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Seasonal investigation of heavy metals in marine fishes captured from the Bay of Bengal and the implications for human health risk assessment

Narottam Saha^{1,2}, M.Z.I. Mollah³, M.F. Alam³, and M. Safiur Rahman^{3,4}

¹ Department of Applied Chemistry and Chemical Engineering, Faculty of Engineering, University of Rajshahi, Rajshahi 6205, Bangladesh.

² School of Earth Sciences, University of Queensland, QLD 4072, Australia.

³ Environmental Analytical Chemistry Laboratory, Institute of Nuclear Science and Technology, Bangladesh Atomic Energy Commission, GPO Box 3787, Dhaka 1000, Bangladesh.

⁴ Atmospheric and Environmental Chemistry Laboratory, Atomic Energy Center, 4-Kazi Nazrul Islam Avenue, Dhaka 1000, Bangladesh (present Address)

Corresponding author: E-mail address: n.saha@uq.edu.au (N. Saha) & safiur_baec@yahoo.com (S. Rahman)

Abstract

To investigate the seasonal contamination levels and to evaluate the potential human health risks, ten heavy metals (As, Cd, Co, Cr, Cu, Mn, Ni, Pb, Se, and Zn) were measured in ten different marine fish species from the Bay of Bengal during four seasons. The metal concentrations varied significantly ($p < 0.05$) among the fish species with maximum and minimum accumulation of Zn (46.47 $\mu\text{g/g}$) and Cd (0.25 $\mu\text{g/g}$), respectively. The fishes captured during summer accumulated higher amount of metals relative to other seasons attributed to washing down the agricultural waste, sewage water and sludge by heavy rainfall and floods. According to estimated daily intake (EDI), target hazard quotient (THQ), total target hazard quotient (TTHQ), and the permissible safety limits prescribed by various agencies, consumption of the examined fishes should be considered as safe for human health. However, the estimation of carcinogenic risk ($\text{CR} > 10^{-5}$) due to exposure to arsenic indicated that consumers remain at risk of cancer.

12

Keywords: Fish; Bay of Bengal; Seasonal fluctuations of heavy metals; Carcinogenic and non-carcinogenic human health risk.

15 1. Introduction

16 Aquatic environments worldwide are being heavily polluted by various contaminants, such as heavy metals
17 (Rahman, Molla, Saha, & Rahman, 2012; and reference therein), polycyclic aromatic hydrocarbons (PAHs) (Y.
18 Li, Li, Ma, Song, Zhou, Han, et al., 2015; and reference therein), polychlorinated biphenyls (PCBs) (Zhang, Li,
19 Wang, Zhang, Zhang, Zhang, et al., 2014, Hosseini, et al., 2008; and reference therein) with rapid economic
20 development and population growth. Pollution from the heavy metals (e.g., Cd, Cr, Pb etc.) has become a
21 serious issue due to their stable, persistent, and nonbiodegradable nature (Saha & Zaman, 2013). Although the
22 metals are naturally occurring constituents in the environment, their concentrations can be exacerbated by
23 anthropogenic activities such as rapid industrialization and urbanization, massive land use changes and
24 associated enhanced terrestrial runoff (Rahman, Molla, Saha, & Rahman, 2012). Another important source of
25 anthropogenic pollution in Chittagong coastal area, Bangladesh is ship breaking activities began since 1969.
26 Various refuse and disposable materials are being discharged and spilled from scrapped ships, and often get
27 mixed with the beach soil and sea water. Under certain conditions, the heavy metals can be bioaccumulated in
28 aquatic organisms (e.g., fish) from the surrounding environments and/or bioamplified to hazardous levels via
29 dietary exposure. Subsequently, these contaminants migrate to human body through diet and results in various
30 adverse health effects such as impaired kidney function, poor reproductive capacity, liver damage, skin and
31 bladder cancer, and even death (Wei, Zhang, Zhang, Tu, & Luo, 2014). However, there are some essential heavy
32 metals, such as Fe, Zn, and Cu required for both aquatic organisms and human body. But prolonged exposure to
33 the excess level of these metals could be toxic. On the contrary, there are some nonessential toxic metals such as
34 As, Cd, and Pb with no known potential benefits for a human being (Wei, Zhang, Zhang, Tu, & Luo, 2014).

35 Food consumption is the main pathway (accounting for > 90%) for human exposure to heavy metals compared
36 to other routes such as inhalation and dermal contact. Among various foodstuffs, fish is widely consumed and a
37 main source of nutrition in many coastal communities. It contributes to a healthy diet by providing high-value
38 amino acids and nutrients and is an excellent source of essential omega-3 fatty acids. Although fish is highly
39 nutritious, higher consumption rate can have significant deleterious effects on human health because of
40 bioaccumulated toxic metals beyond the safe limits. To date, the balance between benefits and risks has been
41 poorly known. However, fishes are often treated as the most suitable bioindicator of aquatic ecosystem since
42 they occupy high trophic level (Rahman, Molla, Saha, & Rahman, 2012). Many researchers globally have made
43 efforts to determine the concentrations of heavy metals in fishes to investigate the eco-environments and
44 potential health risks to the consumers (Ahmed, Shaheen, Islam, Habibullah-al-Mamun, Islam, Mohiduzzaman,

45 et al., 2015; Copat, Arena, Fiore, Ledda, Fallico, Sciacca, et al., 2013; P. Li, Zhang, Xie, Liu, Liang, Ren, et al.,
46 2015; Mendil, Demirci, Tuzen, & Soylak, 2010; Saha & Zaman, 2013; Wei, Zhang, Zhang, Tu, & Luo, 2014).

47 In recent years, the quantification of risk has become important due to the fact that exceedance of recommended
48 levels of contaminants prescribed by various regulatory bodies does not always represent a risk for the human
49 health. The target hazard quotient (THQ) and/or total target hazard quotient (TTHQ) set by US Environmental
50 Protection Agency (USEPA, 2000) are commonly used to evaluate the potential non-carcinogenic health risks
51 associated with variety of metals through fish consumption (Copat, et al., 2013; Saha & Zaman, 2013). USEPA
52 also provided cancer slope factor for arsenic, to determine the carcinogenic risk (CR) over a lifetime exposure to
53 arsenic.

54 Bangladesh is one of the coastal countries of the Bay of Bengal, where 37 to 38 million peoples live in the
55 coastal zone. The country is located between 20°34' to 26°38' north latitude and 88°01' to 92°42' east longitude.
56 Fisheries resources play a vital role in the economy of Bangladesh (4.57% of GDP in 2008-09) and it is the
57 second most important source of the foreign exchange earnings (5.71%). The national and international demand
58 of Bangladeshi fishes is increasing gradually. Recently, CBI (CBI, 2012) reported that in the period of 2002-
59 2011, import of marine fishes by the European Union (EU) has been increased from US\$ 150 m to US\$ 360 m.
60 Thus, the concentrations of heavy metals in the seafood deserves farther investigation not only for their potential
61 ecological impacts but also due to the health risks for the people in Bangladesh and other importer countries.

