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## Renewable Energy and Carbon Management in the Cradle-to-Cradle Certification Limitations and Opportunities

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1 Journal of Industrial Ecology

2 **Type of contribution: Research and analysis**

3

4 **Renewable energy and carbon management in the Cradle-to-Cradle**  
5 **certification: Limitations and opportunities**

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7

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## 11 **Summary**

12 As part of the Cradle to Cradle® (C2C) certification program, the C2C certification criterion  
13 “Renewable Energy and Carbon Management” (RE&CM) focuses on use of electricity from RE  
14 and direct greenhouse gas offsets in the manufacturing stage and to a limited extent on the cradle to  
15 gate only at the highest level of certification. The aim of this study is to provide decision-makers  
16 with a quantified overview of possible limitations of that C2C certification requirement and  
17 potential gains by introducing a full Life Cycle Assessment (LCA) perspective to the scheme.  
18 Scenario analysis was used to perform an LCA of an aluminum can system representing different  
19 levels of the C2C certification criterion RE&CM, considering different strategies to achieve 100%  
20 RE in the manufacturing stage. The adoption of a broader life cycle RE perspective was considered  
21 through the implementation of electricity from renewable sources from cradle to grave. Our results  
22 show that compliance with the current RE&CM certification framework offers limited benefits, i.e.  
23 significant reduction for climate change but negligible reductions for other environmental impacts,  
24 e.g. particulate matter and acidification. However, increasing the share of RE in the primary  
25 aluminum production from a full life cycle perspective can greatly increase the environmental  
26 benefits brought up by the C2C certification, not only for climate change, but for the broader range  
27 of impact categories. In our striving towards environmental sustainability, which often cannot be  
28 approximated by climate change impacts alone, we therefore recommend decision-makers in  
29 industries to combine the C2C certification with LCA when they define strategies for the selection  
30 of renewable energy and raw materials suppliers.

31

32 **Keywords** (max 6): Life Cycle Assessment (LCA), aluminum, packaging, circular economy, C2C,  
33 decision support

## 34 <heading level 1> Introduction

35 With the current political and business emphasis on circular economy, defined as a restorative or  
36 regenerative industrial system by intention and design (EMF 2013; EC 2015), the Cradle-to-Cradle®  
37 (C2C) design framework has gained an increasing visibility in industry (Toxopeus et al. 2015). C2C  
38 is a design framework oriented towards product quality and innovation, aiming to maximize the  
39 overall benefits of products to ecological and economical systems by designing “eco-effective”  
40 solutions. C2C relies on three key principles: “waste equal food”, “use current solar income” and  
41 “celebrate diversity”(McDonough and Braungart 2002). Until now in the circular economy context,  
42 efforts have largely focused on implementing the former, i.e. attempting to shift from a waste  
43 paradigm to a resource one, where waste from some industries can serve as resources for others. To  
44 allow companies to monitor and market their progress in C2C compliance, a certification program  
45 known as the Cradle to Cradle Certified™ Product Standard was established and recently updated  
46 (Cradle to Cradle Products Innovation Institute 2016). Applicants have to comply with a series of  
47 requirements for five categories: material health (MH), material reutilization (MR), renewable  
48 energy and carbon management (RE&CM), water stewardship (WS) and social fairness (SF), each  
49 of them being scored on a 5-grade scaling system, i.e. basic, bronze, silver, gold and platinum,  
50 reflecting an increased stringency in the C2C requirements.

51 As already discussed in past studies (Bjørn and Hauschild 2013; Toxopeus et al. 2015), the trade-off  
52 between resource conservation and energy use is a weakness of the C2C design framework and  
53 therefore of the certification program. For all the grades but platinum in the scaling system, only  
54 electricity use and greenhouse gases (GHGs) emissions in the manufacturing stage of a product are  
55 thus considered. This means that for most of the grades the environmental impacts stemming from  
56 the raw materials extraction, production, construction, and decommissioning of the energy  
57 generation facilities are disregarded even though those stages may be important drivers of the  
58 environmental impacts. This is particularly relevant for energy production based on renewable

59 energy sources e.g. wind (Dolan and Heath 2012; Turconi et al. 2013; Asdrubali et al. 2015) or  
60 solar power (Hsu et al. 2012; Turconi et al. 2013; Asdrubali et al. 2015), for which life cycle  
61 assessment (LCA) showed relatively low environmental impacts in the use/operation stage  
62 compared to their production and disposal stages. Moreover, the need to include a broader range of  
63 impact categories than climate change to gauge the environmental sustainability when shifting  
64 electricity production from the use of fossils to the use of renewables has been pointed out (Laurent  
65 et al. 2012; Hertwich et al. 2014).

66 These gaps in the RE&CM requirement of the C2C certification can induce important biases in the  
67 decision making process for companies, who might not be aware of such limitations and associated  
68 uncertainties. In this study, we therefore aim to provide decision-makers in industry a quantified  
69 overview of possible limitations of the RE&CM requirement and potential gains by introducing a  
70 full LCA perspective, as well as recommendations to alleviate these shortcomings in decision-  
71 making processes.

72 We build on the results of an existing LCA of aluminum beverage cans (termed “AIC system” in  
73 the following) (Niero et al. 2016). We focus on (primary) aluminum production, which belongs to  
74 those sectors, where energy consumption during manufacturing is an environmental hotspot of the  
75 technologies and systems (EAA 2013) and is thus fully relevant from a life cycle perspective (Liu  
76 and Müller 2012). Due to its rapid aluminum industry development in the last decade, China has  
77 become the largest primary aluminum producer in the world and now faces urgent needs to reduce  
78 associated environmental impacts (Sun et al. 2015). As a result, the longtime front-runner in  
79 aluminum production, Europe, is today the second largest producer ([http://www.world-  
80 aluminium.org/statistics/](http://www.world-aluminium.org/statistics/)). Chinese and European aluminum productions differ from each other with  
81 regard to their supporting electricity mixes, which are mainly based on coal and hydropower,  
82 respectively. In the current study, we therefore consider different can systems including either  
83 China or Europe as aluminum-producing countries: (1) the AIC system as commonly in place  
84 (baseline scenario), (2) the AIC system with implementation of the C2C certification requirement at

85 the highest grades for the RE criterion (i.e. gold and platinum) using alternative renewable energy  
86 sources, and (3) the AIC system with adoption of a broader life cycle RE perspective, i.e.  
87 implementation of electricity from renewable sources from cradle to grave.

