Avoiding impacts on biodiversity through strengthening the first stage of the mitigation hierarchy

- 3 4
- Ben Phalan^{1,2*}, Genevieve Hayes³, Sharon Brooks⁴, David Marsh⁵, Pippa Howard⁵,
- 5 Brendan Costelloe⁶, Bhaskar Vira⁷, Aida Kowalska³, Samir Whitaker³
- 6
- 7 ¹ Conservation Science Group, Department of Zoology, University of Cambridge,
- 8 Downing Street, Cambridge CB2 3EJ, UK
- 9 ² Department of Forest Ecosystems and Society, Oregon State University, Corvallis,
- 10 Oregon 97331, USA
- 11 ³ BirdLife International, Wellbrook Court, Girton Road, Cambridge CB3 ONA, UK
- 12 ⁴ UNEP–WCMC, 219 Huntingdon Road, Cambridge CB3 0DL, UK
- 13 ⁵ Fauna & Flora International, Jupiter House, Station Road, Cambridge CB1 2JD, UK
- ⁶ Royal Society for the Protection of Birds, The Lodge, Sandy, Bedfordshire SG19
- 15 *2DL, UK*
- 16 ⁷ Department of Geography, University of Cambridge, Downing Street, Cambridge
- 17 *CB2 3EN, UK*
- 18 *Email: benphalan@gmail.com
- 19
- 20

21 Abstract

- 22
- 23 The mitigation hierarchy is a decision-making framework designed to address
- 24 impacts on biodiversity and ecosystem services through first seeking to avoid
- 25 impacts wherever possible, then minimising or restoring impacts, and finally by
- offsetting any unavoidable impacts. Avoiding impacts is seen by many as the
- 27 most certain and effective way of managing harm to biodiversity, and its position
- as the first stage of the mitigation hierarchy indicates that it should be
- 29 prioritised ahead of other stages. However, despite an abundance of legislative
- and voluntary requirements, there is often a failure to avoid impacts. We discuss
- reasons for this failure and outline some possible solutions. We highlight the key
- 32 roles that can be played by conservation organisations in cultivating political
- 33 will, holding decision-makers accountable to the law, improving the processes of
- 34 impact assessment and avoidance, building capacity, and providing technical
- 35 knowledge. A renewed focus on impact avoidance as the foundation of the
- 36 mitigation hierarchy could help to limit the impacts on biodiversity of large-scale
- developments in energy, infrastructure, agriculture and other sectors.
- 38 39
- 40 Words: 4846
- 41
- 42
- 43

44 Introduction

45

46 The development of mines, infrastructure, buildings and plantations changes

47 landscapes and seascapes profoundly, putting pressure on biodiversity and often

48 reducing the provision of important ecosystem services. The mitigation

- 49 hierarchy is a decision-making framework developed with the aim of preventing
- and remediating environmental impacts from such developments. It requires
- 51 developers to first avoid impacts where possible. Where impact avoidance is not
- 52 possible, developers should strive to minimise impacts, to restore affected areas,

and finally, to offset any remaining residual impacts (McKenney & Kiesecker,

- 54 2010; Clare et al., 2011; BBOP, 2012a; Gardner et al., 2013).
- 55

56 The first part of the mitigation hierarchy – impact avoidance – requires

- 57 developers to "anticipate and prevent adverse impacts on biodiversity before
- 58 actions or decisions are taken that could lead to such impacts" (Ekstrom et al.,
- 59 2015). Impact avoidance is typically identified as the most important stage of the
- 60 mitigation hierarchy (McKenney & Kiesecker, 2010; Clare et al., 2011; Ekstrom et
- al., 2015). In principle, impact avoidance can reduce the need for remediation,
- 62 thus side-stepping problems such as restoration time lags, limits to what can be
- 63 offset, and negative social implications of removing people's access to nature
- 64 locally and attempting to replace it elsewhere (Bendor, 2009; Maron et al., 2010;
- 65 Pilgrim, Brownlie, Ekstrom, Gardner, von Hase, ten Kate, Savy, Stephens, Temple,
- 66 Treweek, Ussher, et al., 2013; Ives & Bekessy, 2015).
- 67

68 In practice, there are concerns that impact avoidance is often ignored,

- 69 misunderstood and poorly applied by developers, impact assessment
- practitioners and regulators (Clare et al., 2011; Villarroya et al., 2014). At the
- same time, there are signs of offsetting being situated in policy as a means to
- 72 legitimise development which would not otherwise have been permitted
- 73 (Walker et al., 2009; Ferreira et al., 2014; Sullivan & Hannis, 2015). Situating
- offsets as a "license to trash" runs counter to the principle of avoiding negative
- impacts to the maximum extent practicable, a core tenet of the mitigation
- hierarchy (US EPA and DA, 1990; McKenney & Kiesecker, 2010; BBOP, 2012a).
- 77

78 We review the incentives and requirements for impact avoidance, identify

challenges for achieving it, and outline some possible solutions, with an

80 emphasis on the ways in which conservation organisations can advocate for

better impact avoidance. The issues discussed here, and a wider range of case
studies, are detailed further by BirdLife International et al. (2015) and Ekstrom

- et al. (2015).
- 84

85 Current incentives, requirements and criteria for impact avoidance 86

- 87 National laws, voluntary sustainability standards, corporate commitments and
- 88 pressure from civil society organisations all have roles to play in creating
- 89 incentives and requirements for impact avoidance. Most countries require
- 90 impact avoidance to be considered as part of the Environmental and Social
- 91 Impact Assessment (ESIA) process (Pope et al., 2013). Sustainability standards
- 92 include those set by financial institutions, such as Performance Standard 6 of the

