

Inverse Modeling of Unsaturated Flow Using Clusters of Soil Texture and Pedotransfer Functions

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1 **Key points:**

- 2 • Two new ways to parameterize vadose zone hydraulic properties based on soil texture are proposed
- 3 and analyzed.
- 4 • One of these preserves heterogeneity with only a few adjustable parameters.
- 5 • The two approaches are compared through application to deep vadose zone experimental data.

6 **Abstract**

7 Characterization of heterogeneous soil hydraulic parameters of deep vadose zones is often difficult
8 and expensive, making it necessary to rely on other sources of information. Pedotransfer functions
9 (PTFs) based on soil texture data constitute a simple alternative to inverse hydraulic parameter
10 estimation, but their accuracy is often modest. Inverse modeling entails a compromise between detailed
11 description of subsurface heterogeneity and the need to restrict the number of parameters. We
12 propose two methods of parameterizing vadose zone hydraulic properties using a combination of k -
13 means clustering of kriged soil texture data, PTFs and model inversion. One approach entails
14 homogeneous and the other heterogeneous clusters. Clusters may include subdomains of the
15 computational grid that need not be contiguous in space. The first approach homogenizes within-
16 cluster variability into initial hydraulic parameter estimates that are subsequently optimized by
17 inversion. The second approach maintains heterogeneity through multiplication of each spatially
18 varying initial hydraulic parameter by a scale factor, estimated *a posteriori* through inversion. This
19 allows preserving heterogeneity without introducing a large number of adjustable parameters. We
20 use each approach to simulate a 95-day infiltration experiment in unsaturated layered sediments at a
21 semiarid site near Phoenix, Arizona, over an area of $50 \times 50 \text{ m}^2$ down to a depth of 14.5 m. Results
22 show that both clustering approaches improve simulated moisture contents considerably in comparison
23 to those based solely on PTF estimates. Our calibrated models are validated against data from a
24 subsequent 295-day infiltration experiment at the site.

25

26 **1. Introduction**

27 Modeling of vadose zone flow and transport processes requires characterization of subsurface
28 architecture and hydraulic properties. Information about lithotype distribution can often be obtained
29 from well logs, ground penetrating radar (GPR) [*Kowalsky et al.*, 2005], electrical resistance
30 tomography (ERT) [*Yeh*, 2002; *Liu and Yeh*, 2004] or seismic tomography [*Nolet*, 1987; *Tromp et al.*,
31 2005]. In deep vadose zones, hydraulic parameters are difficult if not impossible to measure in situ.
32 Acquiring undisturbed samples and determining their hydraulic properties in the laboratory is likewise
33 difficult and expensive. A common alternative is to characterize hydraulic properties indirectly, on
34 the basis of soil composition and/or flow data, through pedotransfer functions (PTFs) and/or inverse
35 modeling.

36 PTFs allow the estimation of soil hydraulic properties using empirical correlations between
37 hydraulic characteristics, soil texture and quantities such as soil bulk density. Soil texture and bulk
38 density are generally easier and less expensive to assess than hydraulic properties [e.g., *Rawls et al.*,
39 1982; *Pachepsky et al.*, 2006]. The combination of a PTF with high-density sampling and/or
40 geospatial modeling of soil texture and bulk density should, in principle, allow one to resolve
41 subsurface hydraulic properties in detail. The accuracy of PTFs is, however, often modest when
42 applied to data collected independently of those employed for PTF calibration [*Schaap and Leij*, 1998].
43 Calibration of PTFs against site-specific data brings about improvement [*Ye et al.*, 2007] but also
44 requires considerable effort and cost. As shown by *Wang et al.* [2003] and suggested further by our
45 study, PTF-derived estimates of hydraulic properties tend to result in systematic errors that can (and

46 we believe should) be reduced by calibrating these properties further against observed state variables,
47 such as moisture content and/or pressure head, via inverse modeling.

48 Inverse modeling in hydrogeology typically entails the following steps [e.g., *Neuman*, 1973;
49 *Carrera and Neuman* 1986b, 1986c; *Neuman*, 2003; *Franssen et al.*, 2009]: 1) proposition of a
50 conceptual model (or a set of alternative conceptual models) of the system under study in terms of
51 geological/sedimentological structure, mathematical rendition of (mass, energy, momentum)
52 conservation principles, initial and boundary conditions, as well as forcing terms; 2) parametrization
53 of the system model with the objective of characterizing (either in a stochastic or deterministic fashion)
54 spatial variability of critical parameters throughout the domain; and 3) estimation of model parameters
55 by minimizing a suitable measure of mismatch between observed and simulated state variables (e.g.,
56 moisture content or pressure heads). In some cases, this measure includes prior information about
57 model structure and parameters. The latter is typically embedded in a regularization (or plausibility)
58 term, which penalizes deviations of estimated parameters from prior values and helps stabilize the
59 inverse solution [*Neuman*, 1973; *Carrera and Neuman* 1986a, 1986b; *Vrugt and Bouten* 2002].
60 Inverse methods can be combined with numerically intensive Monte Carlo techniques to cope with
61 propagation of uncertainty associated with spatial parameter distributions (and/or initial conditions and
62 forcing terms) to state variables of interest [e.g., *Zimmerman et al.*, 1998; *Chen and Zhang*, 2006].

63 Inverse modeling of unsaturated flow is more challenging than that of saturated flow. Notably,
64 functional dependence of soil moisture content and hydraulic conductivity on matric potential leads to
65 high-dimensionality issues in the parameter space, even under conditions where closed-form

66 expressions of these models such as the Brooks-Corey-Burdine [*Burdine*, 1953; *Brooks and Corey*,
67 1964] or van Genuchten-Mualem formulations are used [*Mualem* 1976; *van Genuchten*, 1980].
68 Reviews of inverse methods in the context of vadose zone hydrology are found in *Hopmans and*
69 *Simunek* [1999], *Hopmans et al.* [2002], and *Vrugt et al.* [2008]. Inverse methods based on zonation
70 (subdivision of the flow domain into uniform subdomains [*Emsellem and De Marsily*, 1971; *Neuman*
71 *and Yakowitz*, 1979; *Wildenschild and Jensen*, 1999; *Wang et al.*, 2003; *Vrugt et al.*, 2008]) are used
72 among practitioners due to their relatively straightforward implementation and its flexibility to
73 accommodate geological information. However, the number of zones (and so the resolution of
74 heterogeneity) is limited by the need to avoid overparameterization. A variation on the zonation method
75 to resolve subsurface heterogeneity is to use similar media scaling [*Miller and Miller*, 1956; *Vogel et*
76 *al.*, 1991], which relies on the dependence of hydraulic properties on pore size and pore geometry
77 descriptors. This allows scaling of hydraulic water retention and unsaturated hydraulic conductivity
78 functions of multiple soils to unique reference functions [e.g. *Tuli et al.*, 2001; *Das et al.*, 2005; *Nasta*
79 *et al.*, 2013]. *Zhang et al.* [2004] used a Combined Parameter Scaling and Inversion Technique
80 (CPSIT) to estimate the hydraulic properties of Equivalent Hydraulic Media (EHMs). Their method
81 requires that soil hydraulic parameters at the local scale be determined using the same method as that
82 used for the experimental site and at the same spatial scale, i.e., core size. In their approach, ratios of
83 hydraulic properties in different EHMs relative to hydraulic properties in reference EHMs remain fixed
84 during inversion. Therefore, field scale hydraulic parameters of reference EHMs can be estimated
85 during inversion from local scale values. *Zhang et al.* [2004] recognized this to be a limitation in that

86 any local-scale parameter estimation error transfers to the field scale.

87 Inverse methods combined with geostatistical methods constitute an alternative approach to
88 estimate soil hydraulic parameters. The Pilot Point Method (PPM), introduced by *de Marsily* [1984],
89 is one such approach and consists of calibrating an initial kriged parameter field, generated from the
90 measured values of hydraulic parameters and a set of additional parameter values (which are unknown
91 prior to calibration) at selected unmeasured locations in the simulation domain, called “pilot points”.
92 The location of pilot points can also be incorporated into the inverse problem to find the optimal
93 position using a couple of adjoint sensitivity analysis and kriging [*LaVenue and Pickens*, 1992;
94 *RamaRao et al.*, 1995; *Franssen et al.*, 2009]. The PPM method has mostly been applied to saturated
95 problems and has found little application in vadose zone systems. *Kowalsky et al.* [2005] used the
96 PPM to derive the distribution of permeability using GPR and hydrological measurements collected
97 during a transient flow experiment. *Morales-Casique et al.* [2010] calibrated log permeability and
98 porosity at selected pilot points against observed pressures in two pneumatic injection tests of
99 unsaturated fractured tuff in Arizona.

