

1 **Scenarios for land use and ecosystem services under global change**

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47 **Abstract**

48 Scenarios provide a platform to explore the provision of ecosystem services under global
49 change. Despite their relevance to land-use policy, there is a paucity of such assessments,
50 particularly in developing countries. Central Chile provides a good example from the Latin
51 American realm as the region has experienced rapid transformation from natural landscapes to
52 urbanization and agricultural development. Local experts from Central Chile identified climate
53 change, urbanization, and fire regimes as key drivers of change. Scenarios depicting plausible
54 future trajectories of change were developed to assess the combined effects on carbon storage,
55 wine production, and scenic beauty for the year 2050. Across the region, the action of the
56 drivers reduced the total amount of carbon storage (by 85%) and wine production (by 52%)
57 compared with a baseline scenario, with minor changes incurred for scenic beauty. The carbon
58 storage and wine production had declined by 90% and scenic beauty by 28% when the reaction
59 to changed fire regimes was also taken into account. The cumulative outcomes of climate
60 change and urbanization are likely to place substantial pressures on ecosystem services in
61 Central Chile by mid-century, revealing the need for stronger planning regulations to manage
62 land-use change.

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64 *Keywords:* carbon storage; scenic beauty; wine production; urbanization; climate change; fire.

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95 **1. Introduction**

96 Global efforts to achieve the United Nations Sustainable Development Goals will
97 require an understanding of how the provision of ecosystem services will be affected as a result
98 of global environmental change (Schröter et al. 2005, Rockström et al. 2009, Nelson et al. 2010,
99 González-Varo et al. 2013, Mace 2013). Drivers of environmental change are factors that
100 influence ecosystem services directly (e.g. climate change, land use change, invasive species)
101 or indirectly (policies, science and technology, cultural factors) shaping the direction,
102 magnitude and rate of future global change (MEA 2005, Kosow and Gaßner 2008). The
103 drivers of environmental change do not operate in isolation, necessitating that the combined
104 consequences of multiple drivers be determined (Nelson et al. 2006, Carpenter et al. 2009,
105 González-Varo et al. 2013).

106 Over the past three decades, scenario analyses have played a central role in assessments
107 of the potential effects of global environmental change on land systems at a variety of scales
108 (Nakicenovic et al. 2000, MEA 2005, O'Neill et al. 2008, Van Vliet et al. 2010, Bryan et al.
109 2016). Scenarios explicitly incorporate uncertainty by exploring the outcomes that could arise
110 due to multiple plausible futures. Scenarios are derived from a coherent and internally
111 consistent set of assumptions or storylines (Peterson et al. 2003, MEA 2005, Adams et al.
112 2016), which can be depicted as spatially and temporally-explicit projections of drivers such
113 as land-use and land cover, and climate change (Rounsevell et al. 2006, Rounsevell and
114 Metzger 2010). Such projections enhance the communication of ecosystem services
115 assessments and thus inform the development of robust land-use policies (Dunford et al. 2014,
116 Lamarque et al. 2014).

117 There has been a paucity of studies of ecosystem services assessments under global
118 change [but see (Oteros-Rozas et al. 2015)], particularly in developing countries of Latin
119 America (Seppelt et al. 2011, Runting et al. 2016). The exact nature of the effect of global
120 change in these countries is largely unknown, and their adaptive capacity is expected to be low
121 (Sinivasan 2010). In Latin America, the drivers assessed are climate change and deforestation,
122 parameters that are just a limited subset of global change (Grau and Aide 2008, Martínez et al.
123 2009, Birch et al. 2010, Carreño et al. 2012, Mendoza-González et al. 2012, Nahuelhual et al.
124 2014). Measuring the aftermath of a more-extensive set of global changes on ecosystem
125 services is an important policy-relevant task in Latin American countries (Schröter et al. 2005,
126 González-Varo et al. 2013, Oliveira et al. 2013). Rapid assessments using expert judgment and
127 existing empirical information can be used in initial policy cycle phases to help demonstrate
128 potential possible futures involving the drivers of change and their likely effects. Such
129 assessments are being called for to inform initiatives such as the Intergovernmental Science-
130 Policy Platform on Biodiversity, and Ecosystem Services (IPBES) (Brooks et al. 2014, Kok et
131 al. 2016).

132 In Chile, historical trends for the past 20 years and predictions covering the next century
133 suggest major changes in climate with a decline in rainfall and higher temperatures (Fuenzalida
134 et al. 2007, Falvey and Garreaud 2009). Such changes are expected to have an effect on the
135 distribution of ecological communities (Marquet et al. 2010). Chile also has experienced a rapid
136 process of economic development in the past 30 years, and this has resulted in the extensive
137 urbanization of the Metropolitan Region (Cohen 2004, Banzhaf et al. 2013). In this region, the
138 native Mediterranean vegetation is adapted to repeated cycles of forest fires associated with

139 high temperatures (Castillo et al. 2012b). Forest fires have increased in the past decade
140 resulting from human activity and land use change, and as a consequence, fires have been most
141 prolific in proximity to urban centers and roads (Castillo et al. 2012b, Altamirano et al. 2013).

142 We demonstrate a rapid assessment of the effects of global change on key ecosystem
143 services in the Central Chile region, where data is sparse. Our approach translated expert-
144 derived qualitative scenario storylines into quantitative spatial predictions of the combined
145 impacts of climate change, urbanization and fire on the future provision of carbon storage, wine
146 production and scenic beauty for the year 2050. The three ecosystem services evaluated are
147 critically important for the country's environmental sustainability, its economic activity, and
148 societal well-being (Figueroa 2016). Central Chile provides an exemplary study case of the
149 Latin American context as the region has experienced a long history of land conversion from
150 forest to agriculture, rapid urbanization and a changing climate with consequent effects on fire
151 regimes (Armesto et al. 2010, Schulz et al. 2010).

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153 **2. Methods**

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155 **2.1 Study Area**

156 The Metropolitan Region of Central Chile (33°26' and 34°19'S, Fig. 1) encompasses
157 approximately 15,402 km², with elevation ranging from 0 to 6500 m.a.s.l. Characterized by a
158 Mediterranean climate (warm and dry summers; cool and rainy winters) with mean
159 temperatures ranging from 20°C in summer to 8°C in winter and with an annual precipitation
160 of approximately 350 mm in the central valley, increasing with altitude (Meza et al. 2014).
161 Central Chile is the most densely populated area of the country with almost 7 million people
162 inhabiting the region (or 40% of the country's population), with 97% of people living in urban
163 areas (INE 2012) and producing, in 2014, 44% of Chile's total economic product (Central Bank
164 of Chile 2015). Urban development has occurred mainly on alluvial floodplains, which are also
165 the most fertile soils for agriculture (Puertas et al. 2014), especially for fruit and wine
166 production (Romero and Ordenes 2004). Urbanization is also trending into higher elevation
167 areas (Romero and Ordenes 2004, Romero et al. 2012). The study region has a high incidence
168 of fire events that have caused considerable material and environmental losses (Castillo et al.
169 2012b), and their frequency has increased in the past 20 years with an average of 5,000 fire
170 events per annum (Altamirano et al. 2013).

171

172 [Insert Figure 1 near here.]

173

174 **2.2 Scenario Building Process**

175 We developed and applied a framework for building scenarios that composed four main
176 steps (Schwarz 1991, Metzger et al. 2010): (1) define the scope and the focal questions, (2)
177 identify key drivers, (3) construct qualitative scenario storylines, and (4) quantify and map the
178 provision of ecosystem services under baseline conditions and under projections of land-use
179 and climate change.

180 [Insert Figure 2 near here.]