88

## 89 <heading level 1> **Materials and methods**

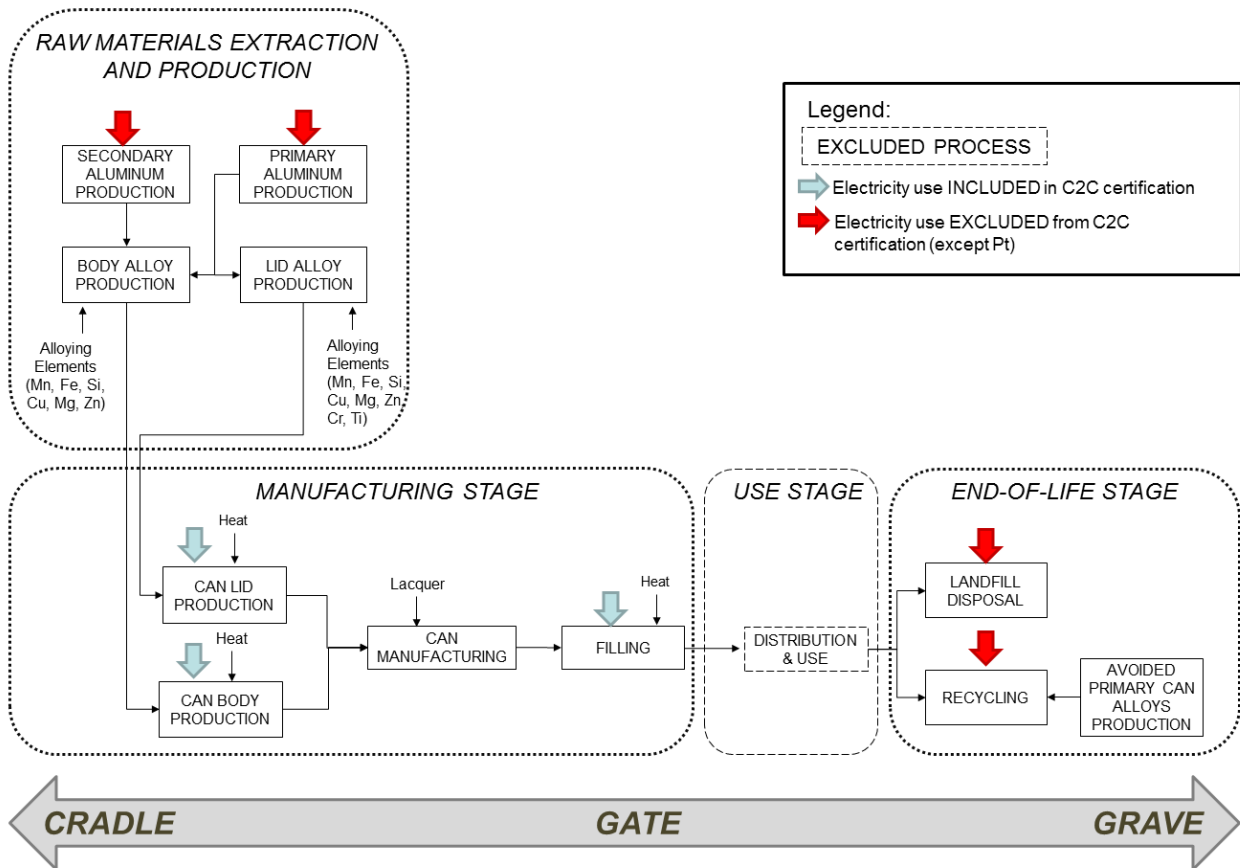
90 Aluminum cans systems have recently been evaluated by means of LCA, e.g. van der Harst and  
91 colleagues (2015). We consider here an AIC system for beer containment in the UK market  
92 previously used to model 20 different scenarios complying to different degrees with two of the C2C  
93 certification requirements, namely RE and MR (Niero et al. 2016). We followed the requirements of  
94 the ISO 14040-44 standards (ISO 2006a, 2006b) and the technical guidance provided by  
95 International Reference Life Cycle Data System (ILCD) handbook (EC-JRC-IES 2010). We also  
96 used the approach from the product environmental footprint (PEF) guide (EC 2013) to model the  
97 end-of-life (EoL) as it is the one recommended in the context of policy support applications  
98 (Allacker et al. 2014).

99

## 100 <heading level 2> **Goal and scope of the LCA**

101 The goal of the LCA study is to compare different AIC systems, some representative of different  
102 level of compliance with the requirements of the C2C certification for the RE&CM criterion, and  
103 some going beyond C2C certification through the inclusion of RE in a life cycle perspective,  
104 considering the average primary aluminum production in either Europe or China. Since the aim of  
105 the LCA is to provide decision support related to product development, with small scale changes in  
106 the background system, i.e. in terms of energy supply and material supplier, then the decision  
107 context is a situation A type according to the ILCD Handbook, i.e. micro-level decision support  
108 (EC-JRC-IES 2010). The considered functional unit is “the containment of 1 hl of beer until the

109 expiry date”, in accordance with the draft version of the PEF category rules for beer published in  
110 the context of the beer PEF pilot (Technical Secretariat for the Beer Pilot 2015). In the case of 33 cl  
111 aluminum cans, with average weight of 13.5 g, 4.22 kg of material per functional unit is required  
112 (Niero et al. 2016). Only the primary packaging, i.e. the materials which come into direct contact  
113 with the product, is considered, being the object of the C2C certification. The product system under  
114 study includes the supply of raw materials, i.e. the aluminum alloys used for the lid (21% of the can  
115 weight) and the body (79% of the can weight), the manufacturing of the lid and body, as well as  
116 their assembly, the filling of the can and its final EoL. The system boundaries are presented in  
117 Figure 1: the main exclusions regard the distribution and use stages, since these are assumed  
118 identical for all compared systems. The influence of transports on the overall environmental impact  
119 of AIC systems cans is usually minor compared to the other life cycle stages, and the use stage, e.g.  
120 refrigeration of the beverage, is typically not included as it is assumed that the beverage is  
121 consumed shortly after the purchase (Amienyo and Azapagic 2016).  
122



123  
 124  
 125

**Figure 1. Life cycle stages of the aluminum can (AIC) system considered, with indication of main inputs and outputs; excluded processes are marked with dashed box.**