100 In this study, we introduce two ways of parameterizing deep vadose zone hydraulic properties based
101 on *k*-means clustering of kriged soil textural data, a pedotransfer function (PTF) and numerical
102 inversion of a vadose zone flow model. In contrast to traditional zonation often employed in vadose
103 zone inverse modeling, a cluster in our model may (and generally does) consist of noncontiguous
104 subdomains. The initial hydraulic parameters at each grid point in a cluster are estimated with a PTF.
105 Our approach admits that these initial PTF estimates entail systematic errors [*Schaap and Leij*, 1998]

106 which are not known *a priori* [Romano, 2004; Chirico et al., 2007; Assouline and Or, 2013]. The
107 purpose of our inverse analysis is precisely to minimize these, and ancillary random, errors in parameter
108 estimates. Our approach is predicated on the belief that it is better to rely on reasonably well founded
109 (if not entirely accurate) PTF-derived initial parameter estimates than on other, less robust alternatives.
110 In the first homogeneous cluster approach, each hydraulic parameter (or its logarithm) is averaged over
111 all grid points in a cluster to yield prior hydraulic parameter estimates; posterior estimates (which are
112 uniform within each cluster but differ between clusters) are then estimated by optimizing the fit
113 between computed and observed moisture contents. All prior and posterior parameter estimates
114 within a cluster are homogeneous. In the heterogeneous cluster approach, prior hydraulic parameter
115 estimates vary from one grid point to another. Posterior estimates in a cluster are expressed as
116 products of corresponding prior estimates and a cluster-specific scaling factor. Scaling factors of all
117 parameters in all clusters are then estimated by the same criteria. Both of our approaches are
118 evaluated by model quality measures. We use our approach to simulate a 95-day infiltration
119 experiment in unsaturated layered sediments at a semiarid site near Phoenix, Arizona, over an area of
120 $50 \times 50 \text{ m}^2$ down to a depth of 14.5 m. We then validate our calibrated models against data from a
121 subsequent 295-day infiltration experiment at the site.

122

123 **1.1. Description of Site and Infiltration Experiment**

124 The data used in this study were collected at the University of Arizona Maricopa Agricultural
125 Center (latitude 33.069478 N, longitude 111.973667 W), Arizona, USA, between 1997 and 2004. Four

126 deep vadose zone infiltration experiments were conducted at this site to test the effectiveness of several
127 vadose zone monitoring instruments and modeling techniques. The site was nominally $60 \times 60 \text{ m}^2$ in
128 the horizontal direction and 15 m in the vertical direction and situated in alluvial valley deposits with
129 a textural composition ranging from gravel to clay. An impermeable pond liner was used to eliminate
130 evaporation and rainfall and an inner area of $50 \times 50 \text{ m}^2$ (Figure 1) was outfitted with 164 irrigation
131 driplines containing emitters spaced 30 cm apart. Major instrumentation included nine neutron
132 thermalization wells with depths down to 14.25 m (numbered 402...445 in Figure 1) and tensiometers
133 placed one meter south of each well at depths of 3, 5 and 10 m. A perched groundwater table was
134 observed at a depth of about 13 m. Detailed descriptions of the site and instrument calibration are
135 provided by *Young et al.* [1999], *Wang et al.* [2002] and *Schaap* [2013].

136 Here we focus on data from Experiment 3 and 4 used, respectively, for model calibration and
137 validation. Experiment 3 started on 17 January 2001 (Day-of-Year 17, hereafter termed as DOY 17)
138 and ended on 28 January 2002 (corresponding to DOY 393). An extensive 800 day drainage period
139 preceded Experiment 3, resulting in a nearly constant soil moisture content profile, as verified by
140 neutron thermalization measurements on DOY 17.5, 47.5, 67.5, and 108.5. Drip irrigation (and
141 associated infiltration) started at noon on 24 April 2001 (DOY 114.5) and ended 28 days later at noon
142 on 22 May 2001 (DOY 142.5). With minor interruptions, metered irrigation was applied six times a
143 day at a mean rate of 27.2 mm/day; about 16 mm of water was applied before DOY114.5 to test the
144 irrigation system. Tensiometer readings indicated that full saturation conditions did not occur at any of
145 the monitored locations.

146 Neutron thermalization was conducted on 42 dates with 0.25 m increments from a depth of 0.25 m
147 down to 12.5 m; neutron count ratios were converted to soil moisture contents using a texture-
148 dependent calibration model presented in *Schaap* [2013]. Sparse data at depths greater than 12.5 m
149 were also available at some wells, but were not used in this study. Data from well 442 (Figure 1)
150 were not considered due to evidence of lateral flow from a flood irrigated field immediately to the north
151 of the site. Data from well 405 at depths of 5.0 - 10.0 m were likewise not used because of
152 anomalously dry readings, presumably due to large air pockets around the PVC well casing. For
153 model calibration data collected between DOY 67.5 and DOY 163.5 were used, which included 29
154 dates with 11,020 individual moisture content observations. There were two observation dates before
155 the start of irrigation (DOY 114.5), while sampling took place every one to three days during irrigation
156 and at approximately weekly intervals during subsequent redistribution. As explained in Section
157 2.2.1, the relatively sparse data before infiltration caused some problems with obtaining physically
158 realistic initial moisture contents.

159 Experiment 4 started on 26 March 2002 (DOY 450) and ended 295 days later on 14 January 2003
160 (DOY 744). Irrigation started on 26 March 2002 (DOY 450) ended 230 days later on 11 November
161 2002 (DOY 680). It was followed by a 65-day drainage period that ended on 14 January 2003 (DOY
162 744). The site was irrigated for 5 minutes, 12 times a day, at a mean rate of 26.8 mm/day. Neutron
163 counts (and, correspondingly, water contents) were measured on 32 dates; 26 dates were measured
164 during the infiltration period and 6 dates during the drainage period. They were conducted at vertical
165 increments of 0.25 m from depth 0.25 m down to 12.5 m in 9 boreholes. Neutron depth coverage was

166 less consistent than in Experiment 3 in that measurements were taken preferentially near the infiltration
167 front, less frequently below it during the infiltration period. Discarding unreliable observations in well
168 442 due to the same reason as in Experiment 3 left us with a total of 9,297 neutron count values for
169 validation.

170

171 **1.2. Geospatial Analysis of Soil Texture and Bulk Density**

172 No reliable measurements of site's hydraulic characteristics are available. For this reason, we
173 estimate these characteristics using PTFs based on a geospatial analysis of texture and bulk density
174 data. *Wang* [2002] performed a three-dimensional geostatistical analysis of 520 texture samples
175 collected at depths down to 5 m. *Schaap* [2013] reanalyzed an extended dataset of 1042 soil texture
176 and 250 bulk density samples down to a depth of 15 m and identified two principal components (PC1
177 and PC2) extracted from measured sand, silt, and clay percentages, with PC1 accounting for 92% of
178 textural triangle variance and PC2 accounting for the remaining 8% of this variance. Residual
179 variograms were obtained upon subtracting spatially-averaged vertical trends from PC1, PC2 and bulk
180 density data (see *Schaap* [2013] for details).

181 Using the same data as *Schaap* [2013], *Guadagnini et al.* [2013] showed that the univariate
182 distribution of texture is non-Gaussian, rendering texture amenable to representation as a sub-Gaussian
183 random field, key statistics of which vary with scale. The same is true for hydraulic parameters
184 estimated from textural data [*Guadagnini et al.*, 2014] using the pedotransfer function Rosetta of
185 *Schaap et al.* [2001]. Based on results obtained by *Guadagnini et al.* [2013] and [*Guadagnini et al.*,

186 2014] we recognize that the current kriging approach (see below) has the potential to produce a bias in
187 the prior estimates, which can be reduced, but not eliminated [*Grondona and Cressie*, 1991], through
188 an iterative approach proposed by *Neuman and Jacobson* [1984]. As a result, the approach followed
189 may yield somewhat biased prior hydraulic parameter estimates because kriging of texture and bulk
190 density provides a smooth estimate of actual spatial variability while sub-Gaussian fields may present
191 a more accurate description of the prior parameter estimate.

192 Because the development of conditional simulation and kriging-based estimation of sub-Gaussian
193 fields of the kind found by *Guadagnini et al.* [2013, 2014] is still in its infancy [*Riva et al.*, 2015;
194 *Panzeri et al.*, 2016], we cannot yet simulate conditional sub-Gaussian random fields. Quantifying the
195 potential bias of the kriging approach is impossible. However, we expect it to impact our prior
196 parameter estimates to a greater extent than our posterior estimates, which depend strongly on
197 additional data (in our case, water contents) and are known to be generally less biased.

198 Our study is thus confined to the kriged three-dimensional (3D) texture and bulk density models
199 that were obtained by *Schaap* [2013] for prior soil hydraulic parameter estimates. The work by
200 *Schaap* [2013] relied on anisotropic Gaussian models with horizontal variogram ranges of 13.1, 5.6
201 and 7.7 m for PC1, PC2 and bulk density, respectively. Vertical range estimates were 0.28 and 0.85
202 m for PC1 and PC2, respectively. No reliable estimate of vertical range was found for bulk density
203 and we (as did *Schaap* [2013]) set bulk density below 5 m depth equal to 1.85 g/cm³. Point kriging
204 was used to obtain PC1, PC2 for the entire 60 × 60 × 15 m domain as well as bulk density for the
205 domain above 5 m depth. We found that a grid with a resolution of 5 × 5 × 0.25 m produced a

206 variability in PC1 and PC2 that was nearly identical to the observations; higher resolution grids did not
 207 yield meaningful improvements. The vertical resolution is further consistent with resolution of
 208 moisture content measurements. Additional details of the experimental setting and geospatial
 209 analysis are given by *Schaap* [2013]. Once kriging was completed, PC1 and PC2 were back-
 210 transformed into sand, silt and clay percentages. Point values of these kriged results formed prior
 211 estimates for purposes of inverse modeling.

