddin et al.
2 (2006), in addition, **As** concentrations of 85 tube-wells out of 102 tubewells exceeded WHO
3 Guideline for Drinking Water Quality (0.01 mg/L) value. Interestingly, in the present study out
4 of 99 tube-wells 50 tubewells were exceeded WHO standard, 47 tubewells showed not detected
5 (BDL) and only 2 samples were below the WHO standard for drinking water. Furthermore, 13
6 samples exceeds the Bangladesh domestic water quality standard (0.05 mg/L) for **As** in drinking
7 water. The unpublished data revealed that the highest concentration of **As** was 0.7 mg/L,
8 although the highest concentration of **As** was found 0.113mg/L in the present study. So
9 concentration of **As** in the present study was lower than that in previous. The socio economic
10 study states that there is no mining region, any manufacturing or anthropogenic source of **As** in
11 the study area. The variation of this concentration may be due to the geo-environmental settings
12 in the area. **As** concentration from the Padma–Jamuna confluence along with the possibility of
13 co-precipitation with other metals. Moreover, the mean concentration of Mn, Fe, and Cu were
14 0.46 mg/L, (0.054-2.72 mg/L), 0.81 mg/L (BDL-19.29 mg/L), 0.002 mg/L (BDL-0.013 mg/L),
15 respectively. The mean concentrations of Mn and Fe were above the WHO Guideline for
16 Drinking water Quality (GDWQ) value for Mn (0.1 mg/L) and Fe (0.8 mg/L). The mean
17 concentrations of NO_3^- , PO_4^{3-} , NH_3 , Na^+ , K^+ , Mg^{2+} and Ca^{2+} were 253.18, 0.10, 10.75, 88.80,
18 4.14, 49.43 and 195.98 mg/L, respectively. Only 3 samples contained Cd. The concentrations of
19 NO_3^- and Na^+ showed higher concentrations mentioned by WHO GDWQ. The high
20 concentration of **As**, Fe, Mn and PO_4^{3-} may be due to the oxic to sub-oxic conditions of the
21 shallow aquifer of the study area. This result can reflects that the reductive dissolution of iron
22 oxyhydroxide throughout the Holocene alluvial aquifers of Bangladesh and West Bengal, India
23 are driving factor for the mobilization of **As** (Anawar et al., 2011 and Nickson et al., 2000)
24

25 *Correlation of As with elements by PCA*

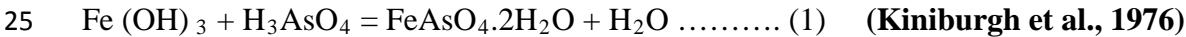
26 Correlation matrix was calculated in order to search for the relationship between **As** and other
27 substances including iron (Table 1). The correlation matrix showed the significant positive
28 correlation (**$r=0.34$**) between arsenic (**As**) and iron (**Fe**) in groundwater in the study area. When
29 the **Fe** concentrations were 0.134, 0.091, 9.262, 1.843 and 19.296 ppm, the **As** concentrations
30 were 0.025, 0.056, 0.041, 0.044 and 0.113 ppm, respectively (fig.3). Many authors found that the
31 **As** mobilization in groundwater of the Bengal delta and other parts of the world is associated
32 with arsenian pyrite (Peters et al., 1999; Raessler et al., 2000; Schreiber et al., 2000). In this
33 study the positive but weak relationship of Fe with **As** concentrations confirms the findings of
34 Safiullah (1998) and McArthur et al. (2001). In groundwater of the Bengal delta, the poor
35 correlation between Fe and total **As** concentration is attributed to the loss of **As** by precipitation
36 of Fe–**As** sulfides, a precursor to arsenopyrites, and siderite (FeCO_3) (Nickson et al., 2000), and
37 this can be a clue on the possible mechanisms controlling **As** mobilization. On the other hand, **As**
38 and **Mn** showed significant positive but very weak correlation (**$r=0.02$**). The reduction of Fe and
39 Mn oxides, a governing process for **As** mobilization has not been supported by this study is due
40 to weak/very weak relationship among **As**, Fe and Mn (Anawar et al., 2003 and Nickson et al.,
41 1998) In addition, Cu and **As** also showed significant positive correlation (**$r=0.49$**). On the other
42 hand **As**- PO_4^{3-} showed significant correlation (**$r=0.32$**). High concentrations of **As**, especially
43 its reduced species of **As(III)** along with Fe, DOC (dissolved organic carbon), NH_4^+ and PO_4^{3-}
44 and low concentrations of dissolved oxygen (DO), NO_3^- and SO_4^{2-} in groundwater, indicating