## 126 <heading level 2> Data collection and system modelling

127 In line with the identified decision context situation A of the ILCD guidelines (see previous  
128 section), the Life Cycle Inventory (LCI) modelling framework chosen is attributional with the use  
129 of system expansion to model process multi-functionality (EC-JRC-IES 2010). The ecoinvent 3.1  
130 database (attributional default) was used to build the LCI (Weidema et al. 2013), considering the  
131 avoided impacts from the average market situation to model recycling at the EoL (EC-JRC-IES  
132 2010). The LCI model was built in the LCA software SimaPro v.8.0.4.30 (PRé 2013). The  
133 foreground system was modelled with primary data, e.g. using electricity and heat consumption data  
134 from the filling facilities (Niero et al. 2016), while secondary data were used for modelling the  
135 background system, i.e. primary aluminum production (Stichling and Nguyen-Ngoc 2009), can  
136 manufacturing (e.g. lacquering (Li and Qiu 2013)) and the EoL management, which includes  
137 recycling and landfilling (Stichling and Nguyen-Ngoc 2009). We considered the current recycling  
138 rate for aluminum cans in UK (65%) (EAA 2015), and an average % of recycled content of 67.8%  
139 (PE Americas 2010). We used the default ecoinvent 3.1 datasets to model the input materials, i.e.  
140 average primary aluminum production both in Europe and China, secondary aluminum production,  
141 lacquer composition and EoL treatment. We modelled the can components, i.e. body and lid,  
142 according to their actual aluminum alloy composition: AA5182 for the lid and AA3004 for the  
143 body, respectively, as suggested in Niero and Olsen (2016). We considered the maximum threshold  
144 values of alloying elements allowed for the two abovementioned alloys (The University of  
145 Liverpool 2015). The main modifications to the default datasets in the scenario analysis consisted of  
146 changed energy input, i.e. the electricity used during manufacturing, primary aluminum production  
147 and recycling.

148

**<heading level 2> The C2C requirement for RE&CM**

150 The intention of the RE&CM category of the C2C certification program is “*to provide a*  
151 *quantitative measure of the percentage of renewably generated energy that is utilized in the*  
152 *manufacture of the product. Purchased electricity and direct on-site emissions associated with the*  
153 *final manufacturing stage of the product, as well as embodied energy associated with the product*  
154 *from Cradle to Gate, are considered, depending on the level of certification*” (Cradle to Cradle  
155 Products Innovation Institute 2016). The product under analysis is indeed graded based on  
156 quantitative parameters, e.g. the proportion of electricity coming from renewable sources (termed  
157 “% RE” in the following) and the proportions of direct GHGs emissions which are offset (named  
158 “% GHGs offset” in the following), and qualitative ones, e.g. the development of strategy for  
159 energy use and carbon management. Direct GHG emissions in scope for this requirement are those  
160 that are either emitted directly during the product’s final manufacture or on-site treatment of process  
161 wastes or associated with purchased heat (Cradle to Cradle Products Innovation Institute 2016). We  
162 focus here on the quantitative aspects and consider the highest levels for the RE&CM criterion, i.e.  
163 gold (G) and platinum (P), which are achieved if the manufacturing stage of the product (see Fig.  
164 1), meets the two following conditions: (i) 50% (for gold) and 100% (for platinum) of purchased  
165 electricity is renewably sourced or offset with renewable energy projects, and (ii) the same  
166 proportions of direct on-site GHG emissions are offset (Cradle to Cradle Products Innovation  
167 Institute 2016). For the platinum level additional requirements apply that comprise the supply chain,  
168 and therefore a “cradle to gate” perspective (see Fig. 1): iii) the accounting of the embodied GHG  
169 emissions; (iv) the definition of a strategy to optimize the embodied GHG of the product; (v) the  
170 coverage of at least 5% of the embodied energy associated with the product (cradle to gate) by  
171 offsets or other mechanisms, e.g. projects with suppliers, product re-design, savings during the use  
172 phase (Cradle to Cradle Products Innovation Institute 2016).

173

174 <heading level 2> **Scenarios definition**

175 In the development of the scenarios we refer to the % RE in the electricity mix used for the life  
176 cycle stages included in the C2C certification up to platinum level, i.e. body and lid manufacturing  
177 and filling, as well as the stages included only partially in the platinum level, i.e. primary aluminum  
178 production and excluded from the certification, i.e. the EoL, as shown in Fig. 1. Since can  
179 manufacturing and filling are assumed to take place in the UK, the electricity mix for UK was used  
180 taking the defaultecoinvent 3.1 unit processes as a starting point. In the C2C certification program,  
181 renewable electricity that is already a standard part of the grid mix does not count toward this  
182 requirement, “*unless the applicant is participating in a voluntary green pricing program or the*  
183 *applicant has verified that their utility is delivering renewable electricity that may be claimed by the*  
184 *utility customer without being double-counted elsewhere in the system*” (Cradle to Cradle Products  
185 Innovation Institute 2016). For the AIC system under study we assumed that the applicant is  
186 involved in a voluntary green pricing program. To take into account of the variability in renewable  
187 energy sources different scenarios, including different mixes of renewable electricity, were  
188 considered for the highest certification level, i.e. platinum. In terms of direct GHGs we accounted  
189 only for those associated with purchased heat by the utility during the manufacturing stage (Fig. 1)  
190 and deducted their contribution from the climate change impact category, according to the  
191 requirements set by the certification level (i.e. 50% for gold and 100% for platinum). No impacts  
192 generating from the actions undertaken to provide the offset are considered, therefore the case  
193 modelled represents the best case scenario. Table 1 provides an overview of all scenarios assessed  
194 in the study. The details of the datasets used in the LCI modelling for the electricity from RE,  
195 primary aluminum production and heat (from natural gas) are reported in Table S1, in the  
196 Supplementary Information (SI).

197

198 **Table 1 Summary of the 16 scenarios considered for the aluminum can (AIC) system, where EU refers**  
 199 **to manufacturing in Europe and CN in China, respectively. The detail of the ecoinvent datasets used is**  
 200 **reported in Table S1, in the Supplementary Information (SI).**

Designation	Scenario description (In brackets the C2C certification level)	Primary Al alloy production	Manufacturing	
			Electricity mix	GHGs offset
1-B-EU	Baseline AIC system in Europe	Default Europe	Current (2015) UK el. mix (21% RE) <sup>a</sup>	-
1-B-CN	Baseline AIC system in China	Default China		
2-C2C/G-EU	AIC in Europe (Gold)	Default Europe	Current (2015) UK el. mix adjusted to 50% RE <sup>b</sup>	50% GHGs from heat
2-C2C/G-CN	AIC in China (Gold)	Default China		
3-C2C/P(2015UK)-EU	AIC in Europe (Platinum) with current UK mix	Default Europe	Current (2015) UK el. mix adjusted to 100% RE <sup>c</sup>	100% GHGs from heat
3-C2C/P(2015 UK)-CN	AIC in China (Platinum) with current UK mix	Default China		
4-C2C/P(solar)-EU	AIC in Europe (Platinum) with 100% solar energy	Default Europe	100% solar energy (single-Si)	100% GHGs from heat
4-C2C/P(solar)-CN	AIC in China (Platinum) with 100% solar energy	Default China		
5-C2C/P(wind)-EU	AIC in Europe (Platinum) with 100% wind energy	Default Europe	100% wind energy (on-shore >3MW)	100% GHGs from heat
5-C2C/P(wind)-CN	AIC in China (Platinum) with 100% wind energy	Default China		
6-LC(2015 UK)-EU	AIC in Europe, current UK mix (100%RE) + life cycle perspective for RE	Europe with 100% RE <sup>d</sup>	Current (2015) UK el. mix adjusted to 100% RE <sup>e</sup>	100% GHGs from heat
6-LC(2015 UK)-CN	AIC in China, current UK mix (100%RE) + life cycle perspective for RE	China with 100% RE <sup>e</sup>		
7-LC(solar)-EU	AIC in Europe, (100% solar) + life cycle perspective for RE	Europe with 100% RE <sup>d</sup>	100% solar energy (single-Si)	100% GHGs from heat
7-LC(solar)-CN	AIC in China, (100% solar) + life cycle perspective for RE	China with 100% RE <sup>e</sup>		
8-LC(wind)-EU	AIC in Europe, (100% wind) + life cycle perspective for RE	Europe with 100% RE <sup>d</sup>	100% wind energy (on-shore >3MW)	100% GHGs from heat
8-LC(wind)-CN	AIC in China, (100% wind) + life cycle perspective for RE	China with 100% RE <sup>e</sup>		