62 The objectives of this study are (i) to report on the fluctuation of concentrations of heavy metals (As, Cd, Co,
63 Cr, Cu, Mn, Ni, Pb, Se, and Zn) in fishes captured from Chittagong coastal zone with respect to seasonal
64 variation, and (ii) to quantify the non-carcinogenic and carcinogenic human health risks associated with these
65 metals.

66

67 **2. Materials and Methods**

68

69 *2.1. Sample collection and preservation*

70 Ten fish species of *Lates calcarifer* (Bhetki), *Pangasius pangasius* (Pangas), *Polynemus indicus* (Lakhua),
71 *Ilisha megaloptera* (Choukya), *Arius cruciger* (Rita), *Pampus chinensis* (Rupchanda), *Setipinna phasa* (Phasa),
72 *Scomberomorus guttatum* (Maitya), *Cirrhina reba* (Beta) and *Arius arius* (Kata mach) were procured from
73 fishermen while they were fishing nearshore sites at Cox's Bazar, Chittagong (Fig. 1) during spring, summer,
74 autumn and winter. Samples were collected in triplicate and in total, 120 individuals (i.e., 30 samples in each

75 season) representing 10 different species were collected. Immediately after collection, fishes were washed with
76 fresh water to remove the mud or other fouling substances, and were wrapped in polyethylene bags to transport
77 into the analytical chemistry laboratory of Institute of Nuclear Science and Technology, Bangladesh Atomic
78 Energy Commission (BAEC). After transportation to the laboratory, the muscle tissue of each sample was
79 removed and chopped into pieces with the aid of a steam cleaned stainless steel knife. The samples were then air
80 dried to remove the extra water and subsequently, oven-dried to a constant weight. Finally, the dried samples
81 were ground, sieved, and stored in clean and dry airtight plastic vials inside desiccators for succeeding uses.

82

83 *2.2. Sample Digestion*

84 For metal analysis, 0.5 g of each powdered sample was digested with 2.5 ml of conc. H_2SO_4 and 4.0 ml
85 conc. HNO_3 . When initial vigorous reaction subsided, the mixture was heated slowly on an oil bath by the
86 addition of 3/4 drops of H_2O_2 . This step was repeated till the solution became clear. Subsequently, the solution
87 was heated for additional 20 min. at about 150 °C and allowed to cool to room temperature. The digested fish
88 sample was then diluted to a total volume of 50 ml with double distilled water. Thereafter, the diluted solution
89 was filtered and stored in 50-ml polypropylene tubes.

90

91 *2.3. Analytical methods and quality control*

92 The concentration of Cd, Co, Cr, Cu, Mn, Ni, Pb, and Zn in digested solutions was analyzed by atomic
93 absorption spectrophotometer (AAS) (Model AA-6800, Shimadzu Corporation, Japan) using air-acetylene flame
94 with digital read out system. However, hydride vapor generators (HVG) were used along with the flame AAS
95 (FAAS) systems to determine As and Se concentrations. The instrument calibration standards were made by
96 diluting standards (1000 mg/l) supplied by Wako Pure Chemical Industry Ltd., Japan.

97 The accuracy of the method was evaluated by analyzing blanks and a certified reference material MA-A-2 (fish-
98 flesh standard from International Atomic Energy Agency (IAEA), Vienna) by the same procedure used for fish
99 samples. Mean recoveries of the analyzed metals were between 95-104 %, indicating a good agreement between
100 the certified and measured values (Table S1).

101

102 *2.4. Statistical analysis*

103 ANOVA test was conducted to investigate the effect of different fish species and seasons on the variation in
104 metal concentrations. In order to quantify the seasonal effects on the variability of metals, regression analyses

were performed using dummy variables for the seasons (winter as reference). In all cases, the level of significance was set at a 95 % (i.e., $\alpha = 0.05$). The data were analyzed statistically using IBM-SPSS Statistics (Version 21) for windows (IBM, USA).

108

2.5. Human health risk assessment

110

2.5.1. Estimated daily intake (EDI)

The estimated daily intake (EDI) for each analyzed heavy metal was calculated in the following way (Saha & Zaman, 2013):

$$EDI = \frac{E_F \times E_D \times F_{IR} \times C_f \times C_m}{W_{AB} \times T_A} \times 10^{-3} \quad (i)$$

Where E_F is the exposure frequency (365 days/year). E_D is the exposure duration, equivalent to an average life time of the Bangladeshi (i.e., 60 years). F_{IR} is the ingestion rate (g/person/day) of fish tissue. The fish production in Bangladesh was unable to keep pace with high population growth and as a result, per capita consumption of fish was decreased from 33 g in 1963-64 to 20.5 g in 1989-90 against the recommended level of 38 g per capita per day. Recently, the per capita fish consumption has increased to 26 g (M. M. Hossain, 2010) and used in the EDI calculations. C_f is the conversion factor ($C_f = 0.208$) to convert fresh weight to dry weight considering 79% of moisture content of the fish fillet (Rahman, Molla, Saha, & Rahman, 2012). C_m is the metal concentration in fish fillet ($\mu\text{g/g}$ - dry weight basis). Since most As present in foodstuffs as less harmful organic forms, the EDI for this metal was calculated assuming 3 % toxic inorganic As in total (Copat, et al., 2013; P. Li, et al., 2015). W_{AB} is the average body weight of an adult (60 kg), and T_A (equal to $E_F \times E_D$) is the average exposure time for non-carcinogens (Saha & Zaman, 2013).

126

2.5.2. Non-carcinogenic risk

The target hazard quotient (THQ), which is used to express the risk of non-carcinogenic effects, is the ratio between the estimated exposure (EDI) and the oral reference dose (RfD, mg/kg bw/day) (USEPA, 2000). The RfD represents an estimate of the daily exposure to which the human population may be continually exposed over a lifetime without an appreciable risk of deleterious effects. The RfDs are based on 0.0003, 0.001, 0.0003, 1.5, 0.04, 0.14, 0.02, 0.004, 0.005, and 0.3 (mg/kg bw/day) for As (inorganic), Cd, Co, Cr, Cu, Mn, Ni, Pb, Se and Zn, respectively. The THQ was calculated based on the following equation (USEPA, 2000):

133

$$134 \quad \text{THQ} = \frac{\text{EDI}}{\text{RfD}} \quad (\text{ii})$$

135 If the THQ value is less than 1, the exposed population should not experience any adverse health hazard.
 136 Conversely, the receptors of concern may experience non-carcinogenic health risks if the THQ is equal 1, with
 137 an increasing probability as the value increases. The method for the determination of THQ was provided in US
 138 EPA Region III risk-based concentration table (USEPA, 2000). The THQ calculations were made based on the
 139 facts that the ingestion dose is equal to the absorbed contaminant dose, cooking has no effect on the
 140 contaminants (USEPA, 1989) and average body weight of an adult is 60 kg (Saha & Zaman, 2013).