181 **2.2.1 Scope of the scenarios**

182 We defined the scope of the scenarios analysis as the exploration of the potential influence
183 of key global change drivers on three ecosystem services: carbon storage, wine production, and
184 scenic beauty, in Central Chile for the year 2050. Carbon storage in the native Mediterranean
185 forest has been identified as an important mechanism for mitigating the burden of climate
186 change (Gibbs et al. 2007, Caparros et al. 2011). Scenic beauty is defined as the aesthetic values
187 derived from the appreciation of natural scenery and scenic views (Bourassa et al. 2004,
188 Bagstad et al. 2014). The Mediterranean mountain landscapes in Central Chile are in demand
189 for leisure activities due to scenic views (De la Fuente et al. 2006, Schirpke et al. 2013). The
190 Mediterranean climate region is also an important region for wine production (Hannah et al.
191 2013), being the fifth largest exporter of wines in the world and the ninth largest producer
192 (Lobos et al. 2014). Central Chile has a large area that is potentially suitable for irrigated high-
193 quality wine production, particularly at the bottom of valleys (Montes et al. 2012).

194 **2.2.2 Identification of key drivers of change**

195 We developed a list of drivers of land-use and land-cover change via semi-structured
196 interviews with local experts (Appendix A). We initially contacted 25 experts by email and
197 completed 10 interviews. The experts were from different disciplines (demography,
198 economics, urban development, climate change, water, ecology, conservation, and
199 biodiversity) and possessed both local and regional-scale expertise of the study region. The list
200 of potential drivers was presented to the experts, and they selected and ranked drivers they
201 considered would have the greatest effect on the landscapes of the region for the year 2050
202 (Appendix A). We selected the two highest-ranked drivers for the development of the
203 storylines: climate change (specifically increasing temperature and decreasing precipitation)
204 and urbanization. Climate change is predicted to reduce the distribution of sclerophyllous and
205 thorny Mediterranean forest and reduce the carbon storage capacity of the landscape (Marquet
206 et al. 2010). Urbanization is being encouraged through regional urban plans (PRMS 2014),
207 which seek to expand the peri-urban limits of cities, especially in the northern and southeastern
208 sectors (Puertas et al. 2014). The ongoing expansion of urban areas is expected to lead to the
209 loss of native vegetation and fertile soils for viticulture and could reduce the scenic beauty of
210 the Andean foothills (Romero and Ordenes 2004, Banzhaf et al. 2013, Puertas et al. 2014).

211 **2.2.3 The scenario storylines**

212 To construct the storylines we developed a scenario matrix and defined assumptions about
213 the possible trends associated with climate change and urbanization (Plieninger et al. 2013),
214 reflecting ranges from low/weak to high/strong. The possible combination of drivers resulted
215 in four scenario storylines (see Fig. 3 for the definition of scenarios A, B, C, and D).

216 To define the assumptions for climate change we considered the greenhouse gas trajectories
217 (RCP: Representative Concentration Pathways) adopted in the Intergovernmental Panel on
218 Climate Change (IPCC) fifth assessment report for the year 2050, describing possible climate
219 futures (IPCC 2013). We focused on the lower and higher greenhouse gas-concentration level
220 trajectories (RCP 2.6 and 8.5 respectively) to encompass the range of uncertainty. According
221 to the climate change driver, scenarios A and B will follow trends defined in RCP 2.6 where
222 predicted emissions are substantially reduced over time (Kay 2013). Under this pathway, the
223 temperature will increase by no more than 2°C and will result in a reduction in precipitation by
224 no more than 10% by 2050 (Fuenzalida et al. 2007). Scenarios C and D will follow trends

225 defined in RCP 8.5 representing the business-as-usual scenario characterized by increasing
226 greenhouse gas emissions over time (Van Vuuren et al. 2011). “Business-as-usual” will result
227 in an increase of temperature by 3.5°C and a 15% reduction in precipitation by 2050, along
228 with an increase in the frequency of long and severe dry seasons (Fuenzalida et al. 2007, IPCC
229 2013, Kay 2013).

230 For urbanization, we considered the Regulatory Plan of the Metropolitan Region of
231 Santiago (PRMS 2014) and identified two opposing trajectories (see Fig. 3). Scenarios A and
232 C maintain urbanization at the current urban limits. This translates to the maintenance of
233 current urban areas and the locating of new dwellings on available land inside the urban radius
234 (23,800 ha of land available for construction). Scenarios B and D follow the new urban limits
235 defined in the regulatory plan PRMS 100 (PRMS 2014). This represents an expansion of the
236 urban radius by approximately 100 km² in eight districts of Santiago and removing construction
237 restrictions above an elevation of 1000 m.a.s.l.

238 Climate change and urbanization are not independent, and system feedbacks magnify the
239 interaction of both drivers and their combined effects (Nelson and Bennett 2005). The projected
240 consequences of climate change, along with the increasing human population density and
241 associated expansion of the road network, are predicted to lead to an increase in the prevalence
242 of fires in the study region. These interactions were included in the ecosystem service
243 assessments (Castillo 2012a, Altamirano et al. 2013).

244 [Insert Figure 3 near here.]

245

246 **2.2.4 Quantitative ecosystem service maps**

247 To translate the storylines into quantitative scenario maps, we identified available spatial
248 models and spatial criteria representing each of the assumptions behind the scenarios. We
249 mapped and modeled ecosystem services under baseline conditions and then developed
250 ecosystem services models representing likely changes in the provision of ecosystem services
251 under projected conditions. Finally, we developed a new set of ecosystem service scenarios
252 incorporating future shifts in the probability of fire, caused by the interaction between climate
253 change and urbanization.

254

255 ***Ecosystem service maps under baseline conditions***

256 ***Carbon storage***

257 To define the level of carbon storage for our study, we considered the carbon present in the
258 native forest where the tree cover density was greater than 10%, excluding exotic tree
259 plantations and harvest areas. Although exotic tree plantations store carbon, we excluded them
260 because in this region native forests have been heavily replaced with exotic plantations that
261 have proved to be incompatible with biodiversity conservation and restoration (Miranda et al.
262 2017). Sixteen forest type categories (Appendix B, Table B2) were identified in the study
263 region by intersecting four potential forest vegetation types (deciduous forest, sclerophyllous
264 forest, sclerophyllous Andean forest and thorny forest) (Luebert and Pliscoff 2006) (Appendix
265 B; Table B1) with the remnant native forest classes (closed >75%, semi-closed 50-75%, open
266 25-50% and very open 10-25% forest) from the land cover map (CONAF-CONAMA-BIRF
267 2014). The current carbon storage (total weight of carbon stored per hectare, Mg C ha⁻¹) of
268 each forest type category was measured as the long-term confinement of aboveground (AGB)

269 and belowground tree biomass (BGB). Aboveground carbon was quantified through a literature
270 review of biomass and carbon estimates for the representative species (Muñoz et al. 2007) and
271 BGB was estimated from ratios drawn from the literature (Aalde et al. 2006) (More details in
272 Appendix B).

273 *Wine production*

274 Wine production was mapped based on the area cultivated with *Vitis vinifera* within the
275 agricultural land cover class and the number of vines planted per area (Larrañaga 2011). The
276 number of vines per hectare was converted to yield (tonnes ha⁻¹ yr⁻¹) assuming that one vine
277 produces 7 kg yr⁻¹ of grapes for wine production (Muñoz et al. 2002). We obtained a map of
278 current wine yield production that we classified in 4 yield categories: 3 to 5ton/ha, 6 to 8ton/ha,
279 9 to 10ton/ha and 11 to 16ton/ha.