201

202 <sup>a</sup> Based on UK-DECC (2014)203 <sup>b</sup> Based on the current (2015) mix distribution (UK-DECC 2014), but adjusted with 50% RE, i.e. wind  
204 (34%), heat and power co-generation from biogas (12%) and biomass (2%), hydro (2%)205 <sup>c</sup> Based on the current (2015) mix distribution, but with adjusted with 100% RE, i.e. wind (67%), heat and  
206 power co-generation from biogas (23%) and biomass (5%), hydro (5%).207 <sup>d</sup> Modelled considering 80% hydropower and 20% wind power208 <sup>e</sup> Modelled considering 10% hydropower and 90% wind power

209

210

### 211 <heading level 3> **Baseline scenarios**

212 For the baseline scenarios (1-B-EU, 1-B-CN) we considered the current UK electricity mix for the  
213 manufacturing processes based on the reference scenario described by the UK Department of  
214 Energy and Climate Change (UK-DECC 2014). This leads to the following distribution: hard coal  
215 (35%), natural gas (26%), nuclear (17%), oil (1%), wind (14%), heat and power co-generation from  
216 biogas (5%) and biomass (1%), hydro (1%). Overall, the share of RE is equal to 21%.

217

### 218 <heading level 3> **Gold scenarios**

219 For the scenarios representing gold certification (2-C2C/G-EU, 2-C2C/G-CN), the relative  
220 distribution of each renewable and non-renewable energy source as in the current electricity mix  
221 were kept and adjusted so that the aggregated contributions of RE and non RE sources amount to  
222 50:50% of the modelled electricity mix, respectively. Therefore, the electricity mix considered is:  
223 hard coal (22%), natural gas (16.4%), nuclear (11%), oil (0.6%), wind (34%), heat and power co-  
224 generation from biogas (12%) and biomass (2%), hydro (2%).

225

### 226 <heading level 3> **Platinum scenarios**

227 With regard to platinum certification, corresponding to 100% RE, different scenarios were built  
228 using: the current UK RE mix (for scenarios 3-C2C/P(2015UK)-EU and 3-C2C/P(2015 UK)-CN);  
229 100% RE from solar energy (for 4-C2C/P(solar)-EU and 4-C2C/P(solar)-CN), represented by the  
230 single-Si panel technology (choice subject to limited data availability) and 100% RE from wind  
231 source (5-C2C/P(wind)-EU; 5-C2C/P(wind)-CN), assuming on-shore wind technology, which  
232 currently is more mature and is foreseen to support a larger share of electricity generation than off-  
233 shore wind power technology (IEA 2013). The selection of wind and solar power was motivated by  
234 their relatively lower reported environmental impacts, their anticipated role in future electricity  
235 generation landscapes and their contribution to energy security (Asdrubali et al. 2015;

236 Hosenuzzaman et al. 2015). Moreover, to model the offset of 5% of the embodied energy associated  
237 with the product, we reduced by 5% the electricity used in the manufacturing stage.

238

### 239 <heading level 3> Life cycle scenarios

240 For including the life cycle perspective from cradle to grave in the AIC system we considered  
241 additional sets of scenarios, built on the platinum certification level, but assuming an increase of RE  
242 for electricity in the raw materials extraction and production (i.e. prior to the manufacturing stage)  
243 and EoL stages (see Fig. 1). Primary aluminum production is very energy intensive, and hence the  
244 location of production plants is often determined by access to large amounts of cheap electricity,  
245 which often results in an electricity mix that is different from the general grid mix of the countries  
246 where the production plants are located, thus aluminum industry specific electricity markets were  
247 used (Moreno Ruiz et al. 2014). Therefore, we modified the electricity mix used in primary ingot  
248 production, as well as in the primary liquid aluminum production (including bauxite mine  
249 operations, Al hydroxide, Al oxide). We assumed that the current fraction of non-renewable energy  
250 (termed “non-RE” in the following) in the aluminum specific electricity markets, i.e. 20% for  
251 Europe and 90% for China, can be substituted with the most competitive RE source, i.e. on-shore  
252 wind power (IEA 2013). An extension to 100% RE from hydropower is not deemed realistic due to  
253 the already high exploitation of hydropower capacity in these regions, e.g. hydropower is already  
254 extensively developed and with little further expansion potential left in Europe (IEA 2012). This  
255 leads to electricity mixes of 80:20% and 10:90% hydropower (current share):wind power sources  
256 throughout primary aluminum production in Europe and China, respectively. These grid mixes, in  
257 particular that of China, should be regarded as explorative mixes as a share of 90% wind power in  
258 the electric mix in China would imply effective grid management systems including storage  
259 capacity, which are not encompassed in this study due to lack of data.