141 It has been reported that exposure to two or more pollutants may result in additive and/or interactive effects.
 142 Thus, in this study, cumulative health risk was evaluated by summing of the individual metal THQ value and
 143 expressed as total THQ (TTHQ) (also called hazard index, HI) as follows:

$$144 \quad \text{TTHQ} = \text{THQ (toxicant 1)} + \text{THQ (toxicant 2)} + \dots + \text{THQ (toxicant n)} \quad (\text{iii})$$

145 The greater the value of TTHQ the greater the level of concern. The TTHQ value above 1 generally indicates
 146 the potential for adverse human health effects and suggests the need for undertaking a farther level of
 147 investigation or possibly remedial action.

149 2.5.3. Carcinogenic risk

150 Carcinogenic risk (CR) indicates the incremental probability of an individual developing cancer over a
 151 lifetime due to exposure to a potential carcinogen. The cancer risk over a lifetime exposure to As was obtained
 152 using cancer slope factor (CSF), provided by USEPA (USEPA, 2000). The equation used for estimating the
 153 cancer risk is as follow (USEPA, 2000):

$$154 \quad \text{CR} = \text{CSF} \times \text{EDI} \quad (\text{iv})$$

155 Where CSF is the carcinogenic slope factor of $1.5 \text{ (mg/kg/day)}^{-1}$ set by USEPA only for inorganic As. EDI is the
 156 estimated daily intake of heavy metals. The acceptable lifetime carcinogenic risk considered by USEPA is 10^{-5}
 157 (risk of developing cancer over a human lifetime is 1 in 100000) and applied in this study.

158

159 3. Result

160 3.1. Levels of metals accumulation in fish fillets

161 Mean concentrations and standard deviations of As, Cd, Co, Cr, Cu, Mn, Ni, Pb, Se, and Zn obtained in the
162 edible portions of ten fish species during spring, summer, autumn, and winter are presented in Table S2 (also see
163 Fig. 2). Although some investigators prefer to determine heavy metals in liver, kidney, and gills, the present
164 investigation was carried out on the edible flesh of the marine fish species, because the dietary habit of the
165 people of Bangladesh excludes other parts of fish. Our results revealed that the ranking order of mean metal
166 concentrations were, Zn (46.47) > Mn (10.61) > Cu (5.21) > Ni (4.85) > Pb (3.57) > Cr (3.06) > As (2.79) > Se
167 (2.66) > Co (0.38) > Cd (0.25) (mean concentration; $\mu\text{g/g}$ - dry weight basis). The heavy metal concentrations
168 detected in this study were also compared with other published data in Fig. 3 as an effort to determine the degree
169 of contamination in the present study area.

170

171 *3.1.1. As*

172 Arsenic is ubiquitous and a potential toxic trace metal in the environment and it originates from both the
173 natural and anthropogenic processes. It has been reported by US Food and Drug Administration (USFDA) that
174 fish and other seafood account for 90% of total human exposure to As (USFDA, 1993). In this study, the highest
175 concentration (4.43 $\mu\text{g/g}$) of As was found in *S. guttatus* during summer, while the lowest concentration (1.15
176 $\mu\text{g/g}$) was in *L. calcarifer* during winter season (Table S2). The maximum permitted concentration of As set by
177 Australia New Zealand Food Standards Code is 2.0 $\mu\text{g/g}$ on a wet weight basis, i.e., 9.6 $\mu\text{g/g}$ on a dry wet basis
178 assuming 79% of moisture content in fish muscles (ANZFA, 2011). As a criterion for human health protection,
179 USEPA has set arsenic tissue residues of 1.3 $\mu\text{g/g}$ on a wet weight basis, i.e., 6.24 $\mu\text{g/g}$ on a dry wet basis
180 (Burger & Gochfeld, 2005). Thus, the concentrations of As recorded in this study were well below the
181 recommended values by ANZFA and USEPA. In addition, Fig. 3 shows that our values were in good agreement
182 with the As levels found in fishes from the Bay of Bengal (Sharif, Alamgir, Krishnamoorthy, & Mustafa, 1993),
183 the Paira River (Islam, Ahmed, Raknuzzaman, Habibullah-Al-Mamun, & Masunaga, 2015), the Bangshi River
184 (Rahman, Molla, Saha, & Rahman, 2012) and various markets throughout Bangladesh (Ahmed, et al., 2015).

185

186 *3.1.2. Cd*

187 Cadmium is a nonessential element that causes severe toxic effects even at a concentration of $\sim 1 \mu\text{g/g}$. Its
188 accumulation in the human body may lead to the occurrence of renal, pulmonary, hepatic, skeletal, reproductive
189 effects and even cancer (Ahmed, et al., 2015). The highest (0.47 $\mu\text{g/g}$) and lowest (0.02 $\mu\text{g/g}$) concentrations of
190 Cd were found in *P. indicus* and *P. chinensis*, respectively (Table S2). Maximum Cd level permitted by Spanish

191 legislation is 1.0 $\mu\text{g/g}$. Australian National Health and Medical Research Council (ANHMRC) standard for Cd
192 in seafood is 2.0 $\mu\text{g/g}$ while Western Australian Authorities has proposed concentration of 5.5 $\mu\text{g/g}$ for Cd
193 (Plaskett & Potter, 1979). Our Cd results, however, were below the standard limits. The measured mean Cd
194 concentration (0.25 $\mu\text{g/g}$) was similar to the mean values found in fishes from the Bay of Bengal (M. S. Hossain
195 & Khan, 2001), the Paira River (Islam, Ahmed, Raknuzzaman, Habibullah-Al-Mamun, & Masunaga, 2015) and
196 the Bangshi River (Rahman, Molla, Saha, & Rahman, 2012) (Fig. 3). However, mean Cd concentration
197 measured by Sharif, Alamgir, Mustafa, Hossain, and Amin (1993), Sharif, Mustafa, Amin, and Safiullah (1993),
198 and Ahmed, et al. (2015) were lower than our mean value, while Ahmad, Islam, Rahman, Haque, and Islam
199 (2010) found higher mean value relative to us in fishes from the Buriganga River which is highly polluted due to
200 municipal sewage discharge (Fig. 3).

201

202 3.1.3. Co

203 Cobalt is an essential transition element in metabolism of many different organisms. Muscle is the main site
204 of accumulation of Co in fish, and the accumulated concentration is low, perhaps due to its higher affinity with
205 organic matter in the water column (Andreji, Stránai, Massányi, & Valent, 2005). Muscular contents of Co in
206 this study were also low with the highest recorded value of 0.56 $\mu\text{g/g}$ in *P. indicus* and the lowest value (0.21
207 $\mu\text{g/g}$) in *C. reba* (Table S2). Unfortunately, the safe level for Co in fish is not defined.

