

Motorboat noise impacts parental behaviour and offspring survival in a reef fish

Sophie L. Nedelec^{1*}, Andrew N. Radford¹, Leanne Pearl², Brendan Nedelec¹,

Mark I. McCormick³, Mark G. Meekan⁴, Stephen D. Simpson⁵

¹School of Biological Sciences, Life Sciences Building, 24 Tyndall Avenue, University of Bristol, BS8 1TQ, U.K.

²Ecology and Evolutionary Biology, University of Colorado Boulder, 1900 Pleasant Drive, 334 UCB, Boulder, CO 80309-0334

³School of Marine and Tropical Biology, James Cook University, Townsville, Queensland 4811, Australia

⁴Australian Institute of Marine Science, Botany Building, The University of Western Australia, Crawley 6009, Western Australia, Australia

⁵Biosciences, College of Life and Environmental Sciences, University of Exeter, Stocker Road, Exeter, EX4 4QD, U.K.

*Correspondence: sophie.nedelec@hotmail.co.uk;

+44 (0) 1392 722714

Now at: Environment and Sustainability Institute, University of Exeter, Penryn, Cornwall TR10 9FE, U.K.

Keywords: anthropogenic noise, parental care, pollution, Lizard Island, motorboat-noise
playback

Anthropogenic noise is a pollutant of international concern, with mounting evidence of disturbance and impacts on animal behaviour and physiology. However, empirical studies measuring survival consequences are rare. We use a field experiment to investigate how repeated motorboat-noise playback affects parental behaviour and offspring survival in the spiny chromis (*Acanthochromis polyacanthus*), a brooding coral reef fish. Repeated observations were made for 12 days at 38 natural nests with broods of young. Exposure to motorboat-noise playback compared to ambient-sound playback increased defensive acts, and reduced both feeding and offspring interactions by brood-guarding males. Anthropogenic noise did not affect the growth of developing offspring, but reduced the likelihood of offspring survival; while offspring survived at all 19 nests exposed to ambient-sound playback, six of the 19 nests exposed to motorboat-noise playback suffered complete brood mortality. Our study, providing field-based experimental evidence of the consequences of anthropogenic noise, suggests potential fitness consequences of this global pollutant.

Introduction

Mounting evidence indicates that anthropogenic noise, a pervasive pollutant, disturbs and has detrimental effects on a wide range of species, including mammals, birds, anurans, fishes and invertebrates (see reviews in [1-6]). Studies showing short-term behavioural and physiological impacts of noise are numerous [7-10]. Some chronic effects of noise, such as altered habitat use and reduced pairing success, have also been identified [11-12]. However, studies that reveal impacts on reproduction or survival via experimental manipulations, with suitable controls and replicates are rare (for exceptions, see [12-14]).

Anthropogenic noise has been shown to affect parental behaviour, including feeding, nest maintenance and defence. Specific examples include reduced time spent tending nests in the damselfish *Chromis chromis* [15], increased latency to visit a nest box in great tits (*Parus major* [9]), and increased missed detections of parents leading to reduced begging in tree swallows (*Tachycineta bicolor* [16]). While noise has clear effects on parental-care behaviour in the short term, there remains the possibility that ongoing exposure would allow animals to habituate, compensate or move away from the source [4,5,17,18]. Therefore, longer-term studies considering offspring survival as well as parental care are required.

We investigated the effects of repeated exposure to anthropogenic noise on male parental behaviour and offspring growth and survival in a coral reef fish, the spiny chromis (*Acanthochromis polyacanthus*). We used playbacks of recordings of motorboat noise, since that is the most common source of anthropogenic noise in shallow reef environments [19]. *A. polyacanthus* exhibits bi-parental care of eggs and larvae at nests within shallow reef habitat in the tropical Western Pacific [20,21]; males contribute more care than females in this species (McCormick, pers. obs.). One of the most vital roles of adults is to guard their brood by chasing away potential predators and competitors [22]. Parental care is energetically expensive [23] and thus it is important that parents feed regularly. Moreover, *A. polyacanthus* parents provide their offspring with mucus, which can contain proteins, hormones, ions, micro-organisms, immunoglobulins and secretocytes undergoing mitosis [24-26]. Mucus is delivered via 'glancing' (also called 'parent-touching' or 'contacting' in other species); parents are relatively passive in this process, but do actively avoid offspring on some occasions. These three key parental-care behaviours (guarding, feeding and glancing) are all easily observed in *A. polyacanthus* in its natural habitat [27,28].

We exposed 38 *A. polyacanthus* nests with recently hatched juveniles to 12 days of playback of either motorboat passes recorded near reefs or natural ambient sound recorded at the same locations.

We collected data throughout the acoustic-exposure period to answer three main questions. 1) Is guarding, feeding and glancing behaviour of brood-guarding males negatively impacted by the addition of motorboat noise? 2) Can an increased frequency of defensive acts by brood-guarding males be explained by changes in the prevalence or behaviour of other local species? 3) Do *A. polyacanthus* offspring at nests experiencing motorboat-noise playback suffer reduced growth or survival compared to control nests with playback of ambient reef sound?