260

### 261 <heading level 2> Life Cycle Impact Assessment and sensitivity analysis

262 The life cycle impact assessment (LCIA) was performed using the ILCD 2011 recommended  
263 methodology v1.05 (Hauschild et al. 2013) as embedded in SimaPro LCA software (PRÉ  
264 Consultants, 2015). The covered impact categories include climate change (CC), stratospheric  
265 ozone depletion (OD), human toxicity, considering both cancer effects (HT-c) and non-cancer  
266 effects (HT-nc), particulate matter (PM), ionizing radiation impacting human health (IR-HH),  
267 photochemical ozone formation (POF), acidification (Ac), terrestrial eutrophication (TE),  
268 freshwater eutrophication (FE), marine eutrophication (ME), freshwater ecotoxicity (FET), water  
269 use (WU), land use (LU) and resource depletion (RD), including mineral, fossil and renewable  
270 resources. To assess water use, significant advances have been made since the review of the  
271 methods leading to the ILCD recommendations was conducted; the water scarcity index (WSI)  
272 method developed by Pfister and colleagues (2009) was thus considered instead. Furthermore, given  
273 the focus on the energy aspect we considered the non RE Cumulative Energy Demand (CED) v1.09  
274 (Frischknecht et al. 2007) as an LCI indicator since it has been proven to provide insights for  
275 product comparison in the beverage packaging sector (Scipioni et al. 2013). To illustrate the  
276 differences observed between the different scenarios for each impact category, we performed  
277 ‘division-by-baseline’ internal normalization, i.e. dividing results obtained for a given impact  
278 category for each scenario by the corresponding impact results of the baseline scenario (thus taken  
279 as a reference) (Laurent and Hauschild 2015). This enables to quantify the impact reductions  
280 brought by the implementation of the different scenarios compared to the baseline scenario. We  
281 assumed a cut-off of 10% to identify a significant difference among the alternatives, following e.g.  
282 Humbert et al. (2009); this cut-off was arbitrarily defined and does not necessarily reflect the actual  
283 uncertainty assessment. As a sensitivity check at the impact assessment level, a different LCIA  
284 methodology, i.e. ReCiPe 2008 midpoint, hierarchist v.1.11 (Goedkoop et al. 2009) was  
285 additionally used.

286

287 <heading level 1> **Results and discussion**

288 The detailed characterized impact scores for the aluminum can system are reported in Table S2 and  
 289 S3 for Europe and China, respectively in the SI. Table 2 reports the normalized results of the  
 290 progression from the baseline to the C2C gold and platinum grades and “life cycle” scenarios (as  
 291 defined in Table 1) for 4.22 kg of aluminum cans manufactured in the UK with primary aluminum  
 292 produced in Europe. Normalized impact results for China are reported in a similar way in Table S4  
 293 in the SI.

294 **Table 2 Normalized impact scores for the aluminum can system (indexed based on the**  
 295 **baseline scenario) according to different C2C certification levels and scenarios defined in**  
 296 **Table 1 (Europe). The color coding is used to indicate the ranking of the scenarios, where the**  
 297 **option with higher environmental impact are marked with red and the one with the lower**  
 298 **environmental impact are marked with green, according to the following legend: above 1.10**  
 299 **(dark red); 1.00-1.09 (red); 0.90-0.99 (orange); 0.80-0.89 (dark yellow); 0.70-0.79 (yellow);**  
 300 **0.60-0.69 (light green); 0.50-0.59 (green); below 0.49 (dark green). Results for China are**  
 301 **provided in Table S4.**

Impact category	1- B	2- C2C/G	3- C2C/P (2015UK)	4- C2C/P (solar)	5- C2C/P (wind)	6- LC (2015UK)	7- LC (solar)	8- LC (wind)
	-EU	-EU	-EU	-EU	-EU	-EU	-EU	-EU
Climate change (CC)	1.00	0.78	0.54	0.54	0.53	0.47	0.47	0.46
Ozone depletion (OD)	1.00	0.99	0.97	0.99	0.96	0.91	0.93	0.90
Human toxicity, cancer (HT-c)	1.00	1.00	1.01	1.00	1.00	1.01	1.01	1.00
Human toxicity, non cancer (HT-nc)	1.00	0.99	0.98	1.01	0.99	0.97	1.00	0.98
Particulate matter (PM)	1.00	0.99	0.96	0.98	0.95	0.86	0.87	0.84
Ionizing radiation, human health (IR-HH)	1.00	0.94	0.83	0.84	0.82	0.62	0.62	0.61
Photochemical Ozone Formation (POF)	1.00	0.98	0.94	0.94	0.92	0.87	0.87	0.85
Acidification (Ac)	1.00	0.97	0.93	0.93	0.91	0.84	0.85	0.82
Terrestrial Eutrophication (TE)	1.00	0.98	0.95	0.92	0.90	0.87	0.85	0.82
Freshwater Eutrophication (FE)	1.00	0.99	0.98	1.00	0.98	0.89	0.91	0.89
Marine eutrophication (ME)	1.00	1.00	1.01	0.91	0.89	0.93	0.83	0.81
Freshwater Ecotoxicity (FET)	1.00	1.02	1.05	1.00	0.99	1.04	0.99	0.98
Land Use (LU)	1.00	1.05	1.14	0.93	0.92	1.13	0.92	0.91
Resource depletion (RD)	1.00	1.00	1.01	1.06	1.00	1.01	1.06	1.01
Non Renewable Cumulative Energy Demand (Non-RE CED)	1.00	0.97	0.91	0.92	0.90	0.82	0.83	0.81
Water Scarcity Index (WSI)	1.00	1.04	1.10	1.01	0.93	1.07	0.98	0.90