208

209 3.1.4. Cr

210 The presence of Cr in the diet is essential due to its active involvement in insulin function and lipid metabolism
211 (Ahmed, et al., 2015). According to Western Australian Food and Drug regulations, the recommended
212 maximum permitted value of Cr is 5.5 $\mu\text{g/g}$ (Plaskett & Potter, 1979). In this study, *A. cruciger* accumulated the
213 highest amount (4.66 $\mu\text{g/g}$) of Cr during summer, while *C. reba* accumulated the lowest amount (1.27 $\mu\text{g/g}$)
214 during winter (Table S2). A comparison of Cr accumulation in fishes from various regions of Bangladesh is
215 presented in Fig. 3.

216

217 3.1.5. Cu

218 Copper (Cu) is an essential element due to its presence in several enzymes and its necessity for hemoglobin
219 synthesis (Sivaperumal, Sankar, & Nair, 2007). However, very high levels of Cu can cause acute toxicity.
220 Human deaths have been known to occur from deliberate ingestion of large quantities of copper sulfate.

221 Different regulatory body established a safe limit for Cu concentration. For instance, ANHMRC and FAO
222 prescribed permissible limit of 30 $\mu\text{g/g}$ on a wet weight basis (Bebbington, Mackay, & Chvojka, 1977).
223 According to UK Food Standards Committee Report, Cu levels in food should not exceed 20 $\mu\text{g/g}$ on a wet
224 weight basis (Cronin, Davies, Newton, Pirie, Topping, & Swan, 1998). Australian Food Standard Code
225 established maximum concentration for Cu at 10 $\mu\text{g/g}$ (wet weight basis). There is also legislation in other
226 countries regulating maximum metal concentration. For example, Spanish legislation proposed the level at 20
227 $\mu\text{g/g}$ (wet weight basis). Our recorded maximum value (8.54 $\mu\text{g/g}$) in *I. megaloptera* during summer was lower
228 than the recommended concentrations by various authorities and the variability of Cu concentrations was also
229 similar to other published studies (Fig. 3).

230

231 3.1.6. Mn

232 Manganese is an essential element and its deficiency might lead to severe skeletal and reproductive
233 abnormalities in mammals (Ahmed, et al., 2015). No maximum limit is specified for Mn in fish samples. In this
234 study, the highest (17.80 $\mu\text{g/g}$) Mn content was found in *A. cruciger* during summer and the lowest (3.63 $\mu\text{g/g}$)
235 was in *P. pangasius* during winter (Table S2). Similar ranges of Mn concentrations were reported in fishes from
236 the Bay of Bengal (M. S. Hossain & Khan, 2001; Sharif, Mustafa, Amin, & Safiullah, 1993), and various
237 markets throughout Bangladesh (Ahmed, et al., 2015) (Fig. 3).

238

239 3.1.7. Ni

240 The concentration of Ni in the environment is normally very low, but can cause a variety of adverse health
241 effects including lung inflammation, fibrosis, emphysema, and tumours. The highest concentration (7.56 $\mu\text{g/g}$)
242 of Ni was measured in *C. reba* during summer and the lowest concentration (1.88 $\mu\text{g/g}$) was found in *P. indicus*
243 during winter. Our values were lower than the established safe level of 5.5 $\mu\text{g/g}$ on a wet weight basis (i.e., 26.4
244 $\mu\text{g/g}$ on a dry weight basis assuming 79% moisture content in fish muscles) by Western Australian Food and
245 Drug Regulations (Plaskett & Potter, 1979). Figure 3 shows that Ni concentrations found in this study were
246 similar to the values in literature. However, the mean Ni content found in the fishes from the Buriganga River
247 (Ahmad, Islam, Rahman, Haque, & Islam, 2010) was two-fold higher than this study, while the values reported
248 by Ahmed, et al. (2015) were several fold lower than us.

249

250 3.1.8. Pb

251 Lead is a nonessential trace metal and the adverse health effects from Pb are well known. Among the
252 individual fish species, *A. arius* accumulated the highest Pb concentration of 6.23 µg/g during summer, while *I.*
253 *megaloptera* accumulated the lowest of 0.80 µg/g during winter (Table S2). ANHMRC has set the maximum
254 permitted concentration of Pb as 2.0 µg/g (wet weight basis, i.e., 9.6 µg/g on a dry weight basis assuming 79 %
255 moisture content) (Plaskett & Potter, 1979). According to UK and Spanish legislation, Pb in fish should not
256 exceed 9.6 µg/g on a dry weight basis (Demirak, Yilmaz, Tuna, & Ozdemir, 2006). The present observation
257 showed that Pb level in the all the analyzed fish species was well within the acceptable limit for human
258 consumption.

259

260 3.1.9. Se

261 Selenium is an essential element in some oxidation-reduction processes and is a component of a cellular
262 enzyme called glutathione peroxidase, which is responsible for converting H₂O₂ to CO₂ and H₂O (Sivaperumal,
263 Sankar, & Nair, 2007). Selenium was found to be a significant element in the prevention of Hg toxicity due to a
264 strong binding affinity between Se and Hg that protects direct sequestration of Hg by Se. However, excessive
265 intake of Se can cause chronic toxicity in human, resulting in a condition termed selenosis, characterized by hair
266 and nail loss and brittleness, gastrointestinal problems, skin lesions, tooth decay, garlic breath odor, and nervous
267 system abnormalities (Yang, Wang, Zhou, & Sun, 1983). Clinical symptoms of Se toxicity may also include
268 severe irritations of the respiratory system, lung edema, rhinitis, and broncho-pneumonia. Due to the narrow
269 margin between beneficial and harmful effects of Se (as little as a factor of two), accurate and precise data on
270 selenium intake are needed and thus, the researchers are being interested in the distribution and speciation of Se
271 in fish samples. The details of the mean Se levels detected in individual fish species during four seasons are
272 given in Table S2.

273 Among individual fish species minimum (1.23 µg/g) Se concentration was observed in *I. megaloptera*
274 during winter and maximum (4.00 µg/g) was in *P. pangasius* during summer. The recorded selenium
275 concentrations were significantly lower than FAO/WHO (FAO/WHO, 1984, 2009) and ASTDR (ATSDR,
276 2003) recommended permissible limits of 9.6 µg/g on a dry weight basis, suggesting that exposure to Se was not
277 a problem. The Se concentrations recorded in this study were also on the low end of the published Se values
278 (Fig. 3).