Methods

General experimental set-up

Data were collected between October and December 2013 at Lizard Island Research Station (14°4'S 145°28'E), Great Barrier Reef, Australia. Thirty-eight *A. polyacanthus* nests with new clutches of juveniles were studied; full details in Supplementary Material. Half of the nests were allocated to the 'Ambient' and half to the 'Boat' sound treatment. Four replicate playbacks were constructed for each treatment. Each replicate used a different recording of either ambient sound or motorboat noise, and was played on a loop (resulting in six boat disturbances per hour in the Boat treatment) at the relevant nest during daylight hours (06:00–18:00). Figure 1 shows examples of mean sound-pressure and particle-acceleration levels from spectral analysis of 60 x 1 s windows (window length = 1024, overlap = 50%) at nests. The mean \pm standard error root-mean-square (RMS) sound-pressure level between 0 and 2000 Hz across 60 s samples was 108.1 ± 0.5 dB re 1 μ Pa at 1 m at the 19 Ambient sites and 128.7 ± 0.2 dB re 1 μ Pa at 1 m at the 19 Boat sites. Further details of recordings and playbacks, including a figure (S1) showing the set-up of playback equipment at nests, are in the Supplementary Material.

Acanthochromis polyacanthus male behaviour

Data on three key behaviours by brood-guarding males were collected. (1) Number of defensive acts (chasing/making aggressive strikes towards other fish; any potential competitors for the territory would also be potential predators of the offspring). (2) Percent time feeding (characteristic short or extended movements in the water column searching for and consuming plankton, and grazing on algae from the substrate). (3) Number of instances of 'glancing' (where offspring eat mucus from the focal male; males do not initiate these interactions by characteristic posing, but they can choose to avoid them).

Each nest was visited by SLN every other day for 12 days between 08:00 and 16:00. Fish were given 1-min settling time to resume normal activity following the arrival of the observer, after which behaviour of the adult male was observed for 3 min at a distance of ~2 m from the nest. In a rigorous examination of consistency of behaviour in any fish, White et al. [29] showed that juvenile damselfish are consistent in behaviour over short (hours) and medium (days) timeframes, and that 3 min is sufficient to obtain a good indication of their behavioural traits. During preliminary observations on our focal species and life-stage, we found that feeding, aggression and glancing could all be observed within 3 min. The adult male was chosen for behavioural observation as he provides a greater proportion of parental care in this species (McCormick, pers. obs.) and is easily identified by his large genital papilla.

Prevalence and behaviour of other fish species

To assess whether any changes in paternal care or offspring survival were the consequence of a change in the local fish community, all fish within a 5 m radius of the nest were counted (by LP) directly after behavioural observations at each site. Fishes from the families Gobiidae and Blenniidae were excluded to avoid potentially unreliable data as species-level identification underwater was difficult. To assess whether the increased number of defensive acts by brood-guarding males was the consequence of a change in predation threat, the number of potentially predatory fish within 5 m and within 1 m of the nest was also calculated on each occasion. Potential predators were those species that had been seen

previously or during this study to strike at juvenile *A. polyacanthus*. Potential predators of adults were seen only rarely. The lists of predators and other fish in the community seen surrounding nests can be found in the Supplementary Material in tables S1 and S2 respectively. The number of aggressive strikes made towards the male *A. polyacanthus* by other fish species (all of which were potential predators of the offspring) was also recorded during the behavioural observation period on *A. polyacanthus* males.

***Acanthochromis polyacanthus* offspring growth and survival**

Three *A. polyacanthus* juveniles from each of the focal nests were removed for measurement by hand net at the beginning of the acoustic-exposure period. It was not possible to collect juveniles from one of the Ambient nests at day 0 due to the morphology of their coral shelter. Removals on day 0 represented between 1.2% and 7.3% of broods from different nests; the percentage removed did not differ significantly between sound treatments (Mann Whitney test: $U=150$, $N_{\text{Ambient}}=18$, $N_{\text{Boat}}=19$, $P=0.523$). Removed juveniles were not returned to the nest after measurements were taken. Twenty more juveniles were removed for measurement at the end of the acoustic-exposure period (i.e. end of day 12) from those nests where broods had survived. Each removed fish was weighed (wet mass) and measured for standard length and body width (cross sectional perimeter at the cloaca, and therefore not influenced by gut fullness, perpendicular to the line from tip of the mouth to middle of the tail used for standard length). Body width is a measure of muscular development; shape was measured as the ratio of body width to standard length. Survival was measured by whether any offspring remained at the nest at the end of the experiment.

Statistical analysis

General linear mixed-effects models (LMMs) fitted by maximum likelihood (Laplace approximation) were used (after log transformation to meet the assumption of normality where necessary), to test for

the effects of sound treatment and number of days of sound exposure (including a possible interaction) on male behaviour, while controlling for the random effects of nest and time of day. Number of aggressive strikes were also included in the model of feeding behaviour to test whether aggression affected time allocated to feeding. Glancing was split into a binomial generalized linear mixed-effects model (GLMM) to test for whether glancing occurred or not, and a GLMM with Poisson errors for counts of glancing when it did occur. Results of interaction terms are presented only if significant. See the Supplementary Material for further details of how these mixed-effects models were used. At two Boat nests, offspring survival was zero before the first parental behaviour observations could take place on day 2, thus these nests were not included in the analysis of paternal-care behaviour.