1 reducing condition in the Holocene aquifer, have resulted in suggestions that reductive
2 dissolution of Fe(III)-oxyhydroxides is the primary mechanism for the release of adsorbed and
3 co-precipitated As (Harvey et al., 2002 and Anawar et al., 2003)

4 Principal component analysis (PCA) is a versatile tool of factor analysis, which can reduce the
5 total number of variables to few factors with most of the variation in the original sets of data and
6 simplify the interpretation (Hair et al., 1998). The number of significant principle component is
7 selected as considering Kaiser Criterion (Kaiser, 1960). PCA on the combined datasets provided
8 two factors with Eigenvalue >1 that can explains approximately 50% of the variability of the
9 data (PC 1 variance of 27.79% and PC 2 variance of 21.58%) (Table 2).

10 From the biplot analysis of PCA (Fig. 4), when two variables are far from the centre and present
11 closely to each others the closely present variables are said significantly positively correlated
12 ($r=1$). Fig. 4 depicts the PC where Ca^{+2} , Mg^{+2} , Na^+ , NH_3 and K^+ may represented for PC 1, and
13 on the other hand, As, Cu, Fe and PO_4^{-3} belongs to PC 2. From the Fig. 4 As is significantly
14 correlated with Cu ($r= 0.49$), Fe ($r= 0.34$) and PO_4^{-3} ($r= 0.32$); however, there is no significant
15 correlations between As-Mn ($r= 0.02$), and insignificant negative correlations between As and
16 Na^+ or K^+ . There was a good correlation between As-Mn, in addition, correlation between As-Fe
17 was also found in the fine particle core sediment sample by Anawar et al. (2003). However, in
18 this study, a significant positive correlation between As-Fe had existed but another study (Ohno
19 et al., 2005) showed no significant correlation between these two metals. However, NO_3^- did not
20 show correlation between two principle components except with Fe ($r= 38$) and PO_4^{-3} ($r= 37$).

21 The mechanism for the accumulation of As in the ground water is adsorption on and co-
22 precipitation with hydrous oxides (**Aggett and O'Brien, 1985**). The As derivatives (H_2AsO_4^- ,
23 HAsO_4^{2-} and H_3AsO_3) in contaminated ground water can easily be adsorbed on to hydrous iron
24 oxides that have a high pH and follow this reaction.



26 The Fig.5 revealed the stepwise linier regreesion between Fe-As, Mn-As, and Cu-As, and
27 revealed there were no significant relationships among them with very small r-square value. Fig.
28 6 also showed the regression with As and onther ions with very small r-square value except with
29 PO_4^{-3} . Thus the dissolved As can be declined to a very low concentration within a short distance
30 from its source. The BGS and DPHE technical report (**DPHE 2000**) have drawn the As Vs Fe
31 map of ground water for the whole country. They reported that often these 2 elements can be
32 strongly correlated. Indeed, on a well by well basis, the **Fe** concentration in well water generally
33 provides a poor predictor of **As** concentration. It is clear that Fe oxides are closely associated
34 with the development of high **As** ground water in Bangladesh.