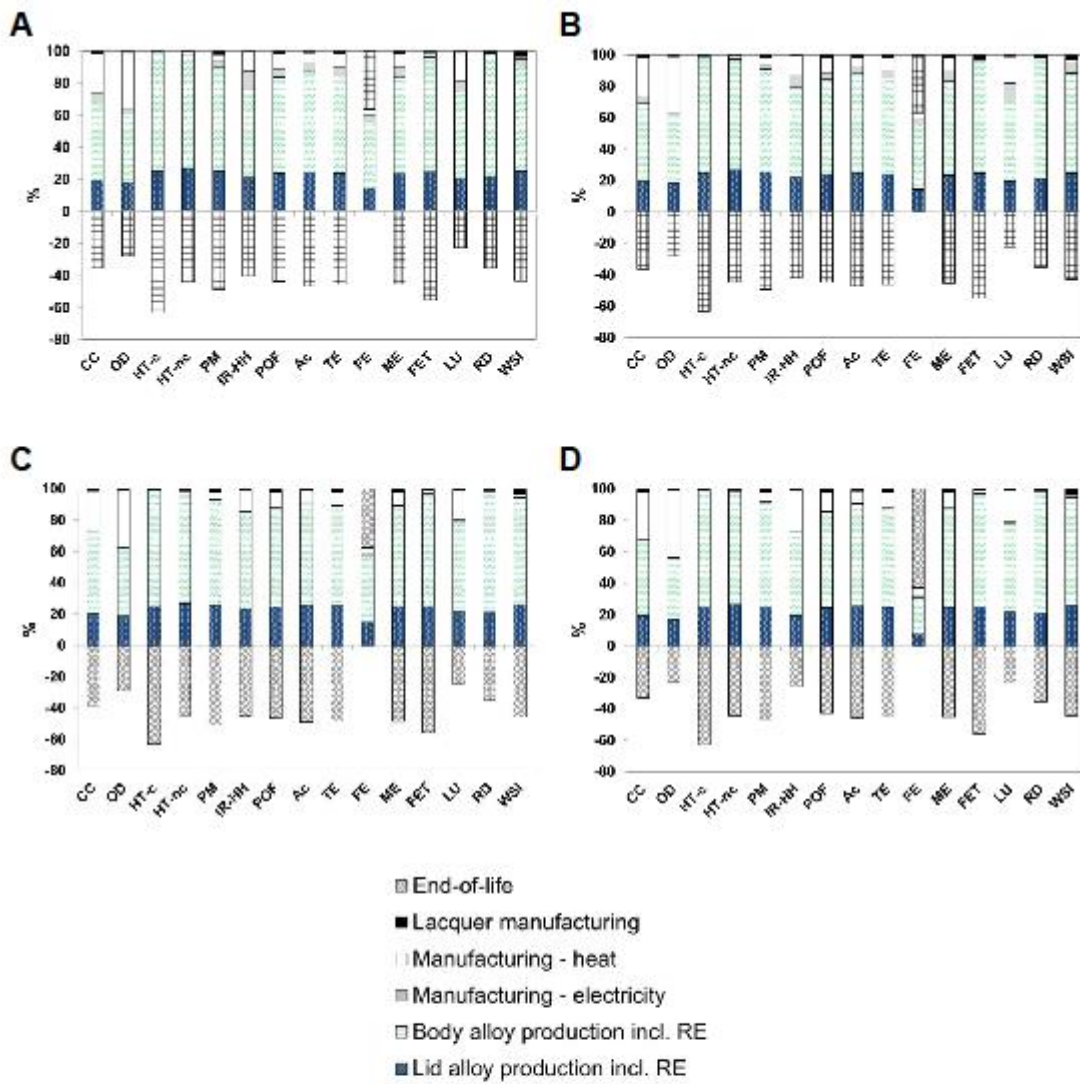
302



303 **<heading level 2> What environmental impact reductions can C2C certification achieve?**

304 Only considering the scenarios relating to the C2C certification, i.e. scenarios 2-5, a common trend  
305 across the different certification scenarios can be identified for the impact categories in Table 2,  
306 which can be divided in three groups. A first group includes IR-HH and CC, which present  
307 significant reductions among scenarios 2-5, i.e. 22% for gold and ca 45% for platinum. A second  
308 group includes OD, PM, POF, Ac, TE, FE, ME, non RE CED which shows a slightly decreasing  
309 but not significant reduction in impact scores from the baseline towards gold and platinum  
310 certification (i.e. below 10%). The third group includes the toxicity related and resource-related  
311 impact categories, i.e. HT-c, HT-nc, FET, LU, RD and WSI, which show a slightly increasing  
312 difference (maximum 14% for LU) towards the higher certification levels, except for the last  
313 platinum scenario (i.e. number 5).

314 The marginal impact reduction (except for CC and IR-HH) across the C2C certification scenarios  
315 can be explained by the contribution analysis per life cycle stage for the baseline scenario (Fig. 2a),  
316 the gold certification scenario (Fig. 2b), and the platinum certification scenario considering 100%  
317 wind energy during manufacturing (i.e. the platinum scenarios with lower impacts; Fig. 2c). The  
318 positive contribution of the electricity use to environmental impacts during manufacturing, i.e. the  
319 one included in the C2C certification, is negligible compared to the contributions from heat during  
320 manufacturing and from lid and body alloys production, which include all the upstream processes  
321 such as alumina refining and electrolysis. These three life cycle stages thus represent the hotspots  
322 across all impact categories, as reported in the most recent LCA of primary aluminum production,  
323 which identified electricity and thermal energy as the factors responsible for the large contribution  
324 of alumina refining and electrolysis to GHGs emissions (Nunez and Jones 2015). The negative  
325 values observed in Fig. 2 refer to the End-of-life phase, namely to the avoided environmental  
326 impacts due to recycling.



327

328 **Figure 2 Contribution analysis for a selection of scenarios relative to Europe A) 1-B**  
 329 **(baseline); B) 2-C2C/G; C) 5-C2C/P (wind); D) 8-LC (wind). Lid and body alloys production**  
 330 **include all the upstream processes, as presented in Figure 1.**

331

332 <heading level 2> **Importance of including a full life cycle perspective in RE&CM criterion**

333 The normalized impact results for scenarios 6-8 indexed on the baseline scenario are presented in  
334 Table 2 (for EU) and Table S4 in the SI (for CN). They reflect the changes when including a full  
335 life cycle perspective, i.e. from cradle to grave, for electricity use in the AIC systems. The same  
336 trends can overall be identified for EU and CN, with some discrepancies. Significant deviations of  
337 the environmental impacts from the baseline (i.e. higher than 10%) are generally observed, with the  
338 exceptions of OD and HT-nc (for EU), for which the decrease is below the cut-off, and HT-c, FET,  
339 RD and LU (except 6-LC(2015UK)-EU), for which relatively negligible increases in the  
340 environmental impacts can overall be observed. For RD, the increase is mainly due to an increased  
341 metal extraction specific to PV, as confirmed by the sensitivity analysis performed with ReCiPe,  
342 which distinguishes between fossil and metal depletion, see Tables S5 and S6. For WSI, the  
343 inclusion of renewable energy highly based on hydropower in the primary Al production (scenarios  
344 6) causes an increase of the environmental impact compared to the other scenarios with lower  
345 shares of hydropower (i.e. scenarios 7 and 8).