280 *Scenic Beauty*

281 Scenic beauty was mapped through a viewshed analysis in ArcMap 10.3.1 (ESRI 2011).
282 The viewshed analysis generated lines of sight between an observer site and the centroid
283 of each 90 m resolution cell of a digital elevation model DEM (Jarvis et al. 2008, Nutsford
284 et al. 2015). An average height of 22 m was then assigned to buildings in urban areas in the
285 capital city and 5 m elsewhere (PRMS 2014). An average height of 5m was assigned to forest
286 and 2 m to shrubs (CONAF-CONAMA-BIRF 2014). Viewpoints were selected for the
287 viewshed analysis to represent each of the populated peripheral provinces in the region except
288 provinces within the center of Santiago, excluded because of the high density of buildings taller
289 than 20 m. We also included viewpoints in conservation areas that are known for recreational
290 uses because of their scenery (see Appendix C for DEM and location of viewpoints).

291 We used one point per peripheral province (107 in the region) and one point per
292 conservation areas (24 in the region). In Appendix H we present the geographic coordinates
293 and characteristics of the 131 viewpoints. The viewpoint was set as the centroid of each
294 commune and conservation area respectively. To calculate the centroid of the communes we
295 used the Feature to Point (Data Management) tool in ArcMap 10.3.1 (ESRI 2011). This method
296 calculates the geometric center of the province or conservation area as a polygon feature,
297 computed using the weighted mean center of all the feature parts.

298 The viewshed analysis produced a visibility raster recording the number of times each
299 area was seen from the viewpoints' locations. We obtained a visibility raster map with values
300 ranging from 1 to 65 (e.g. areas that can be seen up to 65 times from viewpoints). This map
301 was classified in a qualitative scale map using the natural breaks categories of the visibility
302 raster map ranging from: very low (1-12), low (13-25), medium (26-37), high (38-50) and very
303 high (51-65). We also accounted for the contribution of scenic features providing high-quality
304 views, based on features identified in local studies (De la Fuente et al. 2006, De La Fuente and
305 Mühlhauser 2014). We extracted forest, water bodies and snow features from the land cover
306 map and intersected them with the visible area from the visibility raster. We included water
307 bodies and urban parks that were located in the visible area raster, which were obtained from
308 the OpenStreetMap for Chile (OSM 2016). These features were assigned a high scenic value
309 to represent people's preferences.

310

311 *Ecosystem service maps under future conditions*

312 To map the potential change in the distribution of the carbon and wine production under
313 climate change we employed maximum entropy bioclimatic modeling techniques at a
314 100x100 m grid cell resolution using MaxEnt v3.3.3j (Phillips et al. 2004, Phillips et al. 2006).
315 Each model was fitted using a split-sample approach (Guisan and Zimmermann 2000), using a
316 random set of occurrence points and reserving 25% for testing the performance of the model.
317 The minimum distance between the occurrence points for all dependent variables was restricted
318 to a maximum of 1000 m to minimize spatial autocorrelation.

319 The models were trained by establishing a relationship with current climate and the
320 selected environmental predictors at known occurrence points. The predictor variables used in
321 the MaxEnt model for carbon were climate and topography, for wine: climate, soil, and
322 hydrology (see Appendix D). The baseline scenario for carbon and wine production was the
323 output of the application of the MaxEnt model on current observations. The relationship was
324 then projected into the future climate under each of the RCP 2.6 and 8.5 scenarios by allocating
325 each 100x100 m grid cell to the dependent variable with the highest likelihood of prediction.
326 The climate predictor variables included a South American dataset available for baseline
327 climate conditions (1950-2000) obtained from a total of 930 weather stations at 1 km resolution
328 (Plischoff et al. 2014). For future climate projections, the output of a global climate model
329 (HadGEM2-AO (Fuenzalida et al. 2007, Falvey and Garreaud 2009) from the fifth phase of the
330 Coupled Model Intercomparison Project (CMIP5) was employed (Pachauri et al. 2014). The
331 remaining predictor variables used in each model are detailed in Appendix D.

332 The models were initially constructed using all predictor variables and then the
333 variables were sequentially excluded based on their percentage contribution, permutation
334 importance, the relative effect on model performance measured by the area under curve (AUC)
335 scores (see Appendix E for details in the AUC scores and Appendix F for the marginal plots of
336 the resultant models). To test how well the model predictions matched the reality we applied
337 simple linear regression analysis using the t-test and F-test respectively and assessed the
338 goodness-of-fit of the relationship between the current observation of carbon and wine
339 production (predictor) and the predictions obtained from the baseline scenario (dependent
340 variable). The result for current and future climate was a continuous value projection (0 to 1).
341 The continuous probability maps were converted into binary presence/absence maps applying
342 the maximum sensitivity plus specificity logistic threshold (Liu et al. 2005). To map the
343 potential change in the distribution of scenic beauty under climate change, we considered the
344 new potential distribution of forest cover under scenarios RCP 2.6 and RCP 8.5.

345 To map urbanization we employed the cartographic layer representing the new urban
346 regulatory plan (PRMS 2014). For scenarios A and C we maintained the current boundaries of
347 the city (no expansion), and for scenarios B and D we applied the new urban plan layer. We
348 then used these layers to assess the effects of urbanization on the future provision of carbon,
349 wine production, and scenic beauty. To combine both drivers in the final scenario maps the
350 urbanization driver was first applied, and then the climate change driver was applied to the
351 remaining area (see Fig. 4).

352

353 *Ecosystem service maps under future conditions incorporating fire*

354 To predict the future probability of fire occurrence, we applied bioclimatic suitability models
355 based on historical fire data [datasets for the period 1986-2010 from the National Corporation
356 of Forest (CONAF)] and environmental explanatory variables. We considered the dominant
357 environmental factors that influence fire: climatic conditions that affect the length and severity
358 of fire episodes; human activities that have increased the incidence of fires; and the presence
359 of flammable vegetation (Moritz et al. 2012). Climatic conditions were included in the set of
360 bioclimatic variables under scenarios RCP 2.6 and RCP 8.5 and the digital elevation model.
361 Human variables were incorporated through the distance of observed fires to roads and cities
362 considering the current urban plan and the new urban plan. Vegetation was included through
363 land cover categories with vegetation (Appendix D). We developed a new set of scenarios A,
364 B, C and D incorporating the effects of fire on the provision of the ecosystem services
365 according to the climate change and urbanization assumptions in each scenario (see Fig. 4). To
366 quantify the magnitude of change, we compared the percentage of change in the provision of
367 each ecosystem service for the eight future scenarios, relative to the baseline conditions.

368 [Insert Figure 4 near here.]

369

370 **3. Results**

371

372 The simple regression model outputs testing the plausibility of the carbon and wine
373 MaxEnt model predictions matched current observations and showed a robust significant and
374 positive relationship for carbon ($R^2=0.55$, $p<0.0001$, $F=7771$, $DF=6300$) and wine ($R^2=0.35$,
375 $p<0.0001$, $F=136.2$, $DF=254$).

376 The future scenarios revealed profound influence on the provision of ecosystem
377 services relative to the baseline scenario. For carbon storage, the four scenarios predicted a
378 substantial decline: close to 85% of baseline carbon stores and reaching up to 90% when the
379 effects of fire were accounted for (see Fig. 5). For wine production, the four scenarios also
380 predicted a decline, which ranged from 9 to 18% under scenarios A and B, with a pronounced
381 decline of 48 to 52% under scenarios C and D. The decline was even more dramatic when
382 changed fire regimes were accounted for, with total wine production declining by 90% (see
383 Fig. 5). Scenic beauty did not change much under the four scenarios with a slight increase
384 (7%) under scenario A and a slight decrease under scenarios considering urban expansion and
385 a business-as-usual climate. Scenic beauty was projected to decrease by 18 to 28% (see Fig.5)
386 when the outcomes of changes in the fire regimes were taken into account.

387 [Insert Figure 5 near here.]