35 The correlation matrix among As and various ions is summerised in Table 2. A significant
36 correlation revealed between As and PO_4^{-3} (**$r=0.32$**). Like As, PO_4^{-3} is strongly adsorbed by
37 amorphous iron oxide and is released under anaerobic condition (**Jacobs et al., 1970**). Other
38 anions were less effective in displacing As. The order of effectiveness decreasing for As (V) of
39 $\text{H}_2\text{PO}_4^- > \text{H}_2\text{AsO}_4^- > \text{SO}_4^{2-} > \text{CO}_3^{2-}$ and for As (III) of $\text{H}_2\text{PO}_4^- > \text{H}_3\text{AsO}_3 > \text{F}^- > \text{SO}_4^{2-} > \text{CO}_3^{2-}$.
40 **Acharyya et al. (2000; 2005)** hypothesized that As anions sorbet to aquifer sediments were
41 displaced to solution by competitive exchange with phosphate from the fertilizers. The
42 correlation of As and NO_3^- revealed negative but very weak (**$r=-0.02$**), and on the other hand,

1 Na⁺ and K⁺ showed negative weak correlation (**r=-0.15, r=-0.18**). As shown in Fig. 6 showed
2 the stepwise linear regression graph between Mg⁺²-As, NH₃-As, K⁺-As, Ca²⁺-As, NO₃⁻-As and
3 PO₄³⁻-As. They possess insignificant relation with As.

5 **Conclusion**

6 The study revealed that the mean As level in ground water in study area was lower compared to
7 the previous. Excessive withdrawal of ground water might be the reason towards the findings of
8 the study of As. However, due to mobilization of As, it might be deposited to the lower portion
9 of aquifers. Chemical weathering of biotitic and other basic minerals in the Holocene aquifer
10 could be a primary cause of As mobilization in ground water because of positive correlation of
11 As concentration with some nutrient elements. Iron (Fe) may come from ferro hydroxide
12 minerals, pyritic material and other rock forming primary minerals. For hunting safe and As free
13 pure water we need to know the correlation between As and other trace elements. Correlation of
14 As with Fe and Mn may also give an idea about the carrier of these elements. It means that Fe
15 and Mn are great adsorbent of As in solid phase, and ultimately causing the mobilization of As
16 along river flow. However, although in this study the correlation between As and Fe was
17 significant, that between As and Mn showed an insignificant correlation. The concentration may
18 increase or decrease due to some biogeochemical and mechanical factors. It is hard to say about
19 concrete indicator of As in groundwater using metals and other nutrient ions.