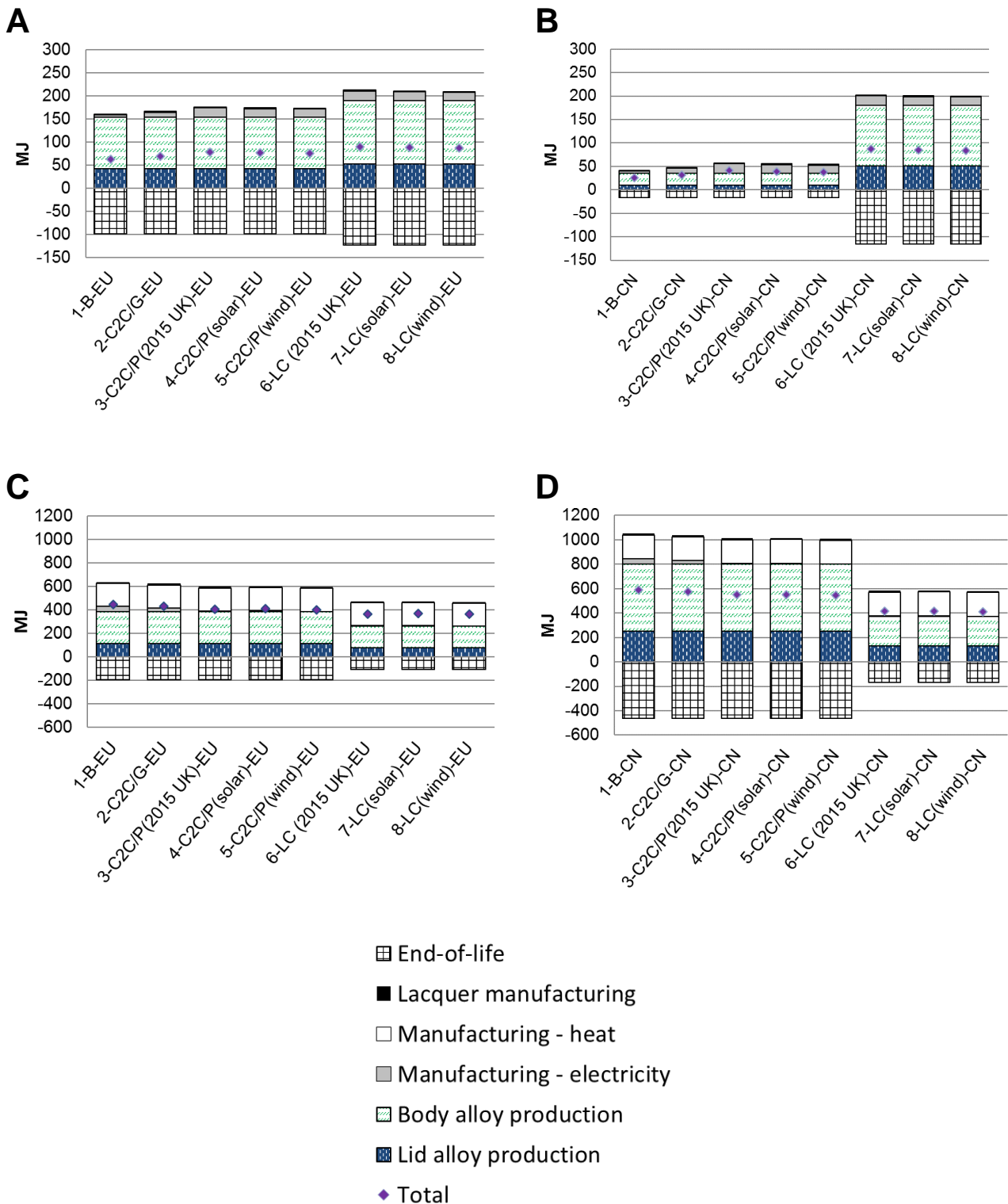
346 For the remaining impact categories, the deviations from the baseline are different between EU and  
347 CN scenarios. For EU, the impact categories PM, POF, Ac, TE, FE, ME and non RE CED show a  
348 moderate reduction (up to 20%) in impact scores from the baseline towards the “LC scenarios”,  
349 meanwhile for CN the decrease is up to 40-60%. For CC the decrease from P-scenarios to LC-  
350 scenarios is negligible (i.e. less than 10%) for EU, but relevant for CN (i.e. around 50%). This  
351 decrease is due to the contribution of the GHGs offset for the thermal energy, which is comparable  
352 in magnitude to the CC impact score from cradle to grave for CN, but not for EU. For IR-HH a  
353 consistent decrease of the impact score is shown for EU, but no differences can be detected for CN  
354 due to the absence of nuclear energy in the electricity mix considered for Chinese primary  
355 aluminum production. No strong influence in terms of RE electricity source used can be detected

356 with regard to the manufacturing stage, since the impact deriving from the other life cycle stages is  
357 much higher.

358 The trend reported in Table 2 confirms the importance of considering a broader range of  
359 environmental impacts alongside climate change, already reported for the specific UK case by  
360 Kouloumpis and colleagues (2015) and also at the global scale by Laurent and Espinosa (2015).  
361 Kouloumpis and colleagues (2015) concluded that in the UK case a decarbonisation of electricity  
362 supply to meet the 2050 carbon targets would lead to a reduction in the majority of the life cycle  
363 impacts by 2070, including climate change, with the exception of resource depletion, which would  
364 increase by 4–145 times on today’s value, and health impact from radiation which would increase  
365 four-fold if nuclear power is used and electricity demand grows strongly (Kouloumpis et al. 2015).  
366 Moreover, the selection of the RE source could have some implications on the potential for impact  
367 reduction, even though in our case the deviations across platinum scenarios and “LC scenarios” are  
368 below 10%. The sensitivity analysis performed using a different LCIA method, i.e. ReCiPe 2008  
369 (midpoint, hierarchist v.1.11), confirmed the results obtained with the ILCD recommended method  
370 for all impact categories (see detailed explanation in Table S5 in the SI)

371 When switching from the baseline to the LC-scenarios, the reduction in terms of non RE CED is not  
372 significant for EU and moderate for CN (see Tables 2 and S4). Figure 3 represents the CED results  
373 for renewable and non-renewable energy results for the 16 scenarios analyzed, in the case of  
374 primary Al from Europe (Fig. 3a and 3c) and China (Fig. 3b and 3d).

375



377 **Figure 3 Contribution analysis in terms of Cumulative Energy Demand (CED) for the AIC**  
 378 **systems according to the scenarios presented in Table 1, with distinction between renewable**  
 379 **energy for (A) EU and (B) CN and non-renewable energy for (C) EU and (D) CN.**

380

381 The contribution analysis displayed in Fig. 3 shows that a considerable share of non RE comes from  
382 the heat used in the manufacturing stage. Despite the use of electricity from RE in the primary Al  
383 production in the LC scenarios, the share of non-renewable energy still dominates the CED, due to  
384 the contribution from the non-fossil component (e.g. natural gas and coal) used for heating and in  
385 the background processes, both for EU and CN. However, the switch to electricity from RE in the  
386 primary aluminum production can considerably contribute to the reduction of the non RE CED in  
387 the case of CN, where the current electricity mix used for aluminum production is mainly based on  
388 non RE sources.

389

## 390 <heading level 1> **Conclusions and recommendations**

391 The outcomes of the LCA-based scenario analysis showed that compliance with the current  
392 RE&CM C2C certification requirement offers limited benefits in terms of reduced environmental  
393 impacts for the AIC system considered. Although there are slight differences between the EU and  
394 CN scenarios, the observed trends are the same: with the exception of CC impacts, the reduction in  
395 potential environmental impacts that can be achieved at the highest certification levels (gold and  
396 platinum) is negligible compared to the potential for reduction that the LC perspective can bring.  
397 The recently-updated RE&CM criterion includes a partial life cycle perspective – only at the  
398 platinum level – and has limited minimum requirements, i.e. coverage of at least 5% of the  
399 embodied energy associated with the product from cradle to gate. Even though our modelling was  
400 incomplete and did not include all background processes (see Section “Life cycle scenarios”), we  
401 showed that increasing the share of electricity from RE in a cradle to grave perspective can greatly  
402 exceed the environmental benefits brought up by the C2C certification, not only for CC, but for the  
403 broader range of impact categories. The impacts from thermal energy are currently dealt in terms of  
404 direct on-site GHGs emissions, which need to be offset to achieve the highest C2C certification

405 level. In our study we have not modelled the impacts originating from the offset activities, which  
406 could further limit the benefits in terms of environmental impacts reduction from the baseline to the  
407 C2C certification scenarios and potentially also lead to increases of impacts.

