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## An assessment of the Norwegian Deep/Skagerrak shrimp stock using the Stock Synthesis statistical framework

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#### Abstract

A length-based stock synthesis statisktical framework was applied to provide a quantitative assessment of the northern shrimp (Pandalus boralis) in Divisions IIIa and IV East (Skagerrak, Northern North Sea in the Norwegian Deep).


The model diagnostics of the assessment showed no major problems and were in line with the assessment of the benchmark 2016. The residuals of the proportion-at-length of the commercial fleet and survey were generally small with no apparent year or cohort effects. Model retrospective analyses showed that the SSB and $R$ have been overestimated and $F$ underestimated the last years.

Spawning stock biomass (SSB) decreased from a high level of 19780 tons in 2008 to the time series' low of 5800 t in 2014, then to increase slightly in 2015, but decrease again to 7100 tons in 2016. This is lower than MSY $B_{\text {trigger }}$ of 9900 tons, but greater than $B_{\lim }$ of 6300 tons. The stock has experienced a series of low recruitment years between 2008 and 2016, with the exeption of the high 2013 year class. Fishing mortality decreased from 0.94 in 2014, which was the highest value of the time series, to the second highest value of 0.78 in 2015. Consequently, the stock is currently exploited at a greater level than $\mathrm{F}_{\mathrm{msy}}$ of 0.62 .

## Background

This stock was benchmarked in January 2016 (ICES 2016) when it was decided that a length based stock systhesis statistical framework should replace the surplus production model (Hvingel 2015) used since 2013, to assess status of the stock and form the basis for advice. Also reference points presented below were defined at the 2016 benchmark (ICES 2016).

## Model description and configuration

A thorough description of the catch data, tuning index, biological information and a description of the stock synthesis model (SS3) and its configuration can be found in the stock annex produced during the benchmark 2016 (ICES 2016, for a link see the references section; see also the benchmark report (ICES 2016) for a model description and SCR Documents 16/53,56 and 57 for a description of the data used).

A summay of the data used in the model and the model settings can be found in Fig. 1 and Table 1, respecitvely. Whereas the assessment formally includes the 2016 year through estimated catches for this year and the survey index from the beginning of the year, the assessment results are interpreted as providing information on stock development up to January $1^{\text {st }} 2016$, i.e. not for the year 2016. Moreover, the survey does not provide information on the 2016 year class, for which abundance is unknown.

## Model performance

Model convergence, log-likelihood values, number of parameters near bounds and estimates of the biological parameters for the final model are presented in Table 2. The modelled catches fitted well with the observed catches by quarter (Fig. 2). The predicted abundances similarly matched the observed survey indices, with the exeption of year 2016 (Figs. $3-5$ ), due to a large decline in the survey index this year. The model estimates also fitted well to the observed length-compositions averaged over the time series, and when fitted on a yearly basis, for the commercial fleet and the surveys (Figs. $7-9$ ). The residuals of the proportion-at-length of the commercial fleet and surveys were generally small, with no apparent year or cohort effects (Fig. 10). The model retrospective analyses of SSB, R and F show that the model overestimated the SSB and R the last three and five years respectively, and underestimated F the last two years (Figs. 14 - 16). The likely reason for this is that the large 2013 year class was unexpectedly not detected as a large number of 3 year old individuals in the survey 2016. The estimated fishery selectivity-at-length for the commercial fleet and for the surveys are shown in Fig. 6.

Sensitivity analysis
The benchmark in 2016 (ICES 2016) recognized the uncertainty in the current assumption of $M=0.75$ to the assessment, which is based on estimates from the Barents Sea in the 1990s (Barenboim et al.1991), and recommended that the sensitivity of model outputs and catch advice to the specifications of $M$ should be explored. Preliminary sensitivity analyses of the assessment model regarding different levels of $M$ carried out at the current NIPAG meeting, showed that $M=0.90$ does not change the perception of the current level of $F$ and $S S B$ relative to the reference points of $F_{m s y}$ and $B_{p a}$ compared with $M=0.75$ (base model; Fig. 17). Using $M$ $=0.90$, the $S S B$ in 2017 will still be under $B_{\lim }$ (the new $B_{l i m}$ ) at the current level of catches, indicating that the advice is rather robust to the assumption of $M$, within this range. However, shrimp in the North Sea/Skagerrak are considered to have a lifespan of only about half of that of shrimp in the Barents Sea and it is therefore likely that $M$ could be substantially higher and outside the 0.75-0.90 range explored. Previous analyses of different $M$-assumptions for this stock (SCR $14 / 66$ ) provide support for this hypothesis. NIPAG was not in a position at this meeting to fully explore the sensitivity to the $M$ assumption used and stresses the importance of further investigations to be conducted no later than during the proposed benchmark in 201819.