212

213 **1.3. Soil Hydraulic Properties**

214 The Rosetta-H3 pedotransfer function model [*Schaap et al.*, 2001] was applied to the three-
 215 dimensional (3D) kriged fields of sand, silt, clay, and bulk density determined in Section 1.2 to obtain
 216 prior estimates of parameters entering into the Mualem-van Genuchten model (*van Genuchten* [1980];
 217 *Mualem* [1976], abbreviated here as VGM) for water retention (1) and unsaturated hydraulic
 218 conductivity (2):

$$219 \quad \theta(h) = \begin{cases} \theta_r + \frac{\theta_s - \theta_r}{[1 + |\alpha h|^n]^m} & h \leq 0 \\ \theta_s & h > 0 \end{cases} \quad (1)$$

$$220 \quad K(S_e) = \begin{cases} K_s S_e^L [1 - (1 - S_e^{1/m})^m]^2 & h \leq 0 \\ K_s & h > 0 \end{cases} \quad (2)$$

221 where

$$222 \quad S_e = \frac{\theta(h) - \theta_r}{\theta_s - \theta_r} \quad (3)$$

223 is effective saturation; θ is volumetric moisture content ($\text{cm}^3 \text{ cm}^{-3}$) at matric potential h (cm, < 0 for

224 unsaturated conditions); θ_r ($\text{cm}^3 \text{ cm}^{-3}$) and θ_s ($\text{cm}^3 \text{ cm}^{-3}$) are residual and saturated moisture contents,
225 respectively; α (> 0 , in cm^{-1}) and n (> 1) are curve-shape parameters; and $m = 1-1/n$. In (2), K_s is the
226 saturated hydraulic conductivity (cm d^{-1}) while L is an empirical parameter with a value of 0.5 (*Mualem*
227 [1976]). The application of Rosetta-H3 to the kriged field of texture and bulk density thus yields 3D
228 distributions of each of the five VGM parameters θ_r , θ_s , α , n and K_s . The purpose of our work is not
229 to compare one PTF with another but to introduce and illustrate two methods of parameterizing vadose
230 zone hydraulic properties based on a (in principle any) PTF and clustering, followed by inversion.
231 Therefore, only Rosetta-H3 PTF is reported in this study.

232

233 **2. Inverse Modeling Approach**

234 To implement our homogeneous and heterogeneous cluster approaches we subdivide the flow
235 domain into clusters using the method of k -means clustering. In the homogeneous cluster approach,
236 initial estimates of the VGM parameters are obtained by averaging over all grid points belonging to a
237 cluster. In the heterogeneous cluster approach, initial VGM estimates vary from one grid point to
238 another, each of which is associated with a cluster-specific scaling factor. We optimize hydraulic
239 parameters during the inversion of the homogeneous approach, while we optimize scale factors in the
240 case of the heterogeneous approach.

241

242 **2.1. Definition of Clusters and Inversion Method**

243 Various ways to decompose a domain into clusters are available, such as grouping spatially varying
244 kriged values of target quantities into stratigraphic [e.g., *Wang et al.*, 2003] or USDA textural classes.
245 Here we adopt *k-means* clustering according to which kriged values are grouped into k clusters by
246 minimizing

$$247 \quad \xi(k) = \sum_{i=1}^k \sum_{j=1}^{n_i} \| \mathbf{x}_{ij} - \boldsymbol{\mu}_i \|^2 \quad (4)$$

248 where n_i is the number of data points belonging to cluster i ; \mathbf{x}_{ij} is a vector of attributes (i.e., kriged sand,
249 silt, clay percentages in this study) of the j^{th} data point in cluster i ; $\boldsymbol{\mu}_i$ is a vector of attribute mean values
250 in cluster i ; and $\| \cdot \|$ denotes Euclidean norm (in data units). To perform *k-means* clustering we used
251 the algorithm of *Hartigan and Wong* [1979] in the statistical package R (version 3.0.2, *Ihaka and*
252 *Gentleman*, 1996; *R development core team*, 2005, <http://www.R-project.org>). A series of preliminary
253 analyses suggested that classification relying on *k-means* clustering of kriged soil texture yielded the
254 most satisfactory results (in terms of root mean square error between computed and observed water
255 contents at all times through the domain of interest). Simulations based on *k-means* clustering of initial
256 soil hydraulic parameters determined in Section 1.3 yielded results of distinctly inferior quality. A
257 reason for this might be that initial soil hydraulic parameter estimates are not very accurate, and
258 posterior estimates are not available prior to inversion. We therefore rely on *k-means* clustering of soil
259 texture. Clustering associates each kriged point with a unique cluster without requiring that points
260 defining a cluster be contiguous in space.

261 As the first principal component (PC1) represents closely the overall soil texture at the site, we

262 illustrate in Figure 2a its spatial variability along an east-west vertical section at $y = 30$ m (see Figure
 263 1), which passes through wells 422, 423 and 425. This is to be compared with $k = 2, 3, \dots, 6$ soil texture
 264 clusters in Figures 2b, 2c, ... 2f defined by the *k-means* method. As expected, increasing the number
 265 k of clusters renders their distribution closer and closer to that of PC1 in Figure 2a. Note that the
 266 clusters are generally not contiguous in space. Three-dimensional versions of these clusters form the
 267 basis for our definition of homogeneous and heterogeneous clusters below.

268 In the homogeneous cluster approach, each VGM parameter within a cluster is a constant. Table
 269 1 lists arithmetic mean sand, silt and clay percentages for clusters associated with various numbers k
 270 of clusters and arithmetic mean values of PTF-derived (using Rosetta-H3) hydraulic parameter
 271 estimates in each cluster. Parameters α , n and K_s in Table 1 are antilogs of average $\log_{10}(\alpha)$, $\log_{10}(n)$
 272 and $\log_{10}(K_s)$ Rosetta estimates. These represent prior hydraulic parameters for the homogeneous
 273 cluster approach. Posteriors are estimated by the simulation procedure defined in Section 2.2.

274 In the heterogeneous cluster approach parameters are expressed as

$$275 \quad \mathbf{p}'_i(\mathbf{x}) = \mathbf{B}_i \times \mathbf{p}_i(\mathbf{x}) \quad (5)$$

276 where $\mathbf{p}_i(\mathbf{x})$ is a vector whose entries are the five VGM parameters θ_r , θ_s , $\log_{10}(\alpha)$, $\log_{10}(n)$, and
 277 $\log_{10}(K_s)$ estimated from texture and bulk density data using Rosetta-H3 at location $\mathbf{x} = (x, y, z)$ in
 278 cluster i ($i = 1, \dots, k$). The square matrix \mathbf{B}_i in (5) is taken to be diagonal, containing scaling factors
 279 initialized to 1 and then optimized by inversion. Vector $\mathbf{p}'_i(\mathbf{x})$ thus represents posterior (inverse)
 280 estimates of the five VGM parameters. We designate forward runs with homogeneous clusters HoC
 281 and those with heterogeneous clusters HeC. Inverse runs are designated by prefix I and suffix k where

282 k denotes number of clusters.

283

284 **2.2. Vadose Zone Flow Simulation and Estimation Criterion**

285 **2.2.1. Numerical Simulation of Flow**

286 Water flow is simulated with the Subsurface Transport Over Multiple Phases (STOMP) code of
287 *White and Oostrom [2006]* that solves the Richards equation using finite differences with Newton-
288 Raphson iteration. Consistent with the geospatial grid, the simulation grid cells measure $5 \times 5 \text{ m}^2$
289 horizontally covering an area of $60 \times 60 \text{ m}^2$ (Figure 1), and 0.25 m vertically, extending down to depth
290 14.5 m. In both Experiment 3 and 4, vertical flux at the top boundary is prescribed to be zero except
291 during irrigation when it is set equal to the daily irrigation rate across the inner $50 \times 50 \text{ m}^2$ area (Figure
292 1). Pressure head at the bottom boundary, at a depth of 14.5 m, is set equal to positive 1.5 m to reflect
293 the presence of a perched water table at a depth of 13 m [*Wang et al., 2003*]. As the irrigated area
294 was surrounded by a tarp-covered collar which helped render flow to be predominantly vertical
295 [*Schaap, 2013*], no flow is allowed to take place across the four lateral sides of the grid during flow
296 simulations. Experiment 3 is used for the model calibrations and flow is simulated over a period of
297 95 days including a 47-day pre-irrigation period from DOY 67.5 to DOY 114.5, 28 days of irrigation
298 from DOY 114.5 to DOY 142.5, and 20 redistribution days from DOY 142.5 to DOY 162.5.
299 Observation data in Experiment 4 is used for model validations, which includes 230-day irrigation
300 period from DOY 450 to DOY 680, and 65 redistribution period from DOY 680 to DOY 744.

301 Initial moisture contents on DOY 67.5 at locations other than the neutron wells [e.g., *Schaap*, 2013]
302 are not available. Due to the prolonged drainage period prior to DOY 67.5, moisture contents at the
303 neutron wells were nearly constant with the zeroth moment of moisture content (see Appendix A)
304 varying only by a few millimeters over a total profile length of 12.5 m between DOY 17.5 and 114.5
305 (See Figure 8.5 in *Schaap* [2013]). Tensiometer pressure head readings at depths 3, 5 and 10 m,
306 ranging between negative 2 and negative 3.5 m, were also nearly constant over time. This implies that
307 total head is not constant throughout the profile (but rather varies with depth) and the system is thus
308 not under static equilibrium. Ambient flow prior to infiltration, and following redistribution, is
309 nevertheless negligibly small due to the low hydraulic conductivity of the profile at its ambient water
310 content and pressure head values. We therefore assumed that there should be a strong correlation
311 between texture and neutron thermalization count ratio (CR) [for details see *Schaap*, 2013]. A
312 stepwise regression between observed CR for DOY 67.5 and observed texture yielded the following
313 expression:

$$314 \quad CR = a \times sand + b \times silt + c \times sand^2 + d \times silt^2 + e \times sand \times silt + f \quad (6)$$

315 where *sand* and *silt* are expressed as percentages, respectively, while $a = 0.3210$; $b = 0.4038$; $c = -$
316 0.0017 ; $d = -0.0030$; $e = -0.0039$; $f = -13.8815$; the Pearson correlation coefficient (R) was 0.71.

