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Strong Delayed Interactive Effects of Metal Exposure and Warming: ² Latitude-Dependent Synergisms Persist Across Metamorphosis

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S Supporting Information 9

ABSTRACT: As contaminants are often more toxic at higher temper-10 atures, predicting their impact under global warming remains a key 11 challenge for ecological risk assessment. Ignoring delayed effects, 12 synergistic interactions between contaminants and warming, and differ-13 ences in sensitivity across species' ranges could lead to an important 14 underestimation of the risks. We addressed all three mechanisms by 15 studying effects of larval exposure to zinc and warming before, during, and 16 17 after metamorphosis in Ischnura elegans damselflies from high- and low-18 latitude populations. By integrating these mechanisms into a single study, 19 we could identify two novel patterns. First, during exposure zinc did not affect survival, whereas it induced mild to moderate postexposure mortality 20 in the larval stage and at metamorphosis, and very strongly reduced adult 21 lifespan. This severe delayed effect across metamorphosis was especially 22 remarkable in high-latitude animals, as they appeared almost insensitive to 23



2.4

zinc during the larval stage. Second, the well-known synergism between metals and warming was manifested not only during the larval stage but also after metamorphosis, yet notably only in low-latitude damselflies. These results highlight that a more 25

complete life-cycle approach that incorporates the possibility of delayed interactions between contaminants and warming in a 26

geographical context is crucial for a more realistic risk assessment in a warming world. 27

INTRODUCTION 2.8

29 Current approaches to assess the risk of contaminants to 30 aquatic ecosystems fail to adequately protect biodiversity.¹⁻³ 31 This may not be surprising as these approaches are developed 32 for single stressors while stressors may interact,⁴ often do not 33 include delayed postexposure effects,^{5,6} and ignore geographic 34 variation in the sensitivity to pollutants.⁷ However, studies 35 integrating these three key aspects are lacking, yet crucial to 36 arrive at a more realistic ecological risk assessment. This is 37 furthermore important in the context of global warming, which 38 represents a serious threat that may strongly magnify the 39 impact of contaminants.^{8–11} Integrating warming into risk 40 assessment of contaminants has therefore been identified as a 41 major challenge for ecotoxicology.^{12,13}

Delayed postexposure effects of contaminants can be equally 42 43 strong or even stronger than the initial effects during exposure 44 (e.g., refs 5, 6, 14-17). In animals with a so-called complex life 45 cycle, such delayed effects may be revealed or magnified at 46 metamorphosis, which has been identified as a sensitive, 47 stressful event ("stressful metamorphosis hypothesis").¹⁸ This is 48 because the process of metamorphosis involves substantial 49 tissue construction and reorganization, and requires a 50 considerable amount of energy.^{19,20} Recent studies demon-

strated severe lethal effects of contaminants in aquatic insects 51 during metamorphosis despite the absence of any effects during 52 the preceding larval exposure period.^{21–23} Yet, also after 53successful metamorphosis, there is still potential for delayed 54 effects.^{14,23,24} While studies are more frequently considering 55 such delayed effects of single stressors across metamorphosis, 56 and despite the increasing concern that interactions between 57 stressors also occur when they are encountered sequen- 58 tially,^{25,26} delayed postmetamorphic effects of combined larval 59 exposure to contaminants and other stressors, particularly 60 warming, are very limited (but see refs 27, 28). Understanding 61 how aquatic stressors affect aquatic insect emergence and their 62 postmetamorphic survival is important for managing both 63 aquatic and riparian ecosystems as aquatic insects subsidize the 64 diets of terrestrial insectivores such as spiders and birds.^{29,30} 65

Populations from the same species may differ widely in their 66 sensitivity to contaminants³¹⁻³³ and to warming^{34,35} because of 67 geographic differences in the evolution of adaptation to these 68

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⁶⁹ stressors. While adaptation to one stressor may ameliorate the ⁷⁰ negative consequences of exposure to another stressor,^{36–38} ⁷¹ this has been poorly studied in the context of warming and ⁷² contaminants, yet crucial to arrive at a better risk assessment in ⁷³ a warming world.⁸ Studies focusing on how differential thermal ⁷⁴ adaptation in populations from different latitudes creates a ⁷⁵ geographic signal in the interactive effects of warming and ⁷⁶ contaminants may be especially revealing. The comparison of ⁷⁷ contaminant effects in high-latitude populations with low-⁷⁸ latitude populations currently experiencing the higher temper-⁷⁹ atures that are predicted to occur at the high latitude under ⁸⁰ global warming, may inform about the impact of gradual ⁸¹ thermal evolution on the toxicity of contaminants in a warming ⁸² world ("space-for-time substitution").³⁹