93 International Finance Corporation, as well as sector-specific standards such as

- 94 those of the Roundtable on Sustainable Palm Oil (RSPO). Companies are
- 95 increasingly adopting commitments to No Net Loss (NNL) or Net Positive Impact
- 96 (NPI), which seek to ensure that negative impacts on biodiversity and ecosystem
- 97 services are balanced (for NNL) or outweighed (for NPI) by impact avoidance,
- 98 minimisation, restoration and offsetting (Gardner et al., 2013; Rainey et al.,
- 99 2015).
- 100

101 In Table 1, we summarise some of the key components (actions and criteria) of

102 standards and laws requiring avoidance, for a set of illustrative instruments.

- 103 Commonly-required actions include consultation, impact assessment,
- 104 consideration of cumulative impacts, and monitoring. Consulting with local
- 105 communities and conservation organisations is an important step in identifying
- 106 impacts that might be considered serious or unacceptable, and which might
- 107 otherwise go un-assessed. Conducting an ESIA is now standard practice in most
- 108 countries for large developments, but it is less common to consider the
- 109 cumulative impacts of multiple developments, including offsite, cryptic and
- secondary impacts (Pope et al., 2013; Raiter et al., 2014). Transparent long-term

111 monitoring and evaluation is essential for demonstrating that commitments to

- 112 avoid and remediate impacts have been successfully upheld.
- 113

Our examples illustrate four recurrent criteria for moving past the avoidance
stage of the mitigation hierarchy. The first is that alternatives are given full
consideration by regulators and developers, both before and during the ESIA
(Table 1). These can include a "no-project" alternative, alternative site locations
(spatial impact avoidance), alternative scheduling of activities (temporal impact
avoidance), and use of technology and planning within a site (design-based
impact avoidance) (Table 2).

121

122 Early consideration of alternatives is advisable, alongside early engagement with 123 a full range of stakeholders to identify appropriate alternatives (Ekstrom et al., 124 2015). The cost of altering a project is lower early in the planning process, before 125 decisions about locations and technologies are locked in and the range of feasible 126 alternatives is narrowed. Planning tools can facilitate access to data on existing 127 conservation designations and thus help to screen alternative sites, such as the 128 Integrated Biodiversity Assessment Tool (http://www.ibatforbusiness.org) and 129 national land-use planning and protected area databases. 130 131 A second common criterion for proceeding past the avoidance stage is that the 132 societal benefits of a project should outweigh its environmental costs. The European Habitats Directive provides some of the clearest guidance here. 133

- 134 Impacts on priority habitats and species are only permitted for reasons of human
- 135 health, public safety, environmental benefit, or if there are "imperative reasons
- 136 of overriding public interest" for the project to proceed (Council of the European
- 137 Commission, 1992; European Commission, 2007). Even here, however, defining
- 138 when societal benefits are sufficient to justify environmental harm is subjective
- and often highly political, a challenge explored further in the next section.
- 140

141 A third common criterion for moving past avoidance is delivery of NNL or a net

142 gain in biodiversity (BBOP, 2012a; Gardner et al., 2013). NNL is typically only

143 assessed for priority biodiversity features such as endangered species, Critical

144 Habitat and areas of High Conservation Value, even though it might be most

- successfully applied to common species and ecosystems (Pilgrim, Brownlie,
- 146 Ekstrom, Gardner, von Hase, ten Kate, Savy, Stephens, Temple, Treweek, &
- 147 Ussher, 2013). There are partial exceptions: the UK National Planning Policy
- Framework, for example, covers wider biodiversity in addition to designated

sites (Department for Communities and Local Government, 2012). However, the
Framework's reliance on the concept of "significant harm", means in practice

- 151 that small, cumulative impacts to common species are likely to be ignored. NNL
- 152 obligations can often be met through a promise of remediation as well as
- 153 through impact avoidance and minimisation (BBOP, 2012b).
- 154

155 Fourth, legal requirements can help to define opportunities for impact

avoidance, such as through identifying protected sites and species. Laws also set

157 requirements for planning and define how ESIA should be carried out, and where

158 they are contravened, they provide the basis for conservation organisations to

159 challenge failures to avoid impacts in the courts.

160

161 **Challenges for effective impact avoidance**

162

163 We surveyed the literature and drew on our experiences to identify challenges for effective impact avoidance in five broad and often overlapping categories: 164 165 political will, regulation, process, capacity, and technical knowledge (Table 3). 166 Political will refers to the perceived importance among decision-makers of 167 avoiding impacts on biodiversity, relative to other concerns. In the absence of political will, laws are less likely to be enforced, expensive alternatives are more 168 169 likely to be ruled "infeasible", and legal protection is at risk of being weakened or 170 corrupted to cater for powerful private interests. Mascia et al. (2014) found over 171 500 cases of downgrading, downsizing, or degazettement of protected areas in 172 57 countries, most commonly to facilitate industrial-scale resource extraction 173 and development. For companies in the public eve, reputational risk is an 174 incentive to develop the political will to ensure that impacts are reported and 175 avoided (Dawkins & Fraas, 2010), but for smaller companies, and those without 176 shareholders for whom environmental issues are important, it may not be. 177 178 The effectiveness of regulation depends both on the quality of legislation, and its 179 implementation in practice. In Indonesia, there is often a mismatch between 180 official maps and the physical reality of landcover, and thus it can be easier for 181 oil palm companies to obtain concessions in primary forest which is classified as 182 "nonforest estate" on official maps, rather than in the millions of hectares of 183 degraded land which - because it is mapped as "forest estate" - is legally 184 unavailable for development (Rosenbarger et al., 2013). This is a failure of the 185 legal system for land classification, arising perhaps from a lack of political will to 186 protect primary forests. A review of 11 European projects affecting sites protected under the Habitats Directive found consistent failures by the European 187 188 Commission in the interpretation and application of the Directive (Kramer, 189 2009). In only three of these cases were alternative locations assessed, as

