

Policy instruments to control Amazon fires: a simulation approach

Abstract

Agricultural fires are a double-edged sword that allow for cost-efficient land management in the tropics but also cause accidental fires and emissions of carbon and pollutants. To control fires in Amazon, it is currently unclear whether policy-makers should prioritize commandand-control or incentive-based instruments such as REDD+. Aiming to generate knowledge about the relative merits of the two policy approaches, this paper presents a spatially-explicit agent-based model that simulates the causal effects of four policy instruments on intended and unintended fires. All instruments proved effective in overturning the predominance of highly profitable but risky fire-use and decreasing accidental fires, but none were free from imperfections. The performance of command-and-control proved highly sensitive to the spatial and social reach of enforcement. Side-effects of incentive-based instruments included a disproportionate increase in controlled fires and a reduced acceptance of conservation subsidies, caused by the prohibition of reckless fires, and also indirect deforestation. The instruments that were most effective in reducing deforestation were not the most effective in reducing fires and vice-versa, which suggests that the two goals cannot be achieved with a single policy intervention.

Keywords: Amazon, fire, environmental policy, agent-based modelling, land use.

1 Introduction

Fire is one of the major socio-environmental challenges facing the humid tropics, including the Amazon Basin. On the one hand, fire is an efficient tool to prepare, weed and fertilize land, and it provides profit and subsistence to a wide range of farmers from smallholders to large cattle ranchers. On the other, it is a source of escaped fires and atmospheric pollutants, a potential cause of soil degradation and a threat to rainforests, biodiversity and farmers' assets and health (Nepstad et al. 2001 and 2007, Mendonça et al., 2004, Chen et al., 2011, Carmenta et al., 2013). Furthermore, the likelihood of disastrous wildfires this century is increased by predicted climate and vegetation changes linked to a higher frequency of extreme droughts, (Malhi et al., 2009, Chen et al. 2011, Coe et al, 2013, Davidson et al, 2012). For example, in 1998 fires in the Brazilian Amazonian state of Roraima affected over 5 million hectares of forest (Cochrane, 2009, p.17), while 2015 was the hottest year in the amazon over the last century (Jiménez-Muñoz et al., 2016).

Preventing an increase in the prevalence of fires in the Amazon and elsewhere requires policies that internalize externalities (Malhi et al., 2009, Sorrensen, 2009, Carmenta et al, 2013). In Brazil, one of the federal government's main responses to the fire problem is the controlled burn law, which replicates the ban-surveillance-sanction approach that proved highly successful for deforestation. Yet, to-date, there is no assessment of the impact of this policy on fires. Interventions are also occurring at local scales, including incentive-based initiatives of payment for avoided deforestation and avoided forest degradation (REDD+), as well as municipal actions supporting mechanized land preparation substituting for slash-and-burn (Simões and Schmitz, 2000, Börner et al., 2007 and 2013, SEMA-AC, 2011).

The evaluation of impacts and limitations of command-and-control and incentive-based approaches to policy requires reliable empirical evidence. However, empirical work cannot provide definite answers without being guided by refutable hypotheses. This paper seeks to contribute with such hypotheses by developing an analytical device that represents the Amazon fire system both in the absence and presence of intervention. This is achievable with an agent-based, spatially-explicit simulation model. Policies, such as agricultural subsidies or payment ecosystem services (PES), aim to influence decisions with supra-individual consequences made by heterogeneous individuals. Policy interventions inevitably trigger a chain of connected processes whose net impact on the key state variables is not easy to intuit from pure reasoning without the support of an analytical tool. It is in this particular sense that a simulation model is useful (Zhao et al, 2012).

The use of spatially-explicit agent-based models for analyzing policies, especially their implications for land use change, is growing in the literature (Kremmydas, 2012, Zhao, 2012). Examples include changes of the German and Italian agricultural subsidies (Happe et al, 2008 Lobianco, 2007) and incentives to adopt water-saving irrigation techniques (Berger, 2001). The focus on agents allows incorporation of interactions between landscape processes and human decisions as well as heterogeneity among decision-makers. In an explicit modelled space, land parcels influence each other being thus subjected to spatial spill-overs that may be engendered by policy.

Two are the main reasons for adopting agent-based modelling. First, its bottom-up approach enables multiple possibilities of individual and collective reactions to policy, including those that would prevent desired outcomes from being achieved or would favor undesired results. Second, it generates results with a level of heterogeneity/variability which reasonably resembles the data available for policy evaluation. However, regarding this second reason, a clarification is needed. Part of the richness of the results is very hard to reduce to refutable hypotheses that may guide policy evaluation. As producing such hypotheses is one of our main goals, we opted for a causal inference approach to simulation analysis (Marshall and Galea, 2014). This means focussing on comparing policy outcome variables in baseline and policy scenarios, rather than exposing the plethora of patterns the variables describe across time and space. The model presented in this paper is a tool to build knowledge on the potential results of policy options to reduce Amazon fires. Due to the scarcity of knowledge on this topic, we opt to focus the modelling effort on detailing a few key components of the Amazon fire system, mainly farmer behavior and policy instruments, and incorporate other aspects in a rather stylized way. This approach strives to maximize the usefulness of the exercise for empirical work, because scant existing evidence (table 1) does not allow for testing of the intricate hypotheses that would be yielded by a more comprehensive model.

There is a further methodological reason for adopting a simple (or stylized) model. A clear trade-off exists between realism (the number and detail of real-world natural and social processes represented) and identification of causal effects (the confidence that observed variations in outcome variables are strictly due to variations in policy). Simulation models are different from models with analytic (pen-and-paper) solutions in that they do not necessarily yield identification. Non-linearity and stochasticity, coupled with endogeneization of most variables, makes it hard to track the causes of the observed behavior of the main variables (Marshall and Galea, 2014). This difficulty grows with realism (El-Sayed et al., 2012, Cederman and Giradin, 2007, Townsley and Birks, 2008). We opt first of all for causal effect identification and pay the cost of reduced realism by greatly simplifying the Amazon fire system. The main benefits are the clarity and the empirical refutability of the hypotheses about the impacts of policy that can be derived from the results.

The policy background is synthesized in the next section and the model is presented in section three. The results are analyzed and interpreted in section four, followed by a brief conclusion.

Table 1 [here]

2 Fire policy in Brazil

2.1 Brief overview

Policy interventions that affect Amazon fires include various initiatives that differ in terms of how directly they impact on fires, the level of government introducing the policy, the targeted social group and the type of policy instrument chosen. Here, we examine three key interventions. First, at the national level, the controlled burn law of 1998 regulates fire use by instituting licensing and monitoring (Brasil, 1998). It is a command-and-control instrument

against agricultural fires that have a high probability of turning into uncontrolled fires and causing major damage (Brasil, 1998, Steil, 2009). However, in practice, permit granting is marginal (Toniolo, 2008, p.193-194, Carmenta et al., 2013, Cammelli, 2014, p. 13, Costa, 2004, p. 184), enforcement is rare (IBAMA-PA, 2015) and recent fieldwork¹ indicated that few state and local governments execute these functions. The main barriers for the farmers are the transaction costs of obtaining the documents demanded by permit requisition, especially the proof of land ownership, travelling often long distances from farms to environmental offices in urban areas (Carmenta et al., 2013, Cammelli, 2014, p.48).

Second, subsidies have been used to reduce fire and offer different routes for promoting the technological transition of smallholders to fire-free agriculture; mechanization and agroforestry. These include subsidies for mechanized land preparation offered by some municipal governments, generally together with extra financial support for agricultural inputs (Börner et al, 2007, Emater, 2015b, Simões and Schmitz, 2000). Alternatively, pilot projects are used to stimulate agroforestry systems, which combine trees, crops and animals in the same plot without resort to fire or inputs. The agroforestry pathway tends to be funded by NGOs and public institutions, and is advocated as "greener" and more sustainable than mechanization (Serra, 2005, Arco-verde, 2008, MMA, 2009). However, progress on these fronts tends to be inhibited by constraints facing the targeted farmers, including lack of access to capital and credit, labor, inputs and rural extension services. Two of these constraints are critical for the shift to agroforestry. First, labor, since agroforestry requires a higher working effort, at least initially (Arco-verde, 2008, p.93). Second, credit, as public and private banks still lack a standardized methodology to estimate the profitability of agroforestry systems with sufficient certainty (Emater, 2015a, Kato, 2015).

Third, PES represent incentive-based instruments to reduce fires. An exemplar scheme was the *Proambiente* program, based on payment for avoided deforestation and multiple related ecosystem services including reduced wildfire risk. This program provided payment and technical support between 2004 and 2008 to enable four thousand smallholder households across the nine states of Legal Amazon to adopt fire-free practices (Hall, 2008, Neto, 2008, p.20, Wunder et al., 2009). Another example is the *Bolsa Floresta* program, which transfers cash to forest-dependent communities conditional on forest conservation and avoided carbon emissions. In some cases, transfers are conditional also on the control and reduction of fire use (Börner et al., 2013). However, at present PES programs in the Amazon are restricted to a few projects with only localized impacts. This contrasts with the growing number of papers arguing that incentive-based instruments are the best way to conserve tropical forests and control fires (e.g. through payments for reduced emissions from deforestation and degradation (REDD+; Barlow et al, 2012; Aragão and Shimabukuro, 2010).

2.2 Simulated policy instruments

¹ In April 2014 and March-April 2015 meetings and interviews with key stakeholders were conducted comprising national coordination of PREVFOGO and also Pará state headquarters, a short interview with Pará state institution on environmental surveillance (IBAMA-PA) and Pará state institutions on agricultural research (EMBRAPA CPATU) and rural extension (EMATER).

The model simulates the impacts of simplified representations of three of the classes of policy instruments, namely the controlled burn law (i.e. command-and-control), a subsidy scheme for transition to fire-free agriculture through mechanization and PES schemes (technical details are found in appendices A and B).

A command-and-control (C&C) instrument was simulated by assuming that the environmental authority bans and sanctions only "reckless fires", i.e., agricultural fires with high probability of running out of control and turning into accidental fires. Accounting for the low spatial resolution (1 km² cells) of remote-sensing fire detections that fire monitoring by the Brazilian government is based on (Vasconcelos et al., 2013, INPE, 2015, PREVFOGO, 2015), the modelled landscape used in our simulations was divided into "monitoring zones" of 1 km². Monitoring can only effectively identify fire-users where a zone intersects a single farm (as opposed to parts of multiple farms occupying the same zone). Reckless fires detected in single-farm zones (herein, enforcement-effective zones) generate a fine of fixed value per hectare which is applied to the owner. It is assumed that farmers know perfectly in which zones enforcement is effective.

Second, subsidy schemes, also referred as "incentive-based instruments", are represented as voluntary contracts of three modalities. Each modality targets the promotion of a specific mix of land use and technology (LUT) that can be developed in the parcels in which landscape is subdivided. This includes not only PES schemes but also the subsidy to fire-free agriculture (table 2). The agroforestry route is not modelled for consisting in complex mixes of crops, trees and cattle which take highly heterogeneous and mostly experimental forms in the literature (Serra, 2005, Arco-verde, 2008, MMA, 2009).

The total annual subsidy received by a farmer is the product of the number of parcels with the target activity by a fixed per-hectare basis subsidy. All contracts have a five-year lifetime and are renewable indefinitely. Payment of subsidies occurs every year and is conditional on the compliance of contractual norms (table 2). If any norm is violated, the farmer must return all annual payments received since the start of the current contract. This stiff penalty of early contract termination is employed to assure time-consistency (Gulati and Vercammen, 2006).

Table 2 [here]

3 The model

3.1 Presentation strategy

Model presentation follows the "Overview, Design, Concepts and Details + Decision" (ODD+D) protocol for description of agent-based models proposed by Müeller et al (2013) and the structure adopted by Arfaoui et al (2014). It is intended that the findings of this paper could be relevant to researchers without formal background on computer modelling, and, also to practitioners, including policy-makers and NGOs. To better communicate with the target audience, the body of the text includes only the details essential for understanding results, limitations and conclusions. We first present the model from a conceptual perspective, emphasizing its goals, main features and theoretical foundations. The operational perspective,

i.e. the translation of concepts into an implementable procedure of algorithms and equations, is left to appendices and supplementary material. This simplified presentation strategy adheres to the fundamental ODD+D principle of gradually introducing the reader to model details.

3.2 Description of the model

Here we present a brief summary of the detailed description found in appendices A and B.

3.2.1 Purpose

The model is a tool to build knowledge on the impacts and limitations of command-andcontrol and incentive-based policy instruments designed to control agricultural and accidental fires in the Brazilian Amazon.

3.2.2 Entities, state variables and scales

Four kinds of entities populate the model: farmers (decision-makers), parcels (spatial units), government and nature.

There are four combinations of land use and technology (LUTs, herein) a land parcel can be allocated with. Three of them are agricultural land uses, namely, agriculture based on "controlled fires" (hereafter "controlled fires"), agriculture based on "reckless fires" and "fire-free agriculture" (herein "fire-free"). Reckless fires are conducted without any measure of control such as firebreak construction, burning against the wind or the avoidance of dry periods of the year. Controlled fires take place with the proper control measures, and fire-free agriculture without the use of fire. The remaining land use is forest. Parcels with agricultural LUTs can be converted to forest. In the year when the conversion is made, the parcel's forest age is set to zero. Therefore, the amount of above-ground forest biomass accumulated in the parcel becomes positive only one year after conversion, which incorporates the delayed regeneration observed in practice (Neeff and Santos, 2005).

Farmers decide the LUT portfolio that prevails on a particular set of parcels (farm). They are characterized by their (i) farm, i.e., the set of parcels controlled, (ii) wealth, (iii) accumulated local data on LUTs and accidental fires, (iv) point estimates for parameters behind the probability of accidental fires which is herein referred to as "risk" (see 3.2.3 below) and (v) status regarding subsidy contracts.

Parcels are the basic units of space. They do not make decisions, but execute natural processes of forest growth and forest degradation by fire. Parcels are characterized by their (i) location, (ii) physical suitability (for agriculture), (iii) LUT, (iv) age of forest, (v) total aboveground forest biomass (tons/hectare) and (vi) inclusion in an enforcement-effective zone. Attribute (ii) was calculated from GIS data, and is a metric for the contribution of three locational factors to parcel-level profit: parcel slope and Euclidian distances to nearest road and municipal capital respectively (SM.1). Nature is an observer or "higher-level controller" (Grimm et al, 2010) that decides which parcels are affected by accidental fires at each time step. It is the only entity that knows the true risk parameters and also the values of the random component of parcels' risk (see "define-burned-parcels in next subsection an on appendix B). The government is also an observer. It sanctions reckless fires in the C&C simulations and also offers and monitors voluntary contracts in the incentive-based simulations.

Regarding scale, one time step represents one year, simulations were run for 40 years, one grid cell represents 1 hectare and the model landscape comprises 100 x 100 hectares, i.e., 100 km^2 .

3.2.3 Process overview and scheduling

There are three classes of simulations. Baseline simulations, in which no policy instrument is active, simulations in which only the C&C instrument is active and simulations in which only incentive-based instruments are active. The last class subdivides into three, each comprising a particular kind of subsidy contract (section 2.2 above).

In baseline and C&C simulations, nine modules are processed in the following order: calculate-profit, implement-LUT-portfolio, LUC-cost-account, define-burned-parcels, sanction-rule-breakers, calculate-actual-profit, update-risk-parameters, store-LUT-portfolio-in-memory, and update-forest-age-after-burn. In incentive-based simulations, the same procedures are processed together with two extra procedures: (i) subsidy-payment-account which is deployed right after "implement-LUT-portfolio" and; (ii) update-contract-duration, deployed right after "LUC-cost-account".

The main features of the modules are described below

Calculate-profit (figure 1) defines the LUT portfolio to be implemented in each time step using whole-farm expected profit optimization. Instead of seeking the best LUT portfolio globally, the algorithm identifies the best LUTs for each parcel locally by taking as given the best LUTs of neighboring parcels (queen criteria of contiguity was adopted). In other words, it makes assumptions about neighboring best LUTs. Since this profit-calculation is done sequentially for all parcels, the best LUTs of some neighboring parcels may not yet be defined (i.e., remain unknown) at the stage where the best LUT of a given parcel is to be defined. To mitigate against this, identification of best LUTs for all parcels is iterated until it stops yielding an increase in whole-farm expected profit. Due to limited wealth and the costs of changing between land-uses, farmers prioritize parcels with the highest degree of physical suitability. In incentive-based simulations, "calculate-profit" is subdivided into two sub-procedures (route B of figure 1); one that imposes compliance to contract norms ("restricted identification") and one that does not ("unrestricted identification"). Both these sub-procedures are executed at every step, generating the information that forms the basis of farmers' decisions on voluntary contracts.

Define-burned-parcels defines the parcels that accidentally burn and is executed by nature. The probability that an accidental fire occurs on a given parcel, also called "risk", is a function of two classes of factors. First, variables indicating the parcel's own and neighboring parcels' LUTs. Second, fixed 'risk parameters' that measure the intensity with which LUTs influence risk. Whether an accidental fire occurs in a parcel is determined by the probability level (risk) and by a standard Gaussian disturbance representing non-observables behind risk. Further details, including the functional form linking LUTs with risk, are found on appendix B, section B.8 (with the complete description of the functional form detailed in appendix C. It is helpful to clarify that fire spread is not modelled. Accidental fire is conceived, for simplicity, as a point event, completely restricted to the parcel where it takes place. Parcels accidentally burned generate zero actual profit. For simplicity, it assumed that the above-ground biomass of forested parcels is fully eliminated by accidental fires (defined as "functional deforestation" by Barlow et al., 2012).