To address these issues in an integrated way we investigated 83 84 the effects of larval exposure to a metal and warming on larval 85 and adult fitness-related traits in replicated high- and low-86 latitude populations of an aquatic insect. More specifically, we 87 documented the combined effects of a metal and warming 88 during and after the metal-exposure period in the larval stage 89 (survival, growth rate, development time), during emergence 90 (emergence ratio), as well as in the terrestrial adult stage (size 91 at emergence, lifespan) in a common garden experiment with 92 the damselfly Ischnura elegans. This allowed testing for delayed 93 mortality caused by larval metal exposure and interactive effects 94 with warming before, during and after metamorphosis. High-95 latitude populations were collected in southern Sweden, where 96 the species is semivoltine (one generation per two years), 97 whereas low-latitude populations were collected in southern 98 France, where the species is multivoltine (three to four 99 generations per year).⁴⁰ This difference in life-history strategy 100 is expected to translate into a different susceptibility to 101 contaminant exposure, with the more fast-lived southern populations being more susceptible, because of a trade-off 102 103 between growth and investment in detoxification and repair 104 mechanisms.⁴¹ We further tested for a signal of thermal 105 adaptation in how damselflies from both latitudes differed in 106 their responses to the metal and temperature treatment (which 107 matched the IPCC 4 °C warming scenario RCP8.5⁴² at the 108 higher latitude). By comparing the performance of the high-109 latitude populations at the local cooler and predicted warmer 110 temperatures, predictions can be made about how high-latitude 111 would perform under global warming in the absence of thermal 112 evolution. Comparing them with the low-latitude populations 113 (which have been locally adapted to warmer temperatures), 114 however, simulates a global-warming scenario that also takes 115 into account the performance change due to gradual thermal 116 evolution.

117 MATERIALS AND METHODS

Sampling and Rearing. Using a common garden rearing 118 119 experiment, we quantified key life-history traits before, during 120 and after metamorphosis in the damselfly Ischnura elegans from two latitudes spanning more than 1500 km. Eggs were collected 121 at three high-latitude locations in southern Sweden (56–57°N) 122 and three low-latitude locations in southern France (43°N) 123 (see Supporting Information, Appendix S1 for details).⁴³ At 124 125 each location between 9 and 22 mated females were collected 126 and placed individually in small jars with wet filter paper for 127 oviposition. The eggs were transferred to the laboratory in 128 Belgium where they hatched. Larvae were placed individually in 129 200 mL vials with dechlorinated tap water and assigned to a 130 rearing temperature treatment of 20 or 24 °C. These

temperatures reflect the mean water temperatures in shallow 131 water bodies during summer in southern Scandinavia and 132 southern France, respectively.⁴⁴ The 4 °C temperature 133 difference between these latitudes matches the predicted 134 temperature increase in southern Scandinavia by 2100 under 135 IPCC scenario RCP8.5.⁴² Therefore, the experimental setup 136 allows a space-for-time substitution whereby the low-latitude 137 populations reared at 24 °C provide a proxy for the 138 evolutionary response of the high-latitude populations to global 139 warming. Temperature was manipulated using water baths, 140 across which rearing vials were regularly reshuffled. Throughout 141 the experiment, the larvae were kept at the assigned rearing 142 temperature and a constant photoperiod of 14:10 h light:dark. 143 During the pre-exposure period, the larvae were fed Artemia 144 nauplii ad libitum 5 days per week. 145