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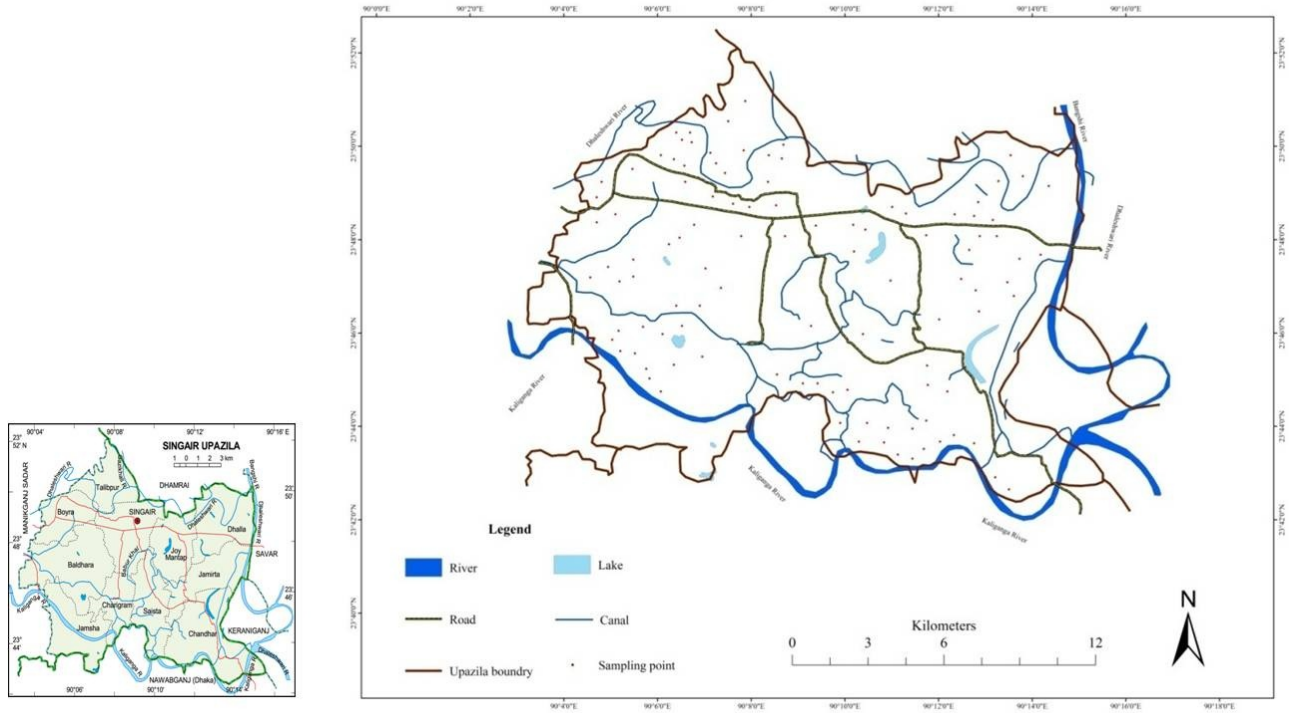
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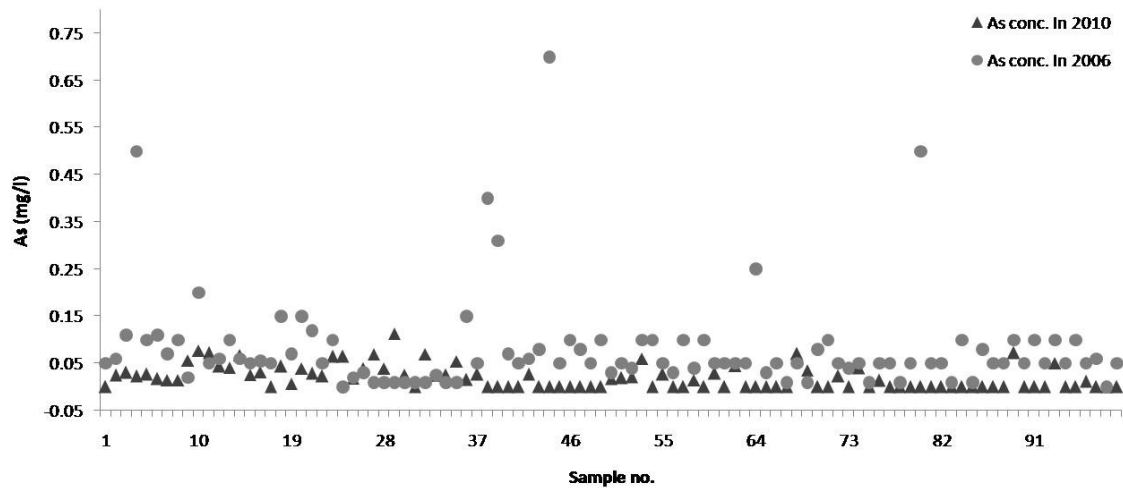


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Fig.1 Map of Singair, Manikganj Bangladesh (Source: Banglapedia, 2006) and Location of the sampling points in the GIS map