Update-risk-parameters is executed by farmers. Such entities are ignorant of true risk parameters but are able to estimate them. For this, they apply a statistical routine to accumulated local data on LUTs and accidental burns. Only their own parcels and parcels within 100 m of farm boundaries are observed by the farmers. At each time step, parameters are re-estimated after data update.

Implement-LUT-portfolio assigns best LUTs for parcels processed by calculate-profit and previous LUTs for remaining parcels. This is preceded, in incentive-based simulations, by the choice between the LUT portfolio designed to comply with contract norms and the unrestricted LUT portfolio. The choice criterion is to pick the option with the highest whole-farm expected profit.

Update-contract-duration is only part of incentive-based simulations. It defines contract status in the basis of (i) LUT portfolio choice made in the previous procedure and (ii) current contract duration. The possible statuses or actions towards contracts are sign or don't sign, keep or break, and renewal or exit

3.2.4 Design concepts

Theoretical and empirical background: profit and risk spillovers

In new economic geography, the allocation of land for alternative uses is driven by "agglomerative" or "attractive" forces and "dispersive" or "repelling" forces (Fujita and Thisse, 2002, p.5, Irwin and Bockstael, 2001, Krugman, 1996). This principle is the basis of LUT choice in the model in our ABM. The forces that drive the agglomeration of agricultural LUTs are positive externalities associated with scale economies (Table 3). They affect LUT choice through the channel of profit (as letter "P" indicates), except for one of the forces that favor forest agglomeration, whose impact passes through the channel of risk (letter "R").

There are (at least) three categories of ecosystem services provided by forests. First, services which support food production, such as water and nutrient supply, soil conservation and climate regulation (Klemick, 2008, Chomitz and Kumari, 1996), as well as pollination and pest control (Tscharntke, 2005). Second, forested land also supports surrounding forested land, through avoiding or reducing edge and isolation effects (Laurance et al, 2006, Ferraz et

al, 2003, Stouffer and Bierregaard, 1995). It therefore seems valid to assume that availability of forest resources) in a forested unit of land increase with the amount of forest in the vicinity. Third, surrounding forests mitigate risk (Brando et al., 2013). The magnitudes in which forest services are provided, and the profit made out of forest products, are assumed to be positively correlated with above-ground biomass of the forest accumulated in the source parcel.

It is assumed that risk increases with the agglomeration of fire-based agriculture. Since accidental fires impose economic losses to farmers, this assumption creates a force that favors dispersion of fire-based agriculture (table 3; better detailed in "expected profit function subsection" below).

Table 3 [here]

Empirical data

The land property structure, initial location of forest and non-forested parcels and values for physical suitability come from empirical GIS data describing the condition in 2010 of a 10 x 10 km square sub-area of Santarém municipality, in the Brazilian Amazon converted into a 100 x 100 cell digital grid (figure 2, details on SM.1). Physical suitability is measured in terms of slope of terrain, distance to roads and urban centers. Land use change cost was estimated from secondary data (SM.2). Forest-growth follows the empirical above-biomass (logistic) growth function estimated by Neeff and Santos (2005).

Figure 1 Flowchart for calculate-profit, baseline and command-and-control simulations (route A) and incentive-based simulations (route B). The agent executing the procedure is indicated with [P] for parcel and [F] for farmer. Procedures are detailed in appendix B.

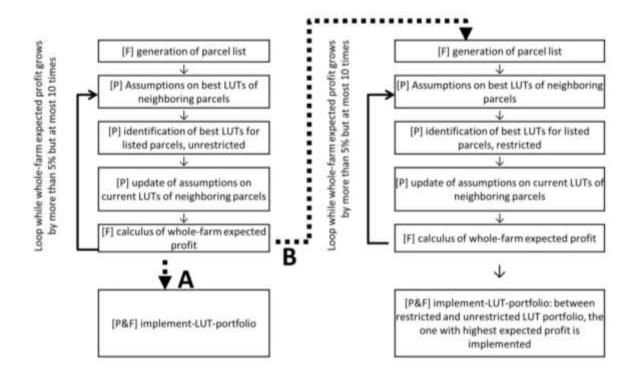
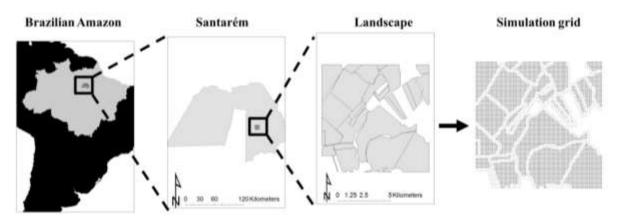


Figure 2 The modelled landscape in the Brazilian Amazon and the gridded version used in simulations



Individual Decision Making: statement of the decision problem

The problem solved each time step by the i-th farmer is:

$$\operatorname{Max}_{\{\tau_k\}_{k=1}^{K_i}} \sum_{k=1}^{K_i} r^e (\eta_k, \tau_k, \tau'_k, \tau^e_k, \hat{\beta}, \gamma) \, \text{s.t.} \sum_{k=1}^{K_i} C(\tau_k, \tau_{k,t-1}) \le W, \hat{\beta} = f(D)$$

Where parcels are indicated by k and the following definitions apply: $r^{e}(.) \equiv$ expected profit function, $\eta_{k} \equiv$ physical suitability, $\tau_{k} \equiv$ k-th parcel's LUT, $\tau_{k}' \equiv$ vector with LUTs of neighboring parcels owned by the i-th farmer, $\tau_{k}^{e} \equiv$ vector with forecasted LUTs of neighboring parcels owned by other farmers, $\hat{\beta} \equiv$ vector with estimated parameters for the accidental fire prediction model (or "risk model"), $\gamma \equiv$ vector with policy variables including LUT restrictions and magnitudes of subsidies and fines, C(.) \equiv LUT change cost (computed by "LUC-cost-account" procedure), $\tau_{k,t-1} \equiv$ previous LUT, W \equiv wealth, f(.) \equiv function representing the estimation of $\hat{\beta}$, D \equiv current data on observed accidental fires (automatically updated; see "update-contract-duration" above).

In a nutshell, farmers choose the LUT portfolio that yields the highest farm-level expected profit, given the norms of prevailing policy (γ), knowledge on the likelihood (risk) of accidental fire occurrence ($\hat{\beta}$), forecasts for neighbors LUT choices (τ_k^e), and the constraints imposed by the wealth (W) and the data on accidental fires (D) that could be accumulated. This optimization problem is not solved once-for-all but repeated at each time step, with the objective function being increased gradually by taking advantage of wealth and data accumulation - as current neighbors' LUTs are forecasted to equal previous period LUTs (table B.1), forecasting may also induce portfolio change. Also, the solution is based in an algorithm (calculate-profit) that proceeds from parcel-level LUTs to farm-level portfolio, cutting through the complexity of profit and risk spill-overs and the implied parcel-level spatial dependence. The algorithm also addressed the need to forecast LUT choices of farmers that control boundary parcels.

The model reduces to a computational implementation of the problem just described. Results generated are a set of solutions to the problem for all agents at all time steps.

Individual Decision Making: solution of the decision problem

The farmer's decision problem is solved with the "calculate-profit" procedure (figure 1, section 3.2.3) which seeks to represent Amazonian farmers' decision-making. Multiple studies attest the influence of capital on land use decisions, here called "wealth", and of the parcel-scale factors determining physical suitability, being them slope of the terrain and distance to roads and urban centers (Deadman et al., 2004, Sorrensen, 2000 and 2004, Moran et al., 2002, Scatena et al., 1996, McCraken et al. 2002). In particular, the wealth allocation principle of giving priority to parcels whose costly conversion is, due to physical factors, more profitable, is supported by empirical evidence that proximity to roads, urban centers and flat terrain have positive effects on deforestation (Pfaff, 1999, Pfaff et al., 2007).

The procedure is also designed as a bounded rationality shortcut to the search for the best among all possible portfolios, which amount to a number of alternatives whose order of magnitude is of 10¹⁸ for a 30-hectare farm, the smallest size considered. Two approaches to land use economics are reconciled by the LUT choice algorithm. The multi-output approach (e.g., Just et al, 1983, Fezzi and Bateman, 2011), for which farmers' choice is guided by whole-farm profit, and the recent spatially-explicit models (e.g., Irwin and Bockstael, 2001, Parker and Meretsky, 2004), which emphasize heterogeneity and spatial externalities at parcel-level.

Expected profit function

Parcel-level expected profit function is $r^e = \eta \theta(1-p) - C(.) + S(\gamma)$, with $\eta \equiv$ physical suitability, $\theta \equiv$ deterministic profit, $p \equiv$ probability of the parcel to be accidentally burned or simply "risk", C(.) = LUT change cost, $S(\gamma) =$ fine (negative value) or subsidy (positive value) assigned by policy. The main forces driving LUT choice in the model, the agglomerative and dispersive forces (table 3), are captured by two components, θ and p. To simplify language, only the product $\theta(1-p)$ is hereafter referred as "expected profit function". The arguments of θ and p are metrics for three classes of variables, (i) own-parcel LUT, (ii) neighboring parcels' LUTs, (iii) forest biomass either (iii.a) in the parcel (if occupied with forest) or (iii.b) in neighboring parcels. Details on the metrics and how they enter the functional forms of θ and p are provided in appendix C. Agglomerative and dispersive forces captured by θ are indicated with "[θ]" in table 3 and those captures by p with "[p]". Expected profit functions are shown in figure 3. Farmers optimize an average of these functions weighted by physical suitability. For non-forest LUTs, the horizontal axis measures the degree of agglomeration in terms of neighboring parcels² with the same LUT as the reference parcel. The higher the average forest biomass in neighboring parcels, the higher the curve. For the case of forest, the agglomeration degree is measured as neighboring parcels' average above-ground biomass of forest. Higher curves represent higher levels of parcel's own forest biomass. As the figure shows, expected profit functions are concave as agglomeration has

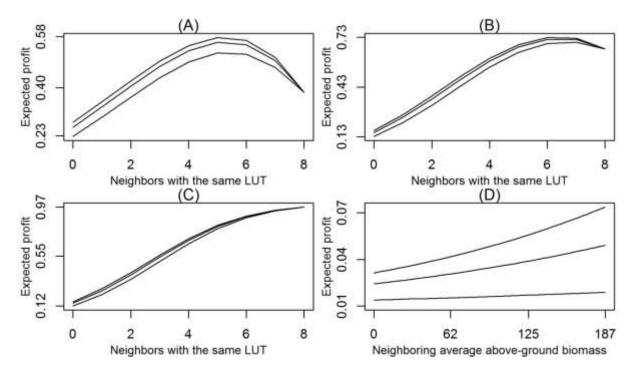
² With squared parcels, queen neighborhoods are made of 3, 5 or 8 parcels.

decreasing returns for all non-forest LUTs, what is in line with standard assumptions in economics (see, for instance, Mas-Colell et al., 1995, p.133-137, Varian, 1992, section 2.1, Doole and Kingwell, 2015, 2.1). This stems from the positive effect of agglomeration on both deterministic profit and risk.

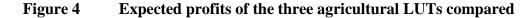
For the case of forest, agglomeration has opposite effects, positive for deterministic profit and negative for risk, yielding, thus, increasing returns. However, the expected profit of forest is concave in forest age, or equivalently, in own-parcel biomass, what stems from the logistic function driving forest growth (obtained from Neeff and Santos, 2002). Such concavity is attested by the bottom left of figure 3, where there is a decreasing distance in between two subsequent curves. Such is also the case for non-forest land uses. Consequently, the expected profits of all LUTs are concave in neighboring average forest biomass.

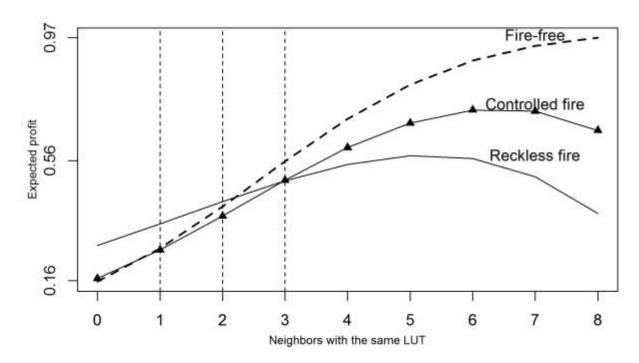
Concavity is merely an expression, in the form of expected profit, of the more general idea of agglomerative and dispersive forces, in particular the forces specified in table 3. The optimal agglomeration level is the lowest for reckless fire as such LUT, by definition, has higher probabilities of turning into an accidental fire, for each agglomeration level. Also influent in such respect is a low level of scale economy (table 3). In contrast, fire-free has the highest optimal agglomeration level due to zero reliance on fire, and thus, low exposure to accidental fires and also to a high level of scale economy (table 3). All non-forest LUTs are supported by forest services, what explains the positive effect of surrounding forest biomass in their expected profits. Forest is subjected to returns from agglomeration as the process increases its capacity to generate services that provide self-support (table 3). The low expected profit forest is assumed to generate accounts for the still relevant deforestation level (~5,000km²/year, Godar et al., 2014) suggesting that forest is still seen as secondary in terms of economic return.

Figure 3 Expected profit of the four LUTs as functions of agglomeration level (reckless fire (A), controlled fire (B), fire-free (C), forest (D))



Note: shifted curves correspond, from down to up, to the following levels of average forest biomass in the neighboring parcels: 56 (age of 11, inflection point of forest growth function, Neeff and Santos, 2002), 144.25 (age of 25), 192.71 (age of 50).





Learning

Farmers learn about true risk parameters by re-estimating them every time step from accumulated local information on LUT and accidental fires.

Interactions

Farmers interact among themselves only indirectly, mediated by parcels, and locally.

3.2.5 Initialization

The initial condition includes 26 farmers with heterogeneous farms. Each farmer has a set of estimated risk parameters and a level of wealth. In C&C simulations, parcels may also differ with regard to inclusion in enforcement-effective zones. All simulations, of all three kinds, have the same initial conditions, which are generated from GIS data on land property and forested and non-forested parcels.

Clarifications on LUT assignment are needed. Initial LUTs are assigned to parcels before simulations were run by a landscape generating code which is separate from the model simulation code. This assures that all simulations depart from the same landscape, i.e., from the same LUTs for each of all ten thousand parcels. Initial LUTs are attributed first as forested or non-forested on the basis of the 2010 land use map of the Brazilian Amazon, developed by INPE-EMBRAPA (2012). In a second step, non-forested parcels are randomly assigned, with equal probability, to one of the three agricultural LUTs or zero-age forest (freshly abandoned land). Random assignment is inevitable as available remotely-sensed data does not allow for distinguishing the three forms of fire here considered namely, reckless, controlled and accidental. Satellite-derived data on fire hotspots (MODIS) only register point or area fire detections without any information on underlying land use or fire control measures (see user guides on UMD, 2016). Additionally, it is also not possible to precisely identify fire-free agriculture in INPE-EMBRAPA (2012). The assignment with uniform probability creates a checkerboard of non-forest LUTs that corresponds to the lowest degree of agglomeration. This gives opportunity for policy instruments to work, as higher degrees of agglomeration would restrict LUT change.

3.2.6. Input Data

The model does not use input data to represent time-varying processes.

3.2.7 Submodels

See appendix B.

4 Results and discussion

This section presents the criteria for analysing results and the analysis itself. Five potential lessons are proposed as hypotheses to guide empirical research. Three of them refer to the impact of the policy instruments, and the four remaining comprehend undesired side-effects. It is detailed how potential lessons stem from simulation results by showing patterns described by the main variables.

4.1 Approach for analysing results

4.1.1 General approach

The impact of policy instruments is conceived as a causal effect (in the sense of Morgan and Winship, 2007, chap.2), at landscape-level, on a set of outcome variables. It can be trivially calculated because the policy simulations (treatment states) differ from the baseline simulation (control state) strictly due to the presence of policy instruments. All exogenous variables, except those that characterize instruments, take the same values across all simulations and the few variables randomly assigned make no difference in outcomes³. Therefore, any difference in endogenous variables found by comparing baseline and policy simulations must therefore be the result of policy interventions.

Additionally, only one voluntary contract is available in a given incentive-based simulation and none of them are available in simulations of command-and-control policy. This "singleinstrument" approach allows for capturing the individual effect of each instrument rather than the mixed effect of multiple instruments.

The dynamics of the model requires a decision on the time window to be taken as the basis for calculating causal effects. Baseline and policy simulations may differ when compared step by step and thus the causal effect could also be calculated step-wise which would yield a short-run appreciation. But this study focusses on the long-run causal effects, which capture the net result, on each outcome variable, of direct and indirect effects of policy. The option for the long-run is in line with the literature on dynamic economic modelling of environmental policy (e.g., Van der Werf and Di Maria, 2012).