Experimental Setup. When a larva moulted into the final 146 instar, it was randomly assigned to a six-day zinc exposure 147 treatment (control or 100 mg L⁻¹ zinc). This resulted in a 148 nested full factorial design with three sampling locations per 149 latitude ×2 latitudes (high/Sweden and low/France) × 2 150 temperatures (20 and 24 $^{\circ}C$) × 2 zinc treatments (zinc present 151 and absent). The chosen zinc concentration was previously 152 shown to reduce growth rate in *I. elegans.*³² Because of the low 153 metal sensitivity of damselflies⁴⁵ and the strictly aqueous 154 exposure route in our experiment (the dominant exposure 155 route for metals is generally through the diet),46,47 this 156 concentration strongly exceeds commonly occurring environ- 157 mentally concentrations (but see ref 48). Note, however, that 158 our aim was not to mimic a realistic field situation, but to test 159 for a proof-of-principle of the idea that susceptibility to a 160 contaminant across metamorphosis can be shaped by temper- 161 ature and latitude. Zinc solutions were prepared from a stock 162 solution of ZnCl₂ (5 g zinc L⁻¹ dissolved in milli-Q water) that 163 was stored in the dark at 4 °C and that was renewed monthly. 164 To prepare exposure media, the stock solution was further 165 diluted with synthetic pond water, which was also used as 166 control medium (for the composition see ref 49). The larvae 167 were exposed to 100 mL of the medium in the same vials as 168 they had been reared in. To minimize concentration changes 169 due to evaporation, the medium was renewed every 2 days. The 170 measured zinc concentrations in the experimental vials when 171 the medium was freshly renewed was 93.10 ± 1.65 mg L⁻¹ 172 (mean \pm SD, n = 5 pooled start samples, taken throughout the 173 experiment); after 48 h (just before renewal of the medium) 174 the concentration was 95.31 \pm 2.84 mg L⁻¹, as verified with 175 ICP-OES. The associated water quality parameters were (mean 176 ± 1 SD, n = 10 samples) the following: conductivity: 387 ± 17 177 μ S/cm, pH: 6.7 ± 0.1, dissolved oxygen: 8.0 ± 0.2 mg L⁻¹, 178 hardness: $132 \pm 4 \text{ mg L}^{-1} \text{ CaCO}_3$ and dissolved organic 179 carbon: 4.9 \pm 1.3 mg L⁻¹. After the six-day zinc exposure, all 180 larvae were transferred to synthetic pond water and kept at 181 their rearing temperature until emergence. This assured that all 182 larvae were exposed to zinc for the same duration. Throughout 183 the final instar, larvae were daily fed Artemia nauplii ad libitum, 184 supplemented with 3 larvae of Chironomus riparius per week. 185

After emergence, the damselflies were kept overnight in their 186 rearing vials to ensure complete hardening of the exoskeleton 187 and wings before handling. Adults that emerged successfully 188 were individually marked by writing a number on one of the 189 wings with a permanent marker (Staedtler Lumocolor 190 permanent F), after which they were transferred to insectaries 191 $(36 \times 30 \times 30 \text{ cm}^3)$ to record lifespan; for logistic reasons this 192 was only done for a subset of the adults. Males and females 193 194 were kept in separate insectaries with maximum 10 adults of 195 similar ages per insectary and were provided with ad libitum 196 fruit flies. The insectaries were kept in a room at 22 °C \pm 1 °C 197 with an air humidity between 70% and 80% and were sprayed 198 twice per day with dechlorinated tap water. By keeping all 199 adults at this intermediate temperature (between the two larval 200 rearing temperatures) any effects of temperature on adult 201 lifespan could be attributed to differences in larval rearing 202 temperature (hence, carry-over effects) without confounding 203 with the temperature experienced in the adult stage.

Initial sample sizes varied between 112 and 133 animals per 204 205 combination of latitude, temperature and zinc treatment (total =969). Between 44 and 102 animals emerged successfully per 206 treatment combination (total =583). Sample sizes for the 207 subset on which lifespan was recorded varied between 18 and 2.08 52 per combination of latitude, temperature and zinc treatment 209 (total =291). The imbalance in sample size was due to 210 211 differential mortality and emergence success in the different 212 treatment combinations (see Results). Sample sizes per 213 treatment combination for each end point are presented in 214 Appendix S2.