190 required by the Directive, and in few if any cases was a robust argument 191 established for the project being of "overriding public interest" (for further details see Kramer, 2009). In these cases, although the law seems to provide 192 193 strong protection in principle, it appears to have been undermined by 194 interpretations that privileged economic development and marginalised 195 environmental protection. In the Democratic Republic of Congo, mining rights have been granted within protected areas, even though such areas are legally 196 197 protected from extractive activities (Javelle & Veit, 2012). This results from 198 contradictory regulations, inconsistent and outdated government information, a 199 lack of cooperation between the two relevant ministries, and ultimately. opposing interests. 200 201 202 Even while acting within the law, regulators and companies have choices about 203 how to pursue the process of identifying and acting on opportunities to avoid

204 impacts. For example, if impact avoidance is not considered until the ESIA.

205 opportunities to fundamentally rethink project alternatives may no longer be

- available. There are incentives, for those who would benefit from a project, to
- 207 ensure that "no-project" alternatives are not considered, and to overlook indirect
- and cumulative impacts. Because ESIA and standards are typically applied at a
 project rather than landscape scale, they are not ideally suited to identifying
- 210 strategic opportunities for impact avoidance. Within standards, requirements
- 211 vary, and there is scope for criteria such as those for identifying High
- 212 Conservation Values to be interpreted differently by different assessors (Senior
- 213 et al., 2015). Responsibility for ensuring permanent protection of avoided areas
- 214 may be unclear: areas avoided during an early phase may be demanded in later
- stages of a project, and areas avoided by one company may not be avoided by
- 216 others. For example, Sakhalin Energy's avoidance plans (Table 2) were
- 217 undermined by Exxon's plans to conduct seismic surveys in the same area
- 218 (Western Grey Whale Advisory Panel, 2015).
- 219

220 Government departments, companies and civil society organisations all too often lack sufficient capacity and resources to properly understand, develop and 221 222 implement sound environmental policies (Ouétier et al., 2014). National and 223 local governments often lack (or do not allocate) sufficient resources to audit 224 compliance with legislation. Small and medium-sized companies may be unable 225 to afford in-house expertise on biodiversity and the mitigation hierarchy. 226 Effective impact avoidance may require upfront investments in assessment and 227 planning at a time when there may be uncertainty about whether a project will 228 proceed. Even large companies may be unwilling to incur these costs. The 229 influence of civil society organisations may be limited if they are poorly-

- resourced and have limited expertise, as is common with local groups.
- 231

Although knowledge is increasing, there are still many gaps in technical

- 233 understanding of the spatial distributions and population status of species,
- especially for plants and invertebrates (Pimm et al., 2014). This makes it difficult
- to identify and prioritise the sites of most importance in advance of
- 236 developments. Information on the success and costs of restoration and offsetting
- efforts is also sparse. Unrealistic assumptions about the capacity and cost of
- restoration and offsetting could result in promises of remediation being a more

attractive option for companies than avoiding impacts early in the project cycle.

A further challenge is that there may be trade-offs between impact avoidance

and other conservation strategies. For example, one way to avoid expanding into

242 natural habitats is to consolidate timber production, farming and infrastructure

243 in existing zones or corridors of disturbance. There is widespread evidence that

244 'sparing' land in this way would be beneficial for many wild species, even if it

- increases the per-hectare impacts of development (Edwards et al., 2014;
- Balmford et al., 2015; Stott et al., 2015). However, some degree of 'sharing' land
- 247 with multiple uses to reduce the per-hectare impacts of land uses is also
- desirable, and clear guidance on how to balance these strategies is not availablein most places.
- 250

251 **Confrontation or compromise?**

252

253 Conservation organisations have strategic choices to make about when to

- collaborate with developers, or oppose them. Campaigning against harmful
- developments is important, and this strategy of 'saying no' can help to

counterbalance the ambit claims (extreme initial demands) of developers

257 (Laurance, 2016). Opposition is thus an important conservation strategy.

258

259 Conservation organisations can also play a role in identifying where

260 developments *should* take place. A key problem with focusing exclusively on

- 261 impact avoidance is the risk of leakage, or displacement of impacts. Development
- avoided at one location is likely to take place elsewhere. Even project
- 263 cancellation is no guarantee that impacts have been avoided, unless regulation
- also constrains the demand drivers incentivising development. Insofar as

development is driven by market demands for minerals, energy and other

266 resources, those demands will continue to incentivise further development

267 (Meyfroidt et al., 2013). Thus, it may be as important for conservationists to get268 involved in defining where development is acceptable, as where it is not (Venter

- et al., 2013: Dinerstein et al., 2015). Mapping priority areas for avoidance often
- also by implication identifies locations for development which may be more
- appropriate (Bright et al., 2008; Martin et al., 2015).
- 272

There are risks involved in both confrontational and collaborative strategies. As
conservation interests are typically less powerful than development interests,
opposition might leave conservation organisations marginalised and without
input into decisions. On the other hand, too much willingness to compromise

277 could result in 'greenwash', conferring an image of environmental responsibility

on companies in exchange for minimal concessions or donations (Robinson,
2012). The best strategy will be context-dependent, and in some situations a

279 2012). The best strategy will be context-dependent, and in some situations a280 combination of principled opposition and pragmatism might have most success.