## Assessment results

The spawning stock biomass (SSB) has been variable over the assessment period 1988 to 2016 (Fig. 11). Since 2008, when SSB was 19780 tons, and which is the second highest SSB estimate/value of the time series, the stock decreased to the time series' low of 5800 tons in 2014. The stock then increased slightly in 2015, but decreased again to 7100 tons in 2016.

The recruitment $(R)$ has, similarly to the SSB, been variable over the assessment period 1988 to 2015 (Fig. 12). A series of low recruitment years between 2008 and 2015, with the exception of year 2013, should be noted. During this period of low recruitment the estimates of SSB were also for some years historically low and below $B_{\text {lim }}$ (Fig. 17). The uncertainties around the estimate of recruitment in 2015 is large. The reason for this is that the model has not yet seen the recruits in the fishery data (data until 2015), only in the survey data (January 2016).

Fishing mortality $(F)$ for ages 1 to 3 remained relatively stable since the beginning of the 1990s to about 2010 (Fig. 13). After 2010, Fincreased steeply to 0.94 in 2014, which is the highest value of the time series, to the second highest value of 0.78 in 2015. The stock is consequently exploited at a greater level than the $F_{m s y}$ of 0.62. (Fig. 17).

A summary of the reference points determined at the benchmark (ICES 2016) is presented in Table 3.

## Quality of the assessment

The benchmark that took place in January 2016 (ICES 2016) agreed on a quarterly length-based model as the basis for the assessment and the provision of catch advice for this Pandalus stock. The length-based model is considered preferable to the previous surplus production model because it makes more use of the available data, including using observed lengths and a quarterly time step to achieve a better representation of the population structure and dynamics. The length-based model is able to better take into account year-to-year changes in recruitment and how these changes influence catch options in the short term.

Input data are considered to be of good quality. However, the survey time series has not been standardised for variability in factors such as swept volume, spatial coverage and trawling speed, which might add uncertainty to the stock estimates. Moreover, the survey indicated a large decline in biomass in 2016, which is not observed in the lpue of the Swedish, Norwegian and Danish fleets.

The assessment and the derived advice depend on the assumption of natural mortality. However, exploration conducted during the benchmark suggests that $M$ might be larger than assumed in the current model.
$B_{\text {lim }}$ was set to $B_{\text {loss, }}$ however, recruitment estimates suggest that large year classes could be produced at levels of $S S B$ close to $B_{l o s s}$ and therefore $B_{\text {lim }}$ might be overestimated. The values below $B_{\text {lim }}$ in the 2016 assessment is due to the SSB curve being shifted slightly downwards compared to in the benchmark assessment, and that the reference points are not updated

Discarding practices in the Norwegian fishery are unknown, and Norwegian discards have been estimated by applying the Danish discard ratio to Norwegian data.

## Catch options

The basis for the catch options are as follows:

| Variable | Value | Source | Notes |
| :--- | ---: | :--- | :--- |
| $\mathrm{F}_{2016}$ | 1.53 | ICES (2016) | Corresponds to the assumed catch in <br> 2016. |
| SSB $_{2016}$ | 7077 t | ICES (2016) |  |
| $\mathrm{R}_{2016}$ | 6838538 <br> thousands | ICES (2016) | GM 2006-2015 |
| Catch $(2016)^{*}$ | 12842 t | ICES (2016) | Equal to projected landings 2016 plus <br> estimated discards. |
| Landings (2016)** | 11085 t | ICES (2016) | Projected landings 2016 |
| Discards $(2016)$ | 1757 t | ICES (2016) | Average discard rate in 2013-2015 <br> (12.5\%) |

* Equal to projected landings corrected for weight loss due to on-board boiling, and with estimated discards added.
** Swedish projected landings 2016 are recorded landings corrected by applying a factor of 1.13 to boiled landings to correct for weight loss due to on-board boiling. Danish and Norwegian projected landings 2016 are not corrected for boiling.