Response Variables. We quantified larval growth rates 215 216 both during the six-day exposure period and during the 217 subsequent first 4 days of the postexposure period. To quantify growth rates, each larva was weighed to the nearest 0.01 mg at 218 219 the start and at the end of the zinc exposure and again 4 days 220 later using an electronic balance (AB135-S, Mettler Toledo, 221 Zaventem, Belgium). Individual growth rates were calculated as 222 ($\ln \text{ mass}_{\text{final}} - \ln \text{ mass}_{\text{initial}}$)/number of days (see ref 50). 223 Rearing vials were checked twice per day for survival and adult 224 emergence. To obtain a detailed picture on when larval 225 mortality occurred we separately considered larval survival 226 during the six-day exposure period and during the remaining 227 postexposure period (until emergence). In addition, we 228 calculated the emergence ratio as the percentage of larvae 229 surviving until the end of the final instar that emerged 230 successfully (i.e., complete emergence without wing malforma-231 tions). To obtain an integrated picture of larval survival and 232 emergence ratio we also quantified the total emergence success 233 as the percentage of individuals initially in the experiment that emerged successfully. Development time was calculated as the 234 number of days between molting to the final instar and adult 235 236 emergence. Adults that emerged successfully were weighed to 237 the nearest 0.01 mg and sexed. Insectaries were checked twice per day to record adult lifespan. 238

Statistical Analyses. Effects of latitude, rearing temper-239 240 ature and zinc exposure on the response variables were 241 analyzed with (generalized) linear mixed-models (GLMMs/ 242 LMMs) using the lme4 package⁵¹ in R.⁵² In each model all 243 interactions between latitude, rearing temperature and zinc exposure (fixed factors) were included, as well as sampling 244 245 location nested in latitude as random effect (but see below for survival). Effects on larval survival were tested with a GLMM 246 with binomial error distribution and a logit link function that 247 also included the fixed factor period (before/after zinc 248 exposure) and its interaction with the other fixed factors. For 249 the analysis of emergence ratio and total emergence success, we 250 ran GLMMs with binomial error distribution and a logit link 251 252 function. The two successive larval growth rates were modeled 253 with a repeated measures LMM that also included the fixed 254 factor period and its interaction with the other fixed factors, and 255 individual as random effect. The models for development time, 256 mass at emergence, and lifespan also included sex as fixed

270

factor, together with its interactions with the other fixed factors. ²⁵⁷ However, as this is not the focus of this manuscript, results on ²⁵⁸ sex differences are not reported. For lifespan a GLMM with a ²⁵⁹ Poisson error distribution and a log link function was used. ²⁶⁰ Because of nonconvergence of the model for larval survival ²⁶¹ when including sampling location as a random factor, we ²⁶² reduced the complexity of this model by including this factor ²⁶³ here as a nested fixed (instead of a random) effect (following ²⁶⁴ refs53, 54). Whenever an interaction between stressors was ²⁶⁵ detected, we calculated the interaction effect size (estimated as ²⁶⁶ Hedges'd) with its 95% confidence interval following Jackson et ²⁶⁷ al.⁵⁵ An interaction effect size larger than zero indicates a ²⁶⁸ synergistic interaction among stressors. ²⁶⁹

RESULTS AND DISCUSSION

Latitude-Wide Delayed Effects of Zinc. While zinc 271 exposure barely affected the survival of the larvae during the six- 272 day exposure period (-1.5%), it considerably reduced survival 273 after the exposure period (-12.4%); period × Zinc; Table 1; 274 th

Table 1. Results of the (G)LMMs Testing for the Effect of Latitude, Rearing Temperature, and Zinc Exposure on Larval Survival and Growth Rate of *Ischnura elegans* in the Final Instar^a

	larval	survival	larval growth rate		
effect	χ^2_1	Р	χ^2_1	Р	
latitude (lat)	0.52	0.47	202.34	< 0.001	
temperature (temp)	5.35	0.021	13.47	< 0.001	
zinc (Zn)	26.13	<0.001	46.01	< 0.001	
lat \times temp	0.59	0.44	70.75	< 0.001	
lat \times Zn	12.12	<0.001	12.38	< 0.001	
temp \times Zn	3.37	0.066	0.35	0.55	
lat \times temp \times Zn	4.39	0.036	6.35	0.012	
period (per)	104.38	<0.001	102.46	< 0.001	
per × lat	3.77	0.052	8.83	0.0030	
per × temp	14.31	<0.001	6.02	0.014	
per × Zn	16.43	<0.001	5.48	0.019	
per \times lat \times temp	0.72	0.40	13.26	< 0.001	
per \times lat \times Zn	4.33	0.037	3.79	0.052	
per \times temp \times zinc	5.21	0.022	2.38	0.12	
per \times lat \times temp \times Zn	0.49	0.48	0.21	0.65	
			1 6		

"Both life-history traits were quantified during and after a six-day exposure to zinc, indicated by the (repeated) factor Period. Significant P values (P < 0.05) are printed in bold.