The long run is assumed to start when the change in aggregate actual profit and land use become negligible (figure 5 below). However, absolute stagnation was never observed with negligible growth prevailing even after a large number (500) of iterations. The long run is assumed to start when aggregate profit has grown for less than 5% in the last five consecutive steps. All simulations reach that point at t = 40 at the latest (but generally well before that, see figure 5). This is, therefore, the reference step for computing long-run causal effects. The rationale for basing analysis on the long run relates to the fact that agents' best responses to policy instruments are observable only after LUT portfolios were optimized. The latter is indicated by profit and landscape stagnation, given the gradual improvement approach farmers follow (section 3.2.4 above). It should also be highlighted that it is in such period that constraints are less stringent, being them either wealth or data on accidental fires, and then response to policy becomes mainly a matter of choice.

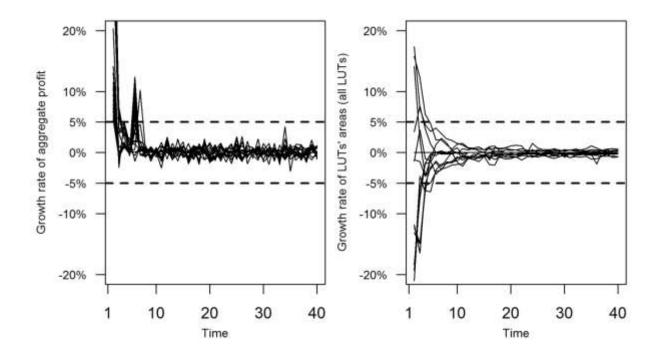
The policy instruments can be implemented in multiple "intensity levels", as varying magnitudes of the fine and the subsidies. The set of intensity levels considered in simulations is $L = \{0.1, 0.2, ..., 1\}$. All monetary values in the model are expressed as shares of the (landscape-wide) maximum parcel-level profit. Thus, a fine (subsidy) of \$0.1/hectare reduces (increases) the profit yielded by a parcel in exactly 10% of the maximum parcel-level profit.

³ There is only one class of exogenous variables assigned with a random number generator, the initial guesses for the eight parameters of the fire-risk model, which are the same, for a given farmer, in all simulations. The random component of accidental fires has zero, and, therefore, zero effect across the landscape.

In summary, the long run causal effect on the v-th outcome variable of the p-th instrument implemented in the l-th intensity level is $\delta_v^{p,l}(t = 40) = W_v^0(t = 40) - W_v^{p,l}(t = 40)$, with the superscript "0" indicating the baseline simulation and W the level of the outcome variable. The three outcome variables considered are counts of parcels with a particular type of fire among (i) accidental fires, (ii) reckless fires and (iii) agricultural fires (reckless and controlled fires). Conclusively, the long run causal effects capture avoided fires of the three kinds detailed.

Two of the instruments, C&C and mechanization, have a reach which is limited, in the spatial and social dimensions respectively. In contrast, all farmers were exposed to the two conservation instruments. To tackle different treatment groups, results are also presented for smallholders (farm size not above 200 hectares, which is the limit for mechanization) and medium-to-large landholders (farm area above 200 hectares; also denoted as "medium-to-large").

Figure 5 Stagnation of aggregate profit (left) and landscape (right) before the long run (t = 40), all simulations.



4.1.2 Sensitivity tests

The robustness of the results to risk and deterministic profit parameters was assessed by introducing percent shocks of -50%, -25%, 25%, 50% and 100% to parameters. A sensitivity simulation is characterized by a parametric vector pair given by $\{\alpha + \Delta \alpha, \beta + \Delta \beta\}$ where α and β are, respectively, the parametric vectors for the deterministic profit and risk and Δ is the percent shock. Changing the ordering of parameters, would change the expected profit

ordering of LUTs, leading to simulations that are excessively different from the one whose results are evaluated in the main text (next subsection). To avoid this, all parameters of a class (risk or LUT) receive the same percent shock. The results are found in SM.3.

4.2 Short-run dynamics

The short-run dynamics of the model is synthesized by the trend of a single workhorse state variable, the average number of neighboring parcels with a given LUT. This measure for the agglomeration level determines the expected profits of the LUTs (section 3.2.4) and, therefore, the long-run LUT portfolios. The short run dynamics is a history of a race that is won in the very beginning (figure 6). The initial condition allocates the four LUTs randomly with uniform probability to non-forested parcels, resulting in an average agglomeration level of 1.10 parcels for agricultural LUTs⁴. For such agglomeration level, the LUT that yields the highest expected profit quickly becomes dominant in the baseline simulation (around t = 5) and remains so in the long run. Such is the case of reckless fire, which is the most profitable LUT up to an agglomeration level of 2 parcels (figure 4).

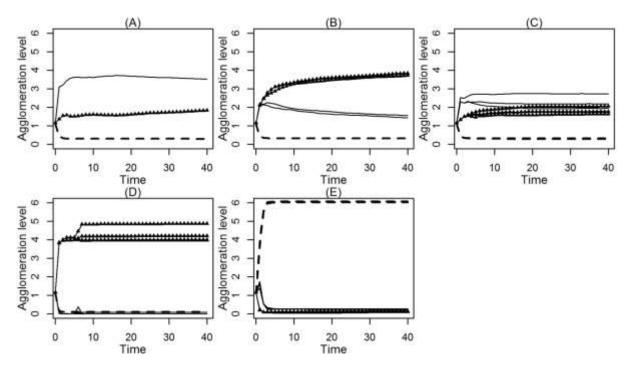
In the baseline, it is observed a feedback in which the agglomeration and spatial diffusion of reckless fire reinforce each other. Such feedback is broken by policy instruments right in the first time step. Fines and subsidies work as exogenous shocks on the expected profit yielded by LUTs, overturning the advantage of reckless fire and opening space for the other two agricultural LUTs. In all policy simulations except Mechanization, which directly incentivizes fire-free, it is controlled fire that dominates in the long run (figure 6). Conservation, the only instrument without a restriction or subsidy against reckless fire, proved to be the least effective in reverting the LUT's dominance.

The baseline trend is consistent with studies advocating the existence of a self-sustaining dominance of fire use (fire lock-in) in the Brazilian Amazon, with emphasis in the dependence of smallholders on slash-and-burn agriculture (Costa, 2004, Nepstad et al, 2001, section 2, Cammelli, 2014, section 4.2.3). There are also widespread claims in the literature that, as our policy simulations show, intervention is needed in order to break the lock-in. This belief, adhering to the originally proposition of technological lock-in (Arthur, 1989), is also confirmed by research on fire use. In Börner's et al (2007) simulations, policies promoting fire-free technologies and also taxing slash-and-burn proved successful. More recently, among smallholders participating in a PES program in Amazonas state, Börner et al (2013, p.56), found weak evidence of a reduction in deforestation, which probably means that the rate of expansion of slashed and burned area decreased. Van Vliet et al (2013) argue that policy based on conditional cash transfers has restrained shifting cultivation.

A novelty of our paper is the process driving the agricultural fire lock-in, which connects the initial land use condition (an historical event in the sense of Arthur, 1982) with agglomeration induced by scale economies (spatially-explicit processes which create increasing returns also in line with Arthur, 1989).

⁴ If forested parcels were not dominant in the initial condition, the number would be around 2 parcels.

Figure 6 Agglomeration level of agricultural LUTs, all simulations (baseline (A), C&C (B), Conservation (C), Conservation+ (D), Mechanization*(E))



Caption: solid line = reckless fires, solid line with triangles = controlled fires, dashed line = fire-free agriculture. Agglomeration level in the vertical axis. Note: For mechanization, only parcels belonging to smallholdings (the instruments' target) were considered.

4.3 Analysis of results

Five main "potential lessons" synthesizing what could be learnt from the results are here presented and discussed. By "lessons" it is not meant recommendations to be put in practice but rather hypotheses about the impacts and limits of policy instruments whose validity is yet to be tested using empirical data. These potential lessons are robust to multiple values of parameters, as attested by sensitivity analysis (SM.3).

4.3.1 Relative effectiveness of instruments

The C&C instrument proved less effective to contain fires than incentive-based instruments (figure 7). This is mainly because C&C has a limited *de facto* spatial reach due to imperfect monitoring. It should not be understood as a proof of ineffectiveness. In fact, for all fine levels, the probabilities with which reckless and accidental fires occur were significantly lower in pixels belonging to enforcement-effective zones (p-value < 0.05 on the permutation test proposed by Röhmel, 1996). Contrariwise, total fires (either controlled or reckless) occurred with significantly higher probability within enforcement-effective zones (p-value < 0.05). This last result shows, just like figure 6, that the ban of reckless fires may increase controlled fires. Additionally, while enforcement-effective zones are circumscribed to 21% of the landscape, partially intercepting only 9 of the 14 medium-to-large-sized farms and no smallholdings, the coverage of conservation contracts is above 70% of the landscape.

Effectively, the coarse spatial resolution of fire-monitoring frees smallholders from costly fines. Smallholders are only impacted by the C&C instrument indirectly, through spatial spill-overs of the reactions of medium-to-large landholders, an effect with negligible magnitude (figure 7). This yields the first potential lesson (PL).

(PL 1) If fire monitoring is based on remote sensing with low spatial resolution, banning reckless fire may exert low impact on accidental fires and on total fire use. Moreover, smallholders may remain unexposed to enforcement.

This finding echoes that of Godar et al (2014), who show that smallholders are the social group least impacted by Brazilian deforestation policy due to the limited spatial resolution of deforestation monitoring and the political acceptability and cost-effectiveness. Additionally, Assunção et al (2013) argue that higher resolution monitoring would increase the effectiveness of deforestation policy. Finally, Börner et al (2015) found no statistically significant impact of field-based enforcement on deforestation of patches below 25 hectares, which is the current resolution of real time monitoring of deforestation in Brazil.

In this study the mechanization subsidy performed best for all three outcome variables (figure 7), considering only the social group exposed to it, namely, smallholders. Stimulating farmers to stop using fire, whether in a reckless or controlled manner, seems to be the most effective path to reduce accidental fires and, obviously, total fire use. Only in avoiding reckless fires, mechanization is matched by any other instrument, in particular conservation+. The latter is the only other incentive instrument that restricts fire use but, in contrast with mechanization incentives, it does not compensate for compliance. Consequently, we conclude that:

(PL 2) Subsidies to shift from fire-based to mechanized agriculture may prove more effective in reducing fires than either C&C or conservation instruments.

Effectiveness of the incentive was proportional to how directly it impacted fire use, as revealed by the impact rank of the three incentive instruments, which, for most of outcome variables and intensity levels, was, in decreasing order, mechanization, conservation+ and conservation (considering only groups exposed to the instrument). Conservation is the least direct instrument due to the absence of clauses regulating fire use, whereas conservation+ prevents reckless fire use only by incentivizing farmers to avoid converting 10 year old forest to reckless fires. Mechanization is the most direct instrument. Such results are compatible with the claim by Ferraro and Kiss (2002) that payments are most effective whether directly remunerating the desired environmental benefit.

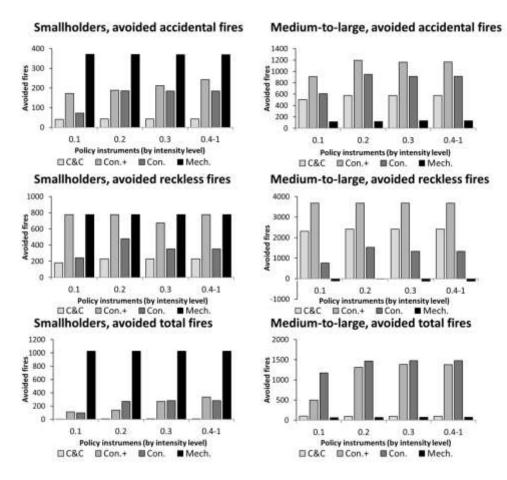
The almost negligible impact of C&C policy on smallholders (Table 4) do not mean ineffectiveness but instead reveal negligible exposure (treatment by policy) driven by limited enforcement. Analogously, the negligible impact of mechanization subsidies on medium-to large-farms is caused by the focus on smallholders.

Interestingly, results show that the effect of policy instruments stops increasing with intensity above 0.2/ha level for C&C policy and above 0.4/ha, for incentive-based policy (figure 7). The stagnation is because it is only possible to change agents' decisions through altering

incentives. The LUT choice is discrete and once a LUT is promoted to the top of the profit rank, further increasing its profitability exogenously does not change the rank and, therefore, incentives remain the same. This is consistent with theory which sustains that policy can only have its effectivity increased while there are open opportunities to change incentives (e.g. Becker, 1968). This "decreasing return" of instruments is especially relevant given that neither of them could reduce the accidental fire rate below 12% of the landscape from a baseline level of 26%. Consequently:

(PL 3) Command-and-control and incentive-based instruments may be effective in altering land use and fire level, but such effectiveness is limited.

Figure 7 Long run causal effects of instruments*, vertical axis: avoided fires measured as counts of parcels (hectares); horizontal axis: instrument intensity (fine or subsidy)



Caption: "C&C" stands for command-and-control policy and "Con+", "Con" and "Mech." for the three incentive-based instruments (section 2.2). In each plot, the effects are shown for intensity levels from 0.1 to 1.

4.3.2 Side-effects of instruments

Conservation subsidies are effective to avoid deforestation but not degradation of forest by accidental fires if unaccompanied by a restriction on reckless fires (figure 8). However, such restriction, as a side effect, increases controlled fire area above the baseline reckless fire area (figure 6). There are two likely explanations for this outcome. First, since replacing reckless

fire is mandatory but not subsidized, farmers face the income loss imposed by the restriction as an opportunity cost of signing a conservation+ contract. Second, reckless and controlled fires are, respectively, the first and second most profitable LUTs in the initial condition of all simulations (section 4.2). Therefore, to compensate for the opportunity cost, reckless fire use and also the lower-ranked fire-free are replaced by controlled fire use.

The compensation through disproportionate expansion of controlled fire use (see figure 6) can be understood by remembering that agglomeration of a LUT is driven by scale economies. Hence, the larger the area of controlled fire is, the more diluted is the fixed cost it incorporates. Making fire-breaks is an example of a fire control measure with fixed cost when aimed at protecting land uses outside the areas to be treated with fire. The relevance of this example is attested by the positive correlation between firebreak investment and value of land use under risk found by Bowman et al. (2008), who analysed data from a protected area in our study region, Santarém.

If agricultural fires are seen as undesirable for other reasons beside accidental fire risk (e.g. smoke and derived air pollutants and GHGs, and soil degradation), results suggest that forest conservation payments are not the most efficient instruments to address the issue⁵. This is also true for C&C instrument as it considerably increased controlled fires (figure 6). Of course, a "conservation++" contract forbidding reckless fire and controlled fire could be designed, and it would probably be more effective at avoiding accidental fires, at least for farmers motivated to sign the contracts. However, the number of farmers entering the scheme would likely be smaller. This is supported by the fact that the total number of farmers willing to sign a forest conservation contract is smaller when reckless fire is forbidden and the payment is below 0.4/hectare (table 6). It is only above this value that the reckless fire restriction has no impact on the total number of signed contracts.

Summing up, the incorporation of restrictions to fire use into PES programs in the Brazilian Amazon is necessary to assure an acceptable "return" for the payments, i.e. the quality of conserved forest. However, the simulations suggest that farmers need to be compensated for the cost of complying with restrictions in order to achieve the double PES goals of an acceptable return and desirable geographical reach. Without this, policy-makers would have to accept both a relevant level of agricultural fires and a reduced amount of land kept as forest. This finding leads to the fourth potential lesson.

(PL 4) The conditioning of conservation payments on the prohibition of reckless fires may be effective to avoid subsidized forest from being accidentally burned. However, as a side-effect, controlled fires may increase considerably and the whole area of conserved forest might fall.

The practical relevance of this potential lesson is attested by Leiva-Montoya's (2013) interviews with participants of the ongoing PES program "Bolsa Floresta". He found that 77% of the interviewees judged the household-targeted payment insufficient to cover the costs of compliance with deforestation and fire use restrictions. In a relevant number of

⁵ Moreover, in practice, during extreme droughts, as observed, for instance, in 2005, 2007 and 2010 (Stosic et al., 2016), controlled fires may become reckless fires, a possibility not modelled.

programs (see Pattanayk et al. 2010, Appendix table 2), compliance monitoring is imperfect and detection of violations is probabilistic. If payments partially cover compliance cost, theoretically (see Ferraro, 2008, p.811), partial compliance would tend to prevail at a level in which its cost balances payments.

Still, that fires may undermine environmental gains brought by incentives to avoid deforestation is argued by Friess et al (2015), Aragão and Shimabukuro (2010) and Barlow et al (2012). Their results show a clear need to introduce fire restrictions in REDD programs. This is already taking place in the Juruá and Rio Negro protected areas of Amazonas (Börner et al, 2013, p.18, Leiva-Montoya, 2013 p.38), where payments for forest conservation are conditional on the adoption of fire control measures (e.g. firebreaks) and on norms restricting the frequency and extent of burning (Börner et al, 2013).

Now, turning to the mechanization instrument, it quadruples fire-free area, promoting the LUT to a degree of diffusion (16%) that has no parallel in the other simulations (figure 6). It is also observed a significant shift from fire-based to fire-free agriculture. In the baseline, the area occupied by the former is 18 times larger than the area occupied by the latter. However, with a mechanization subsidy of R\$0.4, fire-based area is only 3 times larger than that of fire-free.