To examine differences in fish communities surrounding nests, a permutation-based, non-parametric multivariate analysis of similarity (ANOSIM) using the software PRIMER (Plymouth Routines in Multivariate Ecological Research v. 6.1.13; PRIMER-E Ltd, Plymouth Marine Laboratory, Plymouth, UK [30]) was conducted. Further details of this method can be found in the Supplementary Material. The mean number of predatory fish within 1 m and 5 m of each of the focal nests in the two treatments were compared using Mann Whitney U tests, as was the mean number of strikes made by other fish towards the *A. polyacanthus* brood-guarding male.

The number of nests where there was complete brood mortality was compared between sound treatments using a Fisher's exact test. The initial size, shape and mass of juveniles where complete brood mortality occurred was compared with other nests using Mann Whitney U tests. The changes in size, shape and mass of *A. polyacanthus* offspring from day 0 to day 12 were compared between Ambient and Boat nests using Mann Whitney U tests. N was determined by the number of nests where data could be collected at day 12 (i.e. not if offspring survival was zero).

Results

Acanthochromis polyacanthus paternal care behaviour

There was a significant effect of sound treatment on defensive acts made by brood-guarding *A. polyacanthus* males (LMM: $X^2_1 = 5.85$, $p = 0.016$; male ID: variance = 0.18, s.d. = 0.42; time of day: variance = 0, s.d. = 0); there was no significant effect of number of days of sound exposure ($X^2_1 = 0.91$, $p = 0.340$). Boat treatment males made on average twice as many defensive acts (chasing/making aggressive strikes) at other fish compared to males exposed to ambient-sound playback (Fig. 2a). Males at Boat nests also spent 25% less time feeding (displaying characteristic movements in the water column searching for and consuming plankton, or algae from the substrate) than those at Ambient nests (LMM: $X^2_1 = 4.42$, $p = 0.036$; male ID: variance = 414.87, s.d. = 20.37; time of day: variance = 11.47, s.d. = 3.39; Fig. 2b). Time spent feeding also increased with number of days of sound exposure ($X^2_1 = 12.94$, $p < 0.001$) and decreased with increasing number of aggressive strikes made by males ($X^2_1 = 10.94$, $p < 0.001$). Whether offspring glancing (eating mucus from the focal parent) occurred was not significantly affected by sound treatment (GLMM: $X^2_1 = 0.04$, $p = 0.848$; male ID: variance = 0.07, s.d. = 0.26; time of day: variance = 0, s.d. = 0) nor by number of days of sound exposure ($X^2_1 = 0.70$, $p = 0.403$). In cases where offspring glancing did occur, it did so three times less often at nests exposed to motorboat-noise playback compared to those exposed to ambient-sound playback (GLMM: $X^2_1 = 5.07$, $p = 0.024$; male ID: variance = 0, s.d. = 0; time of day: variance = 0.16, s.d. = 0.40; Fig. 2c); there was a non-significant trend for a positive effect of number of days of sound exposure ($X^2_1 = 3.47$, $p = 0.063$).

Prevalence and behaviour of other fish species

The increased number of defensive acts by *A. polyacanthus* brood-guarding males exposed to motorboat-noise playback compared to ambient-sound playback did not appear to be the consequence of a change in the local fish community, because there was no observed effect of sound treatment on community composition surrounding *A. polyacanthus* nest sites (ANOSIM: $R = -0.022$, $P = 0.632$; all

pairwise comparisons, $P > 0.90$). Moreover, the increased number of defensive acts by brood-guarding males did not appear to be the consequence of a change in predation threat because there was no significant difference between sound treatments in the number of predatory fish within 1 m (Mann Whitney test: $U = 98.5$, $N_{\text{Boat}} = 14$, $N_{\text{Ambient}} = 15$, $P = 0.795$) or 5 m ($U = 92$, $N_{\text{Boat}} = 14$, $N_{\text{Ambient}} = 15$, $P = 0.582$) of the focal nest. The increased number of defensive acts by brood-guarding males also did not appear to be the consequence of a change in predatory attacks, because there was no significant difference between sound treatments in the number of attacks made by other fish towards the focal *A. polyacanthus* male ($U = 162.5$, $N_{\text{Boat}} = 18$, $N_{\text{Ambient}} = 19$, $P = 0.729$).

***Acanthochromis polyacanthus* offspring growth and survival**

Complete mortality of broods (survival = 0) was significantly more likely in the Boat treatment (six of 19 nests) compared to the Ambient treatment (zero of 19 nests; Fisher's exact test: $P = 0.020$). A significant difference between treatments was still apparent if the two nests suffering complete mortality in the first two days of motorboat-noise playback were removed from the analysis ($P = 0.040$). The offspring at nests that suffered 100% mortality ($N = 6$) were not significantly different in initial size (Mann Whitney test: $U = 74$, $P = 0.511$), shape ($U = 79$, $P = 0.664$), or mass ($U = 71$, $P = 0.432$) compared with other nests ($N = 31$). At nests that did not suffer complete brood mortality and for which data were available (18 Ambient and 13 Boat nests), there was no significant effect of sound treatment on the change in juvenile fish size (Mann Whitney test: $U = 41$, $P = 0.262$), shape ($U = 67$, $P = 0.601$) or mass ($U = 52$, $P = 0.516$).