The catch options as presented in the ICES advice are as follows:

| Rationale | $\begin{aligned} & \text { Catch } \\ & (2017) \end{aligned}$ | Wanted catch* (2017) | Basis | F catch (2017) | SSB (2018) | \%SSB change^ | \%TAC change ${ }^{\wedge}$ ^ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MSY <br> approach | 2840 | 2485 | $\mathrm{B}_{\text {lim }}$ in 2018 | 0.30 | 6310 | 129 | -84 |
| Zero catch | 0 | 0 | $\mathrm{F}=0$ | 0 | 8054 | 192 | -100 |
| Other options | 5146 | 4503 | $\mathrm{F}_{\text {MSY }}$ | 0.62 | 4991 | 81 | -71 |
|  | 1725 | 1509 | $\begin{gathered} \mathrm{F}=\mathrm{F}_{\mathrm{MSY}} \times\left(\mathrm{SSB}_{2017} / \mathrm{MSY}\right. \\ \left.\mathrm{B}_{\text {trigger }}\right) \\ \hline \end{gathered}$ | 0.17 | 6981 | 153 | -90 |
|  | 9406 | 8230 | $\mathrm{F}_{2016}$ | 1.53 | 2875 | 4 | -48 |

* "Wanted catch" is used to describe shrimp that would be landed in the absence of the EU landing obligation, and has been calculated based on the average discard rates in 2013-2015 (12.5\%).
${ }^{\wedge}$ SSB 2018 relative to SSB 2017.
^^ Wanted catch 2017 relative to TAC 2016.

An extended set of catch options can be found in Annex 1.

At the benchmark in January 2016, a new assessment method and new reference points were agreed. The new assessment model is better able to capture year-to-year changes in stock abundance, in particular in connection with the variable recruitment, and this had substantial impact on the resulting short-term forecast for 2016, which was revised in March 2016.

While there are some differences in assessment results in the present advice compared with the advice issued in October 2015 (i.e. when the production model was used for the assessment) the development of the stock over time as estimated by the current assessment is rather similar to that derived from the production model. The main difference lies in the re-calculation of the reference points, which has resulted in a change in the perception of stock status relative to these reference points.

The 2016 assessment conducted during the benchmark indicated that the stock was below MSY $B_{\text {trigger }}$ in 2015 and was projected to be above MSY $B_{\text {trigger }}$ in 2016. The current advice indicates that fishing mortality in 2015 and 2016 is much larger than $F_{M S Y}$ and that the biomass is below MSY $B_{\text {trigger }}$. This change in the perception of the stock status is caused by the downward revision of the 2013 year class in the 2016 survey and in the current assessment.

Due to the current low level of the $S S B$, and the fact that the fishery is dependent on the incoming year class, in combination with the uncertainty in the magnitude of the decline of the survey estimate between 2015 and 2016, ICES consider that the assessment and the advice should be updated in the beginning of the 2017 after the results of the 2017 survey will be available.

## References

Hvingel, C. 2015. The 2015 assessment of the North Sea/Skagerrack shrimp stock using a Bayesian surplus production model. NAFO SCR Doc. 15/059. Serial No. N6495.

ICES. 2016. Report of the Benchmark Workshop on Pandalus borealis in Skagerrak and Norwegian Deep Sea (WKPAND), 20-22 January 2016, Bergen, Norway. ICES CM 2016/ACOM:39. 20 pp.