388

389 **3.1 Carbon storage**

390 Carbon storage values ranged from 11 Mg ha⁻¹ to 63 Mg ha⁻¹ (see Fig. 6). The forest
391 types with higher carbon values were the closed Mediterranean Andean sclerophyllous,
392 sclerophyllous and deciduous forest types with lower values represented by the very open
393 thorny and sclerophyllous forest (Appendix B, Table B2). Scenarios A and B predicted high
394 carbon storage mainly concentrated on the southwest hills of the region in the coastal range.
395 Under these scenarios, there was a slight increase in the carbon content in the western part of

396 the region, specifically in the hills of the coastal range due to an expansion of closed
397 sclerophyllous vegetation types (Appendix F). There were also important zones of carbon
398 provision under these scenarios in the Andean foothills bordering the eastern part of the city
399 (see Fig. 6a).

400 Scenarios C and D predicted lower values of carbon storage across the region
401 (see Fig. 6b). The reduction was due to an expansion of open Mediterranean Andean
402 sclerophyllous forest, which displaced sclerophyllous forest types. In the south-eastern part of
403 the region, there was also a decrease in carbon storage caused by the displacement of the
404 Mediterranean Andean sclerophyllous forest by very open deciduous forest bordering the city
405 in the Andean range and central valley. Scenarios B and D predicted carbon losses in the
406 north-eastern part of the city due to the expansion of the city limits converting sclerophyllous
407 and Mediterranean Andean sclerophyllous forest to urban land (Appendix F). The expansion
408 of the urban boundary did not have a strong influence on carbon storage, as this land-use
409 change mainly affected agricultural lands bordering the city (which store less carbon). When
410 the probability of fire was incorporated, carbon storage declined by up to 90% compared with
411 the baseline, mainly affecting sclerophyllous forest types in the western coastal hills.

412

413 **3.2 Wine production**

414 The baseline scenario showed that western and southwestern sections of the region
415 closer to the coast had the highest potential for wine production. Interestingly, scenarios A
416 and B predicted some gains distributed along the central valley towards the coast (see Fig. 6a)
417 while scenarios C and D predicted a stronger decline (see Fig. 6b). Areas that would remain
418 with a high potential yield for wine production under these scenarios were located closer to
419 the coast in the southwestern section of the region. The planned expansion of the city
420 following scenarios B and D affected up to 9% of areas suitable for wine production, mainly
421 on those north and southwestern areas bordering the city. The effects of fire were predicted to
422 be severe for wine production because the modeled probability of fire-impacted areas were
423 closer to cities and roads, which were the most suitable for wine production. When the
424 probability of fire was incorporated, total wine production declined by 90% (see Fig. 5).

425

426 **3.3 Scenic beauty**

427 High values of scenic beauty for baseline and future conditions were found bordering
428 the city in higher elevation zones at the foothills of the Andean range (see Fig. 6) particularly
429 the eastern peripheral provinces from north to south (e.g. Colina, Huechuraba, Lo Barnechea,
430 Vitacura, Las Condes, La Reina, Peñalolén, La Florida, Puente Alto and Pirque — see Fig. 1
431 for spatial reference). High values of scenic beauty were found in some western peripheral
432 provinces bordering the city at the foothills of the coastal range (e.g. Lampa, Pudahuel, Isla
433 de Maipo and Paine). Higher elevation areas of the coastal range in the southwest of the
434 region in the Melipilla province presented a high potential for provision of scenic beauty.

435 Conservation areas located in higher elevation zones of the Andean (e.g. Natural
436 Sanctuary “Yerba Loca” and “San Enrique”, National Reserve “Rio Clarillo”) and the
437 Coastal range (e.g. “Cerro el Roble”, “Altos de Cantillana”, “Altos de Chicauma” — see

438 Fig. 1 for spatial reference) presented high scenic beauty values. This was because there are
439 natural features providing high-quality views such as closed sclerophyllous and deciduous
440 forest in the visible area (see Fig. 1 for pictures). Low-elevation areas of the central valley
441 contained most of the zones with low values of scenic beauty.

442 Scenario A predicted an expansion of the sclerophyllous forest in the coastal hills,
443 which increased the total scenic beauty value by 7%. The planned expansion of the city
444 following scenarios B and D would not dramatically affect the provision of scenic beauty
445 (See Fig. 6b). Nevertheless, there was a slight decline in scenic beauty (3%) concentrated in
446 areas set aside for urban expansion in higher elevation zones at the periphery of the city. For
447 example, some western peripheral (Lampa, Pudahuel, Quilicura and San Bernardo) and
448 southern peripheral provinces (Paine and Pirque) of the region suffered a decline in the
449 provision of scenic beauty. Taking the aftermath of fire into account, there was a greater
450 decline, ranging from 18 to 28% of the total provision of scenic beauty, which was mainly
451 due to the loss of coastal sclerophyllous forest (See Fig. 5).

452 [Insert Figure 6 near here.]

453

454 **4. Discussion**

455 Combining scenario analysis with ecosystem service assessments provided a powerful
456 tool for exploring the effects of combined global change drivers on the provision of
457 ecosystem services. Our application was significant because it evaluated the cumulative
458 aftermaths of multiple global change drivers on ecosystem services in a developing country
459 of Latin America, which had rarely been addressed in the literature (Runting et al. 2016).

460 The results demonstrated that global climate change, urbanization, and their
461 interactions in the form of fire dynamics, were likely to place substantial pressures on the
462 provision of carbon storage, wine production and scenic beauty in Central Chile by 2050.
463 This was especially the case for carbon storage and wine production, which suffered major
464 losses when the interactions between drivers were taken into account.

465 Climate change was predicted to have significantly different effects on ecosystem
466 services, with a decline in most, but not all, scenarios. Scenic beauty under the scenario of
467 moderate climate change and no urban expansion was the only service showing a minor
468 increase, and this was due to a localized expansion of the sclerophyllous forest on the hills of
469 the coastal range, which would provide higher quality views in these areas. Carbon storage
470 and wine production were very sensitive to climate change showing a decline under all
471 scenarios relative to baseline conditions. Carbon storage was the most severely affected
472 service, with a pronounced decline of over 85% from the baseline in all scenarios. Wine
473 production saw a small decline under moderate climate change scenarios (A and B), whereas
474 the severe climate change scenarios (C and D) predicted a larger decline. This occurred
475 because the mitigation scenarios (A and B) predicted localized gains at the center of the
476 region from the central valley to the coast.

477 Altitudinal and latitudinal movement of forest types and viticulture responses to the
478 new drier and hotter climate conditions explained the overall decline of carbon storage and
479 wine production. Suitable areas for high-yield wine production were likely to shift towards
480 the coast and southwards, where the temperature was likely to be lower and precipitation
481 higher. This is consistent with previous studies of the strain imposed by climate change on

482 plant communities and agricultural systems in the region (Marquet et al. 2010, Hannah et al.
483 2013). Mediterranean regions were notably vulnerable to climate change as the increase in
484 temperature, and reduced precipitation is expected to extend the duration of severe drought
485 (Schröter et al. 2005). Under these conditions water bodies would most likely be reduced in
486 surface and volume and this could reduce the perception of quality aesthetic values (García-
487 Llorente et al. 2012, Martínez Pastur et al. 2015) and potentially feedback to urban settlement
488 patterns due to pressure on the water supply. We did not include these potential concerns in
489 our analysis.