Even with such major impact on the fire-free area, only marginal unwanted indirect land use change could be found at farms exposed to the subsidy (see SM.3 table SM.3.13). In the case of the mechanization instrument, the unwanted indirect land use change is the replacement of forests by fire-free, which, even though not incentivized, is not ruled out by the contract and could theoretically happen due to the high agglomerative potential of fire-free. But the side-effect is not driven by spill-overs. It is, indeed, very straightforward. Smallholders explore the possibility of converting forest to fire-free in the first period in order to start receiving the mechanization subsidy in the third period. It is exactly what happens in the sensitivity simulation with highest share of indirect land use (table SM.3.13).

This side-effect should not be thought as irrelevant for being caused by a caveat in contractual rules. In theory, the issue could be solved by not remunerating the keeping of fire-free, except in parcels where the LUT replaced fire-based agriculture. However, the fire-based agriculture currently replaced may have taken the place of forest. To solve the problem in practice, recently deforested parcels should be not remunerated, but, still, the definition of "recently" may be a matter of dispute.

In practice, the evidence of deforestation through indirect land use change is more notorious than model's results reveal. Barona et al (2010) and Arima et al (2011) attest the occurrence of indirect deforestation induced by the expansion of fire-free and mechanized soybean growing in Brazilian Amazon. Wunder (2006) considers the possibility of PFES fostering the increase in cattle numbers, one of the main drivers of tropical deforestation, through the channel of capitalization which expands when farmers start being remunerated for leaving land plots idle to grow forests. Nevertheless, our results suggest that indirect deforestation

can be considerably mitigated if only a particular land use change, which does not, obviously, coincides with deforestation, is accurately subsidized. Synthesizing the discussion:

(PL 5) Indirect deforestation induced by mechanization subsidy may be kept low if the conversion of forests to mechanized agriculture is not directly subsidized.

However, mechanization may also have negative social and environmental impacts depending on how it is introduced. If cash flow and credit access are below the levels required for regular and minimum fertilizer application, income will fall in the short term. It will also fall in the long term when soils worn-out by slash-and-burn are not rehabilitated before tractor introduction, which can only lead to further degradation (Reichert et al., 2014). A mechanization subsidy, thus, has to cover the cost of sustainable soil management required by tractor introduction.

Figure 8 Counts of forested parcels accidentally burnt, only subsidized parcels (left) and all parcels (right), Conservation (solid line) and Conservation+ (dashed) contracts.

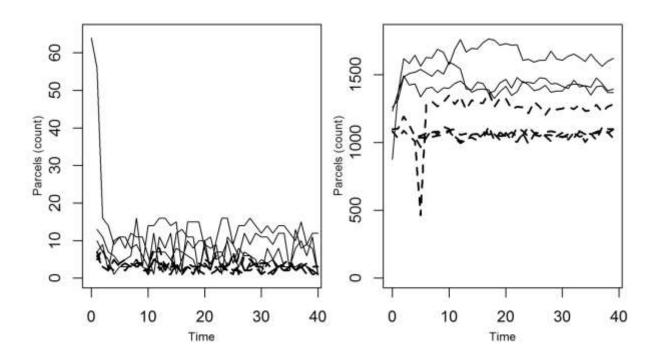


Table 6 [here] Table 7 [here]

5 Conclusions

Without public action, the current number of ignitions by farmers, combined with increased drought hazard and fire-prone vegetation, seems set to lead to disastrous wildfires, biodiversity loss, GHG emissions, and the spread of health-damaging pollutants (Balch, 2014, Jacobson et al., 2014, Malhi et al., 2009, Davidson et al, 2012). Alarmingly, there is no clear evidence that current Brazilian fire policies are promoting a relevant reduction of Amazon fires (see Sorrensen, 2009, Carmenta et al., 2013, Arima et al., 2007 and Costa, 2004). To better understand the impacts and limitations of policy options, we developed and applied an analytical tool to a small fraction of the Brazilian Amazon, identifying five potential lessons. The instruments evaluated cover a relevant fraction of the menu of choices in practice faced by policymakers, ranging from a negative incentive to abandon only reckless fires (C&C instrument), where compliance cost is fully borne by landholders, to a positive incentive where government fully pays for the cost of ceasing to use both reckless and controlled fires (mechanization).

For the social groups exposed to them, all policy instruments proved effective in overturning the predominance of highly profitable and risky fire use and also in decreasing accidental fires. However, none of them were free from imperfections. A ban enforced with fines performed worse than incentive-based instruments due to inadequate monitoring of compliance, leaving smallholders immune to sanctions. Forest conservation subsidies avoided deforestation but not forest degradation by fire. Such subsidies, when made conditional on the avoidance of reckless fires, did ensure reduced forest degradation, but the price paid was increased controlled fire use and less avoided deforestation. A subsidy to shift from fire-based to fire-free agriculture was the most effective instrument to avoid fire use and accidental fires, but it indirectly incentivised deforestation. Furthermore, for most intensity levels, the instruments that were more effective at reducing deforestation were not more effective at reducing fires, and vice versa (table 7). Thus, even within the wide spectrum of policy options examined, an instrument to achieve, with high impact, the double goal of total fire reduction and forest protection could not be found (except for intensity levels 0.1 and 0.2, see table 7). This is likely due to the impossibility of achieving multiple goals with a single instrument, as the Tinbergen rule proposes (Knudson, 2009).

Finally, due to the artificial nature of the data generated by the simulations, the results obtained are far from definitive, and only a point of departure for empirical research aimed at refuting the lessons learned. Nevertheless, the study has contributed a crucial first step towards analysing the impacts of policy on Amazonian fires by providing clear-cut hypotheses for further study.

It must be highlighted that results can only be extrapolated to post-frontier subregions of Amazon characterized by (i) widespread use of fire (lesson 1), (ii) predominance of mediumsized farms of 300-1100 hectares (SM.1), (iii) tenure security (farm boundaries are static across simulations) and (iv) effective enforcement of deforestation policy. The last point follows from the baseline average annual deforestation of 1% (40% in 40 steps), which is approximately equal to the actual municipal average after enforcement intensification (from 2005 to 2013). Smallholders are imprecisely represented with regard to farm geometry (SM.1) and results involving them can only be extrapolated, with care, for the subgroup of farms from between 100 and 200 hectares.

Limitations, to be addressed by future work, remain. First, the policy-implementing agency, i.e., the government, should be properly modelled, especially with regard to its budget and social welfare preferences. Secondly, channels through which climate change impacts on land use profits and fire risk should also be introduced, since the random component of accidental fires is unable to reproduce some of the trends increasingly recognized by the literature as forest flammability amplifiers (see Malhi et al., 2009, Coe et al., 2013, Davidson et al., 2012 and Aragão et al., 2008). Also, it remains to be tested the implications of decision making algorithms based on dynamic optimization. Direct interaction of agents, through communication among neighbors, for instance, is also an avenue to be explored by future work. Both extensions may influence results by increasing agents' ability to anticipate the consequences of policy instruments and to react to them both individually and collectively.

Appendix A ODD+D description of the model

I.Overview

I.i Purpose

<u>Li.a What is the purpose of the study?</u> The model is a tool to build knowledge on the impacts and limitations of command-and-control and incentive-based policy instruments designed to control agricultural and accidental fires in the Brazilian Amazon.

Lii.b For whom is the model designed? Researchers of Amazon fires and policy-makers

I.ii Entities, state variables, and scales

<u>Lii.a What kinds of entities are in the model?</u> Four kinds of entities. First, the decision makers that manage land, called "farmers". Second, spatial units, called "parcels". The latter kind of entity does not make decisions, but executes natural processes (forest growth, forest degradation, etc.) and is employed by farmers to process calculations required by land use and technology (LUT) choice. Third, an observer entity, called "nature", decides which parcels accidentally burn. Fourth, an observer entity, called "government", sanctions reckless fires in the command-and-control policy scenario and offers and monitors voluntary contracts in the incentive-based policy scenarios.

<u>Lii.b By what attributes (i.e. state variables and parameters) are these entities characterized?</u> (immutable initial conditions, which are equal across all simulations, are denoted by "[i.i.c]"; state variables, by "[s]").

(1) Farmers are characterized by: (1.a) Farm, i.e., set of parcels controlled [i.i.c]; (1.b) LUT portfolio choice [s]; (1.c) wealth or accumulated stock of whole-farm profits [s]; (1.d) point estimates for risk parameters [s]; (1.e) accumulated local data on fires and LUTs [s]; (1.f) contract status (regarding incentive-based instruments) [s];

(2) Parcels are characterized by: (2.a) location [i.i.c]; (2.b) farmer in control [i.i.c]; (2.c) physical suitability [p]; (2.c) LUT [s]; (2.d) age of forest, (2.e) above-ground forest biomass (AGB) [s]; (2.f) inclusion in enforcement-effective zone [i.i.c]; (2.g) whether accidentally burned or not [s];

(3) Nature is characterized by: (3.a) true risk parameters [i.i.c]; (3.b) values for standard Gaussian disturbance behind accidental fires [s];

(4) Government is characterized by: (4.a) Active policy instrument [i.i.c]; (4.b) level of intensity for policy (value of fine or subsidy) [i.i.c].

<u>Lii.c What are the exogenous factors / drivers of the model?</u> All attributes of nature and government (see I.ii.b) and the parameters capturing the effect of LUT agglomeration and dispersion on deterministic profit and risk.

<u>Lii.d If applicable, how is space included in the model?</u> With a two dimensional flat landscape whose basic unit is an autonomous processing unit referred as "parcel". The model is spatially explicit and operates in a landscape whose division among private owners comes from real data (Rural Environmental Land Registry, SM.1) and remains fixed across simulations. The initial condition for land use is also partially defined by data.

<u>Lii.e What are the temporal and spatial resolutions and extents of the model?</u> One time step represents one year, simulations were run for 40 years, one grid cell represents 1 ha and the model landscape comprises 100 x 100 hectares.

I.iii Process overview and scheduling

<u>I.iii.a What entity does what, and in what order?</u> (names of procedures are preceded by an indication of the entities that run them, as follows: [F] for farmer, [P] for parcel, [G] for government and [N] for nature)

(A) In baseline and command-and-control policy simulations, the following nine modules are executed in the following order: [P&F] calculate-profit, [P&F] implement-LUT-portfolio, [F] LUC-cost-account, [N] define-burned-parcels, [G] sanction-rulebreakers, [F] calculate-actual-profit, [F] update-risk-parameters, [F] store-LUT-portfolio-in-memory, [P] update-forest-age-after-burn.

The first iteration differs only regarding the absence of the procedure "update-risk-modelparameters" (since farmers have no local data at t = 0).

(B) In incentive-based policy simulations, the same procedures are processed together with two extra procedures: (i) [F] subsidy-payment-account which is deployed right after "implement-LUT-portfolio" and; (ii) [F] update-contract-duration, deployed right after "LUC-cost-account". In procedure (ii), decisions on signing, keeping and renewing a subsidy contract are made. One additional peculiarity of incentive-based simulations is that the "calculate-profit" procedure is subdivided in two sub-procedures, one calculates profit without imposing compliance with contract rules and the other forces compliance (B.1 and B.2 of appendix B).

II. Design concepts

II.i.a Which general concepts, theories or hypotheses are underlying the model's design at the system level or at the level(s) of the submodel(s) (apart from the decision model)? What is the link to complexity and the purpose of the model?

Agglomerative and dispersive forces driving land use, from the theory of new economic geography (Fujita and Thisse, 2002, Irwin and Bockstael, 2001, Krugman, 1996); Forest ecosystem services, including protection from accidental fire, discussed in the literature of land use, ecology and forestry (Klemick, 2008, Chomitz and Kumari, 1996, Tscharntke, 2005, Laurance et al, 2006, Ferraz et al, 2003, Stouffer and Bierregaard, 1995, Peres and Lake, 2003, Brando et al., 2013); Whole-farm profit as the index that guides land allocation decision, assumed by traditional agricultural economics models (Just et al., 1983, Fezzi and

Bateman, 2011); Parcel-level heterogeneity and spatial externalities of land uses (Irwin and Bockstael, 2001, Parker and Meretsky, 2004).

II.i.b On what assumptions is/are the agents' decision model(s) based?

Farmers are boundedly rational and choose LUT portfolio in the basis of a local (parcel-level) optimization procedure adjusted to incorporate farm-level information; Current wealth is a limiting factor of LUT portfolio change due to cost of land use change; It is necessary to form expectations on the LUTs to be developed at third-party parcels in the neighborhood of farm boundaries. This is done by assuming that previous step LUTs will be kept; Farmers are ignorant of true risk parameters and the value of the random disturbance behind accidental fires; Data for estimating risk parameters is collected locally and accumulated stepwise over time.

<u>II.i.c Why is a/are certain decision model(s) chosen?</u> There are two reasons. First, standard global optimization, i.e., finding the best among all possible portfolios is highly computingintensive. Once there are four LUTs, the number of alternative portfolios is equal to four elevated to a power equal to farm's area. For a 30 hectare farm, the smallest farm size in the model, the number of potential portfolios has an order of magnitude of 10¹⁸.Second, empirical studies of Amazon farmers' behavior show that land management decisions are driven by parcel-scale factors and limited by capital/wealth (Deadman et al., 2004, Sorrensen, 2000 and 2004, Moran et al., 2002, Scatena et al., 1996, McCraken et al. 2002). In particular, econometric results (e.g., Pfaff, 1999, Pfaff et al., 2007), reveal that proximity to roads, urban centers and low inclination of land, the three variables captured by the model's physical suitability, positively influence the conversion of forests to agriculture. This causality is the basis of the model's wealth allocation principle of giving priority to parcels whose costly conversion is, due to physical factors, more profitable.

II.i.d If the model / a submodel (e.g. the decision model) is based on empirical data, where does the data come from? The LUT choice algorithm processes information of two basic parcel-level mathematical functions which report levels of risk and of deterministic profit as functions of own-parcel and neighboring parcels' LUTs (appendix C); The parameters of deterministic profit are calculated in the basis of five principles, which refer to spatial spillovers of forest ecosystem services, scale economies, and land use change costs that were estimated from data (see appendix C); The risk parameters are also calculated on the basis of (six) principles grounded in the specific literature (appendix C); Forest-growth follows the empirical above-biomass growth function estimated by Neeff and Santos (2005); The policy instruments are based on concrete policy actions (see section 2.1). They include a simplified version of Brazilian controlled-burn law, two incentive-based instruments inspired by concrete PFES programs and one incentive-based instrument based on municipal mechanization subsidy programs (section 2); The landscape is designed to capture characteristics of a 10km² squared zone of Santarém municipality, Brazilian Amazon (SM.1); GIS data of Santarém municipality define (i) the land property structure, (ii) the initial status of parcels regarding presence and absence of forest and (iii) the physical suitability of parcels for agriculture (SM.1).

<u>II.i.e At which level of aggregation were the data available?</u> Land use change cost data: at the level of production factors (man-days, input quantities/output, etc.), see SM.2; Forest growth model: at the parcel (stand) level; Land property and initial land use data: parcel level, 30 m resolution data.

II.ii Individual Decision Making

<u>II.ii.a What are the subjects and objects of decision-making? On which level of aggregation is decision-making modeled? Are multiple levels of decision making included?</u> Farmers decide on the allocation of parcels among alternative LUTs. Nature decides, based on LUT configuration, which parcels accidentally burn. No other entity is capable of decision making. Government only implements pre-defined rules regarding farmer sanctioning.

<u>II.ii.b What is the basic rationality behind agents' decision-making in the model? Do agents</u> pursue an explicit objective or have other success criteria? Farmers are boundedly rational and seek the portfolio that maximizes whole-farm expected profit. Nature is substantively rational in the sense it does not face barriers for gathering and processing information but has no particular goal. Government does not follow a decision model, it simply implement rules (non-deliberate action).

II.ii.c How do agents make their decisions? Se II.ii.a and II.i.b above and appendix B.

II.ii.d Do the agents adapt their behavior to changing endogenous and exogenous state variables? And if yes, how? ("exo" stands for exogenous variables and "endo" for endogenous).

Yes, farmers adapt to: (1) LUTs of third-party parcels (exo), which create profit and risk spill-overs that influence LUT allocation at the boundary of farms; (2) Knowledge of the true process behind accidental burns, measured by accumulated local data on LUTs and accidental fires (endo. and exo.). This drives change of LUT portfolios due to changed perceived risk associated with LUT mosaics; (3) Own-wealth (endo.), which, being updated by whole-farm profit, becomes less stringent as a constraint on chosen portfolios; (4) Forest growth (exo.), which engenders profit and risk spill-overs that may lead farmers to reconsider LUT portfolios and their statuses regarding conservation contracts.

II.ii.e Do social norms or cultural values play a role in the decision-making process? No, there is no direct interaction among farmers and institutions are abstracted (apart from policy which is immutable during simulations).

<u>II.ii.f Do spatial aspects play a role in the decision process?</u> Yes, a crucial role through spillovers of parcel-level risk and profit and also by signaling to agents that accidental burns emerge (also) from particular spatial configurations of LUTs.

<u>II.ii.g Do temporal aspects play a role in the decision process?</u> Yes, time affects decisions through the dynamics of four stocks, wealth, local data, forest biomass and accumulated payments from subsidy contracts. Also, subsidy contracts have a finite five-year duration.