## ICES 2016 Stock annex:

The stock annes is available on the ICES website Library under the Publication Type "Stock Annexes". Use the search facility to find a particular Stock Annex, refining your search in the left-hand column to include the year, ecoregion, species, and acronym of the relevant ICES expert group.

| Stock <br> ID | Stockname | LAST <br> UPDATED | Link |
| :--- | :--- | :--- | :--- |
| pand_SA | Pandalus in Skagerrak and Norwegian Deep (ICES Division IIIa <br> and IVa East) | January <br> 2016 | $\underline{\text { Pandalus }}$ |

Table 1. Settings and results of the final SS3 assessment model. The table columns show: number of estimated parameters, the initial values (from which the numerical optimization is started), the intervals allowed for the parameters, and the value estimated by maximum likelihood. Parameters in bold were set and not estimated by the model. No priors were used on any parameter.

| Parameter | Number estimated | Initial value | Bounds (Low,hig h) | Value (MLE) |
| :---: | :---: | :---: | :---: | :---: |
| Natural mortality |  |  |  |  |
| M |  | 0.75 |  |  |
| Stock and recruitment |  |  |  |  |
| Ln(R0) | 1 | 6 | $(3,30)$ | 16.19 |
| Steepness (h) |  | 0.99 |  |  |
| Recruitment variability ( $\sigma \mathrm{R}$ ) |  | 0.60 |  |  |
| Ln (Recruitment deviation): 19812014 | 35 |  |  |  |
| Recruitment autocorrelation |  | 0 |  |  |
| Growth |  |  |  |  |
| Linf (cm) | 1 | 2.9 | $(2,4)$ | 2.82 |
| k | 1 | 0.39 | (0.20-0.50) | 0.43 |
| L at minimum age (0.001 years) t0 | 1 | 0 | $(0,4)$ | 0.06 |
| CV of young individuals | 1 | 0.20 | (0.005-0.40) | 0.13 |
| CV of old individuals | 1 | 0.05 | (0.005-0.40) | 0.03 |
| Weight-(kg) at-length (cm) |  |  |  |  |
| A |  | 0.0016 |  |  |
| B |  | 2.7532 |  |  |
| Maturity |  |  |  |  |
| Length (cm) at 50\% mature |  | 1.8 |  |  |
| Length (cm) at 95\% mature |  | 1.9 |  |  |
| Initial fishing mortality |  |  |  |  |
| Commercial trawl fleet | 1 | 0.1 | $(0,4)$ | 2,12 |
| Catchability and selectivity (logistic) |  |  |  |  |
| Commercial trawl fleet |  |  |  |  |
| Time-invariant length-based logistic selectivity | $2$ <br> the parameters are L50\%sel and L95\%sel-L50\%sel (in cm) | 1,1 | $(0,4.5)$ | $(1.63,0.35)$ |
| Norway survey (1988-2002) |  |  |  |  |
| Ln(Q) - catchability | 1 | - | $(0,1)$ | 0.003 |


| Parameter | Number estimated | Initial VALUE | Bounds (Low,high) | Value (MLE) |
| :---: | :---: | :---: | :---: | :---: |
| Extra variability added to input standard deviation | 1 | 0 | $(0,1)$ | 0.13 |
| Time-invariant length-based logistic selectivity | $2$ <br> the parameters are L50\%sel and L95\%sel-L50\%sel (in cm) | 1,1 | $(0,4.5)$ for both parameters | $(1.93,0.80)$ |
| Norway survey (2004-2005) |  |  |  |  |
| $\operatorname{Ln}(Q)$ - catchability | 1 | - | $(0,1)$ | 0.003 |
| Extra variability added to input standard deviation | 1 | 0 |  |  |
| Time-invariant length-based logistic selectivity | $2$ <br> the parameters are L50\%sel and L95\%sel-L50\%sel (in cm) | 1,1 | $(0,4.5)$ for both parameters | $(1.45,0.35)$ |
| Norway survey (2006-2015) |  |  |  |  |
| $\operatorname{Ln}(Q)$ - catchability | 1 | - | $(0,1)$ | 0.001 |
| Extra variability added to input standard deviation | 1 | 0 | $(0,1)$ | 0.16 |
| Time-invariant length-based logistic selectivity | $2$ <br> the parameters are L50\%sel and L95\%sel-L50\%sel (in cm) | 1,1 | $(0,4.5)$ for both parameters | $(1.49,0.58)$ |