Discussion

Defensive and feeding behaviour of *Acanthochromis polyacanthus* brood-guarding males, male-offspring interactions and survival of young were all affected by playback of motorboat noise compared

to ambient-sound playback. We found no evidence of changes in tolerance, habituation, or sensitisation to motorboat-noise exposure over the duration of our 12 day study (cf. [18,31]). Impacts of noise on parental-care behaviour have been shown previously [9,15,16,32]. However, our study also provides experimental evidence of an impact of anthropogenic noise on survival in free-ranging wild animals: motorboat-noise playback resulted in complete brood mortality not seen in ambient-sound playback conditions, although there was no significant difference between sound treatments in offspring growth or shape at surviving nests during our study.

Heightened stress may have caused the higher levels of aggression and chasing of other fish by *A. polyacanthus* brood-guarding males exposed to motorboat-noise playback [33]. Alternatively, stress may have caused distraction or distraction could have occurred without stress, resulting in decision-making errors [34,35]. Distraction could have led males in our study to chase and attack other fish inappropriately when exposed to motorboat-noise playback; for example, chasing fish that were not a predatory threat or chasing threatening fish less efficiently. Our findings that predator presence did not increase, but that offspring survival decreased despite increased parental acts of defence, suggests that parental-care behaviour became less efficient. One consequence of the increased defensive behaviour is reduced time spent by *A. polyacanthus* males on feeding. Motorboat-noise playback may also have impacted foraging directly, as has been seen in various other species [7,36,37]. A reduction in the acquisition of resources combined with higher energy outputs involved in nest defence would be likely to reduce body condition of parents. Measuring changes in parental condition was beyond the scope of our study but should be a focus for future work, as parental condition has previously been associated with increased mortality in offspring of *A. polyacanthus* [38].

We also found a reduction in glancing behaviour of fish exposed to motorboat-noise playback compared to those exposed to ambient-sound playback. While this is an indirect form of provisioning, with parents simply allowing young to eat their mucus, it still requires parents to be present and to

undergo a cost for their offspring, as mucus is energetically expensive to produce [39]. Although the number of glances by juvenile *A. polyacanthus* may not directly determine nutritional state [27], the behaviour is likely to have adaptive functions such as the transfer of growth hormone (tiGH [40]) and building immune function [25,26]. It is possible that reduced glancing could impact growth and survival of offspring beyond the duration of our study.

A number of potential factors could have acted individually or in combination to produce the complete mortality we observed at 32% of the broods exposed to motorboat-noise playback. Parents could have abandoned or cannibalised their offspring [21,22]. Either leaving the territory permanently or stopping looking after their young while still at the territory would constitute abandoning the nest. However, we did not see a decrease in parental-care behaviour prior to nest mortality, and parents were still at the site when we returned to nests multiple times over several days after nest mortality to be sure that offspring were no longer present. Moreover, filial cannibalism is generally rare [22]; we did not observe cannibalism during behavioural observations, although we did observe predation by other fish; and the occasional observations (N = 4) of parental aggression towards offspring in the current study occurred in both sound treatments and not at the nests where mass mortality was recorded. Another possibility – that predation intensity increased in the presence of motorboat noise – also seems unlikely to be the explanation for our results, since greater numbers of predators were not observed in the vicinity of nests nor were attacks by other fish more likely at nests exposed to motorboat-noise playback compared to ambient-sound playback.

Instead, perhaps the most likely explanation for the greater brood failure in the Boat treatment compared to the Ambient treatment is increased risk of predation. There are two mechanisms by which predation risk could have increased. First, although we found no change in size, shape or mass of larvae, it is possible that they suffered impaired predator-avoidance behaviour via stress and/or distraction, as has been seen in juveniles of other damselfish [14]. Second, more chasing of inappropriate species and

at inappropriate times could mean males spent more time focusing attention on other fish and less time in close proximity to the nest, which may have left offspring vulnerable to predatory attack due to reduction in effectiveness of parental defence. An early descriptive study also indicated that motorboat disturbance could increase the vulnerability of fish nests: longear sunfish (*Lepomis megalotis*) were more likely to move away from their nest when a slow-moving motorboat was nearby [41]. Predators that have first located a nest are likely to return, especially if they have been successful at obtaining food from it, and so complete brood mortality could arise. This raises the question of how reproductive output over the length of a whole breeding season may be affected.

Our field study found consequences of chronic-noise exposure on the survival of juvenile *A. polyacanthus* in the wild; direct testing is needed if conclusions are to be drawn about other species. We note the important caveat that our experiment used underwater loudspeakers, which do not broadcast the full range of sounds produced by motorboats. But, it is also possible that our results are therefore conservative with respect to the full impact of motorboat noise, and recent work has found qualitatively similar fitness effects when using playbacks in tanks and real motorboats in open-water conditions [14]. Moreover, other stages of reproduction could also be affected negatively by motorboat noise: one study has indicated, for example, that spawning could be interrupted by the approach of a fast-moving powerboat [42]. Motorboats are found throughout the world wherever humans inhabit coastal areas, and our results suggest that boat noise should be considered in the management of fisheries and protected areas. In an even broader sense, anthropogenic noise is fast becoming an integral part of both marine and terrestrial ecosystems (for example, ship noise can travel for 1000s of km underwater and >80% of land in the USA is within 1 km of a road [43,44]). Nest-defence behaviour is common amongst benthic spawning fishes and parental-care behaviour including defence of offspring is widespread in many other taxa including birds and mammals. Noise-induced increases in mortality due to impaired parental care could therefore be widespread and lead to population-level impacts.