317 Equation (6) was subsequently used to estimate CR for all non-well grid points in the flow domain.

318 Initial moisture contents for all grid points were subsequently estimated using a site-specific neutron
319 thermalization model [*Schaap*, 2013] :

320 $\theta = a \times CR + b \times PCI + c \times PCI^2 + d \times PCI \times CR + e$ (7)

321 where $a = 0.520$; $b = 0.0119$; $c = -7.545 \times 10^{-5}$; $d = -5.097 \times 10^{-3}$; $e = -0.544$.

322 Regardless of inversion method, preliminary simulations consistently resulted in substantial
323 drainage prior to the start of infiltration on DOY 114.5, contradicting observations of nearly constant
324 moisture contents. This may be the consequence of inaccurate initial moisture content derivation
325 from (6) and (7) or inaccurate initial hydraulic parameter estimation by PTF (the estimates being
326 inconsistent with water retention at high pressure or hydraulic conductivity). This problem
327 persisted even after inversion, mainly because of the sparse observations on dates with presumably
328 constant moisture content (2 dates) before the start of infiltration compared to the 26 observations
329 with dynamic moisture content during and after the infiltration period. Ad-hoc approaches, such as
330 assigning large weights to observations before DOY 114.5 did not alleviate the problem.

331 To eliminate the inconsistency we adopted a three-step approach to inversion, applied in each
332 case. The 3-step approach is not applicable to models HoC and HeC, which do not entail inversion;
333 initial water contents for these two models were based on regression results derived from (6) and (7).
334 In Step I, inversion was conducted by simulating the entire 95-day period of the experiment starting
335 with initial moisture contents determined in the above manner. The same initial moisture contents
336 coupled with parameter estimates obtained in Step I were then used to predict, through forward
337 simulation (Step II), moisture contents at the end of a 50-day continued drainage period (i.e. this
338 period did not have any infiltration). The final moisture contents of Step II, and parameters
339 obtained in Step I, were then assigned as initial values on DOY 67.5 in a final 95-day inversion run

340 during Step III.

341 Validation was carried out by running the models forward in time from DOY 67.5 till DOY 744
342 by using initial moisture contents and final parameters from Step III. Simulation results between
343 DOY 450 and 744 were subsequently compared with moisture content observations in Experiment 4.

344

345 **2.2.2. Model Quality Measures**

346 Model inversion is conducted with PEST [Doherty *et al.*, 1994; Doherty, 2003] using Python and
347 Unix-style shell scripts to facilitate data interchange with STOMP. Average soil hydraulic parameters
348 in IHoCk and scaling factors in IHeCk are estimated by minimizing the sum of squared differences
349 between observed and simulated moisture contents at all times through the domain of interest.
350 Estimates are constrained to ensure that $0 < \theta_r < 0.2$, $0.2 < \theta_s < 0.6$, $0.001 < \alpha < 0.1$ (1/cm), $1.1 < n <$
351 5.0 , $0 < K_s < 50,000$ (cm/day). As primary measures of model fit we use the root mean square error

352
$$RMSE = \sqrt{\frac{1}{N_z} \sum_{i=1}^{N_z} (\theta_i - \theta'_i)^2} \quad (8)$$

353 and coefficient of determination

354
$$R^2 = 1 - \frac{\sum_{i=1}^{N_z} (\theta_i - \theta'_i)^2}{\sum_{i=1}^{N_z} (\theta_i - \bar{\theta})^2} \quad (9)$$

355 where N_z is the number of (space-time varying) moisture content observations, $N_z = 11,020$ in model
356 calibration and 9,297 in model validation; θ_i and θ'_i are the i^{th} observed and simulated moisture

357 contents, respectively; and $\bar{\theta}$ is the average of θ_i . *RMSE* is dimensionless (cm³ water per cm³
358 sediment). Other measures of model fit we use include zeroth, first and second temporal moments of
359 observed and simulated moisture contents (Appendix A).

360

361 **3. Results and Discussion**

362 **3.1. Forward and Inverse Modeling Results**

363 Figure 3 shows how *RMSE* varies with number of clusters for various forward and inverse schemes.
364 As one might expect, *RMSE* is largest (0.0688) in the forward homogeneous single cluster case HoC1,
365 dropping down to below 0.045 as the number of clusters increases. No clustering is required for the
366 heterogeneous case HeC to yield a similarly small *RMSE* of 0.0429 without inversion.

367 Inversion is seen to reduce *RMSE* considerably in all cases. In the case of IHoC1 *RMSE*
368 decreases from 0.0688 to 0.0619, declining further to 0.0316 as the number of clusters is increased to
369 2 (IHoC2) down to 0.0260 when this number reaches 4 (IHoC4). Inversion renders the heterogeneous
370 scheme better than the homogeneous scheme: *RMSE* = 0.0309 in the single cluster case IHeC1 and
371 0.0224 in the four-cluster case IHeC4. It is noted that whereas varying the initial hydraulic parameters
372 of forward and inverse models may change the *RMSE* values, it would not affect our overall
373 conclusions in any significant way.

374 Figure 3 suggests that, in all cases, increasing the number of clusters beyond 4 fails to reduce
375 *RMSE* further. Like *Neuman* [1973], we attribute this to overparameterization and adopt four clusters

376 as optimum subdivision of our domain. More sophisticated performance metrics such as *AIC* [Akaike,
377 1974; Ye *et al.*, 2008], *AICc* [Hurvich and Tsai, 1989] and *BIC* [Schwarz, 1978] yielded similar results
378 (not reported).

379

380 **3.2. Interpretation of Results**

381 Table 2 lists estimated scale factors (ratios between posterior and prior values) associated with
382 IHoC4 and IHeC4 inversion and corresponding standard errors. A standard error is calculated in
383 PEST as the square root of parameter estimation variance; the latter constitute diagonal entries of the
384 parameter covariance matrix, computed to lead order of approximation. We note that these scale
385 factors correspond to \log_{10} transformations of α , n , and K_s , as described in Section 2.1. All standard
386 errors are low, suggesting that so is parameter estimation uncertainty. These models also yielded
387 similar patterns in the resulting optimized scale factors, i.e., if one method adjusts a scale factor upward
388 or downward from an initial ratio of 1.0, so does the other method (with only limited exceptions). The
389 ranges of estimated scale factors within each of the two methods are more substantial for some
390 optimized parameters than for others. The largest range is found for θ_t (from 0.64 to 3.00, across both
391 models and clusters), n (from 0.64 to 1.61), and K_s (from 1.06 to 1.66) and moderate for α (from 0.72
392 to 1.13) and θ_s (from 0.66 to 0.99). Actual VGM parameter values (not shown in Table 2) were
393 consistent with limited laboratory measurements on disturbed cores.

394 A visual comparison of simulated moisture contents and observed values corresponding to HeC,
395 IHoC4 and IHeC4 for model calibration and validation is provided in Figure 4. Inversion is seen to

396 improve the quality of this visual comparison markedly in both the homogeneous and heterogeneous
397 clustering cases. Whereas heterogeneous clusters yield better results than do homogeneous clusters,
398 the improvement does not appear to be dramatic for both model calibration and validation.
399 Quantitatively, in model calibration, inversion reduces the *RMSE* (in comparison to HeC of calibration)
400 by 40.8% in the homogeneous and 47.8% in the heterogeneous cases, bringing about an increase in the
401 coefficient of determination (R^2) from 0.66 for HeC through 0.88 for IHoC4 to 0.91 for IHeC4. *RMSE*
402 values associated with models IHoC4 and IHeC4 were lower during the validation period (as they had
403 been during the calibration period) by 43.7% and 49.7%, respectively, than that associated with model
404 HeC; correspondingly, R^2 increased from 0.66 in the case of HeC to 0.83 and 0.87 in the respective
405 cases of IHoC4 and IHeC4. The poor performance of HeC results is likely due to (a) uncertainty in
406 the model used to convert neutron thermalization CR and texture into moisture contents [*Schaap*, 2013],
407 (b) the approximation of initial moisture contents in forward simulations, as discussed in Section 2.2.1,
408 and (c) the assumption of the Gaussian nature of univariate and spatial distributions which is not
409 entirely consistent with findings by *Guadagnini et al.* [2013].

410 The validation results strengthen our conclusion that both clustering approaches improve
411 parameter estimates considerably in comparison to those based solely on PTF estimates from soil
412 texture. The improvement achieved with heterogeneous clusters is slightly better than that obtained
413 with homogeneous clusters.

414 We end by comparing in Figures 5a, 5b, and 5c the ways in which the first three temporal moments
415 of moisture content $M0(t)$, $M1(t)$ and $M2(t)$, defined in Appendix A, evolve when computed on

416 the basis of observations and simulations corresponding to HeC, IHoC4 and IHeC4. Only results for
417 Experiment 3 are shown because the number of neutron count measurements during Experiment 4 was
418 too small to allow computing spatial moments. The depicted moments are averages over seven wells
419 as explained in Appendix A. $M0(t)$ represents incremental moisture content between depth 0.25
420 and 12.5 m, multiplied by this depth (hence given in meters); $M1(t)$ corresponds to mean depth (in
421 meters) of the center of mass of infiltrated water (given in meters); and $M2(t)$ measures the vertical
422 spread of moisture content about its center of mass (in square meters). Because $M1(t)$ and $M2(t)$
423 are normalized by $M0(t)$, which is small prior to DOY 114.5, their values during this period are
424 unstable and therefore not plotted.