Table 2. Results of the SS3 assessment model. Model convergence, total log-likelihood (MLE), lengthfrequency distributions log-likelihood (LFD), survey index log-likelihood (CPUE), recruitment deviations log-likelihood (Recr dev) and number of parameters near bounds.

| CONVERGENCE | TOTAL | LENGHT COMP | SURVEY <br> CPUE | REC DEV | PARA |
| :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  | NEAR BOUNDS |  |
| $\mathbf{0 , 0 0 0 2 2 0 5}$ | $\mathbf{1 4 9 , 7 9 7}$ | $\mathbf{1 7 6 , 9 6 6}$ | $\mathbf{- 1 7 , 8 9 4 8}$ | $\mathbf{- 9 , 3 4 8 4 2}$ | 0 |

Table 3. SS3 final model with quarterly time-step. Summary of the assessment.

| year | REC <br> (age 0) <br> thousands | b total Jandary 1 <br> (TON) | $\begin{aligned} & \text { SSB JANUARY } \\ & \mathbf{1} \\ & \text { (TON) } \end{aligned}$ | $\begin{aligned} & \text { САTCH } \\ & \text { (TON) } \end{aligned}$ | $\begin{aligned} & \text { F } \\ & (1-2.5 \text { см }) \end{aligned}$ | $\begin{aligned} & \text { F } \\ & \text { (1-3 YEARS) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 8618650 | 24992 | 13077 | 12177 | 0.54 | 0.62 |
| 1989 | 12223600 | 20946 | 8695 | 11249 | 0.68 | 0.77 |
| 1990 | 12798300 | 24197 | 7044 | 10239 | 0.52 | 0.58 |
| 1991 | 9209620 | 31508 | 10302 | 11595 | 0.41 | 0.46 |
| 1992 | 17021300 | 33754 | 14588 | 13081 | 0.4 | 0.45 |
| 1993 | 8893240 | 37616 | 14891 | 12753 | 0.38 | 0.43 |
| 1994 | 10458200 | 39699 | 17168 | 11549 | 0.28 | 0.31 |
| 1995 | 13496900 | 37849 | 20342 | 13361 | 0.36 | 0.41 |
| 1996 | 14689900 | 37245 | 16461 | 14149 | 0.42 | 0.47 |
| 1997 | 7449280 | 40171 | 15918 | 15074 | 0.41 | 0.46 |
| 1998 | 9909750 | 37099 | 17824 | 15504 | 0.43 | 0.49 |
| 1999 | 9339180 | 30527 | 15383 | 11254 | 0.39 | 0.44 |
| 2000 | 9054270 | 29610 | 12996 | 11038 | 0.39 | 0.44 |
| 2001 | 16221700 | 28986 | 12994 | 11328 | 0.42 | 0.47 |
| 2002 | 10445000 | 34114 | 12375 | 12474 | 0.41 | 0.47 |
| 2003 | 13507200 | 38008 | 15177 | 13836 | 0.37 | 0.42 |
| 2004 | 9283370 | 38695 | 17629 | 15952 | 0.46 | 0.51 |
| 2005 | 16878400 | 35382 | 15829 | 14208 | 0.42 | 0.47 |
| 2006 | 11806300 | 37741 | 15079 | 14268 | 0.44 | 0.49 |
| 2007 | 11221900 | 40910 | 16289 | 13552 | 0.33 | 0.38 |
| 2008 | 5098770 | 40212 | 19776 | 13554 | 0.34 | 0.38 |
| 2009 | 4547680 | 32865 | 18389 | 11539 | 0.34 | 0.38 |
| 2010 | 4930500 | 23364 | 14638 | 8327 | 0.35 | 0.4 |
| 2011 | 7814620 | 18023 | 9719 | 9044 | 0.58 | 0.67 |
| 2012 | 5618690 | 16597 | 5941 | 8822 | 0.68 | 0.77 |
| 2013 | 12720600 | 16683 | 5813 | 9288 | 0.68 | 0.76 |
| 2014 | 5204940 | 21058 | 5758 | 12341 | 0.84 | 0.94 |
| 2015 | 5079720 | 20991 | 7171 | 12162 | 0.69 | 0.78 |
| 2016 | NA | NA | 7077 | NA | NA | NA |