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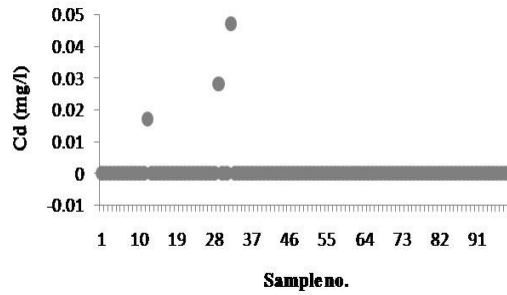
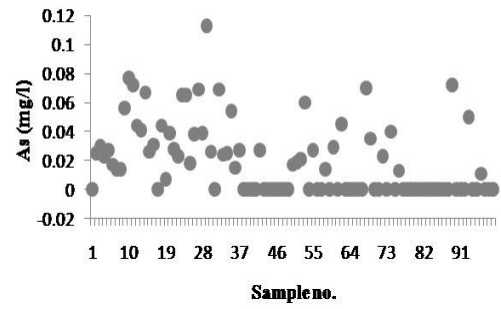
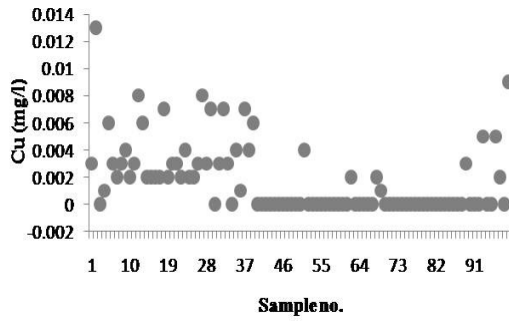
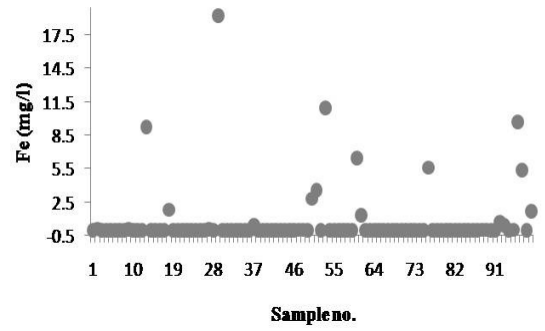
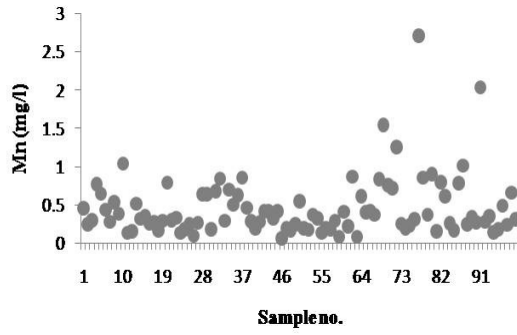
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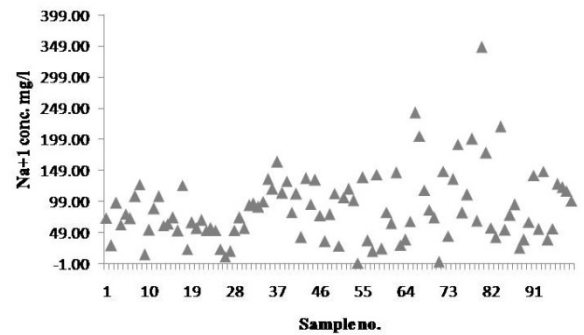
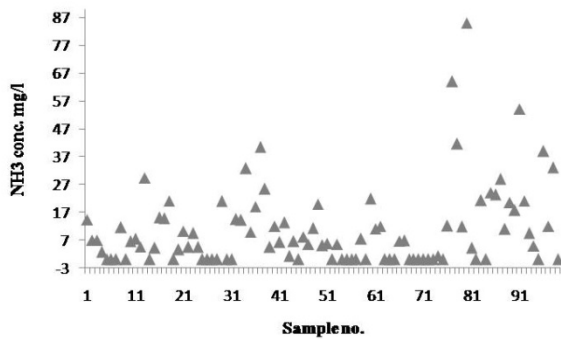
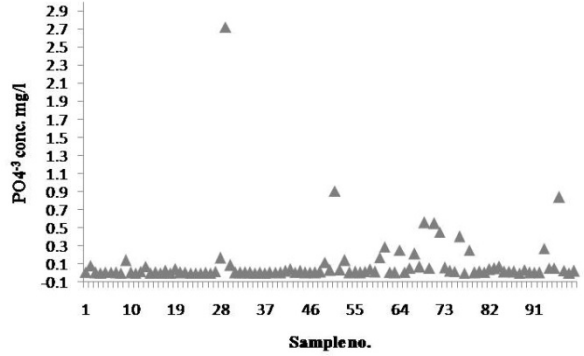
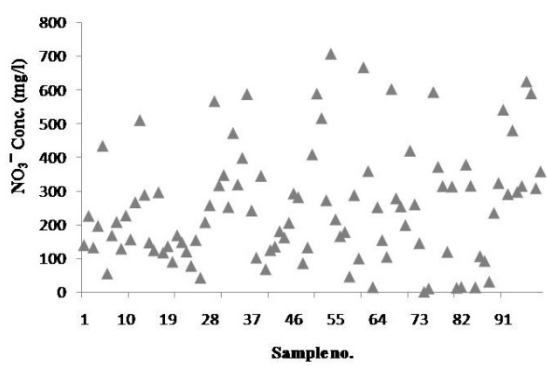
Fig.2 Concentration ranges of As (mg/l) in the study area in 2006 (Uddin et al 2006) and 2010.