279

280 3.1.10. Zn

281 Zinc is present in all organisms and is an essential trace element for metabolic processes. However,
282 prolonged excessive dietary intake of Zn can lead to the deficiencies in iron and copper, nausea, vomiting, fever,
283 headache, tiredness, and abdominal pain. It is also a human skin irritant. However, in this study, the highest and
284 lowest Zn concentrations were found as 74.36 $\mu\text{g/g}$ in *C. reba* during summer and as 13.22 $\mu\text{g/g}$ in *P. chinensis*
285 during winter, respectively. WHO and ANHMRC (Bebbington, Mackay, & Chvojka, 1977) recommended safe
286 level for Zn is 1000 mg/kg, which is much higher than our determined values, indicating no adverse health
287 effects. The ranges of Zn in different fish species found in this study were within the range of variability
288 reported for various fish species in different parts of Bangladesh (Fig. 3).

289

290 4. Discussion

291 The variation of every metal concentration in ten different fish species was significantly different at 95%
292 confidence interval (Table S3). A large number of studies worldwide have also reported interspecific differences
293 in metal concentrations (e.g., Burger & Gochfeld, 2005; Merciai, Guasch, Kumar, Sabater, & García-Berthou,
294 2014; Rahman, Molla, Saha, & Rahman, 2012). The observed differences were might be due to differences in
295 trophic level (P. Li, et al., 2015), variances in size of fishes (Merciai, Guasch, Kumar, Sabater, & García-
296 Berthou, 2014), differences in seasons (Mendil, Demirci, Tuzen, & Soylak, 2010), and tendency of heavy metals
297 to undergo bioamplification in the food chain (Burger & Gochfeld, 2005). However, this study revealed a
298 similar trend in changes of the metal concentrations in all examined fish species (Fig. 2, inset).

299

300 4.1. Seasonal investigation of heavy metals

301 Figure 2 displays the variation of metal levels in the fish muscles with respect to the seasonal (spring,
302 summer, autumn and winter) fluctuation. The metal concentrations during summer were higher followed by
303 spring, autumn and winter. Similar seasonal effect on metal accumulation was reported in fish species from the
304 Nile River (Zayed, Eldien, & Rabie, 1994) and recently from the Black sea (Mendil, Demirci, Tuzen, & Soylak,
305 2010). This observed summer maximum accumulation may be attributed to heavy rainfall and flood events
306 during summer in Bangladesh, which increases metal loads in water bodies by washing down agricultural waste,
307 sewage water, and sludge. Several other factors may be responsible for driving the metal loads toward higher
308 values during summer. For example, cyclonic activity, which is quite common in Bangladesh at the end of
309 spring and summer, may lead to wind associated resuspension of settled sediments, resulting in a higher amount
310 of metals into the water column. Moreover, delivery of higher nutrients to nearshore marine environments

311 during summer can cause a large phytoplankton bloom. Subsequent bacterial decomposition of this organic
312 matters may release metals into the water which in turn accumulate in fish. In addition, the observed maxima
313 may be related to higher fish respiratory rates during summer when they require increased volume of water due
314 to low dissolved oxygen content, and also to a higher feeding rate on plants and grasses during the summer
315 (Zayed, Eldien, & Rabie, 1994). The relative decrease in metal accumulation during autumn and winter months
316 could be attributed to stable hydraulic condition due to low flows or cease-to-flows of terrestrial runoff into the
317 nearshore coastal environments. Suspended particles (major trace metal carrier) also settle down with clear
318 water conditions during dry winters.

319 The variation of metal concentrations in the edible fish fillets in four different seasons were used for
320 ANOVA to know if the observed variances were significant (Table S3). The data obtained from ANOVA
321 clearly demonstrated that seasonal effects on the variability of Cd, Cr, Pb, and Se were significant at 95 %
322 confidence interval while rest of the metals variability were not statistically significant. Moreover, to quantify
323 the seasonal effects on the metals variability found in the studied fish species, three dummy variables were taken
324 for the seasons (winter as reference) when the fish samples were collected (Table S4). The coefficients of
325 dummy variables revealed that on average As concentrations in summer were $\sim 0.41 \mu\text{g/g}$ higher than those in
326 winter while its concentrations in spring and autumn were ~ 0.24 and $\sim 0.11 \mu\text{g/g}$ higher relative to winter.
327 Likewise, Cd accumulation during summer, spring, and autumn were ~ 0.11 , ~ 0.05 , and $\sim 0.03 \mu\text{g/g}$,
328 respectively, higher relative to winter. See Table S4 for rest of the metals.

329

330 *4.2. Risk to humans by consumption of sea-fish*

331 Even though the concentrations of ten analyzed metals in the tested fish samples did not exceed the legal
332 limits set by various agencies (as we discussed in the Result section), the toxic potency of these analytes
333 depends on exposure doses. Thus, combining the determined concentration of contaminant and the estimated
334 daily fish consumption limit is considered as a great tool for evaluating the balance between benefits and risks
335 (carcinogenic and non-carcinogenic), and applied in this study.

336 As a first step of estimation of human health risk, daily intake of every metal through ingestion of the fish
337 was estimated using the formula (1). The highest average value of EDI was observed for Zn (7.73×10^{-3}
338 mg/person/day) while the lowest was for As (1.39×10^{-5} mg/person/day). The EDI results were compared with
339 the respective recommended daily dietary allowance of individual metal suggested by various agencies
340 including Joint FAO/WHO Expert Committee on Food Additive (JECFA) for As, Cd, Cu, Pb, and Zn (JECFA,

341 1982, 1989, 2000), World Health Organization (WHO) for Co and Ni (WHO, 1996, 2006), National Research
342 Council (NRC) for Cr and Mn (NRC, 1989), National Academy of Science (NAS) for Se (ATSDR, 2003) (Fig.
343 4 and also see Table S5). Figure 4, however, displays that the EDI values for all the metals were far below the
344 respective recommended levels, indicating that adverse health effects from exposure to the examined metals
345 were unlikely.

346 Figure 5 represents the THQ values of the selected heavy metals due to consumption of the examined ten
347 fish species by average Bangladeshi adults. The highest average value of THQ was observed for Co (4.18×10^{-2})
348 while the lowest was for Cr (3.39×10^{-4}). None of the metals individually exceeded the hazard quotient threshold
349 of 1. This implies that the level of daily intake of each examined metal by the Bangladeshi adults was lower than
350 that of respective reference dose. Thus, we can suggest that these level of human exposure to the analyzed
351 metals should not cause any deleterious effect during an entire lifetime. Considering that consumption of the
352 fishes involved exposure to a mixture of ten examined metals, we evaluated cumulative health risks by summing
353 the health risks posed by ten studied metals (i.e., total target hazard quotient, TTHQ). The value of TTHQ (i.e.,
354 6.39×10^{-1}) was also less than 1, again indicating no potential significant health risks. In this study, Co and Pb
355 were the two major contributors to TTHQ, accounted for 33.32 % and 23.24 % respectively. Whereas, Cr
356 contributed the least (only 0.05%) to TTHQ.