280 Combination of principled opposition and pragmatism might have most succe 281 For example, conservation organisations engaged with local and national

authorities to develop standards for new housing developments near the

282 authorities to develop standards for new nousing developments near the 283 Thames Basin Heaths in the UK. Disturbance-sensitive bird species were

284 protected by avoiding construction within a buffer zone, while developers were

required to provide dog-walking areas alongside new housing nearby to

285 required to provide dog-walking areas alongside new nousing nearby to

- 286 minimise additional disturbance (BirdLife International, 2015).
- 287

288 **Opportunities for more effective impact avoidance**

289

290 There are multiple ways in which conservation organisations can work towards, 291 and support, more effective impact avoidance (Table 3). They can harness and 292 broadcast public support for conservation, thus creating political space for 293 decision-makers who want to support impact avoidance (Downie, 2015) and increasing the reputational cost on those who ignore it (Ivanova, 2015). When 294 295 they can establish or promote particular ways of thinking about issues ('frames'), 296 they influence the extent to which ecological values are considered in policy 297 development (Sullivan & Hannis, 2015). Conservation organisations could also play a greater role in challenging corrupt or undemocratic institutions that give 298 299 private or political interests excessive influence (Greenwald et al., 2012). The 300 success of high-profile campaigns in persuading companies to adopt 301 sustainability standards and zero-deforestation commitments shows how 302 corporate policies can be influenced (Newton et al., 2013; Gibbs et al., 2016). 303 304 Regulation can be improved, both in the letter of the law and – perhaps more often – in its application. Conservation organisations can advocate for more 305 306 stringent requirements for impact avoidance and clearly-defined legal protection 307 for sites and species. They can also campaign against subsidies and other 308 perverse incentives for development in areas of biodiversity importance, such as 309 the proposal by Brazil to open up to 10% of its strictly-protected areas to mining (Ferreira et al., 2014). Conservation organisations could also campaign to extend 310 stronger protection to common species and habitats, which are sometimes 311 312 overlooked in legislation (Gaston, 2010). Working informally with state agencies 313 and developers might be fruitful: Malcolm and Li (2015) suggest that informal 314 dialogue in advance of submitting project proposals may have helped to reduce 315 the number of proposals submitted in the United States that would put 316 endangered species in jeopardy.

317

318 There are opportunities for conservation organisations to hold governments and 319 companies accountable to the processes and plans they have signed up to. They 320 can track compliance with legislation and standards, especially in jurisdictions 321 where there is limited capacity for public authorities to do so. They can also push for voluntary actions that make success more likely, such as inclusion of indirect 322 323 and other enigmatic impacts in ESIA and genuine consideration of "no-project" 324 options. Perhaps the most important demand that they can make is for decision-325 makers to consider impact avoidance from the earliest stages of the planning 326 process (Ekstrom et al., 2015). Had this been done in the case of the Via Baltica, 327 referenced in Table 2, for example, the case might not have gone to the European Court of Justice and considerable legal costs and delays could have been avoided. 328 329 330 Building the capacity of individuals and institutions plays a pivotal role in the 331 success of conservation efforts (Brooks et al., 2012). Conservation organisations 332 can support the development of biodiversity-inclusive landscape and regional-333 level zoning plans, as well as better policy guidance material to support the implementation and enforcement of legislation on impact avoidance. They can 334

335 contribute to developing voluntary or regulatory mechanisms to ensure avoided

336 areas receive long-term protection – such as developing new legal mechanisms

337 for permanently retiring grazing leases on public land (Leshy & McUsic, 2009).

338 They have also played an important role in the development of voluntary

standards, and will continue to do so. For example, Greenpeace played a key part 339

- 340 in developing a methodology for identifying areas with high carbon stocks that
- 341 should be avoided, in order to address a key gap in the RSPO standard for oil
- 342 palm (Dinerstein et al., 2015).
- 343

344 Conservation organisations can provide tools and technical data to make it easier 345 to conduct cumulative, strategic and environmental impact assessments.

Examples include the Biodiversity Risks and Opportunities Assessment tool, and 346

the Migratory Soaring Bird Sensitivity Map (BirdLife International, 2015). 347

Conservation organisations can work with other civil society organisations to 348

- 349 understand synergies and trade-offs between multiple objectives, and find
- 350 common ground in advocating for impact avoidance. This may require
- understanding complex interactions such as those between conservation and 351
- 352 other human interests. It may also involve identifying places where
- 353 developments might be acceptable, as well as where they should be avoided. A
- final, crucial role is in better evaluation and communication of the success (or 354
- 355 otherwise) of efforts to avoid impacts (Baylis et al., 2016).
- 356

357 Conclusion

358

359 We suggest that a renewed focus on the first stage of the mitigation hierarchy 360 could help to limit the biodiversity impacts of large-scale developments in 361 energy, infrastructure, agriculture and other sectors. Conservation organisations 362 play an important role in cultivating political will, holding decision-makers 363 accountable to the law, improving the processes of impact assessment and 364 avoidance, building capacity, and providing technical knowledge. Ensuring that

365

impact avoidance is considered as early as possible in the planning process, and that it is placed even more firmly at the heart of the mitigation hierarchy in both 366

367 policy and practice should be key demands on their agenda.