Figure 1a,b). Zinc furthermore decreased larval growth rate 275 fl (main effect Zinc, Table 1). Again, this zinc effect was stronger 276 after than during the exposure period (period × Zinc; Table 1; 277 Figure 1c,d). Zinc neither induced a latitude-wide effect on 278 emergence ratio, nor on total emergence success (Table 2; 279 t2 Figure 2a,b), although there were pronounced latitude-specific 280 f2 effects of zinc on these end points (see below). Previous zinc 281 exposure did not lead to a longer development time (Table 2; 282 Figure 3a), but did result in a lower mass at emergence (main 283 f3 effect Zinc; Table 2; Figure 3b). Finally, adult lifespan was 284 strongly reduced (i.e., by ca. 7 days) by larval zinc exposure 285 (main effect Zinc; Table 2; Figure 3c). 286

These results show the presence of strong delayed carry-over 287 effects of larval metal exposure, not only in the larval stage after 288 the exposure period (survival, growth rate) but also after 289 emergence in the adult stage (adult mass and lifespan), and this 290 despite zinc imposing no biologically significant effects on 291



Figure 1. Percentage survival (a,b) and mean (+1 SE) growth rate (c,d) of *Ischnura elegans* larvae in the final instar during and after a six-day exposure to zinc as a function of latitude, rearing temperature and zinc exposure.

Table 2. Results of the (G)LMM Testing for Effects of Latitude, Temperature and Zinc on the Emergence Ratio (% Larvae Surviving until the End of the Final Instar That Emerged Successfully), Total Emergence Success (% Individuals Initially in the Experiment That Emerged Successfully), Development Time (Final Instar Duration), Mass at Emergence, and Adult Lifespan of *Ischnura elegans*^a

	emergence ratio		total emergence success		development time		mass at emergence		adult lifespan			
effect	χ^2_1	Р	χ^2_1	Р	χ^2_1	Р	χ^2_1	Р	χ^2_1	Р		
latitude (lat)	35.59	< 0.001	28.74	<0.001	59.03	<0.001	189.53	<0.001	15.59	< 0.001		
temperature (temp)	35.88	< 0.001	36.08	<0.001	65.55	<0.001	41.76	<0.001	5.60	0.018		
zinc (Zn)	2.98	0.084	18.44	<0.001	0.026	0.87	14.46	<0.001	445.86	<0.001		
lat \times temp	1.31	0.25	4.37	0.037	372.09	<0.001	4.68	0.030	0.72	0.40		
lat \times Zn	20.88	< 0.001	32.57	<0.001	0.43	0.51	2.11	0.15	2.25	0.13		
temp \times Zn	0.10	0.75	1.93	0.16	0.13	0.72	1.24	0.26	0.23	0.63		
lat \times temp \times Zn	0.60	0.44	6.09	0.014	0.030	0.86	0.042	0.84	4.26	0.039		
^a Significant P values ($P < 0.05$) are printed in bold.												

292 survival during the six-day exposure period. These delayed effects and particularly the lethal effects during metamorphosis 293 of larval exposure to metals (see below) confirm the important 294 role of metamorphosis as a survival bottleneck in metal-contaminated environments.^{21,29,56} Besides the important 2.95 296 effects on survival, also the sublethal effect on mass at 297 emergence may result in a lower fitness, as body size is 298 positively related with fecundity in damselflies.⁵⁷ Yet more 299 300 importantly, we showed that larval zinc exposure also caused a 301 remarkably strong reduction in adult lifespan, a crucial fitness 302 measure in damselflies.⁵⁸ This reduction in lifespan would have

particularly strong fitness implications as it reduced adult 303 lifespan to ca. 2 days. Given that sexual maturation takes ca. 6 304 days in the study species, ⁵⁹ no reproduction would occur in the 305 damselflies that emerged successfully from the zinc treatment. 306 This strong postmetamorphic lethal effect of zinc complements 307 the very few documented adverse metal effects across 308 metamorphosis: lowered survival in bryozoans,¹⁴ lowered 309 fecundity in mosquitoes,⁶⁰ lowered immune function and 310 delayed reproduction in flies,²⁷ and reduced body condition in 311 juvenile frogs.²⁸ Although some elimination of metals can 312 happen during metamorphosis (e.g., via the shed larval skins), 313



Figure 2. Emergence ratio (% larvae surviving until the end of the final instar that emerged successfully) (a) and total emergence success (% individuals initially in the experiment that emerged successfully) (b) of *Ischnura elegans* larvae as a function of latitude and previous rearing temperature and zinc exposure in the larval stage.