357 US Food and Drug Administration (USFDA, 1993) has pointed out that dietary intake of inorganic As is
358 strongly associated with wide spectrum carcinogenesis. The CR values of As obtained in this study were ranged
359 from 1.28×10^{-5} to 2.76×10^{-5} (with a mean value of 2.09×10^{-5}) and exceeded the risk of cancer benchmark
360 (i.e., 10^{-5}) (Fig. 5, inset). This implies that the ingesting of fish from the Bay of Bengal would cause a cancer
361 risk for lifetime consumption. However, in this study, we assumed that ingestion dose of heavy metals is equal
362 to the absorbed dose of contaminants, without considering bioavailability of the metals in human body. This
363 assumption might overestimate the As-carcinogenic risk calculation due to the fact that a part of ingested As
364 may be excreted. Thus for more robust estimation of health risks, future study should focus on biomonitoring of
365 metals in blood, urine and hair samples to determine the actual concentrations of bioavailable metals in the local
366 population.

367

368 4.3. Risk management

369 Risk management is the process of systematic identification of environmental hazards, analyzing the
370 likelihood of occurrence and severity of the potential consequences, weighing the policy alternatives and, if

371 required, selecting and implementing appropriate control options, including regulatory measures and subsequent
372 enforcement. In the least developed countries like Bangladesh, anthropogenic pollution of heavy metals in
373 marine environment is the prime concern. A periodic monitoring program and risk assessment approach are
374 necessary to evaluate the scale of anthropogenic impacts on ecological balance and human health. In this
375 communication, we provided information on As, Cd, Co, Cr, Cu, Mn, Ni, Pb, Se, and Zn levels in commonly
376 consumed fish species from the Bay of Bengal as a function of four seasons. Such local data are required to
377 increase the global data inventory and to inform management agencies to facilitate decision making and improve
378 the management strategies of these metals. Though the present study indicates the low concentration of
379 aforementioned elements in the fish species, we should monitor the metal contamination periodically so that the
380 anthropogenic activities can't disrupt the balance of aquatic ecosystem by polluting the marine organisms and to
381 ensure food safety. The estimation of health risks associated with the dietary intake of heavy metals by the
382 consumers is a vital and integral part of regulatory processes. This study estimated the human health risks
383 associated with ten heavy metals due to consumption of ten fish species, and hence further study is
384 recommended taking into account other contaminants, for example, Hg, in a broader range of fish species from
385 the Bay of Bengal. This study deals with fish consumption which constitutes only 3% of per capita per day
386 calorie intake by foodstuffs for the coastal population in Bangladesh. Thus, other foodstuffs such as vegetable,
387 fruits, rice, and cereals are required to be estimated to determine the public health risk via consumption of
388 foodstuffs more robustly and to provide more information to allow the public to maximize the health benefits
389 from food consumption while to minimize the risks from dietary intake of contaminants.

390

391 **5. Conclusions**

392 Accumulation of heavy metals in fish muscles is a threat to global public health since the consumption of
393 contaminated foods is considered as the primary route of human exposure to toxic heavy metals. This study
394 showed that variations in concentrations of As, Cd, Co, Cr, Cu, Mn, Ni, Pb, Se, and Zn in the fish species from
395 the Bay of Bengal were dependent on the seasonal fluctuation. The fishes captured during summer months
396 exhibited higher metal contents followed by spring, autumn, and winter. This summer maxima can be attributed
397 to heavy rainfall and floods during wet summer months in Bangladesh. The concentration levels of all the
398 examined metals remained well below the acceptable limits for human consumption established by various
399 international authorities such as USEPA, ANHMRC, and WHO. From the human health point of view, the
400 estimated daily intake (EDI) of each element was lower than the respective recommended daily dietary

401 allowance. The THQ values for individual element and the TTHQ values for combined elements were lower
402 than 1, indicating no health risks for consumers due to the intake of either individual element or combined
403 elements. However, the risk of cancer due to exposure to arsenic was of concern since the values of
404 carcinogenic risk (CR) were above the acceptable risk limit of 10^{-5} .

405

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411

412 **Conflict of interest**

413 The authors declare that there are no conflicts of interest.

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517

518

Figure captions

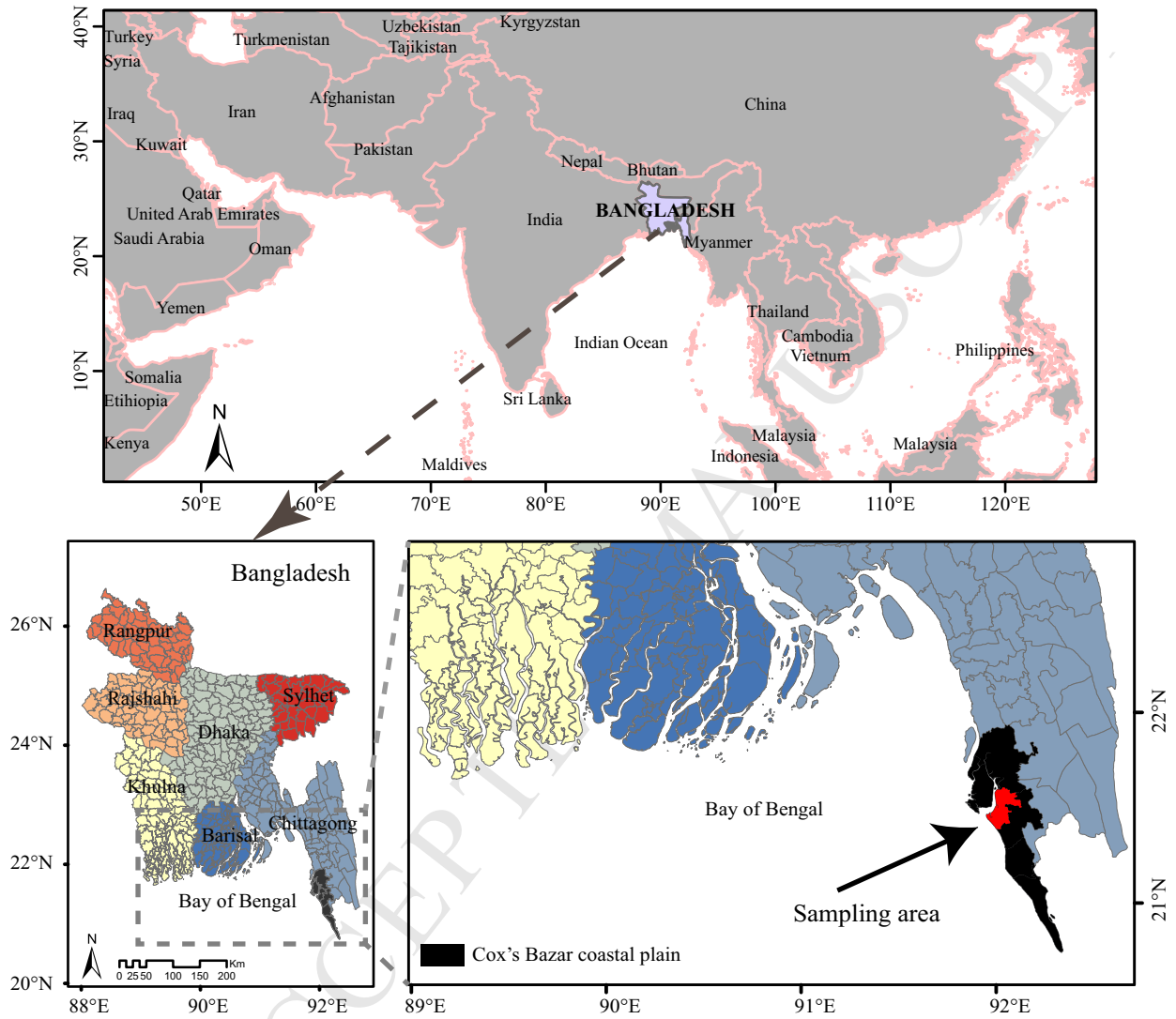
Fig. 1. Photographs of the sampling area, Cox's Bazar, Chittagong, Bangladesh.