<u>II.ii.h To which extent and how is uncertainty included in the agents' decision rules?</u> Agents face two sources of uncertainty while searching for the best LUT portfolio. First, the best LUT portfolio of neighbors is unknown. Second, there is the uncertainty related with the true generating process behind accidental burns coming from (i) ignorance of the true risk parameters and (ii) randomness of unobservables. While (i) is mitigated with the accumulation of local data, (ii) is irreducible, and, therefore, accidental burns are always uncertain events in the model, even after farmers' point estimates for risk parameters have become sufficiently close to true values.

II.iii Learning

II.iii.a Is individual learning included in the decision process? How do individuals change their decision rules over time as a consequence of their experience? Yes, farmers learn about the true process behind accidental burns, as described in appendix B, B.13.

II.iii.b Is collective learning implemented in the model? No.

II.iv Individual Sensing

II.iv.a What endogenous and exogenous state variables are individuals assumed to sense and consider in their decisions? Is the sensing process erroneous? Farmers sense the risk of accidental burns associated with LUT configurations. This sensing is improved with local data accumulation but may prove wrong when estimates of risk parameters do not match the true values which are known only by nature.

<u>II.iv.b What state variables of which other individuals can an individual perceive? Is the sensing process erroneous?</u> Farmers observe the behavior of neighbors, but exclusively with regard to LUTs allocated to parcels bordering their own farm. On the basis of this, they try to forecast current LUTs of such proximate third-party parcels, but such forecasts may prove wrong.

<u>II.iv.c What is the spatial scale of sensing?</u> Local, parcel scale, restricted to own-parcels and third-party parcels within 100 meters from farm boundaries.

II.iv.d Are the mechanisms by which agents obtain information modelled explicitly, or are individuals simply assumed to know these variables?

(1) Farmers: (1.a) The mechanism of accumulation of information on accidental fires and LUTs is modelled explicitly. It is assumed that farmers know the specification of the true function behind accidental burns, which is a standard probit and estimate the parameters with local data; (1.b) Information on LUTs of proximate third-party parcels is obtained through direct observation;

(2) Nature and government know all the information they need to act.

II.iv.e Are costs for cognition and costs for gathering information included in the model? No.

II.v Individual Prediction

<u>II.v.a Which data uses the agent to predict future conditions?</u> Predictions of the future are not part of the model.

II.v.b What internal models are agents assumed to use to estimate future conditions or consequences of their decisions? Parcel-level and whole-farm profits are estimated, since two of the three classes of factors behind accidental fires are only known by nature. Every time step, farmers calculate expected profit for each LUT at each parcel and also for the whole-farm LUT portfolio. This is done by considering a parcel-level expected profit function given by $r^e = \eta\theta(1-p) - C(.) + S(\gamma)$ with $\eta \equiv$ physical suitability, $\theta \equiv$ deterministic profit, $p \equiv$ risk, $C(.) \equiv$ LUT change cost, $S(\gamma) \equiv$ fine (negative value) or subsidy (positive value) assigned by policy (see B.1 and B.2 of appendix B).

<u>II.v.c Might agents be erroneous in the prediction process, and how is it implemented?</u> Yes, point estimates for risk parameters may prove wrong, but it is impossible for agents to know because they are ignorant of the true model. They do not try to improve estimates on the basis of prediction error, but, without deliberation, follow a process of periodic re-estimation based on their stepwise expanded stock of local data.

II.vi Interaction

<u>II.vi.a Are interactions among agents and entities assumed as direct or indirect?</u> Farmers interact only indirectly and locally. Two are the channels through which interactions occur. The first is the forecast of LUTs of proximate neighbors' parcels. The second is the spill-over of dispersive and agglomerative forces across farm boundaries. The profit and risk of a LUT, when developed in a given parcel, depend on the LUTs adopted in all eight (queen) neighboring parcels, whether owned by the same agent or not. Therefore, agglomerative and dispersive forces transcend the limits of land property.

<u>II.vi.b On what do the interactions depend?</u> Proximity, since only neighbors interact (indirectly).

II.vi.c If the interactions involve communication, how are such communications represented? N/A

II.vi.d If a coordination network exists, how does it affect the agent behaviour? Is the structure of the network imposed or emergent? N/A

II.vii Collectives

II.vii.a Do the individuals form or belong to aggregations that affect, and are affected by, the individuals? Are these aggregations imposed by the modeller or do they emerge during the <u>simulation?</u> No, action is purely individual.

II.vii.b How are collectives represented? N/A.

II.viii Heterogeneity

II.viii.a Are the agents heterogeneous? If yes, which state variables and/or processes differ between the agents?

(1) Farmers are heterogeneous regarding: (1.a) Farm, i.e., location and number of parcels controlled; (1.b) Initial condition for wealth and LUT portfolio; (1.c) Initial guesses for parameters of the risk model; (1.d) Share of farm included in enforcement-effective zones (command-and-control policy);

(2) Parcels differ with regard to: (2.a) Location; (2.b) physical suitability; (2.b) Number of neighbors (parcels at the corner of the landscape have less than 8 neighbors in their queen neighborhood); (2.c) Initial condition for above-ground forest biomass (AGB); (2.d) Inclusion in to enforcement-effective zone (command-and-control policy).

II.viii.b Are the agents heterogeneous in their decision-making? If yes, which decision models or decision objects differ between the agents? No.

II.ix Stochasticity

II.ix.a What processes (including initialization) are modeled by assuming they are random or partly random?

There are three sets of variables randomly assigned. (1) Initial LUTs are partly randomly assigned. Forest and non-forest areas are defined by data. Within a non-forest area, LUTs are assigned with a random number generator. This initial configuration is the same across all simulations (including all policy instruments and intensity levels and also sensitivity simulations, SM.3); (2) Initial guesses for the parameters of the risk model are randomly assigned for each farmer, and drawn with uniform probability from intervals of \pm 20% around true values; (3) The random component of parcel-level flammability is drawn from the standard Gaussian cdf at each step with one particular value for each parcel.

II.x Observation

<u>II.x.a What data are collected from the agent-based model (ABM) for testing, understanding,</u> and analyzing it, and how and when are they collected? The data collected comprises measures of causal effects of policy instruments, including three outcome variables each capturing counts of parcels with a particular type of fire (section 4.1). By "causal effects" is understood the difference of outcome variables comparing simulations with and without active policy instrument exclusively in terms of their final step (understood as the long run equilibrium, t = 40). Data is collected with R software ("RNetLogo" package) by running the ABM and storing the main results of all steps of simulations. Thereafter, the causal effects are calculated by comparing data on baseline and policy simulations. By simulation it is understood a set of 40 iterations of the ABM algorithm characterized by particular values of (i) active policy instrument (if any), (ii) policy instrument intensity level, (iii) risk and deterministic profit parameters. The last component is only changed to perform sensitivity tests (SM.3). II.x.b What key results, outputs or characteristics of the model are emerging from the individuals? (Emergence) The patterns described by causal effects lead to conclusions on the impacts and limitations of policies, which can be seen as emergent properties since they are aggregates across the whole landscape and also within two groups of farmers defined by farm size (see section 4.1). The spatial configuration of LUTs, including reckless and controlled fires and also accidental fires is also an emergent property which differs depending on the active policy instrument.

III.Details

III.i Implementation Details

<u>III.i.a How has the model been implemented?</u> The model code is written in NetLogo 5.1.0 and simulations were run with loops programmed in R (RNetLogo package, Thiele, 2014).

<u>III.i.b Is the model accessible and if so where? The</u> model code is sent, in four ASCII files (including a readme file), with this paper as part of the supplementary material (e-content).

III.ii Initialization

<u>III.ii.a What is the initial state of the model world, i.e. at time t=0 of a simulation run?</u> 26 farmers with heterogeneous farms, guesses for risk parameters and wealth. Ten thousand parcels with heterogeneous LUT statuses. In command-and-control simulations the inclusion in enforcement-effective zone is also a characteristic in which parcels differ.

<u>III.ii.b Is initialization always the same, or is it allowed to vary among simulations?</u> Always the same, except for guesses for risk parameters.

III.ii.c Are the initial values chosen arbitrarily or based on data? See II.ix.

III.iii Input Data

III.iii.a Does the model use input from external sources such as data files or other models to represent processes that change over time? No.

III.iv Submodels

III.iv.a What, in detail, are the submodels that represent the processes listed in 'Process overview and scheduling'? See appendix B.

III.iv.b What are the model parameters, their dimensions and reference values? See appendix C.

III.iv.c How were submodels designed or chosen, and how were they parameterized and then tested? See appendices B and C.

Appendix B Submodels

This appendix details the submodels or procedures of the model's algorithm. Sections B.0 to B.15 present all procedures processed by the three categories of simulations. The exact sequences in which procedures are executed in particular simulation categories are presented in section B.16.

B.0 Initialization

B.0.1 [N] setup-landscape

(1) True risk parameters and the deterministic profit parameters are set by following conventions from appendix C;

(2) GIS data on land property and forested and non-forested parcels are imported and incorporated to the landscape. The data embodies adjustments described in SM.1. Further adjustments to eliminate property overlaps are made;

(3) Non-forested parcels have their LUT assigned by randomly drawing with uniform probability from the set $\{0, 1, 2, 3\}$ with 0 standing for reckless fire, 1 for controlled fire, 2 for fire-free and 3 for freshly-abandoned land (forest with age zero);

(4) Forested parcels have their forest age attributed by randomly drawing with uniform probability from the set {5, 10, 25, 50, 100};

(5) Above-ground biomass of forested parcels is calculated from forest age using the empirical growth model proposed by Neeff and Santos (2005).

Note: the results generated by setup-landscape are used, without any alteration, in all simulations. The procedure is, thus, ran only once before all simulations.

B.0.2 [N&F&P] setup-farmers

(1) A number of agents "farmers", equalling the number of properties is created. There is only one farmer holding the code that identifies the set of parcels belonging to a given farm;

(2) Initial guesses for risk parameters are calculated. In the first step the lack of data to conduct estimations is circumvented with initial guesses in which values for parameters are randomly drawn from a uniform distribution, defined to be in the interval [0.8 β j; 1.2 β j], j = 0,...,5. Thus, the maximum error agents start with is ±20%;

(3) Farmers are moved to the centroid of the respective farm and do not move across simulations;

(4) Initial wealth is assigned as the actual profit yielded by the initial LUT portfolio under the assumption that, in t = 0, no parcel is affected by an accidental fire.

B.0.3 [N & P] Delimit-EE-zones (only executed in baseline and command-and-control simulations)

The landscape is subdivided into 100 square zones of 10 x 10 parcels or 1km^2 . Zones intercepted by only one farm are defined as enforcement-effective (EE) zones and it is only in these zones that reckless fire is sanctioned in command-and-control simulations.

B.1 [P&F] calculate-profit (as executed in baseline and command-and-control simulations)

Calculate-profit consists of five sub-procedures, as follows.

B.1.1 [F] generation of parcel list

In each time step, farmers list parcels in descending order of physical suitability, η_k . As described in the next sub-procedures, the best-LUTs of each of the listed parcels are identified. The list is traversed until the accumulated cost of converting parcels to the best LUTs equals the level of available wealth or the end of the list is reached.

B.1.2 [P] Assumptions on best LUTs of neighboring parcels

<u>Procedure:</u> parcels make assumptions on neighbors' current LUTs and AGBs, following two criteria:

(1.i) neighboring parcels that fit the category "inside listed above" (cf. table B.1) are assumed to implement best LUTs and associated AGBs;

(1.ii) remaining neighboring parcels are assumed to implement previous LUTs and one-period-updated above-ground biomasses (AGBs).

Explanation: the agglomerative and dispersive forces behind risk and deterministic profit, as well as the list approach, both make local optimization imprecise due to limited information on the best LUTs of neighboring parcels. First, it is impossible to know best LUTs of "outside" parcels, i.e., parcels belonging to other farmers, since farmers make LUT choice simultaneously. Second, even for "inside" parcels best LUTs may be unknown at a given step of list traversing. There are, in fact, four classes of information availability in which neighboring parcels can be classified (table B.1) and only two of them, "inside listed above" and "unlisted", correspond to available information. The other two, "outside" and "inside listed below", are mitigated by assuming that previous LUTs are kept.

Table B.1 [here]

The need for a hypothesis on parcels of the class "inside listed below" is gradually eliminated by the multiple iterations of the local optimization procedure.

B.1.3 [P] identification of best LUTs for listed parcels

(1) For each of the four LUTs, parcel-level expected profit is calculated as:

$$r^{e}(\eta_{k},\tau_{k},\tau_{k}^{\prime e},\tau_{k,t-1})$$

$$=\eta_{k}\theta(\tau_{k},\tau_{k}^{\prime e})\left(1-\widehat{p_{k}}(\tau_{k},\tau_{k}^{\prime e})\right)-c(\tau_{k,t-1},\tau_{k,t})$$

$$-f1\{\tau_{k}=HRFBA \text{ and } k \in \text{EE zone}\} \qquad (A.1)$$

In equation A.1, τ_k is the LUT potentially adopted at the k-th parcel and τ_k ^e is the vector of LUTs expected to be adopted at the neighboring parcels of k. The term $\theta(\tau_k, \tau_k)^{e}$ is the deterministic profit. The probability with which the k-th parcel accidentally burns is $p_k(\tau_k, \tau_k)^{e}$, referred to as "risk". The term $c(\tau_{k,t-1}, \tau_{k,t})$ is the land use change cost cost. It is defined by table SM.2.9 which contains estimates from concrete data. Land use change cost is a function of three variables, current LUT, previous above-ground biomass (AGB) accumulated in the parcel and current LUT. The four components of whole-farm profit are fractions, belonging to the [0;1] interval.

The last component accounts for fines due to the development of reckless fire in enforcement-effective zones. In command-and-control simulations, farmers committing such transgression are fined with a fixed value, f, for each parcel in which transgression takes place. The indicator function, 1{}, takes unitary value for reckless fire parcels located in enforcement-effective zones. In the baseline experiment, f is set to zero.

The functions $\theta(.)$ and $\widehat{p_k}(.)$ take the form of probit models and their complete specification is found in appendix C.

The parcel-level expected profit is a fraction. It is not an absolute amount of money but the share of the maximum profit level obtainable, i.e., it is a fraction of the profit generation potential of a one-hectare parcel. Consequently, actual profit and its accumulation over time, i.e., wealth, are also measured in "share of profit potential" units, which is the standard for all monetary variables. It is assumed that prices are stable enough over time to not cause relevant changes in monetary values measured as just detailed.

(2) The LUT with the highest parcel-level profit is defined as the "best-LUT";

(3) Above-ground biomass (AGB) of parcels and farm-level land use change cost are updated based on best-LUTs.

B.1.4 [P] update of assumptions on current LUTs of neighboring parcels

After all listed parcels have executed procedures B.1.2 and B.1.5, best-LUTs of listed parcels are known. Listed parcels access this information and replace assumptions on inside-listed below neighbors by actual best-LUTs. Expected profit is recalculated by listed parcels.

B.1.5 [F] calculus of whole-farm expected profit

Whole-farm profit is calculated as the sum of parcel-level profit resulting from B.1.4. For this, best-LUTs are considered to be adopted at listed parcels and previous LUTs to be adopted at unlisted parcels. Procedures B.1.2 to B.1.5 are repeated while (i) repetitions

increase whole-farm expected profit by more than 5% and (ii) the number of repetitions remains below ten.

It must be highlighted that it is only in the first iteration that previous LUTs are assumed to be adopted at neighboring parcels of the class "inside listed below". In subsequent iterations, best-LUTs are available and considered, even if defined in previous iterations.

B.2 [P&F] calculate-profit (as executed in incentive-based simulations)

In incentive-based simulations, the "calculate-profit" procedure is subdivided into two subprocedures. First, "unrestricted-calculate-profit" does not impose compliance to contract norms as restrictions to optimization. Second, "restricted-calculate-profit" forces compliance. The two sub-procedures differ only in one module, "identification of best LUTs for listed parcels" which is the algorithm that incorporates (or not) contract rules.

In addition, the "calculate-profit" procedure differs from baseline and command-and-control simulations with regard to (i) the formula of parcel-level expected profit, (ii) a procedure that incorporates received subsidies and contract breaking penalties to whole-farm expected profit, (iii) a procedure in which resulting portfolio and whole-farm expected profit are stored. In the next two subsections only the altered and additional procedures are detailed. The third subsection presents the structure of "calculate-profit" for incentive-based instruments.

B.2.1 [P] Identification of best LUTs for listed parcels, unrestricted

(1) In incentive-based simulations, the parcel-level expected profit does not contain the component that accounts for fines due to violations of command-and-control policy. This is consistent with the fact that in policy simulations, only one policy instrument is active. Parcel-level expected profit is computed, in incentive-based simulations, as follows.

$$\widetilde{r^{e}}(\eta_{k},\tau_{k},\tau_{k}^{\prime e},\tau_{k,t-1}) = \eta_{k}\theta(\tau_{k},\tau_{k}^{\prime e})\left(1-\widehat{p_{k}}(\tau_{k},\tau_{k}^{\prime e})\right) - c(\tau_{k,t-1},\tau_{k,t})$$
(A.2)

It must be clarified that received LUT subsidy and contract breaking are not accounted for in parcel-level expected profit and, consequently, do not influence the identification of best LUTs. This is consistent with the assumption that contract status is decided by farmers in the basis of the whole-farm expected profits yielded by two LUT portfolios, one that is restricted to fit contract norms and another which is free to violate such norms. It is into the whole-farm expected profits associated with these two portfolios that subsidies and the penalty for contract breaking are incorporated (B.2.4 below).