368

369 Acknowledgments

370

371 This paper is a product of a project on "Strengthening the mitigation hierarchy 372 for greater conservation gains", funded by the Cambridge Conservation Initiative 373 Collaborative Fund. We thank all of the participants in that project, and we are 374 grateful to two anonymous reviewers whose suggestions greatly improved the paper. BP was funded by a Zukerman Fellowship at King's College, Cambridge.

- 375
- 376

377 References

- 378
- 379 AURECON & SLR (2015) Social and Environmental Impact Assessment for the 380 proposed Rössing Uranium Desalination Plant, near Swakopmund, 381 Namibia. Final SEIA Report, Report No. 9408/110914. Rio Tinto Rössing 382 Uranium Limited.
- BALMFORD, A., GREEN, R. & PHALAN, B. (2015) Land for food and land for nature? 383 384 Daedalus, 144, 57-75.

385 BAYLIS, K., HONEY-ROSÉS, J., BÖRNER, J., CORBERA, E., EZZINE-DE-BLAS, D., FERRARO, P., ET 386 AL. (2016) Mainstreaming impact evaluation in nature conservation. 387 Conservation Letters, 9, 58-64. BBOP (2012a) Standard on Biodiversity Offsets. Business and Biodiversity 388 389 Offsets Programme, Washington, D.C., USA. 390 BBOP (2012b) Resource paper: limits to what can be offset. Business and 391 Biodiversity Offsets Programme, Washington, D.C., USA. 392 BENDOR, T. (2009) A dynamic analysis of the wetland mitigation process and its 393 effects on no net loss policy. *Landscape and Urban Planning*, 89, 17–27. 394 BILBY, R.E., SULLIVAN, K. & DUNCAN, S.H. (1989) The generation and fate of road-395 surface sediment in forested watersheds in southwestern Washington. 396 Forest Science, 35, 453-468. 397 BIRDLIFE INTERNATIONAL (2015) Strengthening implementation of the mitigation 398 hierarchy: managing biodiversity risk for conservation gains. BirdLife 399 International, UNEP - WCMC, RSPB, FFI and the University of Cambridge, 400 Cambridge, UK. 401 BRIGHT, J., LANGSTON, R., BULLMAN, R., EVANS, R., GARDNER, S. & PEARCE-HIGGINS, J. 402 (2008) Map of bird sensitivities to wind farms in Scotland: A tool to aid 403 planning and conservation. *Biological Conservation*, 141, 2342–2356. 404 BROOKS, J.S., WAYLEN, K.A. & MULDER, M.B. (2012) How national context, project design, and local community characteristics influence success in 405 406 community-based conservation projects. *Proceedings of the National* 407 Academy of Sciences, 109, 21265-21270. 408 CLARE, S., KROGMAN, N., FOOTE, L. & LEMPHERS, N. (2011) Where is the avoidance in the implementation of wetland law and policy? *Wetlands Ecology and* 409 410 *Management*, 19, 165–182. 411 COUNCIL OF THE EUROPEAN COMMISSION (1992) Council directive 92/43/EEC of 21 412 May 1992 on the conservation of natural habitats and of wild fauna and 413 flora. In pp. 7–49. 414 DAWKINS, C. & FRAAS, J.W. (2010) Coming clean: the impact of environmental performance and visibility on corporate climate change disclosure. 415 416 *Journal of Business Ethics*, 100, 303–322. DEPARTMENT FOR COMMUNITIES AND LOCAL GOVERNMENT (2012) National Planning 417 418 Policy Framework. United Kingdom. 419 DINERSTEIN, E., BACCINI, A., ANDERSON, M., FISKE, G., WIKRAMANAYAKE, E., MCLAUGHLIN, 420 D., ET AL. (2015) Guiding agricultural expansion to spare tropical forests. 421 Conservation Letters, 8, 262–271.