490 The urbanization driver did not dramatically affect the total provision of the
491 ecosystem services we explored, but it may have localized effects. The expansion of the city
492 was predicted to affect agricultural areas (including wine production) on the periphery of the
493 city. Scenic beauty was also expected to be affected by this driver, mainly in the mountainous
494 areas of the city periphery where expansion is planned to occur (Romero and Ordenes 2004,
495 De la Fuente et al. 2006, Romero et al. 2012, Banzhaf et al. 2013, De La Fuente and
496 Mühlhauser 2014, Puertas et al. 2014). Urban development at the foothills of the Andes and
497 in the coastal range had been facilitated by a lack of regulations to protect natural areas and
498 the ecosystem services they provided. The impact on ecosystem services was amplified when
499 the results of fires were taken into consideration. This was decidedly evident for wine
500 production because the occurrence of fire in this area was largely explained by human-
501 induced variables, with a high probability of fire affecting areas close to the city and roads in
502 locations that were suitable for wine production. These results were consistent with other
503 studies modeling the occurrence of fire in Central Chile (Castillo 2012a, Altamirano et al.
504 2013) and highlighted the importance of fire as a major driver of change in the spatial
505 patterns and overall provision of the selected ecosystem services.

506 Our research contributed to the literature on ecosystem services scenarios (Birch
507 et al. 2010, Haines-Young et al. 2011, Swetnam et al. 2011, Goldstein et al. 2012, Lamarque
508 et al. 2014, Lawler et al. 2014, Byrd et al. 2015) in that we rapidly and inexpensively estimated
509 the impact of interacting global and regional drivers on the provision of ecosystem services.
510 There were few scenario studies to date that had assessed future changes in ecosystem services
511 under climate change (e.g. Bryan et al. 2010, Bryan et al. 2011, Bryan et al. 2014, Lamarque
512 et al. 2014), fewer that had assessed the outcome of climate change and urban sprawl (e.g.
513 Bohensky et al. 2011, Shaw et al. 2011, Hoyer and Chang 2014, Byrd et al. 2015), and even
514 fewer that had taken into account the considered interactions between drivers (e.g. Oliveira et
515 al. 2013). For example, Shaw et al. (2011) examined the impact of climate change on the
516 production and value of ecosystem services in California. Byrd et al. (2015) developed climate
517 and land-use change scenarios based on the IPCC narratives to understand the effect on
518 ecosystem services and Hoyer and Chang (2014) mapped the provision of freshwater
519 ecosystem services under urbanization and climate change scenarios.

520 Our study differed from previous studies in that we developed future scenarios
521 highlighting that climate change and urbanization led to an overall decline in the provision of
522 carbon, wine and scenic beauty, which was exacerbated by land-use interactions with climate.
523 Considering the combined consequences was a significant advance over studies that focused
524 on the trajectories of independent drivers (Bryan 2013, Bryan and Crossman 2013).
525 Nonetheless, many challenges remained. There was a need to deepen our knowledge of the

526 emergent properties, complexities, interconnections and synergistic interactions among
527 multiple drivers of change and ecosystem services (Liu et al. 2015). While scenario analysis
528 was an important tool for exploring alternative futures arising from uncertainties in the drivers
529 of change, it did not encompass all the different sources of uncertainty in modeling future
530 outcomes. For example, in this study, in the case of wine production, we did not consider all
531 the possible socioeconomic drivers of vineyard distribution, or the possibility that under severe
532 climate change conditions less favorable environmental conditions would arise for wine
533 production (e.g. aspect, soil moisture, nutrient availability). The models developed could be
534 improved with finer parametrization under future conditions as ecosystem services productivity
535 was quite likely to change unevenly across space according to biophysical and socioeconomic
536 parameters. Ideally, we would have incorporated these uncertainties and others, such as those
537 arising from model parameters and model structure (Refsgaard et al. 2007), which would be
538 likely to lead to further variation in the results presented here. More effective integration also
539 required using more powerful tools than those presented in this study case (e.g. markov
540 decision-making, supply chain analysis, multilevel modeling, agent-based modeling), to be
541 able to predict the emergence of unexpected threats to ecosystem services (Liu et al. 2015).

542 The decision-making processes of governments typically ignore the consequences of
543 global change on the long-term sustainability of ecosystem services (Liu et al. 2015). To
544 address this critical situation, international and national policy needs to include strategies to
545 protect and manage ecosystem services despite the substantial uncertainties in future conditions
546 (Kok et al. 2016). Scenarios of ecosystem services are an important component of the
547 Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES),
548 demonstrating their utility at multiple scales (Díaz et al. 2015). IPBES has identified the
549 development of scenarios as a key tool for helping decision-makers identify potential impacts
550 of different policy options on biodiversity and ecosystem services. The panel needs to engage
551 with the great diversity of local contexts that are linked to global scale scenarios to improve
552 the policy relevance of future IPBES scenarios (Kok et al. 2016). South America is a data
553 sparse region in terms of ecosystem services knowledge (Boerema et al. 2016, Runting et al.
554 2016). The local scenarios developed in this study case have the potential to inform the IPBES
555 Americas section by providing a rapid and inexpensive assessment of the possible effects of
556 drivers on the productivity of ecosystem services that are key to local people.

557

558 **5. Conclusion**

559 Central Chile is a particularly sensitive area for climate change due to severe dry
560 conditions predicted by business-as-usual scenarios. Scenarios depicting plausible future
561 trajectories of change predicted that interactions between land-use and climate will give rise to
562 favorable conditions for fire propagation, putting substantial pressure on the ecosystem
563 services studied and especially on wine production, an important economic activity of the
564 region. This information contributes to our growing understanding of the influence of global
565 change on ecosystem services and highlights the urgent need for institutional responses better
566 able to steer us towards a more desirable future.