Table 4. Summary of the reference points determined at the benchmark in 2016 (ICES 2016).

|  | Type | Value | Technical basis |
| :---: | :---: | :---: | :---: |
| MSY <br> Approach | MSY Btrigger | 9900 t | 5th percentile of equilibrium distribution of SSB when fishing at $\mathrm{F}_{\text {MSY }}$, constrained to be no less than $\mathrm{B}_{\mathrm{pa}}$ |
|  | $\mathrm{F}_{\mathrm{MSY}}$ | 0.62 | F that maximises median equilibrium yield (defining yield as the total catch) |
| Precautionary <br> Approach | Blim | 6300 t | Bloss (lowest observed SSB) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 9900 t | $\mathrm{Blim}^{*} \exp (1.645 * \sigma)$, where $\sigma=0.27$ |
|  | Flim | 1.00 | F that leads to 50\% probability of SSB $<\mathrm{Bl}_{\mathrm{lim}}$ |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.68 | $\mathrm{F}_{\text {lim }} * \exp (-1.645 * \sigma)$, where $\sigma=0.23$ |



Fig. 1. Data by type and year for each fleet and survey used in the SS3 model of the northern shrimp in the northern part of ICES Division IIIa (Skagerrak) and the eastern part of Division IVa (Norwegian Deep).

Landings by season and fleet (fitted:black; observed: one colour per season)


Fig. 2. Observed and fitted catches (in tons) in the SS3 final model by quarter. Observed data are colour coded as follows: Quarter 1: red, Quarter 2: green, Quarter 3: blue, Quarter 4: sky blue.


Fig. 3. Predicted abundances (blue line) and observed survey index (dots) with estimated $95 \%$ probability intervals of the survey index (first survey period).


Fig. 4. Predicted abundances (blue line) and observed survey index (dots) with estimated $95 \%$ probability intervals of the survey index (second survey period).


Fig. 5. Predicted abundances (blue line) and observed survey index (dots) with estimated 95\% probability intervals of the survey index (third survey period).


Fig. 6. Fishery selectivity-at-length estimated for the commercial fleet and for the surveys in the SS3 model.


Fig. 7. Average model fits to the length-composition data estimated for the commercial fleet and for the surveys (shown as observed and predicted length-frequency distributions, averaged across of all data years) for the SS3 model.


Fig. 8a. Yearly model fits to the length-composition data estimated for the commercial fleet (shown as observed and predicted length-frequency distributions by year) for the SS3 model.
length comps, whole catch, Fishery


Fig. 8b. Yearly model fits to the length-composition data estimated for the commercial fleet (shown as observed and predicted length-frequency distributions by year) for the SS3 model.


Fig. 9a. Yearly model fits to the length-composition data estimated for the first survey period (shown as observed and predicted length-frequency distributions by year) for the SS3 final model.

## length comps, whole catch, NorwaySurvey1



Length (cm)
Fig. 9b. Yearly model fits to the length-composition dada estimated for the second survey period (shown as observed and predicted length-frequency distributions by year) for the SS3 final model.


Fig. 9c. Yearly model fits to the length-composition dada estimated for the third survey period (shown as observed and predicted length-frequency distributions by year) for the SS3 final model.


Fig. 10. Pearson residuals for the proportions-at-length in the commercial fleet and surveys of the SS3 final model.


Fig. 11. Time trajectories of spawning biomass (January $1^{\text {st }}$ ) $+/-1$ standard deviation of the estimates as estimated by the model.


Fig. 12. Time trajectories of (age 0 ) recruitment (January $1^{\text {st }}$ ) $+/-1$ standard deviation of the estimates as estimated by the model.


Fig. 13. Time trajectories of fishing mortality (ages 1-3) estimated by the model.


Fig. 14. Retrospective analysis of the SS3 model: SSB (January 1).


Fig. 15. Retrospective analysis of the SS3 model: Recruitment (age 0 abundance on January 1).