425 Zeroth moments computed on the basis of observations and simulations correspond closely to that
426 of cumulative infiltration in Figure 5a until DOY 130. Following this date, they first drop below the
427 latter and, following the end of the infiltration period on DOY 142.5, decline with time. This, as
428 explained in Appendix A, is due to the infiltration front's arrival at depth 12.5 m on DOY 130. It also
429 explains the stabilization of $M1(t)$ and gradual decrease in $M2(t)$ seen, respectively, in Figures 5b
430 and 5c. Whereas HeC simulations underestimate observation-based $M1(t)$ significantly at all times,
431 results based on IHoC4 and IHeC4 represent the latter closely and consistently. The poor
432 performance of HeC results can be attributed in part, as noted previously, to the poor definition of
433 initial moisture contents in this forward simulation. Whereas IHoC4 results overestimates
434 observation-based $M2(t)$ significantly at all times, IHeC4 results underestimate the latter at all but
435 intermediate time. It is difficult to tell on the basis of Figure 5 which of these two inverse approach

436 represent observation-based moments better.

437

438 **4. Conclusions**

439 Our work leads to four major conclusions:

440 1. Whereas it is possible to estimate deep vadose zone hydraulic parameters on the basis of soil
441 texture data with the aid of a pedotransfer function (PTF), as many soil and climate modelers tend to
442 do, we find it necessary to improve upon these estimates by conditioning them on observed system
443 variables such as moisture content through the adoption of a suitable inverse method. The same
444 conclusion was reached previously by *Wang et al.* [2003].

445 2. We proposed two ways of parameterizing vadose zone hydraulic properties on the basis of soil
446 texture data by utilizing PTF and *k*-means clustering. In contrast to traditional zonation often
447 employed in hydrologic inverse modeling, a cluster in our model may (and generally does) consist of
448 noncontiguous subdomains. In both of our two approaches hydraulic parameters at each grid point in
449 a cluster are estimated initially with the aid of a PTF. The heterogeneous cluster approach preserves
450 heterogeneity without introducing more adjustable parameters.

451 3. Upon applying our approach to experimental data from a deep vadose zone site near Maricopa,
452 Arizona, we found clustering combined with inversion improved estimates of moisture contents
453 considerably in comparison to those based solely on soil texture data. The optimum number of
454 clusters in both cases was found to be the same (four). In terms of root mean square errors, the

455 improvement achieved with heterogeneous clusters was slightly better than that obtained with
456 homogeneous clusters. Moment analysis revealed little differences between the two methods.

457 4. The calibrated model was validated against an independent infiltration experiment, producing
458 results of essentially the same quality as those obtained during calibration.

459

460 **Acknowledgments**

461 Site construction and field observations were supported by the US Nuclear Regulatory Commission
462 under contract number NRC-04-97-056 (1997-2002). The data collection for and the development of
463 the geospatial model was supported in part by NSF-EAR grant 0737945 (2005-2009). The current work
464 was supported in part through a contract between the University of Arizona and Vanderbilt University
465 under the Consortium for Risk Evaluation with Stakeholder Participation (CRESP) III, funded by the
466 U.S. Department of Energy. The authors acknowledge previous related work by Mr. Zhufeng Fang.

467

468 **References**

469 Akaike, H. (1974), A new look at the statistical model identification, *IEEE Trans. Autom. Control*,
470 19(6), 716-723.

471 Assouline, S., and D. Or (2013), Conceptual and Parametric Representation of Soil Hydraulic
472 Properties: A Review, *Vadose Zo. J.*, 12, 1-20.

473 Brooks, R., and A. Corey (1964), Hydraulic properties of porous media, *Hydrol. Pap. Color. State*

474 *Univ.*, 3.

475 Burdine, N. T. (1953), Relative Permeability Calculations From Pore Size Distribution Data, *J. Pet.*
476 *Technol.*, 5, 71-78.

477 Carrera, J., and S. P. Neuman (1986a), Estimation of Aquifer Parameters Under Transient and Steady
478 State Conditions: 1. Maximum Likelihood Method Incorporating Prior Information, *Water*
479 *Resour. Res.*, 22(2), 199-210.

480 Carrera, J., and S. P. Neuman (1986b), Estimation of Aquifer Parameters Under Transient and Steady
481 State Conditions: 2. Uniqueness, Stability, and Solution Algorithms, *Water Resour. Res.*, 22(2),
482 211-227.

483 Carrera, J., and S. P. Neuman (1986c), Estimation of Aquifer Parameters Under Transient and Steady
484 State Conditions: 3. Application to Synthetic and Field Data, *Water Resour. Res.*, 22(2), 228-242.

485 Chen, Y., and D. Zhang (2006), Data assimilation for transient flow in geologic formations via
486 ensemble Kalman filter, *Adv. Water Resour.*, 29(8), 1107-1122.

487 Chirico, G. B., H. Medina, and N. Romano (2007), Uncertainty in predicting soil hydraulic properties
488 at the hillslope scale with indirect methods, *J. Hydrol.*, 334(3), 405-22.

489 Das, B. S., N. W. Haws, and P. S. C. Rao (2005), Defining geometric similarity in soils, *Vadose Zo.*
490 *J.*, 4(2), 264-270.

491 de Marsily, G., G. Lavedan, M. Boucher, and G. Fasanino (1984), Interpretation of interference tests

492 in a well field using geostatistical techniques to fit the permeability distribution in a reservoir
493 model, *Geostatistics Nat. Resour. Charact. Part, 2*, 831-849.

494 Doherty, J. (2003), Ground water model calibration using pilot points and regularization, *Ground*
495 *Water*, 41(2), 170-177.

496 Doherty, J., L. Brebber, and P. Whyte (1994), PEST: Model-independent parameter estimation,
497 *Watermark Comput. Corinda, Aust.*, 122.

498 Emsellem, Y., and G. de Marsily (1971), An automatic solution for the inverse problem, *Water*
499 *Resour. Res.*, 7(5), 1264-1283.

500 Franssen, H. J. H., A. Alcolea, M. Riva, M. Bakr, N. Van der Wiel, F. Stauffer, and A. Guadagnini
501 (2009), A comparison of seven methods for the inverse modelling of groundwater flow.
502 Application to the characterisation of well catchments, *Adv. Water Resour.*, 32(6), 851-872.

503 Grondona, M. O., and N. Cressie (1991), Using Spatial Considerations in the Analysis of
504 Experiments, *Technometrics*, 33(4), 381-392.

505 Guadagnini, A., S. P. Neuman, M. G. Schaap, and M. Riva (2013), Anisotropic statistical scaling of
506 vadose zone hydraulic property estimates near Maricopa, Arizona, *Water Resour. Res.*, 49(12),
507 8463-8479.

508 Guadagnini, A., S. P. Neuman, M. G. Schaap, and M. Riva (2014), Anisotropic statistical scaling of
509 soil and sediment texture in a stratified deep vadose zone near Maricopa, Arizona, *Geoderma*,

510 214-215, 217-227.

511 Hartigan, J. A., and M. A. Wong (1979), Algorithm AS 136: A K-Means Clustering Algorithm, *J. R.*
512 *Stat. Soc. C*, 28(1), 100-108.

513 Hopmans, J. W., and J. Šimůnek (1999), Review of inverse estimation of soil hydraulic properties,
514 Proc. Int. Workshop Characterization and Measurement of the Hydraulic Properties of
515 Unsaturated Porous Media, University of California, Riverside, CA.

516 Hopmans, J. W., J. Simunek, N. Romano, and W. Durner (2002), Inverse modeling of transient water
517 flow, *Methods Soil Anal. Part 1, Phys. Methods*, 2.

518 Hurvich, C. M., and C. L. Tsai (1989), Regression and time series model selection in small samples,
519 *Biometrika*, 76, 297-307.

520 Ihaka, R., and R. Gentleman (1996), R: A Language for Data Analysis and Graphics, *J. Comput.*
521 *Graph. Stat.*, 5(3), 299-314.

522 Kowalsky, M. B., S. Finsterle, J. Peterson, S. Hubbard, Y. Rubin, E. Majer, A. Ward, and G. Gee
523 (2005), Estimation of field-scale soil hydraulic and dielectric parameters through joint inversion
524 of GPR and hydrological data, *Water Resour. Res.*, 41(11), 1-19.

525 LaVenue, A. M., and J. F. Pickens (1992), Application of a coupled adjoint sensitivity and kriging
526 approach to calibrate a groundwater flow model, *Water Resour. Res.*, 28(6), 1543-1569.

527 Liu, S., and T.-C. J. Yeh (2004), An Integrative Approach for Monitoring Water Movement in the

528 Vadose Zone, *Vadose Zo. J.*, 3, 681-692.

529 Miller, E. E., and R. D. Miller (1956), Physical theory for capillary flow phenomena, *J. Appl. Phys.*,
530 27(4), 324-332.

531 Morales-Casique, E., S. P. Neuman, and V. V Vesselinov (2010), Maximum likelihood Bayesian
532 averaging of airflow models in unsaturated fractured tuff using Occam and variance windows,
533 *Stoch. Environ. Res. risk Assess.*, 24(6), 863-880.

534 Mualem, Y. (1976), A new model for predicting the hydraulic conductivity of unsaturated porous
535 media, *Water Resour. Res.*, 12(3), 513-522.

536 Nasta, P., N. Romano, S. Assouline, J. A. Vrugt, and J. W. Hopmans (2013), Prediction of spatially
537 variable unsaturated hydraulic conductivity using scaled particle-size distribution functions,
538 *Water Resour. Res.*, 49(7), 4219-4229.

539 Neuman, S. P. (1973), Calibration of distributed parameter groundwater flow models viewed as a
540 multiple - objective decision process under uncertainty, *Water Resour. Res.*, 9(4), 1006-1021.

541 Neuman, S. P. (2003), Maximum likelihood Bayesian averaging of uncertain model predictions,
542 *Stoch. Environ. Res. Risk Assess.*, 17(5), 291-305.