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8. Figures

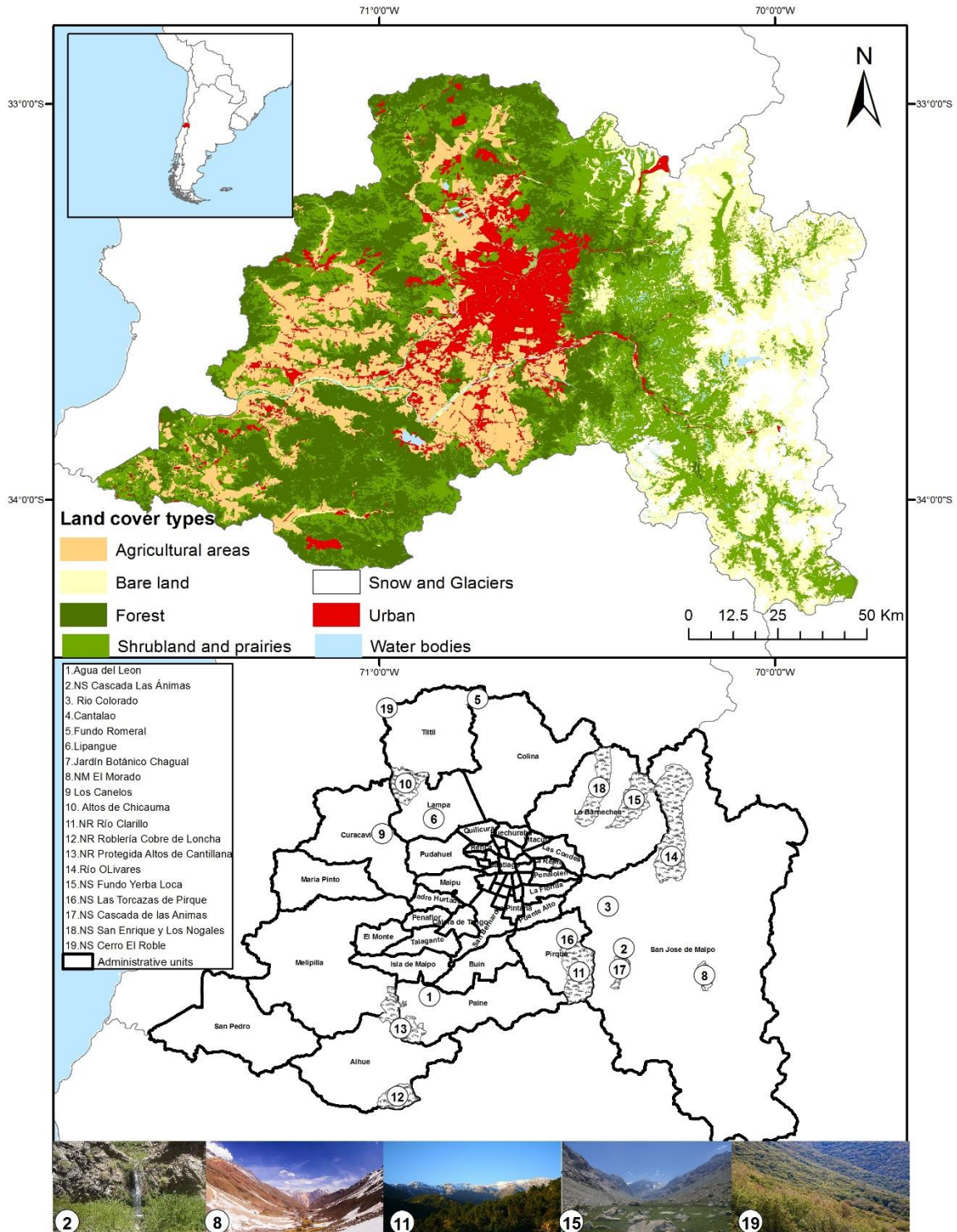


Figure 1. Study region in Central Chile depicting the main land cover types (above) administrative division of the provinces and the location of protected areas (below). All photos licensed by CC BY-NC-ND 2.0. Photos taken by: (2) Leonardo Needham, (8) Jose Letelier Hernandez, (11) Rodrigo Tejada, (15) Hixaga and (19) Jorge Barahona.

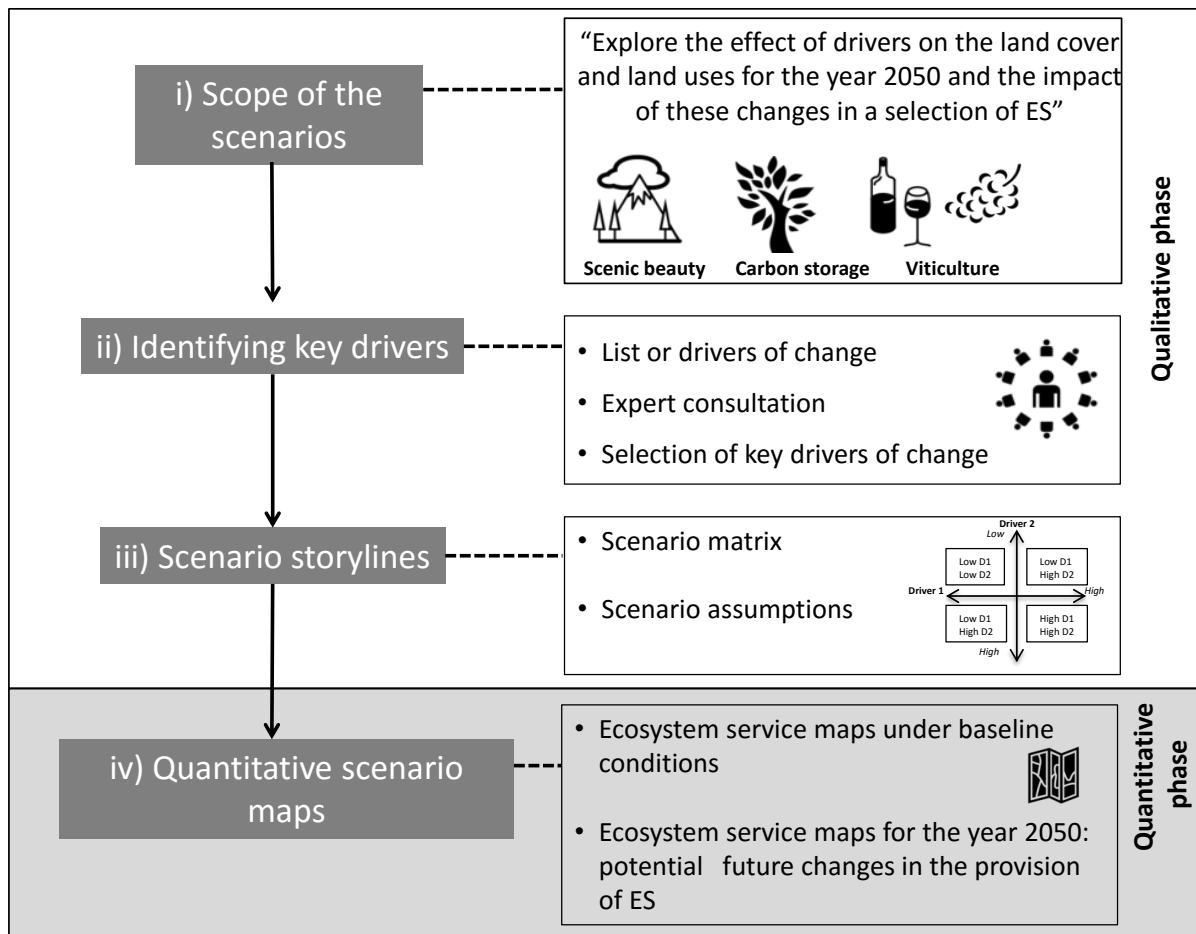


Figure 2. Methodological framework of the scenario building process.