Ethics

This research adhered to the Association for the Study of Animal Behaviour/Animal Behavior Society Guidelines for the Use of Animals in Research, the legal requirements of the country (Australia) in which the work was carried out, and with collection permits and ethics approval from: Great Barrier Reef Marine Park Authority, Australian Institute of Marine Science, Lizard Island Research Station, James Cook University (A2081), the University of Exeter (2013/247) and the University of Bristol (UIN/13/036).

Data

The datasets supporting this article have been uploaded as part of the Supplementary Material.

Competing interests

We have no competing interests

Authors' contributions

SLN, SDS, ANR, MIM and MGM designed the study; SLN, LP, BN and SDS collected the data; SLN analysed and interpreted data with ANR, SDS and MIM; SLN wrote the manuscript and ANR, MIM, MGM and SDS helped to edit the manuscript.

Acknowledgements

We thank the staff at Lizard Island for logistical support, Julius and Kate Piercy for field assistance, Geospectrum Technologies for providing the accelerometer, and Innes Cuthill, Hans Slabbekoorn and three anonymous referees for constructive manuscript comments.

Funding

This work was supported by a NERC Knowledge Exchange Fellowship (for S.D.S.; NE/J500616/2), the UK Department for Environment Food and Rural Affairs (S.D.S. and A.N.R.; ME5207), funds from ARC Centre of Excellence for Coral Reef Studies (M.I.M; EI140100117), an EPSRC studentship and Subacoustech (S.L.N.).

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Figure Captions

Figure 1. Power spectral density for a) sound-pressure levels and b) x-axis particle-acceleration levels.

Original recordings of motorboat noise and ambient sound are compared with playbacks of these recordings at experimental sites (mean of 60 s samples, window length=1024, overlap=50%). Sound-pressure playbacks were recorded at 19 Boat and 19 Ambient sites while particle acceleration could only be recorded at one Boat and one Ambient site. Playbacks reveal a peak in sound level around 2000 Hz and troughs around 800 and 1500 Hz (artefacts of the loudspeakers used), but for both sound pressure and particle acceleration, motorboat noise and motorboat-noise playbacks were louder than ambient sound and ambient-sound playbacks at all sites, at frequencies produced by the speaker (>100 Hz). Also, real motorboats were louder than motorboat-noise playbacks, but real ambient sound was quieter than ambient-sound playbacks, making our experimental playback levels a conservative representation of reality.

Figure 2. Behavioural responses to playback of motorboat noise compared with playback of ambient sound: a) brood-guarding males made more defensive acts per min; b) males spent less time feeding; c) glancing behaviour was rarer. Boxes represent interquartile ranges and lines within boxes represent the median across 19 Ambient and 17 Boat nests. Whiskers represent ± 1.5 x interquartile range. Open circles denote any data points that fall outside of the range of the whiskers. N determined by number of nests, data within nests averaged over duration of exposure, two Boat nests suffered complete mortality prior to first observation.