Fig. 16. Retrospective analysis of the SS3 model: F(ages 1-3).


Fig 17. Summary assessment output: Total catch, including estimated discards since 2008 (tons) and F, SSB and R assessment results. SSB and R depicted with $90 \%$ confidence intervals. The assumed recruitment value for 2016 is unshaded.


Fig. 18. Northern shrimp in Skagerrak and Norwegian Deep: $F$ and $S S B$ assessment results for natural mortality $M=0.75$ (base model, black) and $M=0.90$ (red). Straight lines indicate $B_{p a}$ (left figure panel) and $F_{m s y}$ (right figure panel) based on the different runs.

Annex 1.

|  |  | F |  |  |  |  |
| :--- | :---: | :--- | :--- | :--- | :--- | :--- |
| SSB |  | F | (1- | Catch | Land | SSB |
| $(2016)$ | Rec proj | $(1-2.5 \mathrm{~cm})$ | 3years $)$ | $(2016)$ | $(2016)$ | $(2017)$ |
| 7077 | 6838538 | 1.36 | 1.53 | 12842 | 11085 | 2755 |


| Fmult | Fcatch $(1-2.5 \mathrm{~cm})$ | Fland $(1-2.5 \mathrm{~cm})$ | Fdisc <br> (1- <br> $2.5 \mathrm{~cm})$ | Fcatch <br> (1- <br> 3years) | Catch (2017) | Land (2017) | $\begin{aligned} & \text { Disc } \\ & (2017) \end{aligned}$ | $\begin{aligned} & \text { SSB } \\ & (2018) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8054 |
| 0.1 | 0.0733 | 0.0733 | 0 | 0.0827 | 855 | 715 | 140 | 7517 |
| 0.2 | 0.1467 | 0.1467 | 0 | 0.1654 | 1648 | 1373 | 276 | 7027 |
| 0.3 | 0.22 | 0.22 | 0 | 0.2481 | 2387 | 1979 | 408 | 6580 |
| 0.4 | 0.2934 | 0.2934 | 0 | 0.3308 | 3075 | 2537 | 538 | 6171 |
| 0.5 | 0.3667 | 0.3667 | 0 | 0.4135 | 3717 | 3052 | 665 | 5796 |
| 0.6 | 0.44 | 0.44 | 0 | 0.4962 | 4317 | 3529 | 788 | 5453 |
| 0.7 | 0.5134 | 0.5134 | 0 | 0.5789 | 4879 | 3970 | 909 | 5138 |
| 0.8 | 0.5867 | 0.5867 | 0 | 0.6616 | 5405 | 4378 | 1027 | 4849 |
| 0.9 | 0.66 | 0.66 | 0 | 0.7443 | 5899 | 4757 | 1142 | 4583 |
| 1 | 0.7334 | 0.7334 | 0 | 0.827 | 6363 | 5108 | 1255 | 4338 |
| 1.1 | 0.8067 | 0.8067 | 0 | 0.9097 | 6801 | 5435 | 1365 | 4113 |
| 1.2 | 0.8801 | 0.8801 | 0 | 0.9924 | 7213 | 5740 | 1473 | 3904 |
| 1.3 | 0.9534 | 0.9534 | 0 | 1.0751 | 7602 | 6024 | 1579 | 3711 |
| 1.4 | 1.0267 | 1.0267 | 0 | 1.1578 | 7971 | 6289 | 1682 | 3533 |
| 1.5 | 1.1001 | 1.1001 | 0 | 1.2406 | 8319 | 6536 | 1783 | 3368 |
| 1.6 | 1.1734 | 1.1734 | 0 | 1.3233 | 8650 | 6768 | 1882 | 3214 |
| 1.7 | 1.2467 | 1.2467 | 0 | 1.406 | 8964 | 6985 | 1979 | 3071 |
| 1.8 | 1.3201 | 1.3201 | 0 | 1.4887 | 9263 | 7188 | 2074 | 2938 |
| 2 | 1.4668 | 1.4668 | 0 | 1.6541 | 9818 | 7559 | 2259 | 2698 |