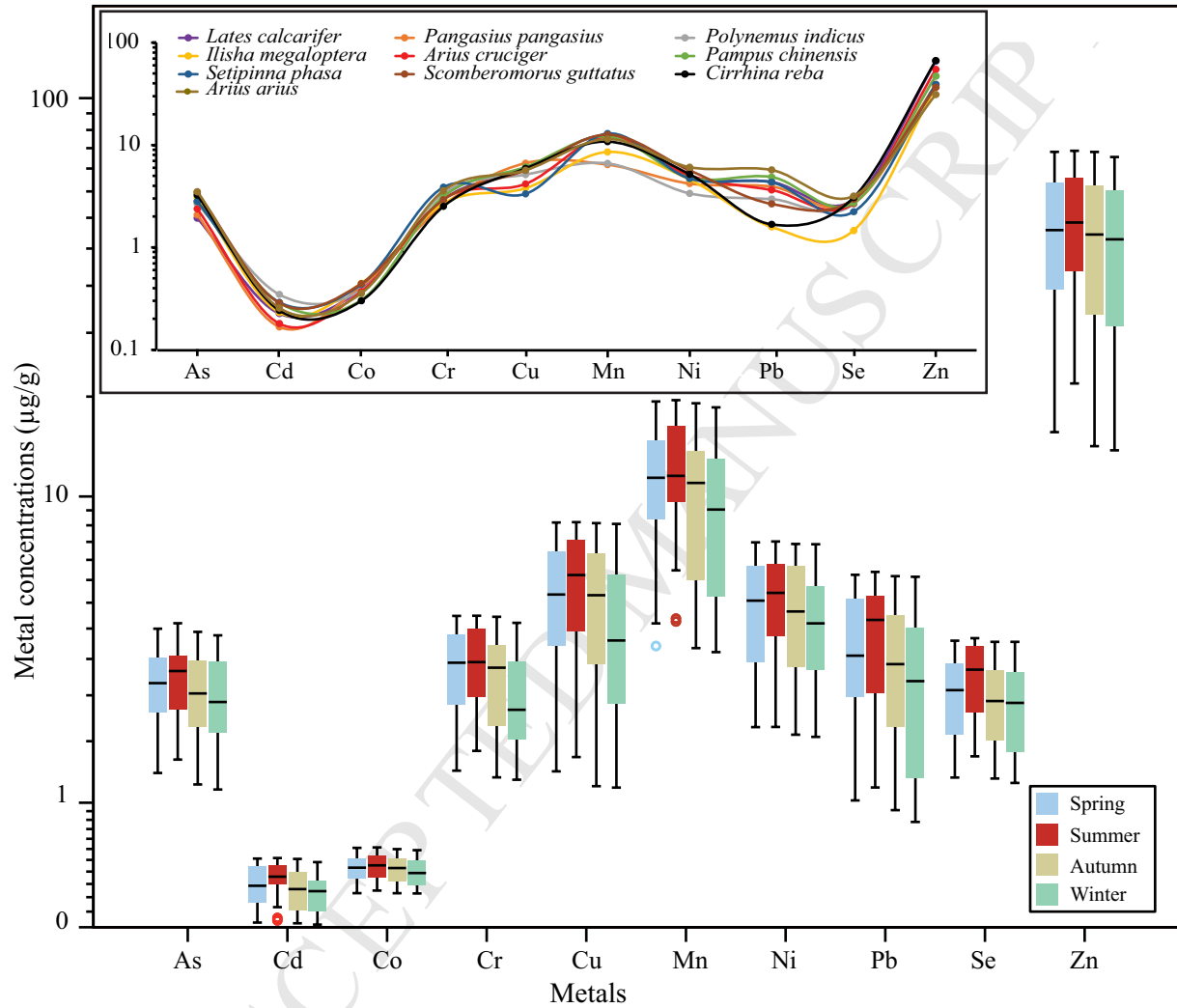
Fig. 2. Box plots show the seasonal variation of metals in the examined fish species. The inset displays the trend in changes of the mean metal concentrations in ten different fish species.