543 Neuman, S. P., and E. A. Jacobson (1984), Analysis of nonintrinsic spatial variability by residual
544 kriging with application to regional groundwater levels, *J. Int. Assoc. Math. Geol.*, 16(5), 499-
545 521.

546 Neuman, S. P., and S. Yakowitz (1979), A statistical approach to the inverse problem of aquifer
547 hydrology, 1. theory, *Water Resour. Res.*, 15(4), 845-860.

548 Nolet, G. (1987), Seismic wave propagation and seismic tomography, in *Seismic tomography*, pp. 1-
549 23, Springer.

550 Pachepsky, Y. A., W. J. Rawls, and H. S. Lin (2006), Hydropedology and pedotransfer functions,
551 *Geoderma*, 131(3-4), 308-316.

552 Panzeri, M., M. Riva, A. Guadagnini, and S.P. Neuman (2016), Theory and generation of conditional,
553 scalable sub-Gaussian random fields, *Water Resour. Res.*, 52.

554 RamaRao, B. S., A. M. LaVenue, G. De Marsily, and M. G. Marietta (1995), Pilot point methodology
555 for automated calibration of an ensemble of conditionally simulated transmissivity fields: 1.
556 Theory and computational experiments, *Water Resour. Res.*, 31(3), 475-493.

557 Rawls, W. J., D. L. Brakensiek, and K. E. Saxton (1982), Estimation of Soil Water Properties, *Trans.*
558 *ASAE*, 25(5), 1316-1320 & 1328.

559 Riva, M., S. P. Neuman, and A. Guadagnini (2015), New scaling model for variables and increments
560 with heavy-tailed distributions, *Water Resour. Res.*, 51(6), 4623-4634.

561 Romano, N. (2004), Spatial structure of PTF estimates, in *Development of pedotransfer functions in*
562 *soil hydrology*, vol. 30, edited by Y. A. Pachepsky and W. J. Rawls, pp. 295-319, Elsevier, New
563 York.

564 Schaap, M. G. (2013), Description, Analysis, and Interpretation of an Infiltration Experiment in a
565 Semiarid Deep Vadose Zone, in *Advances in Hydrogeology*, pp. 159-183, Springer.

566 Schaap, M. G., and F. J. Leij (1998), Using neural networks to predict soil water retention and soil
567 hydraulic conductivity, *Soil Tillage Res.*, 47(1-2), 37-42.

568 Schaap, M. G., F. J. Leij, and M. T. Van Genuchten (2001), Rosetta: A computer program for
569 estimating soil hydraulic parameters with hierarchical pedotransfer functions, *J. Hydrol.*, 251(3-
570 4), 163-176.

571 Schwarz, G. (1978), Estimating the dimension of a model, *Ann. Stat.*, 6(2), 461-464.

572 Tromp, J., C. Tape, and Q. Liu (2005), Seismic tomography, adjoint methods, time reversal and
573 banana-doughnut kernels, *Geophys. J. Int.*, 160(1), 195-216.

574 Tuli, A., K. Kosugi, and J. W. Hopmans (2001), Simultaneous scaling of soil water retention and
575 unsaturated hydraulic conductivity functions assuming lognormal pore-size distribution, *Adv.*
576 *Water Resour.*, 24(6), 677-688.

577 van Genuchten, M. T. (1980), A Closed-form Equation for Predicting the Hydraulic Conductivity of
578 Unsaturated Soils, *Soil Sci. Soc. Am. J.*, 44(5), 892.

579 Vogel, T., M. Cislerova, and J. W. Hopmans (1991), Porous Media With Linearly Variable Hydraulic
580 Properties, *Water Resour. Res.*, 27(10), 2735-2741.

581 Vrugt, J. A., W. Bouten, H. V. Gupta, and S. Sorooshian (2002), Toward improved identifiability of

582 hydrologic model parameters: The information content of experimental data, *Water Resour. Res.*,
583 38(12), 48-1-48-13.

584 Vrugt, J. A., P. H. Stauffer, T. Wöhling, B. A. Robinson, and V. V. Vesselinov (2008), Inverse
585 Modeling of Subsurface Flow and Transport Properties: A Review with New Developments,
586 *Vadose Zo. J.*, 7(2), 843.

587 Wang, W. (2002), Uncertainty analysis of groundwater flow and solute transport in unsaturated-
588 saturated porous medium: Maricopa case, Ph.D diss, The University of Arizona.

589 Wang, W., S. P. Neuman, T. Yao, and P. J. Wierenga (2003), Simulation of Large-Scale Field
590 Infiltration Experiments Using a Hierarchy of Models Based on Public, Generic, and Site Data,
591 *Vadose Zo. J.*, 2, 297-312.

592 White, M. D., and M. Oostrom (2006), STOMP Subsurface Transport Over Multiple Phases Version
593 4.0 User's Guide, Richland, Washington.

594 Wildenschild, D., and K. H. Jensen (1999), Numerical modeling of observed effective flow behavior
595 in unsaturated heterogeneous sands, *Water Resour. Res.*, 35(1), 29-42.

596 Ye, M., R. Khaleel, M. G. Schaap, and J. Zhu (2007), Simulation of field injection experiments in
597 heterogeneous unsaturated media using cokriging and artificial neural network, *Water Resour.*
598 *Res.*, 43(7).

599 Ye, M., P. D. Meyer, and S. P. Neuman (2008), On model selection criteria in multimodel analysis,

600 *Water Resour. Res.*, 44(3).

601 Yeh, T.-C. J. (2002), A geostatistically based inverse model for electrical resistivity surveys and its
602 applications to vadose zone hydrology, *Water Resour. Res.*, 38(12), 1-13.

603 Young, M. H. et al. (1999), Results of Field Studies at the Maricopa Environmental Monitoring Site,
604 Arizona, University of Arizona.

605 Zhang, Z. F., A. L. Ward, and G. W. Gee (2004), A combined parameter scaling and inverse technique
606 to upscale the unsaturated hydraulic parameters for heterogeneous soils, *Water Resour. Res.*,
607 40(8).

608 Zimmerman, D. A. et al. (1998), A comparison of seven geostatistically based inverse approaches to
609 estimate transmissivities for modeling advective transport by groundwater flow, *Water Resour.*
610 *Res.*, 34(6), 1373.

611

612 **Appendix A: Temporal moment analysis**

613 For a given borehole the zeroth order temporal moment of moisture contents is defined as

$$614 \quad M0(t) = \sum_{z=0.25}^{12.5} \theta_{diff}(t, z) \times \Delta z \quad (A1)$$

615 where θ_{diff} is the difference between observed or simulated moisture contents at time t and their initial
616 values, z being depth and $\Delta z = 0.25$ m a depth increment. $M0(t)$ represents the incremental moisture
617 content between depth 0.25 and 12.5 m, multiplied by this depth (hence given in meters). In this study
618 we calculate $M0(t)$ at each of seven neutron wells and average the results (some values measured in
619 well 405, and all values measured in well 442, are considered to be unreliable; we therefore exclude
620 these two wells from our analysis of moments). The temporal evolution of this average $M0$ should
621 follow closely the actual cumulative amount of irrigation water in the absence of horizontal and vertical
622 drainage losses.

623 The first temporal moment is calculated as

$$624 \quad M1(t) = \frac{1}{M0} \sum_{z=0.25}^{12.5} \theta_{diff}(t, z) \times z \times \Delta z \quad (A2)$$

625 and given in square meters. $M1(t)$ represents the mean depth (in meters) of the center of mass of
626 infiltrated water in a borehole; as in the case of $M0(t)$, we average it over seven wells. When water
627 drainage below a depth of 12.5 m is not negligible (in which case $M0(t)$ does not coincide with the total
628 volume injected), $M1(t)$ provides information only about the center of mass of infiltrating water
629 above this depth.

630 The second temporal moment,

631
$$M2(t) = \frac{1}{M0} \sum_{z=0.25}^{12.5} \theta_{diff}(t, z) \times z^2 \times \Delta z - M1^2, \quad (A3)$$

632 measures the vertical spread of moisture content about its center of mass.

633

634 **Tables**

635

636 **Table 1.** Soil clusters based on textural data, corresponding mean sand, silt and clay percentages and
 637 PTF derived mean hydraulic parameters.