422 423 424	DOWNIE, C. (2015) Prolonged international environmental negotiations: the roles and strategies of non-state actors in the EU. <i>International Environmental</i> <i>Agreements: Politics, Law and Economics</i> .
425 426 427 428	EDWARDS, D.P., GILROY, J.J., WOODCOCK, P., EDWARDS, F.A., LARSEN, T.H., ANDREWS, D.J.R., ET AL. (2014) Land-sharing versus land-sparing logging: Reconciling timber extraction with biodiversity conservation. <i>Global Change Biology</i> , 20, 183–191.
429 430 431	Екstrom, J., Bennun, L. & Mitchell, R. (2015) A cross-sector guide for implementing the Mitigation Hierarchy. Cross-Sector Biodiversity Initiative.
432 433	EUROPEAN COMMISSION (2007) Guidance document on Article 6(4) of the 'Habitats Directive' 92/43/EEC.
434 435 436	Ferreira, J., Aragão, L.E.O.C., Barlow, J., Barreto, P., Berenguer, E., Bustamante, M., et al. (2014) Brazil's environmental leadership at risk. <i>Science</i> , 346, 706– 707.
437 438 439	GARDNER, T.A., VON HASE, A., BROWNLIE, S., EKSTROM, J.M.M., PILGRIM, J.D., SAVY, C.E., ET AL. (2013) Biodiversity offsets and the challenge of achieving no net loss. <i>Conservation Biology</i> , 27, 1254–1264.
440	GASTON, K.J. (2010) Valuing common species. Science, 327, 154–155.
441 442 443	GIBBS, H.K., MUNGER, J., L'ROE, J., BARRETO, P., PEREIRA, R., CHRISTIE, M., ET AL. (2016) Did ranchers and slaughterhouses respond to zero-deforestation agreements in the Brazilian Amazon? <i>Conservation Letters</i> , 9, 32–42.
444 445	GREENWALD, D.N., SUCKLING, K.F. & РІММ, S.L. (2012) Critical habitat and the role of peer review in government decisions. <i>BioScience</i> , 62, 686–690.
446 447 448 449 450	 HANCE, J. (2011) Cambodian prime minister cancels titanium mine project citing impact on biodiversity and local people. <i>Mongabay</i>. Http://news.mongabay.com/2011/04/cambodian-prime-minister-cancels-titanium-mine-project-citing-impact-on-biodiversity-and-local-people/ [accessed 11 March 2016].
451 452	IVANOVA, M.R. (2015) Influencing corporations through shareholder activism: the case of three NGO-led campaigns in the UK. PhD thesis, Cardiff University.
453 454	IVES, C.D. & BEKESSY, S.A. (2015) The ethics of offsetting nature. <i>Frontiers in Ecology and the Environment</i> , 13, 568–573.
455 456 457	JAVELLE, AG. & VEIT, P.G. (2012) Managing land for mining and conservation in the Democratic Republic of Congo. Africa Biodiversity Collaborative Group.

458 459 460	KLEINSCHROTH, F., HEALEY, J.R. & GOURLET-FLEURY, S. (2016) Sparing forests in Central Africa: re-use old logging roads to avoid creating new ones. <i>Frontiers in Ecology and the Environment</i> , 14, 9–10.
461 462	KRAMER, L. (2009) The European Commission's Opinions under Article 6(4) of the Habitats Directive. <i>Journal of Environmental Law</i> , eqn028.
463 464 465	LAURANCE, B. (2016) The world's forests will collapse if we don't learn to say 'no'. <i>The Conversation</i> . Http://theconversation.com/the-worlds-forests-will- collapse-if-we-dont-learn-to-say-no-53979 [accessed 15 March 2016].
466 467 468	LESHY, J.D. & McUSIC, M. (2009) Where's the beef? Facilitating voluntary retirement of federal lands from livestock grazing. SSRN Scholarly Paper, Social Science Research Network, Rochester, NY.
469 470 471	MALCOM, J.W. & LI, YW. (2015) Data contradict common perceptions about a controversial provision of the US Endangered Species Act. <i>Proceedings of the National Academy of Sciences</i> , 112, 15844–15849.
472 473 474	MARON, M., DUNN, P.K., MCALPINE, C.A. & APAN, A. (2010) Can offsets really compensate for habitat removal? The case of the endangered red-tailed black-cockatoo. <i>Journal of Applied Ecology</i> , 47, 348–355.
475 476 477	MARTIN, C.S., TOLLEY, M.J., FARMER, E., MCOWEN, C.J., GEFFERT, J.L., SCHARLEMANN, J.P.W., ET AL. (2015) A global map to aid the identification and screening of critical habitat for marine industries. <i>Marine Policy</i> , 53, 45–53.
478 479 480 481	MASCIA, M.B., PAILLER, S., KRITHIVASAN, R., ROSHCHANKA, V., BURNS, D., MLOTHA, M.J., ET AL. (2014) Protected area downgrading, downsizing, and degazettement (PADDD) in Africa, Asia, and Latin America and the Caribbean, 1900– 2010. <i>Biological Conservation</i> , 169, 355–361.
482 483 484	MCKENNEY, B.A. & KIESECKER, J.M. (2010) Policy development for biodiversity offsets: a review of offset frameworks. <i>Environmental Management</i> , 45, 165–176.
485 486 487	MEYFROIDT, P., LAMBIN, E.F., ERB, KH. & HERTEL, T.W. (2013) Globalization of land use: distant drivers of land change and geographic displacement of land use. <i>Current Opinion in Environmental Sustainability</i> , 5, 438–444.
488 489 490	NEWTON, P., AGRAWAL, A. & WOLLENBERG, L. (2013) Enhancing the sustainability of commodity supply chains in tropical forest and agricultural landscapes. <i>Global Environmental Change</i> , 23, 1761–1772.
491 492 493	NIEDZIAŁKOWSKI, K., PAAVOLA, J. & JĘDRZEJEWSKA, B. (2012) Governance of biodiversity in Poland before and after the accession to the EU: the tale of two roads. <i>Environmental Conservation</i> , 40, 108–118.
494 495	Pilgrim, J.D., Brownlie, S., Ekstrom, J.M.M., Gardner, T.A., von Hase, A., ten Kate, K., Savy, C.E., Stephens, R.T.T., Temple, H.J., Treweek, J., Ussher, G.T., et al.