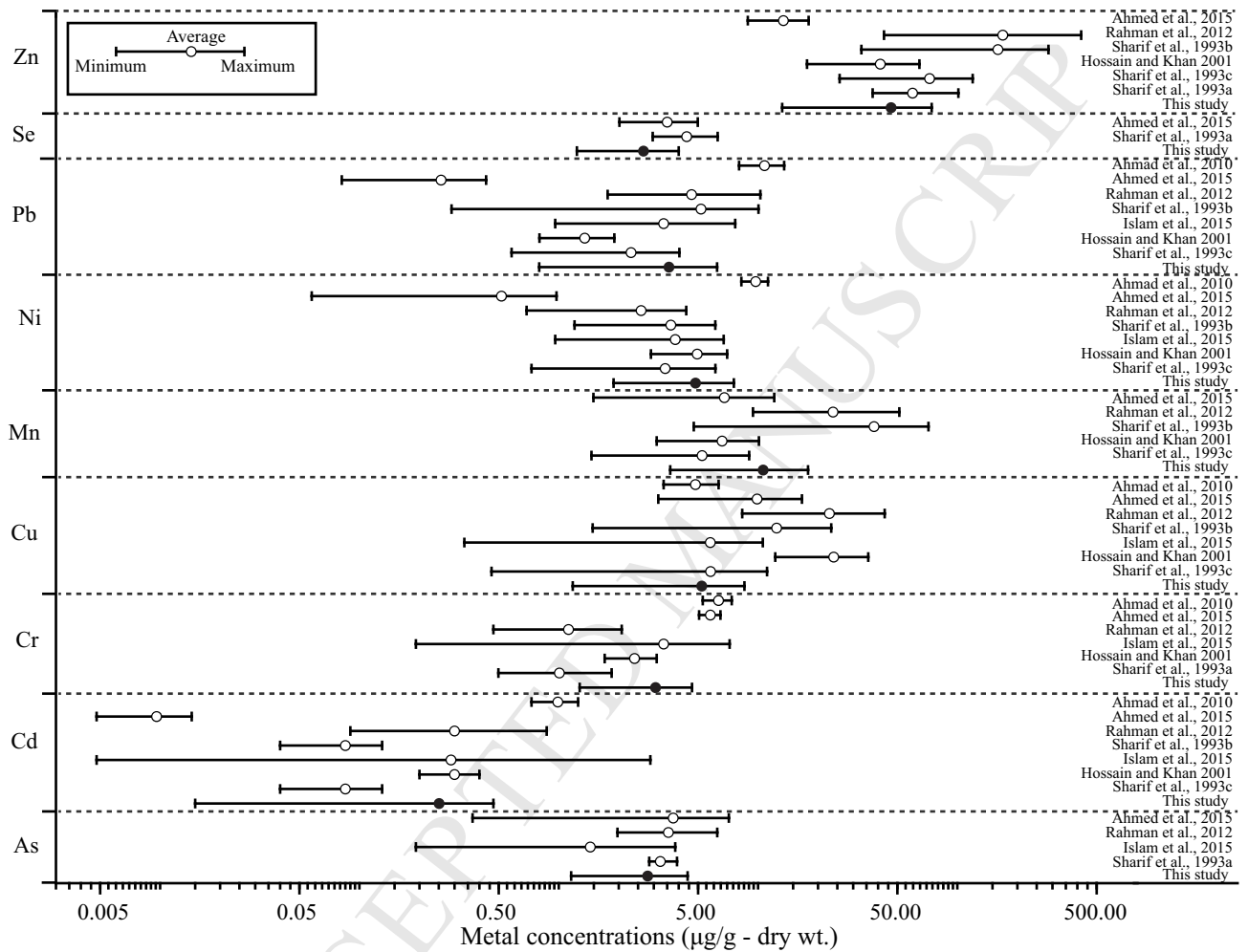
Fig. 3. Comparison of fish metal concentrations ($\mu\text{g/g}$ - dry weight basis) found in this study with published data from the Bay of Bengal (Hossain & Khan, 2001; Sharif, Alamgir, Krishnamoorthy, & Mustafa, 1993 (in Fig. Sharif et al. 1993a)); Sharif, Mustafa, Amin, & Safiullah, 1993 (in Fig. Sharif et al. 1993c)); the Paira River (Islam, Ahmed, Raknuzzaman, Habibullah-Al-Mamun, & Masunaga, 2015); Soari Ghat, the Gumti and Turag Rivers (Sharif, Alamgir, Mustafa, Hossain, & Amin, 1993 (in Fig. Sharif et al. 1993b)); the Bangshi River (Rahman, Molla, Saha, & Rahman, 2012); various markets (Ahmed, Shaheen, Islam, Habibullah-al-Mamun, Islam, Mohiduzzaman, et al., 2015); and the Buriganga River (Ahmad, Islam, Rahman, Haque, & Islam, 2010) of Bangladesh. The concentrations of metals in fishes from the Paira River (Islam, Ahmed, Raknuzzaman, Habibullah-Al-Mamun, & Masunaga, 2015) and various markets (Ahmed, et al., 2015) were wet weight basis and for comparison the values were transformed to dry wet assuming 79% of moisture content.

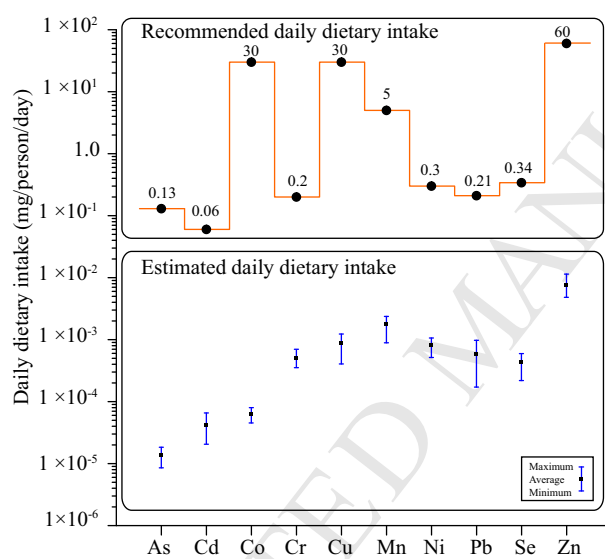
Fig. 4. Estimated daily dietary intake of the metals and their comparison with the respective recommended daily dietary intake. For more information see the Table S5 provided in supplementary materials.

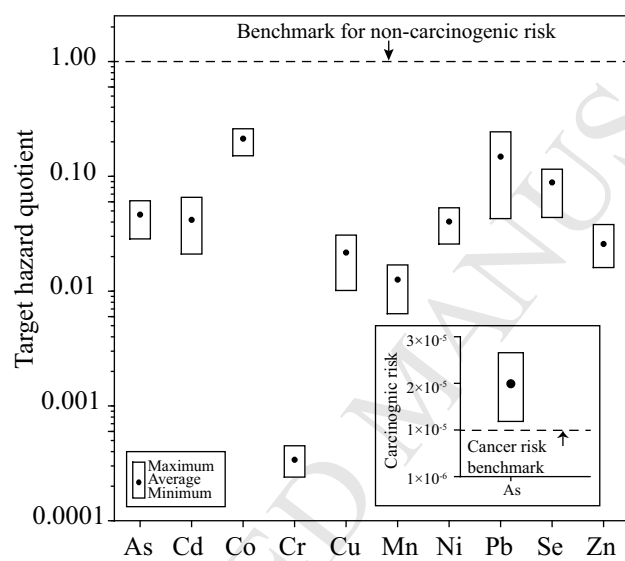
Fig. 5. Non-carcinogenic risks due to intake of ten examined metals. The inset shows the carcinogenic risk of As. For more information see the Table S5 provided in supplementary materials.











Highlights

- Ten metals in 10 fish species from the Bay of Bengal were estimated in 4 seasons.
- Significant interspecies variations were observed for all metals.
- Metal accumulation was higher during wet summer season.
- Consumption of fish was safe for human health.
- Consumers were exposed to arsenic with carcinogenic risks.