Figure 1

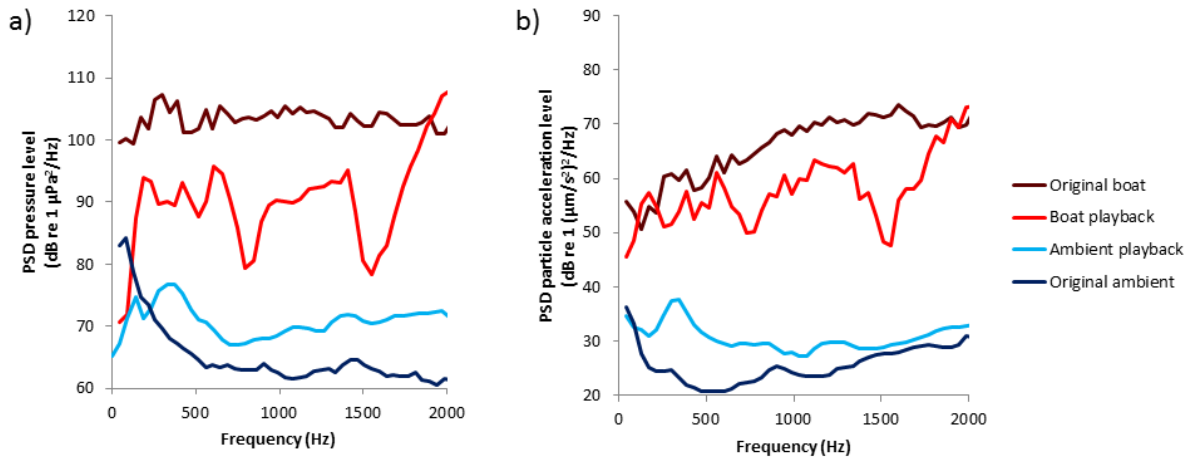
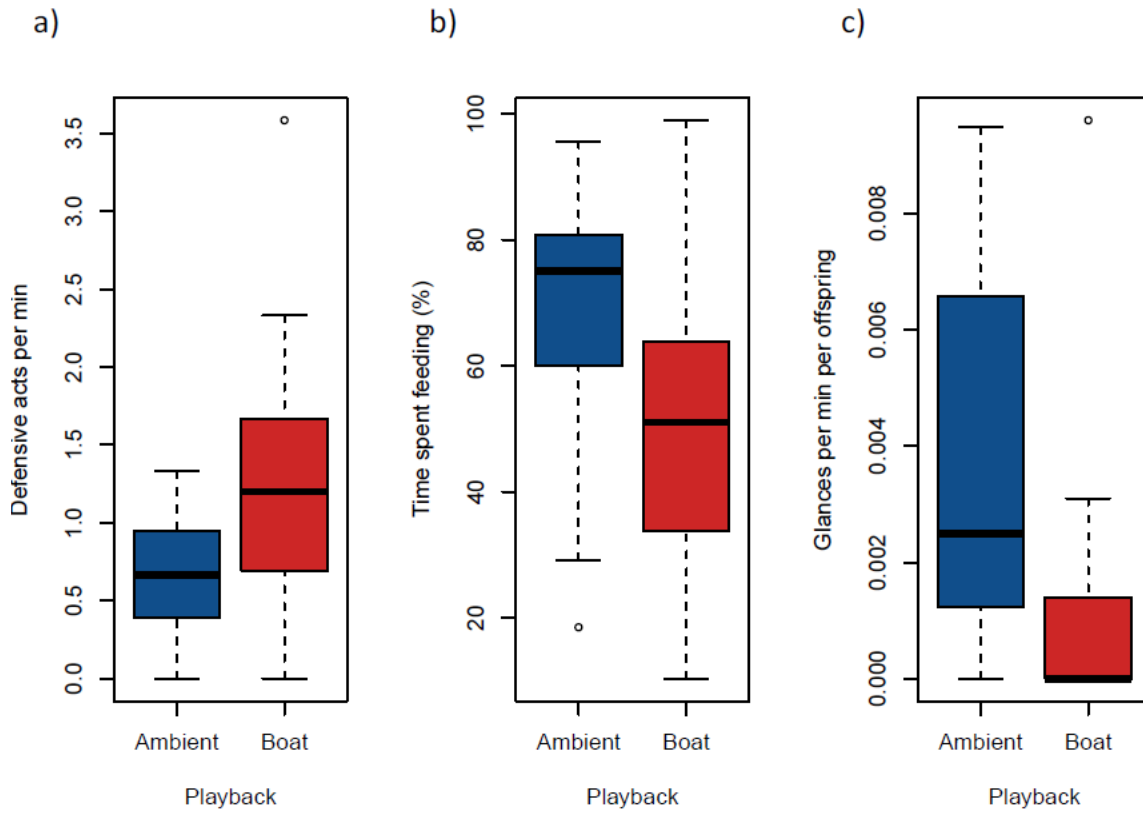


Figure 2



Motorboat noise impacts parental behaviour and offspring survival in a reef fish

Sophie L. Nedelec, Andrew N. Radford, Leanne Pearl, Brendan Nedelec,

Mark I. McCormick, Mark G. Meekan, Stephen D. Simpson

Supplementary Material

Supplemental Experimental Procedures

Study site and species

Juveniles in experimental nests had a standard length <14 mm, meaning they were <10 days post-hatching. Nests were located along the reef (2.4–3.3 m depth at mid tide) in the lagoon between Lizard, Palfrey, and South Islands and Seabird Islet in the Lizard Island group; an area covering ~1 km². Nests were separated by at least 40 m to ensure independence (home ranges of parents at nests is <10 m and parents spend most of their time within 2 m (McCormick, pers. comm.)).

Nineteen of the 38 *A. polyacanthus* nests with offspring were allocated to the Ambient treatment and 19 were allocated to the Boat treatment; allocation alternated between treatments and nests allocated to the two sound treatments did not differ significantly in initial brood size (overall mean±SE=132±8; range=41–247; Mann Whitney test: U=158, N_{Ambient}=N_{Boat}=19, P=0.511).