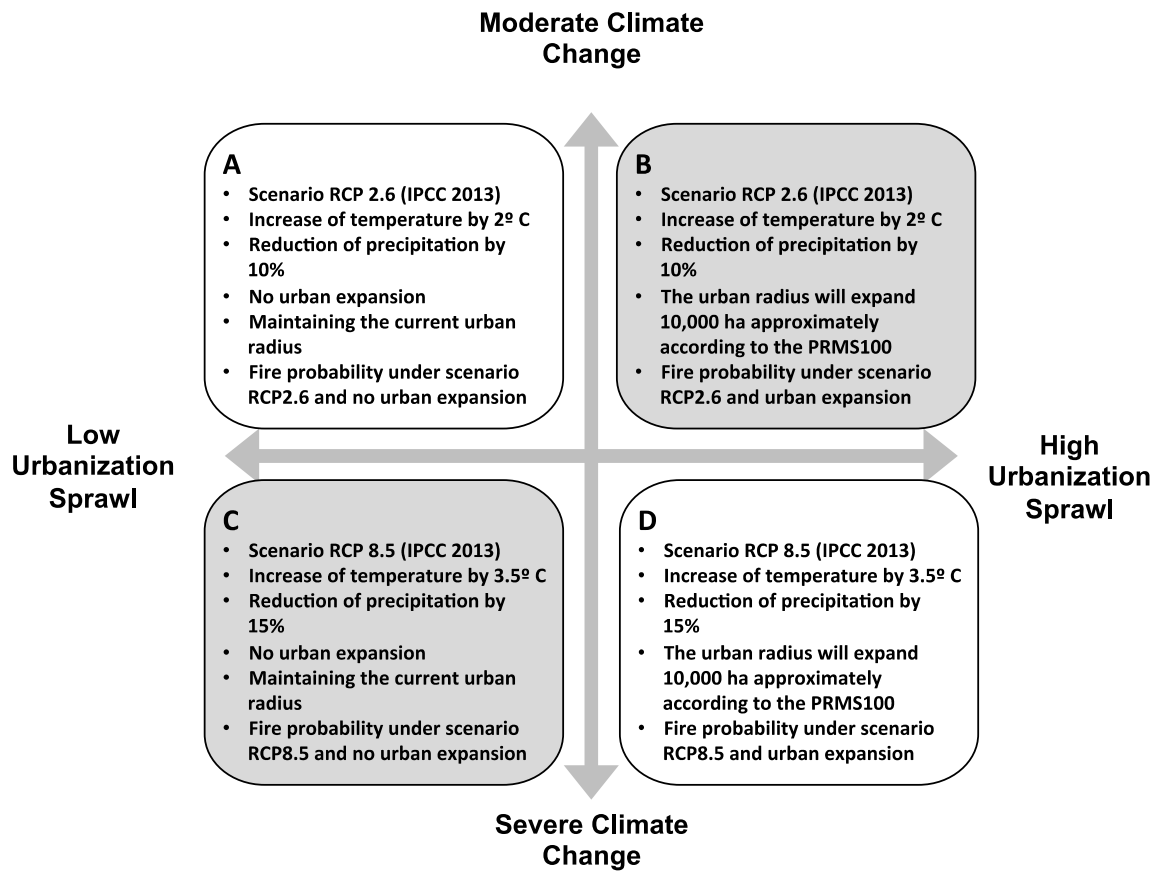
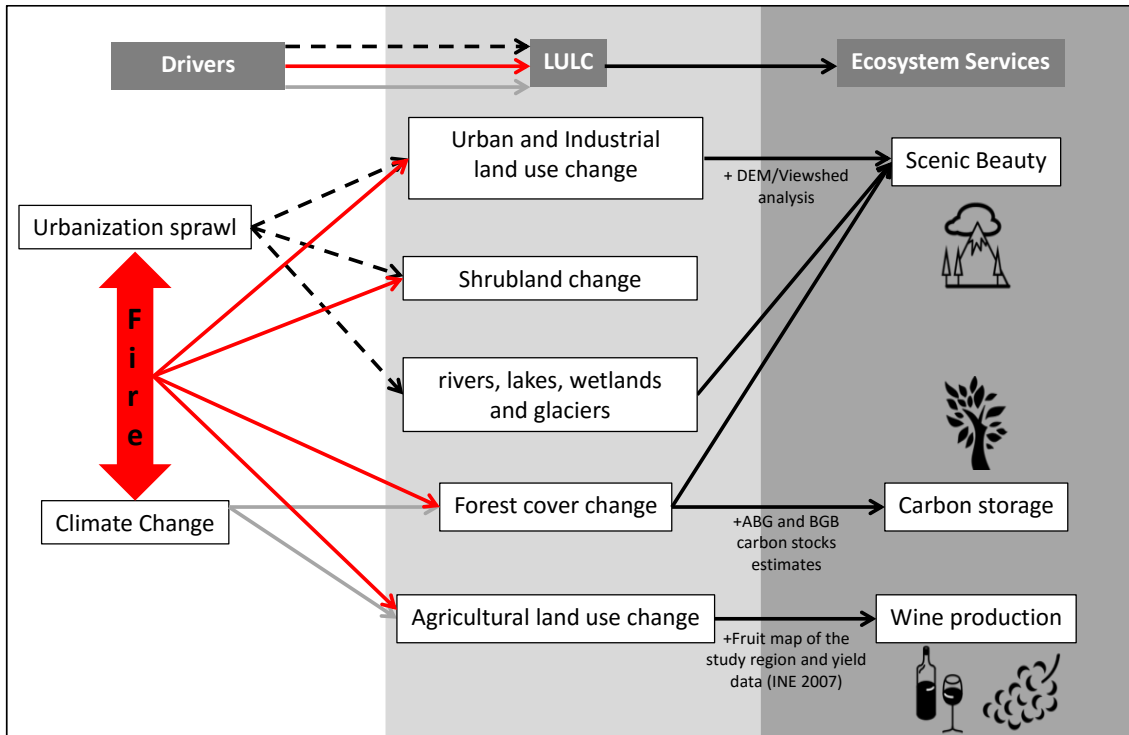


Figure 3. Scenario storylines according to climate change and urbanization drivers. All scenarios were implemented with and without fire probability.



- -> Spatial rules applied to urban regulatory plan (more restrictions and no restrictions)
 -> Application of climate distribution models (according to the RCP 2.6 and 8.5, from the IPCC fifth report)
 -> Effects of fire under climate change and urbanization assumptions
 -> Inputs of the ecosystem services models

Figure 4. Translation of the trajectories of the drivers defined in the storylines into ecosystem service scenario maps.

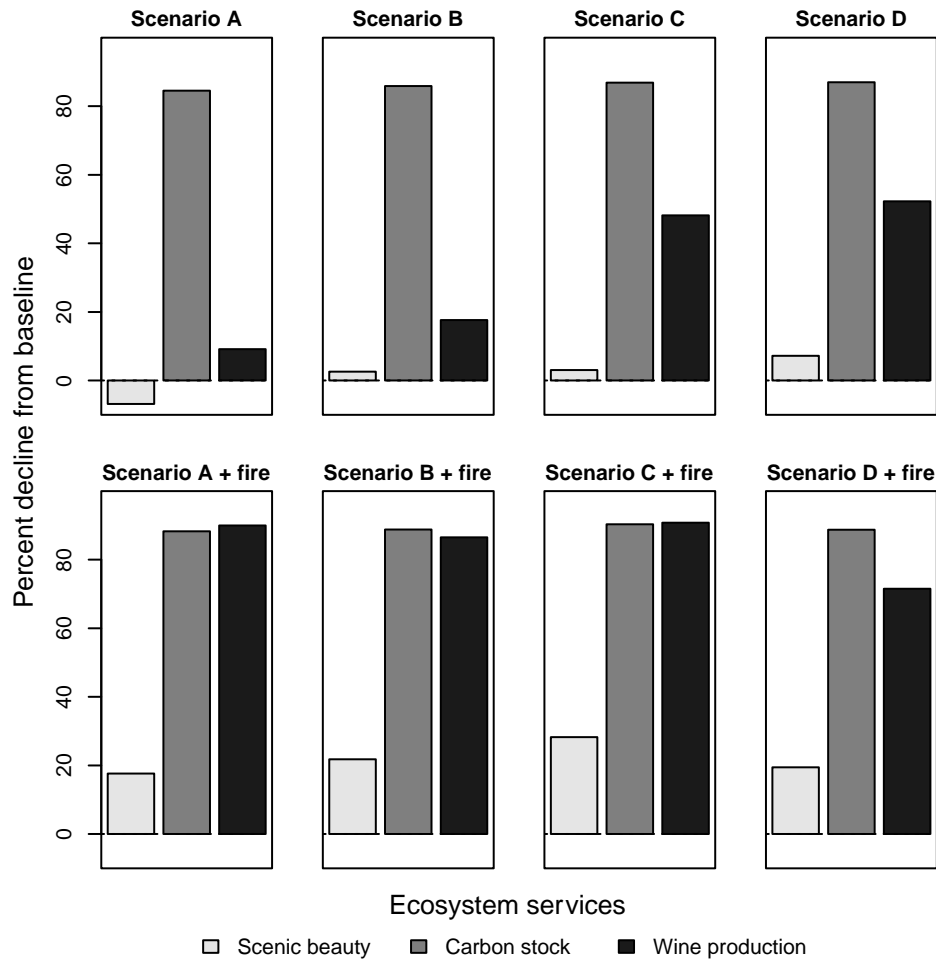


Figure 5. Percentage decline in the total provision of carbon storage, wine production and scenic beauty for all scenarios considering climate, urbanization and fire pressures compared with baseline conditions.