314 metal body burdens after metamorphosis can still remain very 315 high.^{61,62} Moreover, in addition to the stress that metamor-316 phosis itself creates, the remobilization of stored metals during 317 tissue reorganization at metamorphosis has been suggested to 318 magnify the effects of metals.^{56,63}

Latitude-Specific Delayed Effects of Zinc. The survival 319 of French larvae was more strongly reduced by zinc than that of 320 321 Swedish larvae, especially after the exposure, where zinc 322 reduced the survival of the French larvae by ca. 20% and that of Swedish larvae by ca. 6% (period \times latitude \times Zinc, Table 1, 323 Figure 1a,b). Also the zinc-induced reduction in larval growth 324 rate was stronger in French damselflies (latitude × Zinc, Table 325 1, Figure 1c,d). While the emergence ratio of Swedish 326 damselflies was not affected by previous zinc exposure, that 327 of French damselflies was reduced by ca. 18% (latitude \times Zinc; 328 Table 2, Figure 2a). This, in combination with the different 329 330 larval survival rates between the latitudes, resulted in a strong 331 zinc-induced decrease in total emergence success in French, but 332 not in Swedish damselflies (latitude \times Zinc; Table 2; Figure 333 2b). Development time and mass at emergence were similarly 334 affected by previous zinc exposure across latitudes (Table 2). 335 [Results on life-history differences in the control treatment 336 between damselflies from both latitudes are presented and 337 discussed in Appendix S3.]



Figure 3. Mean (+1 SE) development time (number of days in the final instar) (a), mass at emergence (b), and adult lifespan (log transformed) (c) of *Ischnura elegans* as a function of latitude, previous rearing temperature and zinc exposure in the larval stage. All adults were kept under identical conditions.

20°C

24°C

The negative impact of zinc thus strongly differed between 338 latitudes with all effects being more pronounced (and 339 sometimes only present) in French larvae. As French showed 340 adaptation to higher temperatures (see below), this pattern of 341 an increased vulnerability to contaminants has been predicted 342 as a population-level trade-off between tolerance to warming 343 and tolerance to toxicants by Moe et al.⁸ This might 344 furthermore indicate a role of energy-based life-history trade-345 offs shaping sensitivity to contaminants.⁴¹ In such case, fast 346 growth is traded off against a lower investment in detoxification 347 and repair mechanisms, which would make fast-lived individuals 348 more susceptible for contaminants. Furthermore, fast growth 349 ³⁵⁰ often covaries with a higher food intake, a smaller size (hence a ³⁵¹ higher surface-to-volume ratio), and a faster metabolic ³⁵² rate, ^{64–67} traits that can lead to a higher uptake or accumulation ³⁵³ of contaminants.^{9,68,69}

Although zinc-induced effects were consistently stronger in 354 355 French damselflies, even in Swedish adults, which showed only 356 a low zinc-induced reduction in larval survival (6.2% 357 postexposure) and no zinc-induced mortality during meta-358 morphosis, zinc reduced adult lifespan by ca. 71%. This strong 359 reduction happened despite a potentially long recovery period 360 between the end of the exposure and emergence (i.e., on 361 average 27 days at 20 °C and 43 days at 24 °C). This finding 362 particularly underscores the importance of covering the full life 363 cycle in ecological risk assessment of pollutants.^{70,71} If our study would have been limited to premetamorphic life stages 364 (e.g., ref 32) or had ended just after metamorphosis (e.g., ref 365 366 21), we would have erroneously concluded that Swedish 367 animals would only suffer negligible zinc-induced effects, whereas the strong delayed effect on lifespan implicates a 368 complete failure to reproduce. 369