Acoustic recordings

Sound pressure was recorded using a hydrophone (HiTech HTI-96-MIN with inbuilt preamplifier; sensitivity -165 dB re 1V/ μ Pa; frequency range 2 Hz–30 kHz; calibrated by manufacturers; High Tech Inc., Gulfport MS) and a hand-held recorder (PCM-M10, 48 kHz sampling rate, Sony Corporation, Tokyo, Japan). An accelerometer (M20L; sensitivity following a curve over the frequency range 0–3 kHz; calibrated by manufacturers; Geospectrum Technologies, Dartmouth, Canada) was also available for making limited numbers of field recordings of particle motion. When using the accelerometer, both sound components were recorded via a sound card (MAYA44, ESI Audiotechnik GmbH, Leonberg, Germany) onto a laptop (Techra R840-12F, Toshiba). When recording acceleration, the x-axis of the accelerometer was oriented towards the sound source. The soundcard and laptop were calibrated using pure sine wave signals generated in SAS Lab (Avisoft, Germany), played on an MP3 player and measured with an oscilloscope. Four separate 5-min recordings of ambient sound were made in front of the Lizard Island Research Station, 20 m from the nearest coral reef where the water was 5 m deep and the hydrophone and accelerometer were at a depth of 2 m. Four different typical outboard motorboats (with 5 m long aluminium hulls and 30 hp Suzuki 2-stroke outboard motors) were also recorded in the same location for 10 min each; an approach from the shore (600 m away) over 2.5 min was followed by 5 min intense activity where the motorboat was driven around the hydrophone and accelerometer, followed by 2.5 min where the motorboat returned to the shore.

Experimental playback

Boat-noise playback was the experimental treatment and ambient-sound playback was necessary as a control treatment because ambient reef sound was always present during boat recordings. The Boat playback regime meant that there were six periods of nearby motorboat noise per hour. Playback soundtracks were allocated to the nests in a randomised blocked design (four soundtracks and two treatments in each block of eight nests). Playback exposure lasted 12 d as this was

sufficient time to demonstrate an impact of playback on behaviour and fitness, but avoided juveniles growing to a size where they may begin to migrate from the nest (>30 d), allowing assessment of potential impacts on offspring growth.

Playbacks were broadcast from underwater loudspeakers (UW-30, frequency response 0.1–10 kHz, University Sound, Columbus, USA) connected to MP3 players (Sansa Clip+, SanDisk, Milpitas, CA, USA) via a 40 W amplifier (Kemo, Langen Germany). MP3 players, amplifiers and 12 V batteries were housed in waterproof boxes (Peli1200, Peli products, Barcelona, Spain) on the seabed. Loudspeakers were fixed by two metal poles at 1 m pointing towards a nest (Fig. S1). Occasional research station motorboat activity would have occurred within 100 m of the nests, although this would be balanced between the Ambient and Boat treatments as trials were running in parallel. Visual disturbance by motorboats is unlikely as visibility in the lagoon rarely exceeds 10 m. Our own research boat approached the study site slowly and anchored away from nests (>20 m) prior to observation periods.

Sound pressure and particle acceleration were recorded (using the recording equipment described above) during playback at treatment sites 1 m away from the speaker in the area of the water column that the young fish tended to use. Sound-pressure playbacks were measured at all 19 Ambient and 19 Boat sites but, due to accelerometer availability, it was only possible to measure particle motion at one site for each treatment. Illustrative power spectral densities are provided in Fig. 1 in the main paper. Other damselfish are known to hear frequencies up to around 1200 Hz [1]; we present acoustic data up to 2000 Hz as the hearing abilities of the focal species are currently unknown. Acoustic analyses were performed in MATLAB v2010a following the method described in [2]. The mean \pm standard error root-mean-square (RMS) sound-pressure level between 0 and 2000 Hz across 60 s samples was 108.1 ± 0.5 dB re 1 μ Pa at 1 m at the 19 Ambient sites and 128.7 ± 0.2 dB re 1 μ Pa at 1 m at the 19 Boat sites. Playbacks of motorboat noise could not be heard above local ambient sound from other nest sites (verified with the hydrophone described above).

Behavioural observations

The observer for the study (SLN) was initially trained in the standardised quantification of the three behaviours by a very experienced fish observer (MIM), who has conducted 100s of hours of behavioural observations on *A. polyacanthus* and trained numerous others in this regard. Videos were then made of 12 active *A. polyacanthus* nests that were not used in the experiment. After two videos were used to explain the behaviours, the repeatability of the behavioural scoring of defensive acts and feeding was tested by comparing scores by SLN with those from two other experienced observers who scored the videos independently. There were high levels of repeatability between observers for both defensive acts (intra-class correlation: 0.901, 95% confidence interval (CI): 0.744–0.972) and feeding (0.853; 95% CI: 0.637–0.957). Since glancing could not easily be observed from the video footage and the focal fish sometimes disappeared from the field of view, a decision was made to collect data in the experiment in real-time. A snorkeler has a wide-scale view (for defensive acts) whilst also being able to collect fine-scale accurate data (on feeding and glancing); instances where the focal fish was out of sight during the observation period were also greatly reduced compared to video recordings.

At Boat sites, observations were only made during playback of motorboat noise (as playbacks were intermittent; see above). All three relevant behaviours of the brood-guarding male were recorded during observations. Rare occasions where the focal fish briefly entered a cave or chased an intruder out of sight would not have affected the behavioural scoring as they would not be chasing, feeding or glancing inside a cave and if out of sight because of chasing an intruder the behaviour (chasing) was already known.