(2) Same as B.1.3(2), best-LUT is defined as the LUT with highest profit.

(3) Same as B.1.3(3), update of AGB and land use change cost.

B.2.2 [P] Identification of best LUTs for listed parcels, restricted

(1) Same as B.2.1(1).

(2) Same as B.1.3(2), best-LUT is defined as the LUT with highest profit.

(3) Best-LUT is changed depending on its value, on previous LUT and on the active subsidy contract, as follows.

Under Conservation contract

If previous LUT is forest aged at least 10 years, best-LUT is changed to forest

In all remaining cases, best-LUT is not changed.

Under Conservation+ contract

If best-LUT and previous LUT are reckless fire, best-LUT is changed to controlled fire

If best-LUT is reckless fire and previous LUT is not reckless fire, best-LUT is changed to previous LUTIf previous LUT is forest aged at least 10 years, best-LUT is changed to forest

In all remaining cases, best-LUT is not changed.

Under Mechanization contract

If previous LUT is an agricultural LUT (reckless fire, controlled fire or fire-free), best-LUT is changed to fire-free

In all remaining cases, best-LUT is not changed.

(4) Same as B.1.3(3), update of AGB and land use change cost.

B.2.3 [F] Storage of LUT portfolios and whole-farm profits

The two LUT portfolios generated, the "unrestricted" and "restricted", are stored in farmers' memory as well as the associated whole-farm profits.

B.2.4 [F] Incorporation of received subsidies and penalty

(1) Total received subsidy, S, is calculated as the product of a fixed payment per hectare, s, and the number of parcels with subsidised (target) activity, N. Thus, S = sN;

(2) The penalty for contract breaking is calculated as accumulated S since the beginning of the current contract;

(3) Whole-farm expected profit yielded by unrestricted portfolio is updated by the deduction of the contract breaking penalty (which is zero if contract duration is zero or 5 years);

(4) Whole-farm expected profit yielded by restricted portfolio is updated by the addition of S.

B.2.5 [F] Storage of updated whole-farm profits

Whole-farm profits from restricted and unrestricted LUT portfolios are, after being updated with subsidies and penalty, stored in memory in place of previous values.

B.2.6 Structure of calculate-profit

- (1) Unrestricted-calculate-profit
- (1.a) [F] Generation of parcel list
- (1.b) [P] Assumptions on best LUTs of neighboring parcels
- (1.c) [P] Identification of best LUTs for listed parcels, unrestricted
- (1.d) [P] Update of assumptions on current LUTs of neighboring parcels
- (1.e) [F] Calculus of whole-farm expected profit
- (1.f) [F] Storage of unrestricted portfolio and associated whole-farm profit
- (2) Restricted-calculate-profit
- (2.a) [F] Generation of parcel list
- (2.b) [P] Assumptions on best LUTs of neighboring parcels
- (2.c) [P] Identification of best LUTs for listed parcels, restricted
- (2.d) [P] Update of assumptions on current LUTs of neighboring parcels
- (2.e) [F] Calculus of whole-farm expected profit
- (2.f) [F] Storage of restricted portfolio and associated whole-farm profit
- (3) [F] Incorporation of received subsidies and penalty

(4) [F] Storage of updated whole-farm profits

B.3 [P&F] implement-LUT-portfolio (as executed in baseline and command-and-control simulations)

The best-LUTs are adopted at each parcel. Unlisted parcels adopt previous LUTs. Afterwards, AGB is updated.

B.4 [P&F] implement-LUT-portfolio (as executed in incentive-based simulations)

Farmers choose between unrestricted and restricted LUT portfolios the one with the highest whole-farm expected profit. The chosen portfolio is implemented with the best-LUTs it specifies being adopted at listed parcels. Unlisted parcels adopt previous LUTs. Afterwards, AGB is updated.

B.5 [F] subsidy-payment-account (only executed in incentive-based simulations)

(1) If a restricted LUT portfolio was implemented, total received subsidy, S, is calculated as the product of a fixed payment per hectare, s, and the number of parcels with subsidised (target) activity, N. Thus, S = sN. The penalty for contract breaking is set to zero;

(2) If an unrestricted portfolio is implemented, total received subsidy is set to zero. The penalty for contract breaking is calculated as accumulated S since the beginning of the current contract.

B.6 [F] LUC-cost-account

Land use change (LUC) cost is calculated for the whole farm, after LUT portfolio choice.

B.7 [F] update-contract-duration (only executed in incentive-based simulations)

(1) Contract status is updated after the decision between restricted and unrestricted LUT portfolios by processing the rules that follow, which are also summed up in the flowchart at the end of this submodel.

(1.a) If contract duration is <u>zero</u> and:

(1.a.i) an unrestricted portfolio was implemented, contract status is set to "do not adhere/sign";

(1.a.ii) a restricted portfolio was implemented, contract status is set to "adhere/sign";

(1.b) If contract duration is above zero and below 5 periods and:

(1.b.i) a restricted portfolio was implemented, contract status is set to "keep";

(1.b.i) an unrestricted portfolio was implemented, contract status is set to "break";

(1.c) If contract duration is <u>exactly 5 periods</u> and:

(1.c.i) a restricted portfolio was implemented, contract status is set to "renewal";

(1.c.i) an unrestricted portfolio was implemented, contract status is set to "exit";

In sum, whenever restricted portfolio yields a higher profit (being "best"), contract is signed, kept or renewed. Contrariwise, contract is not signed, broken or not renewed (figure B.1 below).

(2) Update of contract duration and accumulated total received subsidies (payments), acc.S, is pursued by applying the following rules:

(2.a) If contract status is "adhere/sign", contract duration is set to one period and $acc.S_t = S_t$ where S_t is the current value of total received subsidies;

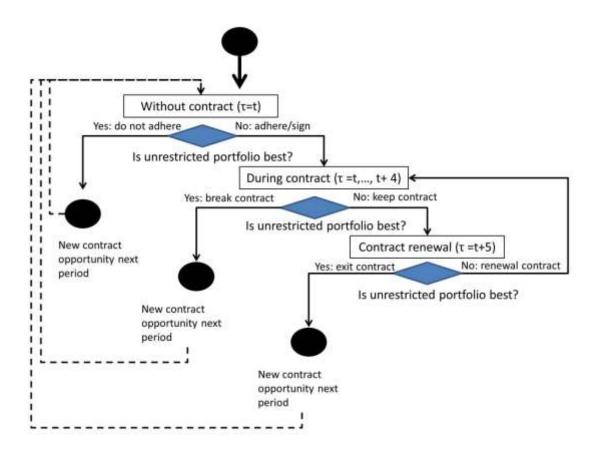
(2.b) If contract status is "keep", contract duration is increased by one period and acc. $S_t = acc.S_{t-1} + S_t$;

(2.c) If contract status is "break", the contract duration and $acc.S_t$ are both set to zero;

(2.d) If contract status is "renewal", contract duration is set to one period and $acc.S_t = S_t$

(2.d) If contract status is "exit", contract duration and $acc.S_t$ are both set to zero.





B.8 [N] define-burned-parcels

Procedure

Nature defines the parcels that accidentally burn on the basis of the components of a parcellevel latent flammability index. Parcels with accidental burns generate null actual profit.

The latent flammability model

The occurrence of an accidental fire in the k-th parcel is assumed to follow a latent variable probit model (Wooldridge, 2002, section 15.3). I_k^* is the latent (unobservable) flammability index such that $y_k = 1\{I_k^* > 0\}$ and $I_k^* = E[I_k^*|X_k] + u_k \sim X_k\beta + u_k$, where 1{} is the indicator function and X_k is a vector of observables. Predictors of I_k^* that are non-observable to farmers are captured by u_k , a standard Gaussian disturbance. Consequently, $p_k(\tau_k, \tau'_k^e) \equiv$

 $P(y_k=1|X_k) = G(X_k\beta)$ with G(.) being the standard Gaussian cdf. Whenever $I_k = X_k\beta + u_k > 0$, and, thus, $y_k = 1$, an accidental fire occurs at the k-th parcel.

The matrix X_k contains independent variables which capture LUTs conducted in the parcel and in its queen neighbourhood (complete specification in appendix C). Above-ground biomass (AGB, tons of biomass / hectare) is taken as a proxy for the ability of a forested pixel to contain fire spreads (see, for instance, Brando et al, 2013).

The unobservables synthesized by u_k can be thought of as physical and climate time-varying conditions such as wind velocity and stochastic (unpredictable) components of local precipitation and temperature.

B.9 [G] sanction-rulebreakers (only executed in command-and-control simulations)

In command-and-control simulations, farmers conducting reckless fire in parcels belonging to enforcement-effective zones are fined with a fixed value, f, for each parcel in which transgression takes place.

B.10 [F] calculate-actual-profit (as executed in baseline and command-and-control simulations)

(1) Whole farm profit after accidental burns is calculated as follows:

$$r(\eta_{k}, \tau_{k}, \tau'_{k}, \tau_{k,t-1}) = \sum_{k=1}^{K_{i}} \eta_{k} \theta(\tau_{k}, \tau'_{k})(1 - y_{k}) - \sum_{k=1}^{K_{i}} c(\tau_{k,t-1}, \tau_{k,t}) - \sum_{k=1}^{K_{i}} f1\{\tau_{k} = HRFBA \text{ and } k \in EE \text{ zone}\}$$
(A.3)

With the occurrence of an accidental fire at the k-th parcel being indicated with value 1 for the binary y_k (and non-occurrence with 0).

(2) Wealth is updated by adding to the current actual profit to its previous value;

B.11 [F] calculate-actual-profit (as executed in incentive-based simulations)

(1) Whole farm profit after accidental burns is calculated as follows:

$$r(\eta_{k}, \tau_{k}, \tau'_{k}, \tau_{k,t-1}) = \sum_{k=1}^{K_{i}} \eta_{k} \theta(\tau_{k}, \tau'_{k})(1 - y_{k}) - \sum_{k=1}^{K_{i}} c(\tau_{k,t-1}, \tau_{k,t}) + 1\{i \text{ holds}\} \left(1\{i \text{ kept}\} sN - 1\{i \text{ broke }\} \sum_{\tau=\tau_{0}}^{t-1} sN_{\tau} \right)$$

$$(A.4)$$

Where "i holds" indicates that the i-th farmer holds a contract in current period and "i kept" and "i broke" indicate, respectively, whether the contract is kept or broken. The number of pixels with target activity is N, τ_0 is the year the contract started and "t" is the current period.

(2) Wealth is updated by adding the current actual profit to its previous value;

B.13 [F] update-risk-parameters

Procedure

(1) After LUT portfolios are implemented and accidental burns occur, the current statuses of parcels regarding these two characteristics are incorporated into farmers' databases. This is done respecting the restriction that farmers observe only the statuses of own parcels and parcels within 100m of the boundaries of the farmer's own farm;

(2) The expanded database is used to re-estimate risk parameters. Estimation is pursued by calling the statistical software R^6 with Rserve extension for NetLogo (Thiele and Grimm, 2011). Farmers then employ the generalized linear model (GLM) routine to estimate a probit model based on available data;

(3) Estimations may be inconclusive for parameters which capture the effect of LUTs not developed on the farm or within 100m of it in the current and previous years. In this case, farmers do not update the parameters' values;

(4) The set of new point estimates generated is stored in memory to be used in the next period for estimating probabilities of parcels being accidentally burned.

Further details

Farmers are ignorant of the two drivers of accidental fires, β and u_k , but try to estimate the former with data on X_k and y_k collected from observations made within a radius of sight that comprises own-farm-parcels and parcels within 100m of farm boundaries. I.e., agents see one pixel beyond farm boundaries. Risk parameters are re-estimated at each step from pooled cross section data covering all previous periods.

B.14 [P] store-LUT-portfolio-in-memory

Current LUTs, forest age and AGBs are stored in memory, in order to be used by calculations that require information about the previous period (e.g., LUC cost calculation).

B.15 [P] update-forest-age-after-burn

A 100% loss of above-ground biomass at forested parcels that accidentally burn is assumed. This procedure defines, for such parcels, forest age to be zero.

B.16 Flows of procedures by simulation category

Table B.2 [here]

⁶ https://cran.r-project.org/

Table B.3 [here]

Appendix C Deterministic profit and risk functions

C.1 Deterministic profit

The functional form of the deterministic component of parcel-level expected profit is as follows.

$$\begin{aligned} \theta(\tau_k, \tau_k') &= G(\alpha_0 + \alpha_1 d_- \tau_- 0_k + \alpha_2 d_- \tau_- 1_k + \alpha_3 d_- \tau_- 2_k + \alpha_4 A G B_k + \alpha_5 d_- \tau_- 0_k * N_- \tau_- 0_k \\ &+ \alpha_6 d_- \tau_- 1_k * N_- \tau_- 1_k + \alpha_7 d_- \tau_- 2_k * N_- \tau_- 2_k + \alpha_8 d_- \tau_- 0_k * w_- A G B_k \\ &+ \alpha_8 d_- \tau_- 1_k * w_- A G B_k + \alpha_8 d_- \tau_- 2_k * w_- A G B_k + \alpha_9 A G B * w_- A G B_k), k \\ &= 1, \dots, K (2) \end{aligned}$$

Where G(.) is the standard Gaussian cdf, d_{τ}_0 , d_{τ}_1 and d_{τ}_2 are binaries indicating, respectively, whether reckless fire ($\tau = 0$) or controlled fire ($\tau = 1$) or fire-free ($\tau = 2$), are developed in the k-th parcel. N_t τ_0 , N_t τ_1 and N_t τ_2 are the counts of parcels with the LUTs just mentioned at the queen-neighborhood. The above-ground biomass of forest accumulated in the parcel is denoted by AGB and measured in tons of biomass/hectare. It is assumed that both own-parcel AGB and the average AGB of neighbouring parcels, w_AGB, are determinants of the profit from forest. AGB grows with forest age according to the logistic function estimated by Neeff and Santos (2005; tables 1, 2 and 3, figure 6) on the basis of data from secondary forests of Tapajós region, Central Amazon. Parcels allocated to nonforest LUTs always have zero AGB.

From equation (2) it is possible to obtain the deterministic profit of the j-th LUT, as follows.

$$\underline{\theta}(\mathbf{j}, \mathbf{\tau}_{\mathbf{k}}^{*}) = G(\alpha_{0} + \alpha_{j+1} + \alpha_{j+5}N_{-}\mathbf{\tau}_{-}j_{k} + \alpha_{8}w_{-}AGB_{k}), k = 1, ..., K, \mathbf{j} = 0, 1, 2$$

$$\underline{\theta}(\mathbf{j}, \mathbf{\tau}_{\mathbf{k}}^{*}) = G(\alpha_{0}), \mathbf{j} = 3 \text{ (forest with zero age or "freshly abandoned land")}$$

$$\underline{\theta}(\mathbf{j}, \mathbf{\tau}_{\mathbf{k}}^{*}) = G(\alpha_{0} + \alpha_{4}AGB_{k} + \alpha_{9}AGB_{k} * w_{-}AGB_{k}), \mathbf{j} = 4 \text{ (forest)}$$

C.2 Deterministic profit parameters

The assignment of the parameters of the deterministic profit follows the principles P1-P5 below, which further detail on how the agglomerative and dispersive forces behind profit work.

(P1) Forest products principle. Forest provides direct benefits with timber and non-timber forest products. The aggregated quantity of products supplied increase with forest age;

(P2) Ecosystem services principle. Forest increases water availability and soil quality, yielding benefits which cross parcel boundaries (table 3 in the main text). The magnitudes of the benefits brought by such positive externalities are positively correlated with accumulated AGB. For the sake of simplicity and lack of precise information about the environmental

service effect of forest on agricultural land uses, it is assumed that the three agricultural land uses have their profit increased by the same magnitude for each increment of the average AGB of surrounding parcels.

(P3) Forest fragmentation principle. An "island" of forest provides less ecosystem services (water and soil quality) and products than a "sea" of forest, as studies of forest fragmentation show (Laurance et al, 2006, Ferraz et al, 2003, Stouffer and Bierregaard, 2006). This principle is roughly captured by the average AGB of neighboring parcels, since deforested neighboring parcels are counted in the denominator of the average even holding with zero amount of AGB.

(P4) Agglomeration principle. The profit of a LUT increases with the number of neighboring parcels with the same LUT (see table 3). The list of LUTs in descending order by agglomeration externalities is (1) fire-free (capital-intensive), (2) controlled fire (incorporates fixed cost of fire control), (3) reckless fire, (4) Forest.

(P5) There are particular neighboring LUT mosaics (configurations) for which the conversion of forest to non-forest LUTs pays off.

The five principles can be quantified in several alternative manners in order to generate the values of the parameters. In this paper, the conventions adopted are presented in table C.1.

Table C.1 [here]

The conventions generate an exactly determined linear system of 10 equations and 10 unknowns in the form $x\alpha = G^{-1}(\Theta)$, with x being the covariates of the deterministic profit function, $G^{-1}(.)$ being the inverse Gaussian cdf and Θ the vector with profit levels (second column of table above).