496 497	(2013) A process for assessing offsetability of biodiversity impacts. <i>Conservation Letters</i> , 6, 376–384.
498	PILGRIM, J.D., BROWNLIE, S., EKSTROM, J.M.M., GARDNER, T.A., VON HASE, A., TEN KATE, K.,
499	SAVY, C.E., STEPHENS, R.T.T., TEMPLE, H.J., TREWEEK, J. & USSHER, G.T. (2013)
500	Offsetability is highest for common and widespread biodiversity:
501	response to Regnery et al. <i>Conservation Letters</i> , 6, 387–388.
502	Рімм, S.L., Jenkins, C.N., Abell, R., Brooks, T.M., Gittleman, J.L., Joppa, L.N., et al.
503	(2014) The biodiversity of species and their rates of extinction,
504	distribution, and protection. <i>Science</i> , 344, 1246752.
505	POPE, J., BOND, A., MORRISON-SAUNDERS, A. & RETIEF, F. (2013) Advancing the theory
506	and practice of impact assessment: Setting the research agenda.
507	<i>Environmental Impact Assessment Review</i> , 41, 1–9.
508	QUÉTIER, F., REGNERY, B. & LEVREL, H. (2014) No net loss of biodiversity or paper
509	offsets? A critical review of the French no net loss policy. <i>Environmental</i>
510	<i>Science & Policy</i> , 38, 120–131.
511	RAINEY, H.J., POLLARD, E.H.B., DUTSON, G., EKSTROM, J.M.M., LIVINGSTONE, S.R., TEMPLE,
512	H.J. & PILGRIM, J.D. (2015) A review of corporate goals of No Net Loss and
513	Net Positive Impact on biodiversity. <i>Oryx</i> , 49, 232–238.
514	RAITER, K.G., POSSINGHAM, H.P., PROBER, S.M. & HOBBS, R.J. (2014) Under the radar:
515	mitigating enigmatic ecological impacts. <i>Trends in Ecology & Evolution</i> , 29,
516	635–644.
517 518 519	ROBINSON, J.G. (2012) Common and conflicting interests in the engagements between conservation organizations and corporations. <i>Conservation Biology</i> , 26, 967–977.
520	Rosenbarger, A., Gingold, B., Prasodjo, R., Alisjahbana, A., Putraditama, A. &
521	Tresya, D. (2013) How to change legal land use classifications to support
522	more sustainable palm oil in Indonesia. World Resources Institute,
523	Washington DC, USA.
524	SAKHALIN ENERGY (2009) Biodiversity action plan. Sakhalin Energy Investment
525	Company Ltd.
526 527 528 529	SENIOR, M.J.M., BROWN, E., VILLALPANDO, P. & HILL, J.K. (2015) Increasing the scientific evidence-base in the 'High Conservation Value' (HCV) approach for biodiversity conservation in managed tropical landscapes. <i>Conservation Letters</i> , 8, 361–367.
530	SHELL (2014) Corrib Development Biodiversity Action Plan 2014-2019. Shell E&P
531	Ireland Limited.
532 533 534	STOTT, I., SOGA, M., INGER, R. & GASTON, K.J. (2015) Land sparing is crucial for urban ecosystem services. <i>Frontiers in Ecology and the Environment</i> , 13, 387–393.

535 536 537	SULLIVAN, S. & HANNIS, M. (2015) Nets and frames, losses and gains: Value struggles in engagements with biodiversity offsetting policy in England. <i>Ecosystem Services</i> , 15, 162–173.
538 539 540 541 542	US EPA AND DA (1990) Memorandum of agreement between the Environmental Protection Agency and the Department of the Army concerning the determination of mitigation under the Clean Water Act Section 404(b)(1) guidelines. US Environmental Protection Agency & US Department of the Army.
543 544 545	VENTER, O., POSSINGHAM, H.P., HOVANI, L., DEWI, S., GRISCOM, B., PAOLI, G., ET AL. (2013) Using systematic conservation planning to minimize REDD+ conflict with agriculture and logging in the tropics. <i>Conservation Letters</i> , 6, 116–124.
546 547	VIDAL, J. (2016) Major Amazon dam opposed by tribes fails to get environmental licence. <i>The Guardian</i> .
548 549 550	VILLARROYA, A., BARROS, A.C. & KIESECKER, J. (2014) Policy development for environmental licensing and biodiversity offsets in Latin America. <i>PLoS</i> <i>ONE</i> , 9, e107144.
551 552	WALKER, S., BROWER, A.L., STEPHENS, R.T.T. & LEE, W.G. (2009) Why bartering biodiversity fails. <i>Conservation Letters</i> , 2, 149–157.
553 554 555 556	WESTERN GREY WHALE ADVISORY PANEL (2015) WGWAP Statement of concern with respect to proposed seismic activity on the Sakhalin shelf in 2015. Http://cmsdata.iucn.org/downloads/finalcleanwgwapstatement_08_may _final.pdf.
557 558	