Fish community

Table S1: Table of predators observed within 5 m of *Acanthochromis polyacanthus* nests:

<i>Scolopsis bilineatus</i>
<i>Acanthochromis polyacanthus</i>
<i>Pomacentrus moluccensis</i>
<i>Pomacentrus wardi</i>
<i>Pomacentrus sp.</i>
<i>Pomacentrus philippinus</i>
<i>Pomacentrus grammorhynchus</i>
<i>Neoglyphidodon melas</i>
<i>Thalassoma lunare</i>
<i>Abudefduf sexfasciatus</i>
<i>Amblyglyphidodon curacao</i>
<i>Neoniphon sammara</i>
<i>Myripristis berndti</i>
<i>Plectorhinchus chrysotaenia</i>
<i>Plectorhinchus lineatus</i>
<i>Lutjanus russellii</i>
<i>Lethrinus obsoletus</i>
<i>Lethrinus argentimaculatus</i>
<i>Octopoda</i>
<i>Cephalopholis argus</i>
<i>Cephalopholis boenak</i>
<i>Cephalopholis microprion</i>

Table S2: Table of fish that were not predators in the community observed within 5 m of *Acanthochromis polyacanthus* nests:

<i>Neopomacentrus cyanomos</i>
<i>Stegastes fasciolatus</i>
<i>Chromis viridis</i>
<i>Chromis atripectoralis</i>
<i>Dichistodus perspicillatus</i>
<i>Dischistodus melanotus</i>
<i>Dischistodus pseudochrysopoecilus</i>
<i>Hemigymnus melapterus</i>
<i>Labridae sp.</i>
<i>Cheilinus chlorourus</i>
<i>Stegastes apicalis</i>
<i>Scaridae sp.</i>

<i>Chaetodon sp.</i>
<i>Chaetodon eureofasciatus</i>
<i>Chaetodon ephippium</i>
<i>Chaetodon vagabundus</i>
<i>Chaetodon auriga</i>
<i>Chaetodon aureofasciatus</i>
<i>Chaetodon lunulatus</i>
<i>Chaetodon lunula</i>
<i>Chaetodon plebeius</i>
<i>Chaetodon melannotus</i>
<i>Acanthuridae sp.</i>
<i>Caesio teres</i>
<i>Pterocaesio marri</i>
<i>Premnas biaculeatus</i>
<i>Amphiprion melanopus</i>
<i>Tetraodontidae</i>
<i>Pomacanthus sexstriatus</i>
<i>Balistidae</i>
<i>Apogon angustatus</i>
<i>Apogon properupta</i>
<i>Pomacentrus amboinensis</i>
<i>Pomacentrus simsiang</i>
<i>Plectropomus leopardus</i>
<i>Chrysiptera cyanea</i>
<i>Dascyllus aruanus</i>
<i>Epinephelus hexagonatus</i>
<i>Siganus doliatus</i>
<i>Siganus puellas</i>
<i>Teuthida</i>
<i>Zebrasoma veliferum</i>
<i>Mulloidichthys minicus</i>
<i>Lutjanus quinquelineatus</i>
<i>Dascyllus reticulatus</i>

<i>Mullidae sp.</i>
<i>Ostracion cubicus</i>
<i>Dasyatis kuhlii</i>
<i>Labroides dimidiatu</i>
<i>Corythoichthys sp.</i>
<i>Chaetodon auriga</i>
<i>Cephalopholis microprion</i>
<i>Apogon properupta</i>
<i>Chromis leucura</i>
<i>Cheilio Inermis</i>
<i>Meiacanthus grammistes</i>
<i>Sargocentron spiniferum</i>

1

2 **Statistical methods**

3 General linear mixed effects models

4 Any effects such as slight water quality variations between nests were controlled for by the
5 statistical models. The variance caused by and standard deviation of the variance for each
6 random effect are presented alongside the results of models. To establish the best-fitting model,
7 terms were eliminated one by one from a maximal model. Simplified models were compared
8 with more complex ones using maximum likelihood ratio tests that employ chi-square statistics
9 to establish whether a simpler model is significantly worse at explaining the data than a more
10 complex one. If a simpler model is not significantly worse when a term is removed, the simpler
11 model is deemed better and thus the term is dropped. If a simpler model is significantly worse,
12 the term is maintained in the model [3]. The degrees of freedom from maximum likelihood tests
13 presented in the Results of the main paper are the difference between the degrees of freedom of
14 the simpler and the more complex models. All potential interactions of fixed effects were
15 examined and are only presented where their exclusion from the model made the model
16 significantly worse at explaining the data at the level $p < 0.10$.

17 ANOSIM methods: A frequency matrix (species by nest) was created, the data were log-
18 transformed to reduce the influence of very abundant species, and Bray-Curtis similarity
19 coefficients between pairs of nests were computed [4]. The ANOSIM procedure was carried out
20 on the similarity matrix. ANOSIM generates an R statistic, which varies between 0 (similarities
21 within and between treatments are the same) and 1 (all samples within treatments are more
22 similar to each other than to any sample across treatments) and is tested for difference from
23 zero with a permutation test (in this study, $N=999$ permutations). A one-way ANOSIM was used
24 to compare fish communities among the two sound treatment types.



25

26 **Figure S1**

27 A photograph of one of the 38 experimental nests to illustrate the set-up; a male parent spiny
28 chromis (middle front) can be seen guarding a brood of offspring (swimming nearby). The
29 underwater loudspeaker used to expose this nest to the experimental playback can be seen
30 attached to two vertical poles. Photo credit: Sophie Nedelec.

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