Latitude-Specific Delayed Synergistic Effects between 370 371 Zinc and Warming. While in Swedish larvae the effect of zinc 372 on survival was similar across temperatures, the survival of French larvae was more strongly reduced by zinc at 24 °C than 373 at 20 °C (latitude × temperature × Zinc; Table 1; Figure 1a,b). 374 This indicated a synergistic interaction between the temper-375 ature and zinc treatments in French larvae, which was 376 supported by the positive interaction effect size both during 377 378 (Hedges' d = 0.049, 95% CI: [0.030, 0.068]) and especially after 379 the exposure period (Hedges' d = 0.36, 95% CI: [0.34, 0.38]). 380 Similarly, only in French larvae growth rate was more strongly 381 affected by zinc at 24 °C than at 20 °C (period × latitude × 382 Zinc; Table 1; Figure 1c,d; interaction effect size during the 383 exposure period: Hedges' d = 0.023, 95% CI: [0.0016,0.044]; after the exposure period: Hedges' d = 0.037, 95% CI: 384 385 [0.014,0.060]). Likewise, only in French damselflies, total 386 emergence success was more strongly reduced by zinc at 24 °C (ca. 52%) than at 20 °C (ca. 23%) (interaction effect size: 387 388 Hedges' *d* = 0.47, 95% CI: [0.45,0.48]) (latitude × temperature $_{389} \times \text{Zinc}$; Table 2; Figure 2b). Finally, the strong effect of zinc on adult lifespan was more pronounced at 24 °C, but again only in 390 French damselflies (latitude \times temperature \times Zinc). This 391 indicated a synergistic interaction between the temperature and 392 zinc treatments in French adults (interaction effect size: 393 394 Hedges' d = 0.17, 95% CI: [0.09, 0.24]).

We thus observed the expected synergism of metals being 395 ³⁹⁶ more toxic at higher temperatures for larval survival and ³⁹⁷ growth, emergence success and adult lifespan.^{9,10} Importantly, 398 the synergistic interactions were mainly delayed (occurred after 399 the exposure period ended) and had a strong geographic signal 400 as these were only present in French damselflies. Despite the concern that trace metals become more toxic at higher 401 temperatures^{9,10} and the potential of strong delayed effects of 402 403 metals during and after metamorphosis (see above), delayed interactive effects between metals and warming have been 404 largely ignored. The only exception are two studies showing 405 406 delayed sublethal effects of previous exposure to metals being magnified under warming (for adult body mas: ref 27, for 407 juvenile body condition: ref 28). Here, we showed that also 408 409 across metamorphosis delayed effects of a metal can be 410 magnified under warming, and even translated in a synergistic 411 lethal effect by shortening adult lifespan. The strong geographic 412 signal in the delayed synergistic interaction also indicated that

thermal adaptation of French damselflies to 24 $^{\circ}$ C (see below) 413 did not alleviate the adverse effects of zinc. Possibly, at 24 $^{\circ}$ C 414 the French damselflies allocated a higher percentage of 415 resources to growth than at 20 $^{\circ}$ C, and this at the expense of 416 investment in detoxification and repair, whereas the Swedish 417 damselflies, which grew even slower at the higher temperature 418 than at the lower temperature, might have been better able to 419 maintain investment in detoxification and repair at an optimal 420 level. 421

Signals of Thermal Adaptation. We found widespread 422 evidence of latitude-associated thermal adaptation (e.g., refs 423 72-74), with the French damselflies being adapted to higher 424 environmental temperatures than the Swedish damselflies. 425 While French larvae had a higher growth rate at 24 °C (the 426 current mean water temperature in southern France⁴⁴) than at 427 20 °C, Swedish instead had a reduced growth rate at 24 °C than 428 at 20 °C (the current mean water temperature in Sweden.⁴⁴) 429 (latitude × temperature; Table 1). French damselflies had a 430 high total emergence success in the control treatment at both 431 temperatures, but that of Swedish damselflies was strongly 432 reduced (by ca. 43%) at the higher temperature (latitude \times 433 temperature \times Zinc; Table 2; Figure 2b). This was partly due 434 to a lower larval survival at the higher temperature, but mainly 435 to a lower emergence ratio. Furthermore, while French 436 damselflies had a ca. six-day shorter development time at 24 437 °C than at 20 °C, the Swedish increased their development 438 time at the higher temperature by ca. 18 days (latitude \times 439 temperature, Table 2, Figure 3a). The differential thermal 440 responsiveness of larval growth and development resulted in a 441 lower mass at emergence in damselflies reared at 24 °C 442 compared to 20 °C (main effect temperature), and this mass 443 reduction at the higher temperature was stronger in French 444 than in Swedish damselflies (latitude × temperature; Table 2, 445 Figure 3b). 446