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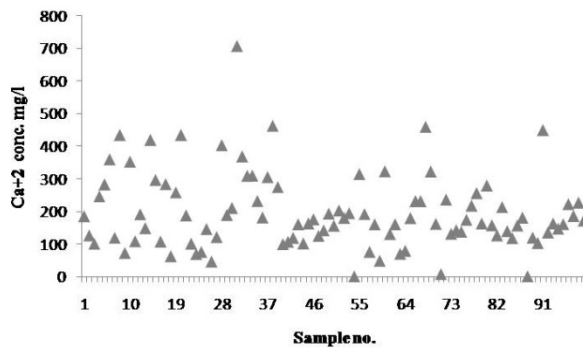
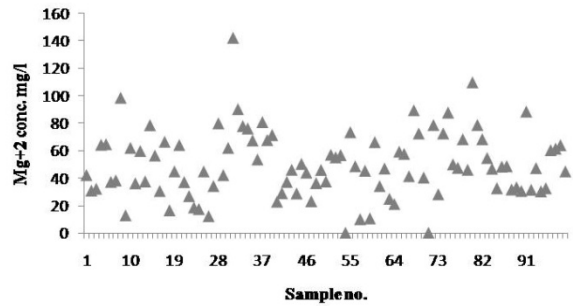
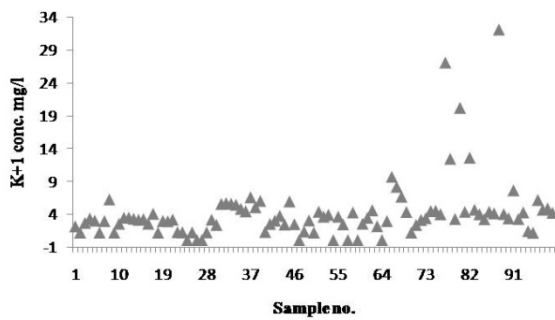


Fig.3 The range of concentration of a. metals (Mn, Cu, Fe, As and Cd) and b. ions (NO_3^- , PO_4 , NH_3 , Na^+ , K^+ , Mg^{2+} and Ca^{2+}) in different tube-well water samples in the study area.