Table 1. Examples of voluntary standards and national legislation which 559

set requirements for impact avoidance by defining actions and criteria. 560

Parentheses indicate cases where a requirement is acknowledged but not 561

- clearly defined. 562
- 563

	Standard or law		Act	ions		Crite pa	eria fo ist av sta	or mo oidan age	ving ce
		Consult with stakeholders	Assess environmental and social impacts	Consider cumulative impacts	Implement long-term monitoring	No viable lower-impact alternative	Overriding public interest	No net impact on critical biodiversity features	Compliance with the law
Ø	Business and Biodiversity Offsets Programme: Standard on Biodiversity Offsets (2012)	1	1	√	V			1	1
d guidance	Cross Sector Biodiversity Initiative: A cross- sector guide for implementing the Mitigation Hierarchy European Bank for Reconstruction and	V	√	V	V	V		V	V
ls an	Development: Performance Requirement 6	√	√	\checkmark	√	√	\checkmark	\checkmark	√
andard	International Finance Corporation: Performance Standard 6*	√	√		V	V		V	V
Sta	World Bank: proposed Environmental and Social Standard 6	\checkmark	1	√	\checkmark	\checkmark		\checkmark	\checkmark
Legislation	Australia: Environment Protection and Biodiversity Conservation Act, environmental offsets policy British Columbia (Canada): Policy for Mitigating		1		(√)	V		V	√
	Impacts on Environmental Values	(√)	\checkmark	\checkmark	(√)	√			√
	European Union: Habitats Directive 92/43/EEC, EIA Directive	√	√			V	√	1	√
	United Kingdom: National Planning Policy Framework 2012	√	√	V		V	V	V	V

564 565 566

*Requirements for consultation and impact assessment are established in IFC Performance Standard 1

567 **Table 2. Different types of impact avoidance, with examples. Inclusion of**

568 projects is solely to illustrate the range of actions that can be taken to avoid

569 impacts on biodiversity, and should not be interpreted as endorsement,

570 nor as a suggestion that best practice was necessarily followed.

571

Type of impact avoidance	Where appropriate	Example	References
No-project	Irreplaceable features with no viable alternatives, and where offsets unlikely to	Development permit refused for São Luiz do Tapajós dam in Brazil	(Vidal, 2016)
	succeed	Titanium mine in Cardamom Mountains of Cambodia cancelled	(Hance, 2011)
Spatial avoidance	Lower-impact alternative locations can be identified	Site for desalination plant in Namibia selected to avoid tern colony	(Aurecon & SLR, 2015)
		Via Baltica road re-routed to avoid Rospuda Valley and other protected sites in Poland	(Niedziałkowski et al., 2012)
Temporal avoidance	Time periods when activities will not affect vulnerable features can be identified	Construction and seismic surveys suspended during breeding season of Steller's Sea Eagles and seasonal presence of Gray Whales in Okhotsk Sea, Russia	(Sakhalin Energy, 2009)
		Logging activities in United States scheduled during dry periods to avoid erosion and sediment runoff	(Bilby et al., 1989)
Design- based avoidance	Technology and planning can be used to modify project components to avoid specific impacts	Tunneling equipment used to install pipeline underground below estuary in Ireland	(Shell, 2014)
		Logging operations to re- use old access roads instead of creating new ones in Central Africa	(Kleinschroth et al., 2016)

Table 3. Reasons for the failure of plans and policies to avoid impacts on biodiversity and ecosystem services, and some possible solutions.

	Reason for failure	Possible solutions				
Political will	Lack of political will to support impact avoidance	Harness and broadcast public support for conservation; expose conflicts of interest; reform institutions giving private interests undue influence				
	Culture within planning authorities of not valuing biodiversity	Make biodiversity education mandatory for all staff of planning authorities				
Regulation	Legal protection insufficient to ensure impact avoidance	Incorporate mitigation hierarchy principles into legislation; resist efforts to weaken legislation				
	Ineffective judicial frameworks for holding decision makers to account	Make full use of those judicial frameworks which are effective; lobbying for stronger legislation				
	Failure to avoid impacts to biodiversity that is not considered "important"	Set avoidance requirements for biodiversity of all kinds, including common species and habitats				
	Weak requirements for restoration and offsetting make remediation more attractive than impact avoidance	Enforce detailed, stringent requirements for restoration and offsetting, including higher bond requirements and penalties for failure to remediate				
-	Impact avoidance not considered until ESIA	Make early stakeholder engagement the industry norm; assess biodiversity risks before ESIA				
	Failure to anticipate and identify likely impacts	Audit impact assessments; require assessment of indirect and cumulative impacts				
ces	"No-project" option not considered	Require assessment of "no-project" option				
Pro	Poor communication between ecologists, engineers, other technical consultants	Require direct cooperation between consultant teams as part of ESIA contract				
	Failure to adhere to plans	Hold governments, companies accountable to plans				
	Decision to proceed is made on basis that remediation will compensate for impacts	Separate the decision to proceed from any assessment of remediation possibilities				
	Lack of resources and ecological expertise	Dedicate resources to create ecologist roles within				
ţ	within planning bodies	planning bodies; improve planner-ecologist liaison				
aci	Poor coordination between conservation	Provide resources to integrate conservation planning				
Cap	and planning authorities	Develop voluntary or regulatory mechanisms to				
	areas	ensure avoided areas receive long-term protection				
	Biodiversity data inaccessible or difficult	Improve data availability through platforms that				
	to use	increase ease of use by non-specialists				
echnical knowledge	Important biodiversity not prioritised and identified before development	Comprehensive assessments of important biodiversity at local, regional and national levels				
	Limited understanding of trade-offs	Incorporate trade-off analysis into ESIA				
	Perception that impact avoidance is too	Neutral analysis of costs and benefits of impact				
	costly	avoidance, including non-monetary				
	Discounting of future costs relative to costs today	Estimation and communication of future costs and limitations of restoration and offsetting				
F	Unrealistic assumptions about technical	Collate evidence on efficacy of restoration and				
	capacity to restore makes remediation more attractive than avoidance	offsetting; communicate limits of remediation; use offset multipliers commensurate with uncertainties				