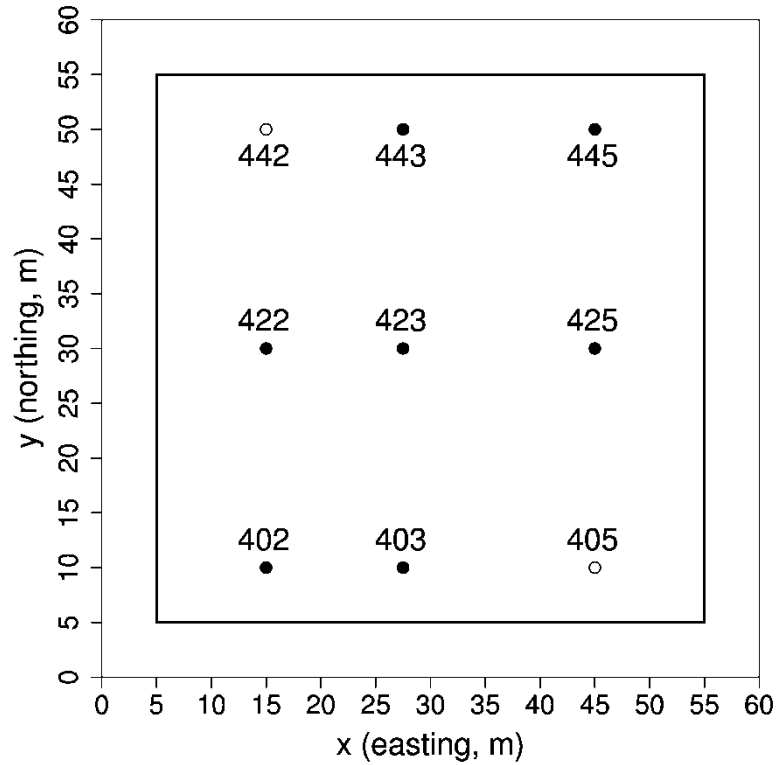
Number <i>k</i> of clusters	Cluster	Sand %	Silt %	Clay %	θ_r , cm ³ cm ⁻³	θ_s , cm ³ cm ⁻³	α , 1/cm	<i>n</i>	<i>K_s</i> , cm/day
1	a	81.034	12.633	6.333	0.041	0.316	0.040	1.867	64.845
2	a	71.279	19.135	9.587	0.040	0.328	0.043	1.429	24.912
	b	88.488	7.665	3.846	0.042	0.308	0.038	2.292	134.694
3	a	68.026	21.280	10.694	0.041	0.333	0.042	1.376	20.858
	b	89.901	6.850	3.249	0.042	0.306	0.037	2.440	165.878
	c	79.063	13.751	7.186	0.041	0.318	0.044	1.609	40.933
4	a	92.769	5.149	2.083	0.040	0.302	0.036	2.814	261.011
	b	66.602	22.072	11.327	0.041	0.336	0.040	1.363	19.750
	c	85.501	9.401	5.098	0.043	0.312	0.040	1.981	84.388
	d	75.766	16.291	7.944	0.040	0.321	0.046	1.498	31.308
5	a	77.282	14.861	7.857	0.040	0.321	0.045	1.539	35.178
	b	92.829	5.125	2.046	0.040	0.302	0.036	2.823	263.575
	c	65.888	19.041	15.071	0.049	0.371	0.029	1.403	24.892
	d	85.859	9.211	4.931	0.043	0.311	0.040	2.008	88.397
	e	69.443	23.309	7.248	0.033	0.302	0.055	1.355	18.064
6	a	92.927	5.116	1.957	0.040	0.301	0.036	2.840	267.898
	b	62.320	23.681	14.000	0.045	0.353	0.032	1.357	18.383
	c	85.233	6.696	8.071	0.047	0.328	0.036	1.881	81.146
	d	70.488	20.305	9.207	0.038	0.324	0.047	1.387	22.141
	e	86.240	10.699	3.061	0.041	0.302	0.043	2.094	94.359
	f	77.581	14.791	7.628	0.040	0.320	0.045	1.546	35.484

638

639 **Table 2.** *Optimized values of scale factors and standard errors for IHoC4 and IHeC4 models. Scale*
640 *factors for IHoC4 were obtained upon dividing optimal VGM parameters by their initial estimates in*
641 *Table 1 (4 clusters).*

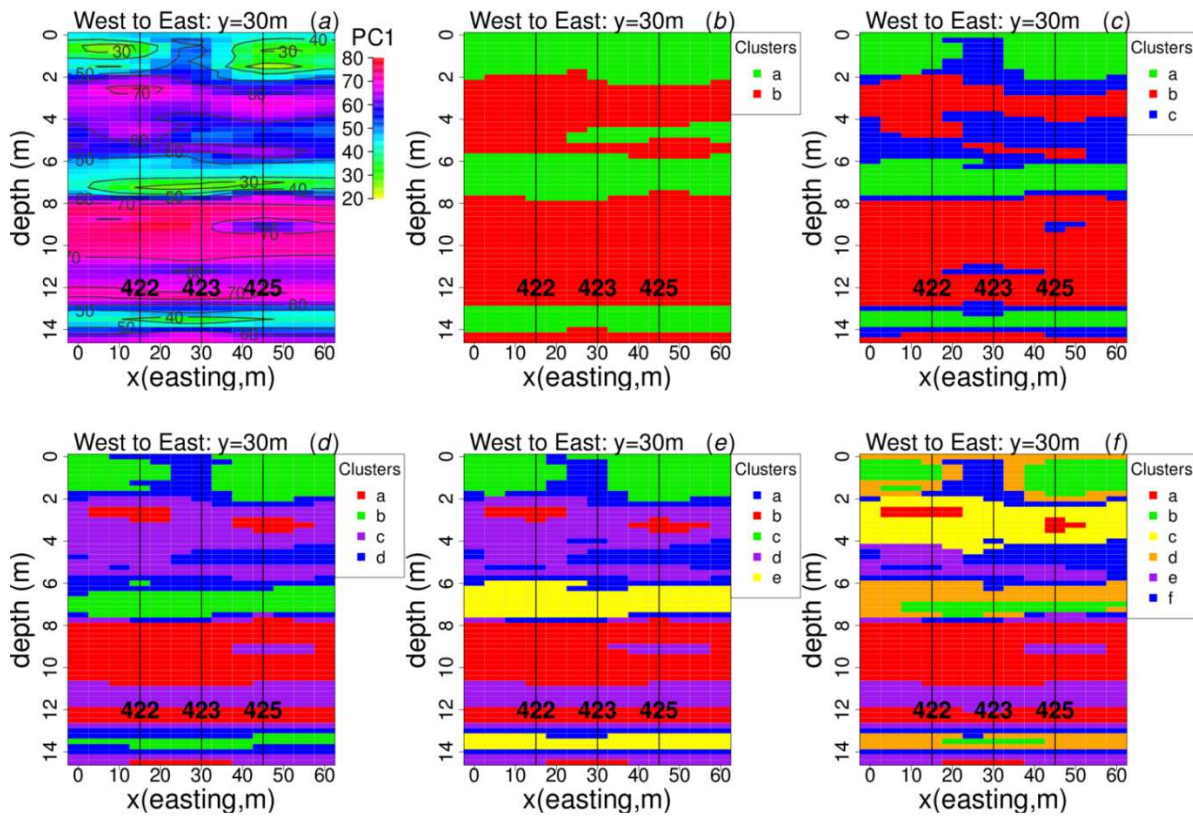
Cluster	Parameters	IHoC4		IHeC4	
		Scale factor	Standard error	Scale factor	Standard error
a	θ_r	2.565	0.0023	2.8816	0.0064
	θ_s	0.9868	0.0025	0.6649	0.0052
	α	1.043	0.0098	0.9766	0.0023
	n	1.5454	0.003	1.196	0.0025
	K_s	1.3122	0.0226	1.0555	0.002
b	θ_r	0.8878	0.0004	0.6438	0.0053
	θ_s	0.8708	0.0005	0.825	0.0028
	α	0.7153	0.0004	0.9405	0.0017
	n	0.6492	0.0006	0.7354	0.0018
	K_s	1.2449	0.0069	1.5657	0.0018
c	θ_r	2.6837	0.0004	3.0034	0.0048
	θ_s	0.9385	0.001	0.7127	0.0051
	α	1.0039	0.005	0.9133	0.0027
	n	1.4952	0.0033	1.0051	0.0025
	K_s	1.3257	0.0018	1.6599	0.0017
d	θ_r	2.6325	0.0006	1.0241	0.0093
	θ_s	0.9153	0.0007	0.9259	0.0025
	α	1.1303	0.0051	0.8413	0.0021
	n	1.6109	0.0024	0.6401	0.0026
	K_s	1.242	0.0067	1.4387	0.0033

642



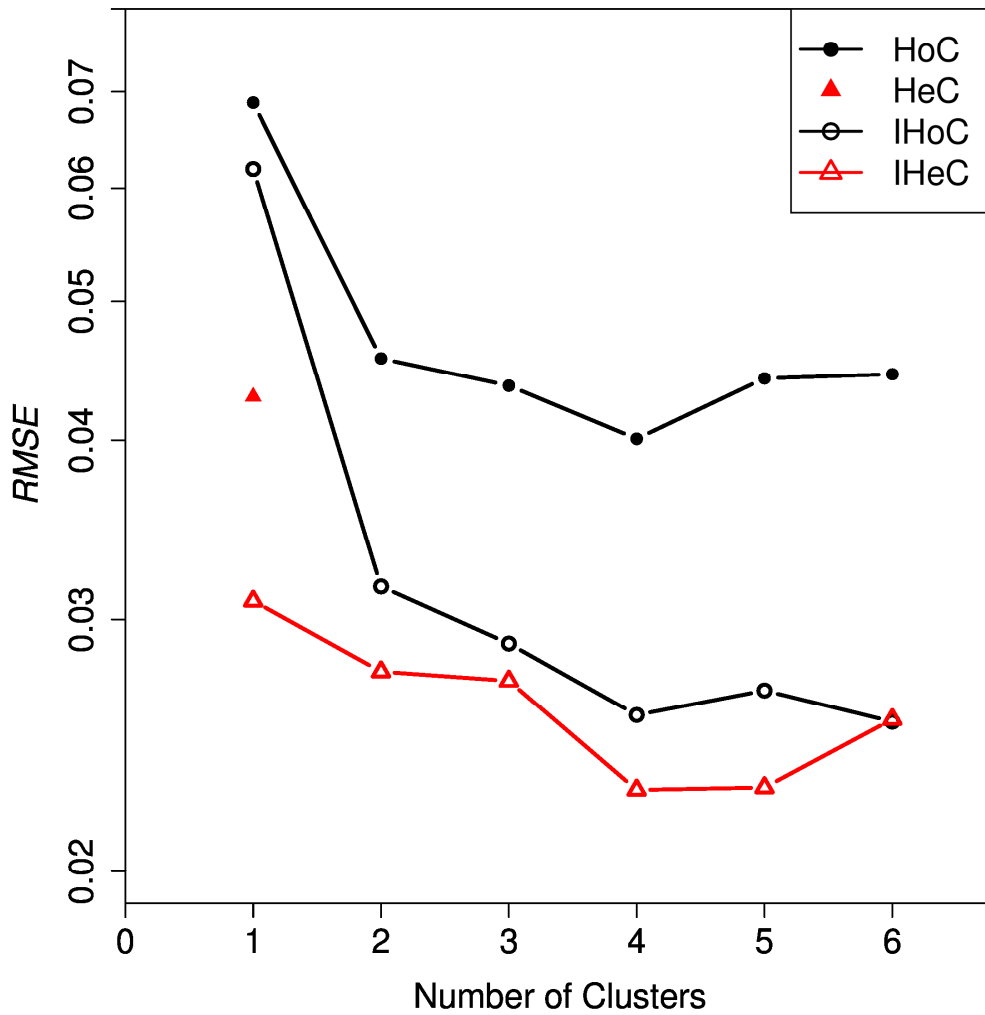
644

645 **Figure 1.** Location of nine monitoring boreholes at Maricopa site. All moisture content data from
 646 wells designated by solid circles were employed during inversion; all or some such data in wells
 647 designated by open circles were considered unreliable and omitted (see text). The 60×60 meter
 648 outer solid square was covered by tarp to prevent evaporation; the inner 50×50 meter square was
 649 drip irrigated.