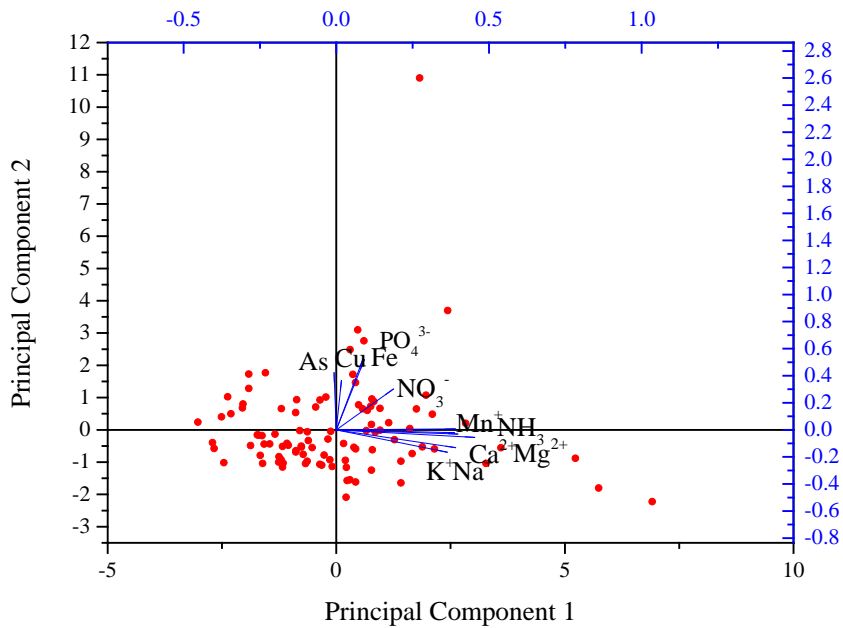
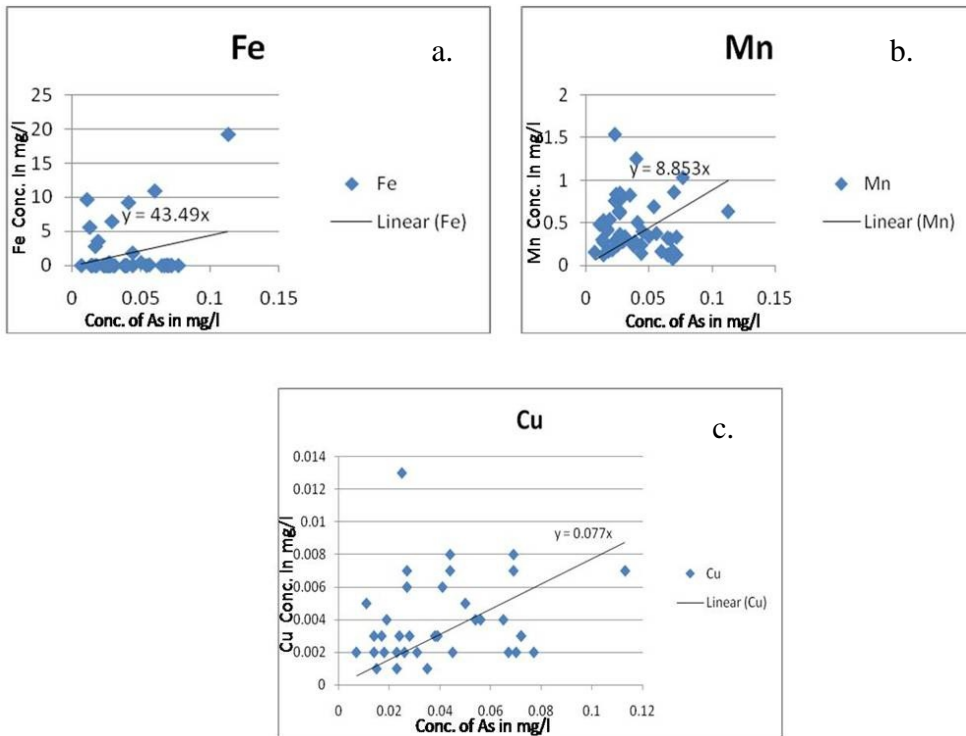
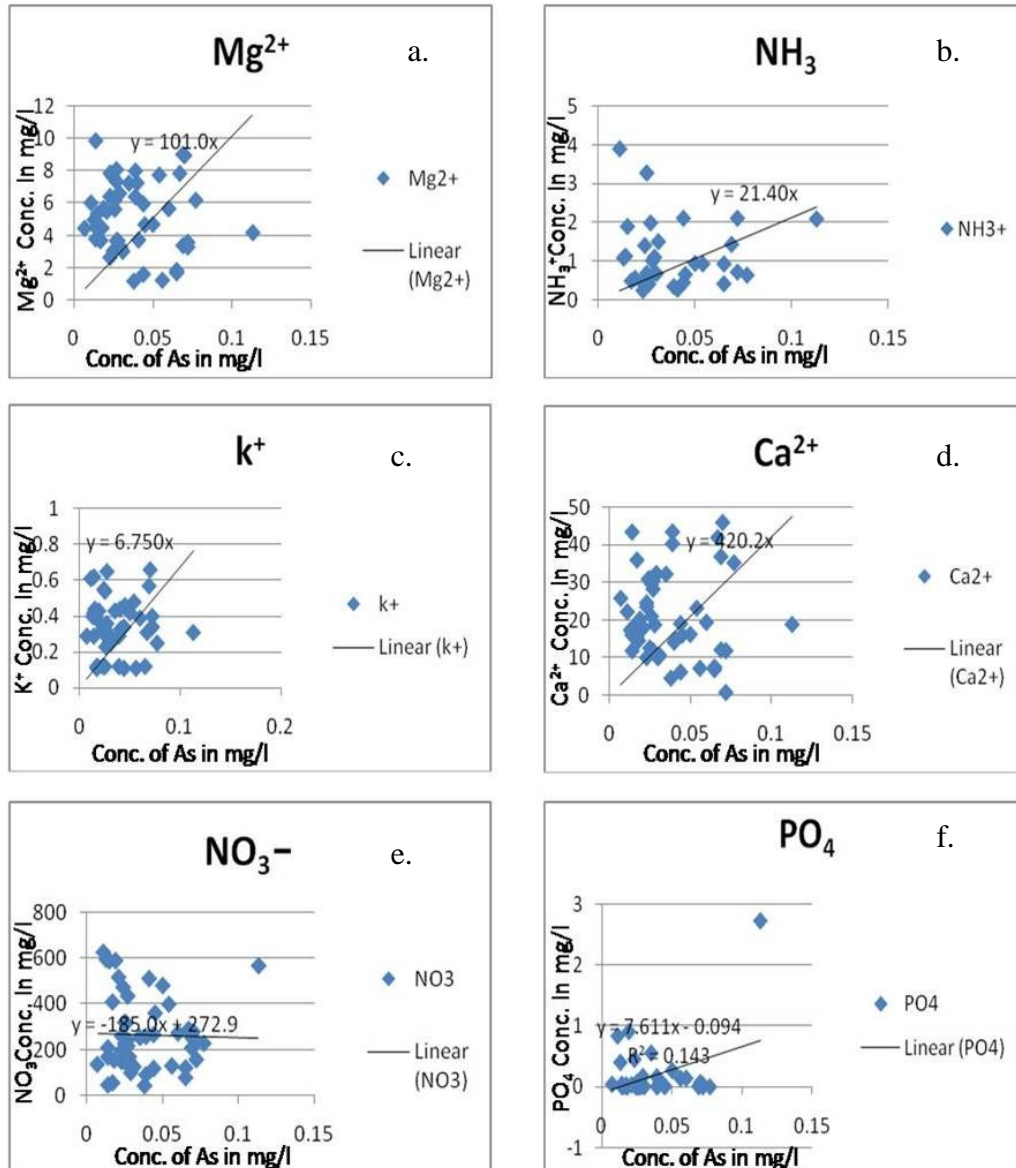


Fig.4 PCA biplot based on correlation matrix



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 2 Fig.5 Stepwise linear regression; a) Fe Vs As, b) Mn Vs As and c) Cu Vs As coefficients were
 3 insignificant with r^2 value 0.079, 0.001 and 0.040 respectively.
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25 Fig.6 Stepwise linear regression; a) Mg²⁺ Vs As, b) NH₃ Vs As, c) K⁺ Vs As d) Ca²⁺ Vs As e) NO₃⁻ Vs As f) PO₄³⁻ coefficients were insignificant except PO₄³⁻.