Implications for Risk Assessment in a Warming 447 World. In conclusion, our integrated study not only identified 448 three key mechanisms challenging current risk assessment (i.e., 449 delayed postexposure effects,⁵ interactions between stressors,⁴ 450 and geographic variation in the sensitivity to pollutants⁷), but 451 more importantly, also documented interactions between these 452 mechanisms. A key observation was that the effects of zinc were 453 very mild during the exposure period, yet were strongly 454 magnified after the exposure period and carried over to the 455 adult stage. We further observed a strong geographic signal in 456 the sensitivity to zinc, with the French damselflies consistently 457 more strongly affected than the Swedish, especially when the 458 larvae had experienced a higher rearing temperature. Moreover, 459 the synergistic interaction between the metal and the warming 460 treatment in the French damselflies bridged metamorphosis 461 and strongly affected (post-) metamorphic traits. However, 462 even though Swedish damselflies experienced only mild effects 463 of zinc on larval and metamorphic end points, they did show a 464 very strong reduction in adult lifespan when exposed to zinc in 465 the larval stage. 466

Predicting and understanding the impact of contaminants 467 under warming remains a key challenge at the interface of 468 ecotoxicology and global change biology.^{8,10} Even though the 469 zinc concentration used in our study is much higher than those 470 that would be encountered in realistic scenarios, our study 471 provided a proof of principle, adding two key related insights 472 that may contribute to the much needed development of a 473 more realistic risk assessment,^{1–3} especially with regard to the 474 impact of contaminants in a warming world.^{12,13} First, our 475 476 results refined the well-known pattern that trace metals are 477 more toxic at higher temperatures by showing this synergistic 478 interaction may persist across metamorphosis and may show a 479 strong geographic signal (being only present in French 480 damselflies). This geographic signal was consistent with a 481 trade-off scenario between thermal adaptation and tolerance to 482 toxicants,⁸ illustrating the complexity of predicting the impact 483 of metals in a warming world. Using a space-for-time 484 substitution our results thereby tentatively suggest that 485 warming at the high latitude would not make zinc more toxic 486 to the study species, unless high-latitude populations show 487 gradual thermal evolution and convert to the current French 488 phenotype.

Second, our study refined the hypothesis of aquatic insect 489 490 emergence as a sensitive indicator of aquatic metal toxicity²¹ in two important ways that are directly relevant for ecological risk 491 492 assessment. (i) Metamorphosis success may strongly depend on latitude as we only observed a zinc-induced mortality during 493 metamorphosis in French animals and not in Swedish animals. 494 (ii) Postmetamorphic survival may even be more important 495 496 than larval survival and emergence success; this result again 497 showed the importance of the geographical context, as this 498 pattern was especially striking for the Swedish animals. Taken 499 together our results thereby challenge current water quality 500 standards (such as EQS, EU Directive 2008/105/EC) which 501 assume that protection of aquatic stages will translate into the 502 protection of adults (but see ref 71), and ignore geographical 503 patterns in sensitivity to contaminants. Aquatic insects are ⁵⁰⁴ important vectors of transfer of nutrients from aquatic to ⁵⁰⁵ terrestrial ecosystems, ^{30,75} and aquatic ecosystems are especially ⁵⁰⁶ vulnerable to warming^{76,77} and pollutants.⁷⁸ Combined with current findings this highlights that risk assessment for aquatic 507 systems can highly benefit from a more complete life-cycle 508 509 approach addressing delayed synergisms under warming in a 510 geographical context in order to obtain more ecologically 511 relevant water quality standards.

512 **ASSOCIATED CONTENT**

S13 Supporting Information

s14 The Supporting Information is available free of charge on the s15 ACS Publications website at DOI: 10.1021/acs.est.6b04989.

Information on the sampled locations (S1), sample sizes
(S2), and presentation of the results and discussion on
baseline life-history differences between the latitudes
(S3) (PDF)

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