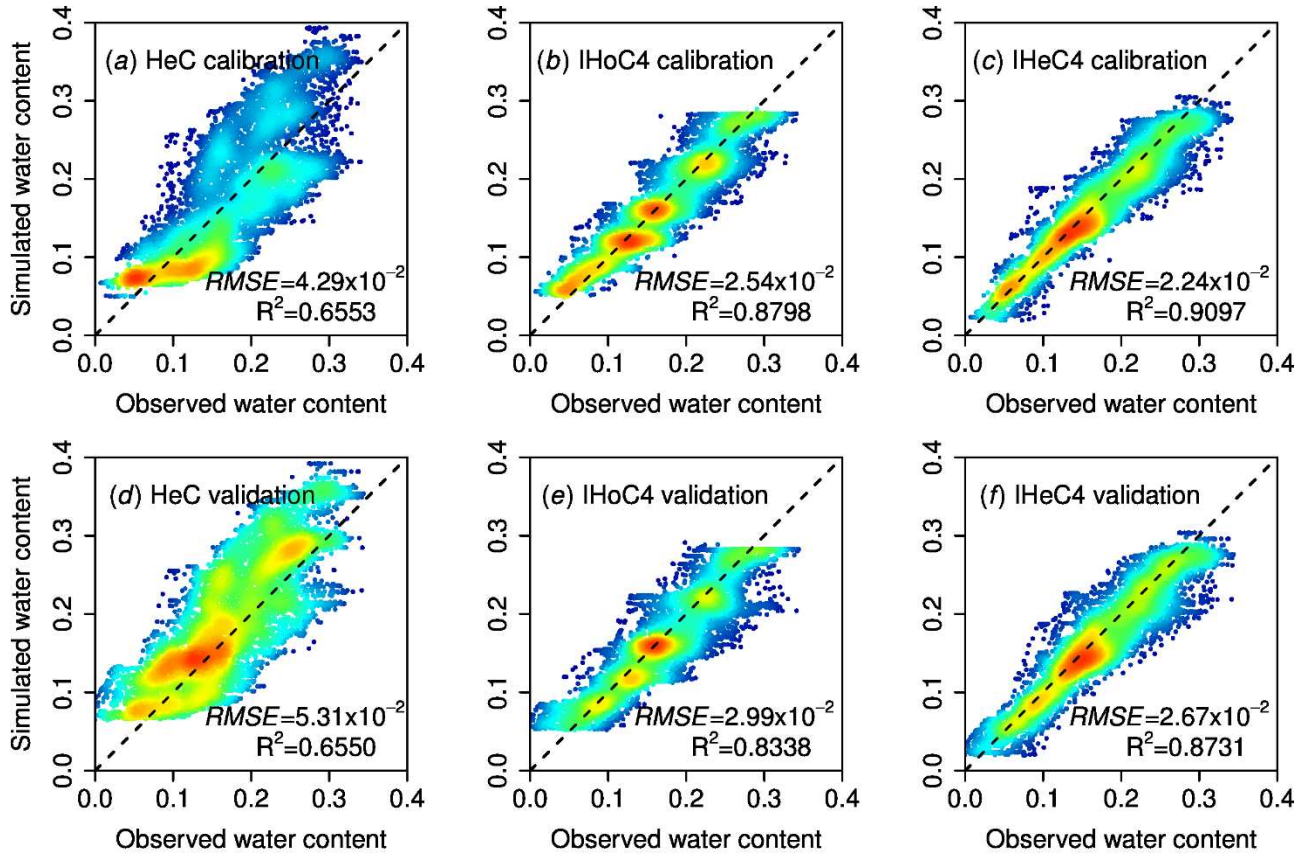
650

651 **Figure 2.** (a) cross sectional depth profile (at $y = 30\text{m}$ in Figure 1) of first principal component
 652 (PC1) extracted from soil texture data with labeled contours of PC1 values. PC1 measures the
 653 coarseness of soil similar to that of sand ; (b-f) clusters of kriged soil texture data with $k = 2-6$.
 654 Numbers at bottom designate well numbers in Figure 1.



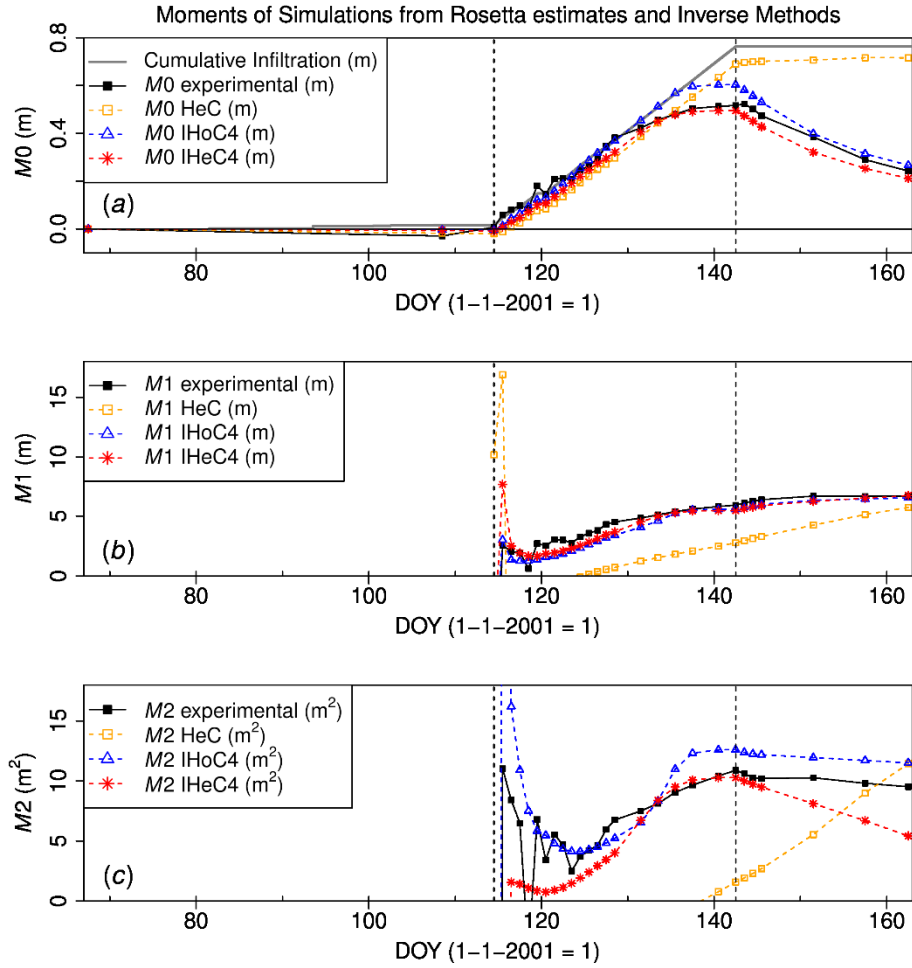
655

656 **Figure 3.** RMSE versus number of clusters and model type (vertical axis is in logarithmic scale).
 657 HoC is initial simulation of IHoC inversion; HeC indicates simulation of full heterogeneous domain
 658 with hydraulic parameter estimates from Rosetta-H3 at all grid points; IHoC represents
 659 homogeneous cluster inversion; IHeC indicates heterogeneous cluster inversion.



660

661 **Figure 4.** Comparison between observed and simulated moisture contents using calibration data for
 662 (a) HeC, (b) IHoC4 and (c) IHeC4 and validation data for (d) HeC, (e) IHoC4 and (f) IHeC4. Red
 663 represents high data density, blue low density.



664 **Figure 5.** Comparison of (a) zeroth (M_0), (b) first (M_1) and (c) second moment (M_2) of measured
 665 and simulated moisture front in wells based on experimental data, HeC, IHoC4 and IHeC4. Vertical
 666 dashed lines indicate start (DOY 114.5) and end (DOY 142.5) dates of infiltration.

667 **Table Captions**

668 **Table 1.** Soil clusters based on textural data, corresponding mean sand, silt and clay percentages and
669 PTF derived mean hydraulic parameters.

670 **Table 2.** Optimized values of scale factors and standard errors for IHoC4 and IHeC4 models. Scale
671 factors for IHoC4 were obtained upon dividing optimal VGM parameters by their initial estimates in
672 Table 1 (4 clusters).

673 **Figure Captions**

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