C.3 Risk functional form

The probability of a parcel to accidentally burn, as estimated by agents, takes the functional form below.

$$p_{k}(\tau(k),\tau(k)') = G(\beta_{0} + \beta_{1}d_{-}\tau_{-}0_{k} + \beta_{2}d_{-}\tau_{-}1_{k} + \beta_{3}d_{-}\tau_{-}2_{k} + \beta_{4}AGB_{k} + \beta_{5}N_{-}\tau_{-}0_{k} + \beta_{6}N_{-}\tau_{-}1_{k} + \beta_{7}N_{-}\tau_{-}2_{k} + \beta_{8}w_{-}AGB_{k})$$

k = 1, ..., K.

C.4 Risk parameters

True parameters of the risk function are fixed in values according with five principles.

(P1) Minimum risk principle. A parcel faces the lowest level of probability⁷ of being accidentally burned when covered with 10 year old forest and surrounded by queen-neighboring parcels which, in average, are covered by 10 year old forest;

(P2) Fire accumulation principle: the larger the number of fire sources in the proximity of the parcels, including the parcel itself, the larger the probability of the parcel to be accidentally burned;

(P3) Fire control principle: controlled fires are less prone to escape than reckless fires;

(P4) Distance principle: the closer to the parcel a fire source is, the higher is the probability with which the parcel accidentally burns. Therefore, the two fire-based LUTs, reckless fire and controlled fire, impose a lower probability of accidental burn when conducted in the neighborhood than in the own parcel;

(P5) Fire protection principle: forest exerts a negative influence on the probability of accidental fires, a positive externality. The higher the average AGB accumulated in neighboring parcels, the lower the risk of the parcel to be accidentally burned. The "protective" effect of forest increases with accumulated AGB;

(P6) fire-free principle: the presence of fire-free in the parcel and in the neighborhood increases sensibly the probability of accidental fires. Even though not representing an ignition source, such technology, as any land management technology, is established through the removal of forest cover and this leads to the accumulation of flammable debris and local dryness.

To quantify the principles above, the conventions in table C.2 are adopted.

Table C.2 [here]

⁷ The description of how deforestation increases fire-proneness of Amazon forest, by Brando et al (2014), makes clear how forest fragmentation affects the probability of fires. "First, by reducing canopy cover and evapotranspiration, deforestation increases average dry-season land-surface temperatures (...), which in turn promotes air movement between open fields and neighboring forests. Consequently, fuels along forest edges are expected to become drier, leading to increased fire intensity (...). Second, deforestation fragments the landscape, creating a greater perimeter of forest edges (...). Third, tree mortality associated with previous logging, fire, severe drought, or edge effects can contribute to coarse fuel loads for multiple years as the twigs and branches of standing dead trees gradually decay and fall to the ground."

Question	Answer
	Unknown, available remote-sensing
Which is the share of remote-sensing fire detections related	data comprehends (i) point detections
with:	or "hotpixels" and (ii) "burned areas".
(1) Agricultural fires	With such information it is only
(1.a) Deforestation;	possible to know the approximate
(1.b) Fallow-based agriculture;	location of fires and path followed, but
(1.c) Pasture management and restoration;	not the finalities with which fires were
(2) Accidental fires.	started.
With which probability does an agricultural fire run out of	
control, turning into an accidental fire, and how does this	
depend on surrounding land use and fire control practices?	Unknown
Which are the economic returns of the following	
alternatives to fire:	
(1) Mechanized land preparation, conducted in small plots	
(3 hectares at most);	
(2) Green land preparation (with fast-growing-N-fixing	A few field-based studies have
species and/or mulching);	produced cost and revenue data, but the
(3) Agroforestry (integrated crop and forestry).	information remains anecdotal.
Which is the rote of illegal fire users identified and	Reports of these events are dissipated
Which is the rate of illegal fire users identified and sanctioned?	across the three levels of government.
sancuoneu	No comprehensive assessment is
Which is the rate of sanctioned farmers among the ones that	available. The number of undetected
have accidentally burned neighbors' land?	occurrences seems to be high for most
	Brazilian Amazon states due to lack of
Which is the rate of identified and sanctioned farmers	monitoring and the difficulty of
among the ones that have caused wildfires?	identifying fire starters.

Table 1 Main uncertainties regarding Amazon fires

Source: authors' research experience.

Feature/ contract	Conservation	Conservation +	Mechanization
Target activity	Conservation of forests aged at least 10 years	Conservation of forests aged at least 10 years	Conversion of fire- based to fire-free agriculture and keeping the latter
Target social group	All farmers	All farmers	Only smallholders (farm area <= 200 ha)
Forbidden LUT	None	reckless fire	None
Forbidden land use change	Conversion of forests aged at least 10 years	Conversion of forests aged at least 10 years	None
Inspiration for contract rules	Costa Rica's Forest Conservation (PFES) program ^a	Same as Conservation, expanded to exclude reckless fire ^b	Subsidies to mechanization by local governments

Table 2Incentive-based instruments

^a as detailed by Sánchez-Azoeifa et al. (2007); ^b in accordance with Barlow et al (2012); ^c see Simões and Schmitz (2000) and Börner et al (2007).

Land use and technology (LUT)	Forces favoring agglomeration	Forces favoring dispersion
Agriculture based on reckless fires (reckless fire)	Labor economies on burnings ^a $[\theta]$	 Accidental fire risk [p] Ecosystem services
Agriculture based on controlled fires (controlled fire)	Scale economies on fire control practices (eg, firebreaks) ^b $[\theta]$	 provided by forest to agriculture [θ] Accidental fire risk
Fire-free agriculture	Scale economies on machinery and input use [θ]	mitigation service provided by forest [p]
Forest	.Edge effects [θ] .Accidental fire risk mitigation service provided by forest [p]	None

Table 3Forces that drive agglomeration and dispersion of LUTs

"P" denotes forces which affect the profit of LUTs and "R" the forces which affect probability of accidental burns (risk).

a see Righi et al (2009) and Sorrensen (2000, 2004)

b Bowman et al (2008)

Table 4Areal share of reckless fire among small and medium-to-large holders, $t = 40^{*}$

Smallholders Intensity]	Medium	ium-to-large			
(\$/ha)	C&C	Con.+	Con.	Mech.	C&C	Con.+	Con.	Mech.
0	69%	69%	69%	69%	51%	51%	51%	51%
0.1	53%	0%	47%	0%	19%	0%	40%	52%
0.2	48%	0%	26%	0%	17%	0%	30%	51%
0.3	48%	9%	38%	0%	17%	0%	32%	52%
0.4-1	48%	0%	38%	0%	17%	0%	32%	52%

*"C&C"indicates command-and-control policy. Intensity level zero is the baseline simulation.

Intensity	Smallholders		Medium large		Total	l
(\$/ha)	Contracts	Area	Contracts	Area	Contracts	Area
0.1	1	-15	1	385	2	370
0.2	3	133	0	-93	3	40
0.3	1	11	0	-192	1	-181
0.4 - 1	0	-48	0	-189	0	-237

Table 6Difference of conservation and conservation+ contracts on counts ofcontracts and area of subsidized 10 year forest, t = 40

Intensity level ^a	Avoided fires rank ^b	Avoided deforestation rank ^c
0.1	Con.	Con.
0.1	Mech.	Con+
0.1	Con+	C&C
0.1	C&C	Mech.
0.2	Con.	Con.
0.2	Con+	Con+
0.2	Mech.	C&C
0.2	C&C	Mech.
0.3-1	Con.	Con+
0.3-1	Con+	Con.
0.3-1	Mech.	C&C
0.3-1	C&C	Mech.

Table 7Ranks for long run causal effects of instruments on fires (avoided fires)and forest (avoided deforestation)

^a The intensity level is the magnitude of the fine and subsidy in the simulations;

^b All fires are considered, i.e., accidental fires, reckless fire and controlled fire, without double-counting;

^c Calculated as the difference in the count of forested parcel between baseline and policy simulations.

Neighboring	Owner	Reason for unavailability of best LUT	Assumption made
parcel class Outside	Other	Farmers choose	Best LUT = previous LUT
Inside listed	farmers Own-	simultaneously	*
above	farmer	Available	Not needed
Inside listed below	Own- farmer	Listed below current parcel	(Needed only in the first iteration) Best LUT = previous LUT
Unlisted	Own- farmer	Listed below wealth- limitation cut-point	Not needed: unlisted parcels remain with previous LUTs

Table B.1Classification of neighboring parcels by availability of information on
best LUTs

Order	Procedure
0	[N&F&P] Initialization
1	[P&F] calculate-profit
2	[P&F] implement-LUT-portfolio
3	[F] LUC-cost-account
4	[N] define-burned-parcels
5	[G] sanction-rulebreakers
6	[F] calculate-actual-profit
7	[F] update-risk-parameters
8	[P] store-LUT-portfolio-in-memory
9	[P] update-forest-age-after-burn

Table B.2Flow of procedures, baseline and command-and-control simulations

Order	Procedure
0	[N&F&P] Initialization
1	[P&F] calculate-profit
2	[P&F] implement-LUT-portfolio
3	[F] subsidy-payment-account
4	[F] LUC-cost-account
5	[F] update-contract-duration
6	[N] define-burned-parcels
7	[F] calculate-actual-profit
8	[F] update-risk-parameters
9	[P] store-LUT-portfolio-in-memory
10	[P] update-forest-age-after-burn

Table B.3 Flow of procedures, incentive-based simulations

Convention	Profit level x 1,000	Previous LUT	Parcel-own LUT	Neighboring LUTs
PC 1	10.00	Any	Freshly abandoned	Any configuration
PC 2	20.00	Any	18 year forest	Freshly abandoned
PC 3	30.00	Any	18 year forest	18 year forest, only
PC 4	280.01	Forest	reckless fire	8 x 25yr forest
PC 5	562.48	Forest	reckless fire	3 x reckless fire + 2 x 75 yr forest
PC 6	280.04	Forest	controlled fire	1 x controlled fire + 7 x 30yr forest
PC 7	561.06	Forest	controlled fire	3 x controlled fire $+ 2 \times 80$ yr forest
PC 8	440.01	Forest	fire-free	2 x fire-free + 6 x 35yr forest
PC 9	740.02	Forest aged above 50 years	fire-free	4 x fire-free + 4 x 100yr forest
PC 10	180.00	Controlled/ reckless fire	fire-free	8 x 35yr forest

 Table C.1
 Conventions for assigning deterministic profit parameters

Note: conventions PC4-PC10 define neighboring LUT configuration in which the shift from the previous to the so-called parcel LUT is feasible (P5). This means the following condition is satisfied: $\eta_0 \theta(\tau_{k,t}, \tau'_{k,t}) \left(1 - p_k(\tau_{k,t}, \tau'_{k,t})\right) - c(\tau_{k,t-1}, \tau_{k,t}) = 0$; with $\eta_0 = 0.5$ (median landscape value), $\tau_{k,t}$ being the parcel-own in the current period and $\tau_{k,t'}$ the current LUT configuration at the neighborhood. Profit levels where calculated as $\theta_0 = \frac{1}{\eta_0} \frac{c(\tau_{k,t-1}, \tau_{k,t})}{(1-p_k(\tau_{k,t}, \tau'_{k,t}))}$.

Convention	Risk level	Parcel-own LUT	Neighboring LUT
RC 1	0.0001	Ten year forest	Ten year forest, only
RC 2	0.0002	Five year forest	Ten year forest, only
RC 3	0.0003	Five year forest	Forest with zero years, all 8 neighbors
RC 4	0.0004	Ten year forest	7/8 10 year forest & 1/8 fire-free agriculture
RC 5	0.0005	Fire-free agriculture	Ten year forest, only
RC 6	0.0006	Ten year forest	7/8 10 year forest & 1/8 controlled fire
RC 7	0.0007	Controlled fire	Ten year forest, only
RC 8	0.0008	Ten year forest	7/8 10 year forest & 1/8 reckless fire
RC 9	0.0009	Reckless fire	Ten year forest, only

 Table C.2
 Conventions for assigning risk parameters

Acknowledgements

We thank Diana Weinhold for crucial thoughts and suggestions on the basis of which the paper was structured and also Petterson Vale and Tahia Devisscher. We also thank Sergio Rivero, Rossano Ramos, Lara Steil and Toby Gardner for discussing earlier drafts. This research was funded by Darwin Initiative Fellowship project EIDPS039 conducted from July 2014 to September 2015.

References

Aragão, L.E.O.C, and Shimabukuro, E., 2010. The incidence of fire in Amazonian forests with implications for REDD. Science 328.5983 (2010): 1275-1278.

Aragão, L.E.O.C., Malhi, Y., Barbier, N., Lima, A., Shimabukuro, Y., Anderson, L., & Saatchi, S., 2008. Interactions between rainfall, deforestation and fires during recent years in the Brazilian Amazonia. Philosophical Transactions of the Royal Society B: Biological Sciences, 363(1498), 1779-1785.

Arco-Verde, M.F. e Amaro, G., 2012. Cálculo de Indicadores Financeiros para Sistemas Agroflorestais. Boa Vista, RR: Embrapa Roraima, 2012. 48p. (Documentos / Embrapa Roraima, 44).

Arfaoui, N., Brouillat, E., Saint Jean, M., 2014. Policy design and technological substitution: Investigating the REACH regulation in an agent-based model. Ecological Economics, 107, 347-365.

Arima, E. Y., Simmons, C.S., Walker, R.T., Cochrane, M.A., 2007. Fire in the Brazilian Amazon: a spatially explicit model for policy impact analysis. Journal of Regional Science 47:541-567.

Arima, E. Y., Richards, P., Walker, R., Caldas, M. M., 2011. Statistical confirmation of indirect land use change in the Brazilian Amazon. Environmental Research Letters, 6(2), 024010.

Arthur, W. B. 1989. Competing technologies, increasing returns, and lock-in by historical events. The economic journal, 116-131.

Balch, J. K. (2014). Atmospheric science: Drought and fire change sink to source. Nature, 506(7486), 41-42.

Barlow, J., Parry, L., Gardner, T. A., Ferreira, J., Aragão, L. E., Carmenta, R., Berenguer, E., Vieira, I.C.G., Souza, C., Cochrane, M. A., 2012. The critical importance of considering fire in REDD+ programs. Biological Conservation, 154, 1-8.

Barona, E., Ramankutty, N., Hyman, G., & Coomes, O. T., 2010. The role of pasture and soybean in deforestation of the Brazilian Amazon. Environmental Research Letters, 5(2), 024002.

Becker, G. S. (1968). Crime and punishment: An economic approach. In The Economic Dimensions of Crime (pp. 13-68). Palgrave Macmillan UK.

Berger, T., 2001. Agent-based spatial models applied to agriculture: a simulation tool for technology diffusion, resource use changes and policy analysis. Agricultural economics 25.2 (2001): 245-260.

Börner, J., Mendoza, A., Vosti, S.A., 2007. Ecosystem services, agriculture, and rural poverty in the Eastern Brazilian Amazon: Interrelationships and policy prescriptions. Ecological Economics 64.2 (2007): 356-373.

Börner, J., Kis-Katos, K., Hargrave, J., & König, K., 2015. Post-Crackdown Effectiveness of Field-Based Forest Law Enforcement in the Brazilian Amazon. PLoS ONE 10(4): e0121544. doi:10.1371/journal.pone.0121544.

Börner, J., Wunder, S., Reimer, F., Bakkegaard, R. Y., Viana, V., Tezza, J., Pinto, T., Lima, L., Marostica, S., 2013. Compensação por serviços ambientais, meios de vida e conservação: o Programa Bolsa Floresta. http://fas-amazonas.org/versao/2012/wordpress/wp-content/uploads/2014/02/BF_report_PORT_web.pdf

Bowman, M. S., Amacher, G. S., Merry, F. D., 2008. Fire use and prevention by traditional households in the Brazilian Amazon. Ecological Economics 67 (2008) 117 – 130.

Brando, P. M., Balch, J. K., Nepstad, D. C., Morton, D. C., Putz, F. E., Coe, M. T., Silvério, D., Macedo, M. N., Davidson, E. A., Nóbrega, C.C., Alencar, A., Soares-Filho, B. S., 2014. Abrupt increases in Amazonian tree mortality due to drought–fire interactions. Proceedings of the National Academy of Sciences, 111(17), 6347-6352.

Brasil, 1998. Decreto nº 2.661, de 8 de julho de 1998. Presidência da República.

Cammelli, F.,2014. Smallholders' collective action and fire risk in the Brazilian Amazon. Master thesis. University of Firenze, Italy.

Carmenta, R., Vermeylen, S., Parry, L., & Barlow, J., 2013. Shifting cultivation and fire policy: insights from the Brazilian Amazon. Human ecology, 41(4), 603-614.

Cederman, L. E., & Girardin, L., 2007. Toward realistic computational models of civil wars. In presentation at the Annual Meeting of the American Political Science Association, Chicago, IL.

https://www.researchgate.net/profile/Luc_Girardin2/publication/228337651_Toward_realistic_computational_models_of_civil_wars/links/550b322f0cf285564097013c.pdf

Chen, Y., Randerson, J. T., Morton, D. C., DeFries, R. S., Collatz, G. J., Kasibhatla, P. S., Giglio, L., Jin, Y., Miriam E. Marlier, M.E., 2011. Forecasting fire season severity in South America using sea surface temperature anomalies. Science, 334(6057), 787-791

Chomitz, K. M., & Kumari, K., 1998. The domestic benefits of tropical forests: a critical review. The World Bank Research Observer, 13(1), 13-35.