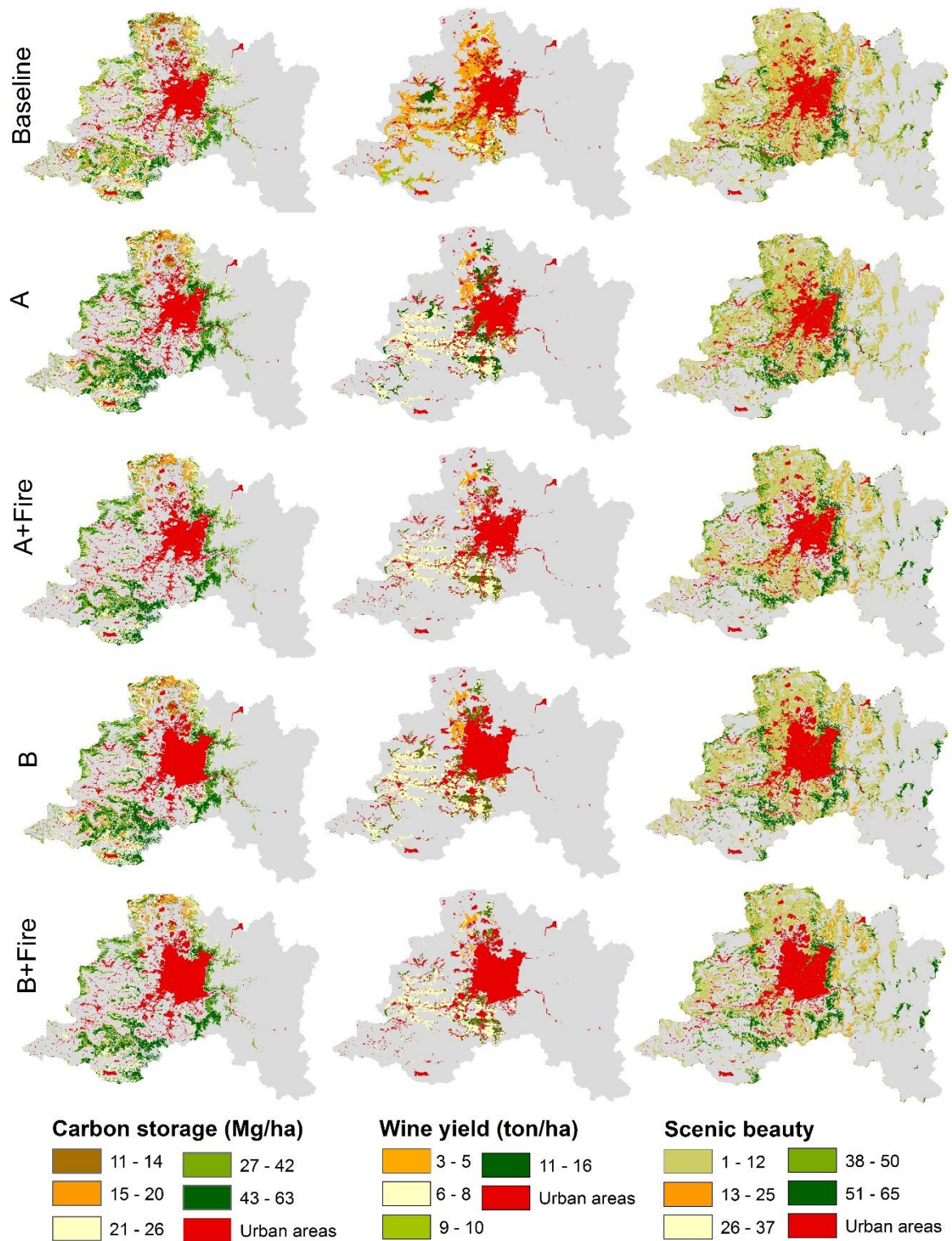


Figure 6a. Maps representing the ecosystem services scenarios part 1.

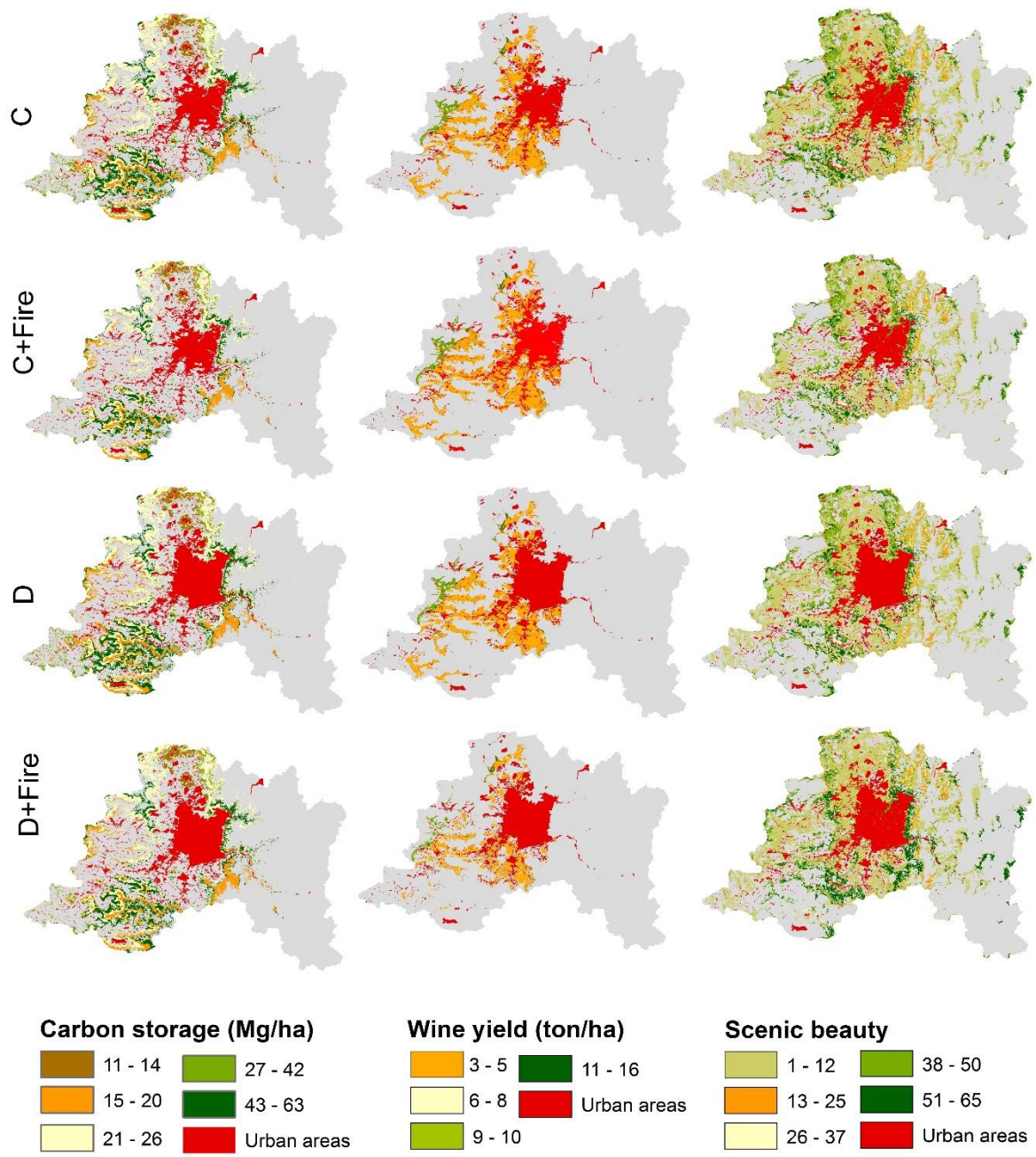


Figure 6b. Maps representing the ecosystem services scenarios part 2.