408 Our findings show that for product systems where most of the environmental impacts come from  
409 raw material extraction and production, the RE share in the upstream processes need to be taken  
410 into account in the product optimization strategy, not only in terms of electricity but also thermal  
411 energy production, which has a significant contribution in terms of non RE CED (see Fig.3c and  
412 Fig. 3d). From a company perspective, this means that the knowledge of the raw material supplier  
413 location is crucial for achieving better environmental performances and higher certification levels.  
414 However, when the C2C certification refers to a specific market, knowing the location at country  
415 level might be insufficient, e.g. for China, where the different Chinese provinces present a wide  
416 range of grid mixes and disparities in GHG emission intensities from primary aluminum production  
417 (Hao et al. 2015). If the location of the plant is known, the recommendation is thus to adjust the  
418 level of details during data collection to be able to model the local energy mix, if relevant, e.g.  
419 considering ecoinvent v3.2, which includes electricity production at the province level for China  
420 (Moreno Ruiz et al. 2015). Moreover, a huge margin for reducing the electricity consumption for  
421 aluminum production, is provided by the use of secondary aluminum (Hao and colleagues 2015),  
422 since only 5% of the energy used for primary aluminum is required to make secondary aluminum  
423 (EAA 2013). Particularly for Chinese Al production, GHG mitigation strategies should be based on  
424 developing secondary aluminum industry, improving energy mix and optimizing resource  
425 efficiency of production (Liu et al. 2016). Therefore, an increase in the recycled content of  
426 aluminum products (measured in terms of MR requirement in the C2C certification) could  
427 positively affect the RE criterion (Niero et al. 2016).

428 According to the C2C certification program, all eligible renewable energy sources , i.e. solar, wind,  
429 hydropower, biomass (when not in competition with food supplies), geothermal and hydrogen fuel  
430 cells, are given the same preference although these renewable energy sources differ in their

431 environmental performances, see e.g. Asdrubali et al. (2015). In the case of aluminum cans, due to  
432 the negligible contribution of electricity during manufacturing, no significant differences between  
433 the different RE sources were detected (see results for the different platinum scenarios). However,  
434 in other settings the selection of an energy source might be influential on the overall environmental  
435 performances of the system; such influences need to be further investigated. In the decision making  
436 process, the limitations of the LCA results should not be overlooked, e.g. the uncertainty associated  
437 to results due to assumptions made and to the LCIA characterization models.

438 As shown by Haas and colleagues (2015) in their assessment of material flows, waste production,  
439 and recycling in the European Union and the world in 2005, reducing the consumption of fossil  
440 energy carriers is necessary to further raise the degree of circularity of the economy. Our results  
441 confirm that the role of an energy transition from fossil to renewable energy resources should not be  
442 neglected in shifting towards a circular economy. Decisions at the company level present  
443 repercussions on the global scale: switching from fossil to renewable energy sources for aluminum  
444 production could induce effects on electricity infrastructures and other industrial sectors, and  
445 potentially lead to burden shifting due to the constrained supply of renewable energy. The  
446 environmental consequences of such a large scale change should therefore be assessed by means of  
447 a broader LCA incorporating the changes with structural market implications beyond the  
448 foreground- system, considering decision context situation B of the ILCD guidelines. Deploying  
449 renewable energy sources to produce electricity and heat for use at industrial scale is in the long  
450 term the only possible solution to implement the ideal C2C platinum scenario, with 100%  
451 renewable energy use. The possibility to use green certificates to achieve the C2C requirement  
452 stimulates the demand for renewable energy, but this is not a consistent long term solution to put  
453 into practice the use of “current solar income” principle. The selection of the perspective on the use  
454 of renewable energy is crucial to avoid burden shifting and assure a true environmental impact  
455 reduction, although the technical challenges at large production scale should be incorporated in the  
456 decision making process. In the case of products, such as aluminum cans, where most of the



457 environmental impacts do not originate from the manufacturing stage of the final products, but from  
458 the raw material extraction and production, the efforts should be directed to the upstream processes.  
459 This conclusion can be extended for all products where raw materials extraction is the most  
460 impacting life cycle stage, e.g. laminated carton containers (Scipioni et al. 2013), plastic containers  
461 (Madival et al. 2009), stainless steel building components (Ibbotson and Kara 2013), metallic  
462 furniture (Babarenda Gamage et al. 2008), etc.

463 LCA allows modeling the consequences of decisions not only at the product level but also on a  
464 large scale, e.g. through consequential LCA. Moreover, since the overall environmental  
465 performances of a product depend not only on the manufacture of the product itself, but on the  
466 whole life cycle, see e.g. the impact from refrigeration of beverages in certain geographical  
467 contexts, the limited focus of the C2C certification on the product can therefore benefit from the  
468 inclusion of a broader life cycle perspective. There has recently been a discussion whether C2C  
469 certified products are better from an environmental point of view, see Llorach-Massana and  
470 colleagues (2015) and the rebuttal by Kausch and Klosterhaus (2015). The primary aim of the C2C  
471 design framework is to provide guidelines for product quality and innovation, and not strictly to  
472 communicate the environmental issues associated with a product. However, considering that the  
473 C2C certification program is used also as a means of marketing towards consumers, caution is  
474 needed when communicating the environmental performances of C2C certified products. We  
475 believe that the current focus of the RE&CM criterion on the CC impact is too narrow for  
476 modelling the actual environmental impact deriving from the use of renewables. Our results  
477 confirmed that GHG emissions could not be used as a single indicator to represent the  
478 environmental performance of a system or technology (Turconi et al. 2013; Laurent et al. 2012).

479 Even though the use of LCA in a C2C certification context is constrained by the limited availability  
480 of datasets representing future energy technologies, therefore preventing the modelling of long term  
481 forecasting scenarios, our main recommendation is to combine the C2C certification with LCA to  
482 make the second C2C principle operational. The adoption of scenarios analysis in an LCA context

483 can support the C2C certification program with a tool to compare the environmental performances  
 484 of alternative improvement strategies which can be implemented in the progression towards higher  
 485 certification levels.

486

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492

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