Cochrane, M., 2009. Tropical fire ecology: climate change, land use and ecosystem dynamics. Springer Science & Business Media.

Coe, M. T., Marthews, T. R., Costa, M. H., Galbraith, D. R., Greenglass, N. L., Imbuzeiro, H. M., ... & Powell, T. L., 2013. Deforestation and climate feedbacks threaten the ecological integrity of south–southeastern Amazonia. Philosophical Transactions of the Royal Society B: Biological Sciences, 368(1619), 20120155.(doi:10.1098/rstb.2012.0155).

Costa, L. M., 2004. Sob o fogo cruzado das campanhas: ambientalismo, comunicação e agricultura familiar na prevenção ao fogo acidental na Amazônia. PhD dissertation. Belém-NAEA.

Davidson, E. A., de Araújo, A. C., Artaxo, P., Balch, J. K., Brown, I. F., Bustamante, M. M., ... & Wofsy, S. C., 2012. The Amazon basin in transition. Nature, 481(7381), 321-328.

Deadman, P., Robinson, D., Moran, E., & Brondizio, E.,2004. Colonist household decisionmaking and land-use change in the Amazon Rainforest: an agent-based simulation. Environment and Planning B, 31, 693-710.

Doole, G. J., & Kingwell, R. (2015). Efficient economic and environmental management of pastoral systems: Theory and application. Agricultural Systems, 133, 73-84.

El-Sayed, A. M., Scarborough, P., Seemann, L., & Galea, S. 2014. Social network analysis and agent-based modeling in social epidemiology. Epidemiologic Perspectives and Innovations, 9(1), 9.

Emater, 2015a. Semistructured interview with senior extensionist of the Belem unit, Para state, March 2015.

Emater, 2015b. Semistructured interviews with chief-extensionists of Santarém unit, Para state, April 2015.

Ferraro, P. J., & Kiss, A., 2002. Direct payments to conserve biodiversity. Science, 298(5599), 1718.

Ferraro, P. J., 2008. Asymmetric information and contract design for payments for environmental services. Ecological economics, 65(4), 810-821.

Ferraz, G., Russell, G. J., Stouffer, P. C., Bierregaard, R. O., Pimm, S. L., & Lovejoy, T. E., 2003. Rates of species loss from Amazonian forest fragments. Proceedings of the National Academy of Sciences, 100(24), 14069-14073.

Fezzi C., Bateman, I.J. 2011. Structural agricultural land use modeling for spatial agroenvironmental policy analysis, American Journal of Agricultural Economics, vol. 93, pp. 1168-1188.

Friess, D. A., Phelps, J., Garmendia, E., & Gómez-Baggethun, E., 2015. Payments for Ecosystem Services (PES) in the face of external biophysical stressors. Global Environmental Change, 30, 31-42.

Fujita, M., Thisse, J. F., 2013. Economics of agglomeration: cities, industrial location, and globalization. Cambridge university press.

Godar, J., Gardner, T. A., Tizado, E. J., & Pacheco, P., 2014. Actor-specific contributions to the deforestation slowdown in the Brazilian Amazon. Proceedings of the National Academy of Sciences 111.43 (2014): 15591-15596.

Grimm, V., Berger, U., DeAngelis, D.L., Polhill, J.G., Giske, J., Railsback, S.F., 2010. The ODD protocol: a review and first update. Ecol. Model. 221, 2760–2768.

Gulati, S., & Vercammen, J., 2006. Time inconsistent resource conservation contracts. Journal of environmental economics and management, 52(1), 454-468.

Hall, A., 2008. Better RED than dead: paying the people for environmental services in Amazonia. Philosophical Transactions of the Royal Society B: Biological Sciences, 363(1498), 1925-1932.

Happe, K., Balmann, A., Kellermann, K., & Sahrbacher, C., 2008. Does structure matter? The impact of switching the agricultural policy regime on farm structures. Journal of Economic Behavior & Organization, 67(2), 431-444.

IBAMA-PA, 2015. Personal communication with the deputy superintendent of IBAMA in Para State, March 2015.

INPE-EMBRAPA, 2012. Land use map for Brazilian Amazon. Terraclass project. Available at: http://www.inpe.br/cra/projetos_pesquisas/terraclass2010.php

INPE, 2015. Frequently asked questions on fire database. Brazilian Space Research Institute. http://www.inpe.br/queimadas/faq.php

Irwin, E. G.,Bockstael, N.E.,2001. The problem of identifying land use spillovers: measuring the effects of open space on residential property values. American journal of agricultural economics (2001): 698-704.

Jacobson, L. D. S. V., de Souza Hacon, S., de Castro, H. A., Ignotti, E., Artaxo, P., Saldiva, P. H. N., & de Leon, A. C. M. P. (2014). Acute effects of particulate matter and black carbon from seasonal fires on peak expiratory flow of schoolchildren in the Brazilian Amazon. PloS one, 9(8), e104177.

Jiménez-Muñoz, J. C., Mattar, C., Barichivich, J., Santamaría-Artigas, A., Takahashi, K., Malhi, Y., ... & van der Schrier, G. (2016). Record-breaking warming and extreme drought in the Amazon rainforest during the course of El Niño 2015–2016. Scientific Reports, 6.

Just, R. E., Zilberman, D., Hochman, E., 1983. Estimation of Multicrop Production Functions. American Journal of Agricultural Economics, Vol. 65, No. 4 (Nov., 1983), pp. 770-780

Kato, O. Researcher of the Brazilian Agricultural Research Corporation and coordinator of "Tipitamba" project. Interview conducted at Embrapa CPATU April 2015.

Klemick, H., 2011. Shifting cultivation, forest fallow, and externalities in ecosystem services: Evidence from the Eastern Amazon. Journal of Environmental Economics and Management, 61(1), 95-106.

Knudson, W. A., 2009. The environment, energy, and the Tinbergen rule. Bulletin of Science, Technology & Society, 29(4), 308-312.

Kremmydas, D., 2012. Agent based modeling for agricultural policy evaluation. AUA Working Paper Series No.2012-3. http://aoatools.aua.gr/RePEc/excel/2012-3/2012-3_kremmydas.pdf

Krugman, P. R., 1996. The self-organizing economy (pp. 122-p). Oxford: Blackwell.

Laurance, W. F., Nascimento, H. E., Laurance, S. G., Andrade, A., Ribeiro, J. E., Giraldo, J. P., ... & D'Angelo, S., 2006. Rapid decay of tree-community composition in Amazonian forest fragments. Proceedings of the National Academy of Sciences, 103(50), 19010-19014.

Leiva-Montoya, R., 2013. An Assessment of the Introduction of REDD+ in Brazil: A case study of the Bolsa Floresta Programme in the Rio Negro Sustainable Reserve.Norwegian University of Life Sciences. Master Thesis. http://brage.bibsys.no/xmlui/handle/11250/187935

Lobianco, A., 2007. The effects of decoupling on two Italian regions. Associazone Alessandro Bartola, PhD studies. https://lobianco.org/antonello/_media/academic:pubs:phdstudies2.pdf

Malhi, Y., Aragão, L. E., Galbraith, D., Huntingford, C., Fisher, R., Zelazowski, P., Sitche, S., McSweeneya, C., Meir, P., 2009. Exploring the likelihood and mechanism of a climatechange-induced dieback of the Amazon rainforest. Proceedings of the National Academy of Sciences, 106(49), 20610-20615.

Marshall, B. D., & Galea, S., 2014. Formalizing the role of agent-based modeling in causal inference and epidemiology. American journal of epidemiology, kwu274

Mas-Colell, A., Whinston, M. D., & Green, J. R. (1995). Microeconomic theory (Vol. 1). New York: Oxford university press.

McCracken, S. D., Siqueira, A. D., Moran, E. F., Brondízio, E. S., Wood, C. H., & Porro, R., 2002. Land use patterns on an agricultural frontier in Brazil: insights and examples from a demographic perspective. Deforestation and land use in the Amazon, 162-192.

Mendonça, M. J. C., Diaz, M. D. C. V., Nepstad, D., da Motta, R. S., Alencar, A., Gomes, J. C., Ortiz, R. A., 2004. The economic cost of the use of fire in the Amazon. Ecological Economics, 49(1), 89-105.

MMA, 2009. Amazônia sem fogo: programa de formação técnica sobre as alternativas ao uso do fogo no processo de desenvolvimento sustentável da Região Amazônica / Ministério do Meio Ambiente e Embaixada da Itália em Brasília. – Brasília: MMA, 2009.

Moran, E. F., Brondizio, E. S., McCracken, S. D., Wood, C. H., & Porro, R., 2002. Trajectories of land use: soils, succession, and crop choice. Deforestation and land use in the Amazon, 193-217.

Morgan, S. L., & Winship, C., 2007. Counterfactuals and causal inference. 1st edition. Cambridge University Press.

Müller, B., Bohn, F., Dreßler, G., Groeneveld, J., Klassert, C., Martin, R.,Schlüter, M., Schulze, J.,Weise, H., Schwarz, N.,2013. Describing human decisions in agent-based models–ODD+ D, an extension of the ODD protocol. Environmental Modelling & Software, 48, 37-48.

Neeff, T., and Santos, J.R., 2005. A growth model for secondary forest in Central Amazonia. Forest Ecology and Management 216.1 (2005): 270-282.

Nepstad, D. C., Carvalho, G. Barros, C., 2001. Road paving, fire regime feedbacks, and the future of Amazon Forest. Forest Ecology and Management 154 (2001) 396-407.

Nepstad, D.C., 2007. The Amazon's Vicious Cycles - Drought and Fire in the Greenhouse. Report to the World Wide Fund for Nature, WWF. http://assets.panda.org/downloads/amazonas_eng_04_12b_web.pdf

Neto, P. S. F.,2008. Avaliação do Proambiente. Relatório de consultoria. Ministério do Meio Ambiente e Ministério do Desenvolvimento Agrário. Brasília. 2008. Disponível em http://www.mma.gov.br/estruturas/sds_proambiente/_arquivos/33_05122008040536.pdf.

Parker, D. C., & Meretsky, V.,2004. Measuring pattern outcomes in an agent-based model of edge-effect externalities using spatial metrics. Agriculture, Ecosystems & Environment, 101(2), 233-250.

Pattanayak, S. K., Wunder, S., & Ferraro, P. J.,2010. Show me the money: Do payments supply environmental services in developing countries?. Review of Environmental Economics and Policy, req006.

Peres, C. A., Lake, I. R., 2003. Extent of nontimber resource extraction in tropical forests: accessibility to game vertebrates by hunters in the Amazon basin. Conservation Biology, 17(2), 521-535.

Pfaff, A. S., 1999. What drives deforestation in the Brazilian Amazon?: Evidence from satellite and socioeconomic data. Journal of Environmental Economics and Management, 37(1), 26-43.

Pfaff, A., Robalino, J., Walker, R., Aldrich, S., Caldas, M., Reis, E., Perz, S., Boher, C., Arima, E., Laurance, W., Kirby, K., 2007. Road Investments, Spatial Spill-overs and Deforestation in the Brazilian Amazon. Journal of Regional Science, 47(1), 109-123.

PREVFOGO, 2015. Personal communication with staff from national coordination of the Brazilian fire prevention and control program. Brasilia, Brazil. May 2014 and March 2015.

Reichert, J. M., Bervald, C. M. P., Rodrigues, M. F., Kato, O. R., & Reinert, D. J.,2014. Mechanized land preparation in eastern Amazon in fire-free forest-based fallow systems as alternatives to slash-and-burn practices: Hydraulic and mechanical soil properties. Agriculture, Ecosystems & Environment, 192, 47-60.

RIGHI, C. A.; LIMA, GRAÇA, P.M.L.A., C., C. C.; FEIGL, B. J., FEARNSIDE, P. M.,2009. Biomass burning in Brazil's Amazonian "arc of deforestation": Burning efficiency and charcoal formation in a fire after mechanized clearing at Feliz Natal, Mato Grosso. Forest Ecology and Management 258 (2009) 2535–2546.

Röhmel, J. (1996). Precision intervals for estimates of the difference in success rates for binary random variables based on the permutation principle. Biometrical journal, 38(8), 977-993.

Sánchez-Azoeifa, G. A., Pfaff, A., Robalino, J. A., & Boomhower, J. P.,2007. Costa Rica's payment for environmental services program: intention, implementation, and impact. Conservation Biology, 21(5), 1165-1173.

Scatena, F. N., Walker, R. T., Homma, A. K. O., de Conto, A. J., Ferreira, C. A. P., de Amorim Carvalho, R., Rocha, A.C.P.N., Santos, A.I.M., Oliveira, P. M.,1996. Cropping and fallowing sequences of small farms in the "terra firme" landscape of the Brazilian Amazon: a case study from Santarem, Para. Ecological Economics, 18(1), 29-40.

Secretaria de estado de meio ambiente do Acre (SEMA-AC), 2011. Plano integrado de prevenção, controle e combate às queimadas e aos incêndios florestais do estado do Acre.

Serra, A. B., 2005. Indicadores de sustentabilidade do solo em sistemas alternativos ao uso do fogo, baseados nos princípios da agroecologia, desenvolvidos por agricultores familiares na região da rodovia transamazônica-oeste do Pará. PhD. dissertation, Universidade Federal do Pará.

Simões, A., & Schmitz, H. H.,2000. Intensificação de sistemas de produção através da mecanização na região da Transamazônica: limites e possibilidades. Novos Cadernos Naea, 3(2).

Sorrensen, C. L., 2000. Linking smallholder land use and fire activity: examining biomass burning in the Brazilian Lower Amazon. Forest Ecology and Management, 128(1), 11-25.

Sorrensen, C.L.,2004. Contributions of fire use study to land use/cover change frameworks: understanding landscape change in agricultural frontiers. Human Ecology 32:395-419.

Sorrensen, C., 2009. Potential hazards of land policy: Conservation, rural development and fire use in the Brazilian Amazon. Land Use Policy, 26(3), 782-791

Steil, L., 2009. Legislação Ambiental pertinente ao Tema Fogo (Brazilian fire legislation). Document delivered through eletronic mail by the coordinator of PREVFOGO. May 2014.

Stosic, T., Telesca, L., da Costa, S. L. L., & Stosic, B. (2016). Identifying drought-induced correlations in the satellite time series of hot pixels recorded in the Brazilian Amazon by means of the detrended fluctuation analysis. Physica A: Statistical Mechanics and its Applications, 444, 660-666.

Stouffer, P.C., Bierregaard Jr., R.,1995. BiologyEffects of Forest Fragmentation on Understory Hummingbirds in Amazonian Brazil. Conservation Biology, Vol. 9, No. 5 (Oct., 1995), pp. 1085-1094

Thiele, J. C., 2014. RNetLogo package. https://cran.rproject.org/web/packages/RNetLogo/index.html

Toniolo, M. A., 2004. The role of land tenure in the occurrence of accidental fires in the Amazon region: case studies from the national forest of Tapajos, Para, Brazil. (Doctoral dissertation, University of Indiana).

Townsley, M., & Birks, D. J., 2008. Building better crime simulations: systematic replication and the introduction of incremental complexity. Journal of Experimental Criminology, 4(3), 309-333.

T. Tscharntke, A.M. Klein, A. Kruess, I. Steffan-Dewenter, C. Thies., 2005. Landscape perspectives on agricultural intensification and biodiversity – ecosystem service management. Ecology letters, 8(8), 857-874. Ecology Letters, 8 (2005), pp. 857–874

Van der Werf, E., & Di Maria, C.,2012. Imperfect environmental policy and polluting emissions: the green paradox and beyond. International Review of Environmental and Resource Economics, 6(2), 153-194.

van Vliet, N., Adams, C., Vieira, I. C. G., & Mertz, O.,2013. "Slash and Burn" and "Shifting" Cultivation Systems in Forest Agriculture Frontiers from the Brazilian Amazon. Society & Natural Resources, 26(12), 1454-1467.

Varian, H. R. (1992). Microeconomic analysis. Norton & Company.

Vasconcelos, S. S., Fearnside, P. M., de Alencastro Graça, P. M. L., Dias, D. V., & Correia, F. W. S., 2013. Variability of vegetation fires with rain and deforestation in Brazil's state of Amazonas. Remote Sensing of Environment, 136, 199-209.

Wunder, S. 2006. Are direct payments for environmental services spelling doom for sustainable forest management in the tropics. Ecology and Society, 11(2), 23.

Wunder, S., Börner, J., Tito, M. R., & Pereira, L., 2009. Pagamentos por serviços ambientais: perspectivas para a Amazônia Legal. Ministério do Meio Ambiente.Série Estudos 10.http://www.bibliotecaflorestal.ufv.br/bitstream/handle/123456789/12379/Livro_Pagament os-por-servi%C3%A7os-ambientais-Amaz%C3%B4nia-

Legal_MMA.pdf?sequence=1&isAllowed=y

Zhao, M., Tong, D., & Zhao, G., 2012. Modelling and Simulation the Environmental Policy using Multi-agent Platform. Journal of Convergence Information Technology, 7(12).