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Effect of pH and pressure on uranium removal from drinking water using NF/RO membranes

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1	Effect of pH and pressure on uranium removal from drinking water using
2	NF/RO membranes.
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17 ABSTRACT ART



20 Abstract. Groundwater is becoming an increasingly important drinking water source. 21 However, the use of groundwater for potable purposes can lead to chronic human 22 exposure to geogenic contaminants, e.g. uranium. Nanofiltration (NF) and reverse 23 osmosis (RO) processes are used for drinking water purification, and it is important to 24 understand how contaminants interact with membranes since accumulation of 25 contaminants to the membrane surface can lead to fouling, performance decline and possible breakthrough of contaminants. During the current study laboratory 26 27 experiments were conducted using NF (TFC-SR2) and RO (BW30) membranes to 28 establish the behaviour of uranium across pH (3-10) and pressure (5-15 bar) ranges. 29 The results showed that important determinants of uranium-membrane sorption 30 interactions were (i) the uranium speciation (uranium species valence and size in 31 relation to membrane surface charge and pore size) and (ii) concentration polarisation, 32 depending on the pH values. The results show that it is important to monitor sorption 33 of uranium to membranes, which is controlled by pH and concentration polarisation, 34 and, if necessary, adjust those parameters controlling uranium sorption.

35

36 Key words: sorption, concentration polarisation, nanofiltration, reverse osmosis,
37 speciation, uranium

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Introduction. Groundwater is an increasingly important source of drinking water,¹ 41 especially for developing nations in *e.g.* Africa² but also in European countries, where 42 20-100% of drinking water is sourced from groundwater.³ As a consequence of 43 hydrological and geochemical processes uranium is naturally present in groundwater 44 at concentrations from below detection to hundreds of µg L⁻¹. ^{4–6} Indirect release of 45 46 uranium into water may also occur through e.g. phosphate ore processing, phosphate fertilizer use, gold and tin-mining.⁷⁻¹⁰ Uranium is above all chemically toxic, and 47 although chronic exposure to uranium is not well understood,¹¹⁻¹⁴ studies have 48 49 identified kidney, bone, liver, heart and brain as potential targets following exposure through ingestion of uranium-containing water.^{11,15–18} The maximum acceptable 50 51 concentrations (MAC) for uranium vary between public authorities, e.g. Canada uses an interim MAC of 20 µg L⁻¹, whilst the US has adopted the same limit as the WHO 52 provisional drinking water limit of 30 µg L^{-1.19} However, concentrations as low as 2 53 µg/L may be hazardous for infants.²⁰ Membrane technology such as nanofiltration 54 (NF) and reverse osmosis (RO) are frequently used for drinking water purification²¹⁻²⁴ 55 since they efficiently remove a wide range of contaminants, including uranium.²⁵ 56 57 Membrane technology is also being considered for application in remote sites due to 58 the modular and flexible configuration and, if coupled with renewable energy, their independence from intermittent or absent energy supplies.^{26–28} The performance of a 59 membrane system using ultrafiltration, NF or RO membranes, coupled with a 60 61 renewable energy supply was tested in the field. The results showed that although the 62 system performed well in terms of producing permeate with low concentrations of most analytes e.g. Ca²⁺, Mg²⁺, Mn²⁺, SO₄²⁻, Na⁺ and Cl⁻, the behaviour of uranium in 63 64 the system differed from that of most other inorganic species found in the water.

Moreover, uranium strongly interacted with the membranes.²⁹ It was deduced that this 65 was largely due to the complex chemical behaviour of uranium, since its speciation 66 67 varies widely depending on physico-chemical parameters such as available inorganic or organic ligands and pH of the water.²⁹⁻³² For instance, the presence of calcium 68 69 affects uranium speciation and is likely to affect uranium interaction with the membranes.²⁹ A variety of processes can be responsible for the observed uranium 70 71 uptake, including precipitation of uranium to the membrane, physi-sorption (such as 72 hydrogen bonding or electrostatic attraction) or chemi-sorption resulting in bond 73 formation between uranium and membrane functional groups. Before the exact 74 mechanisms are determined, the more general terms "sorption" or "uranium-75 membrane interaction" will be used in this paper. To establish the processes involved 76 in the interaction of uranium with the membrane, controlled laboratory experiments 77 are needed, isolating one factor at a time: individual membranes, solution 78 composition, pH and pressure ranges. The present study focused on uranium in the 79 absence of other major ions, e.g. calcium, in order to determine the membrane 80 interactions due to uranium alone. Although there have been studies investigating uranium retention using NF and RO³³⁻³⁶ none have investigated the specific 81 82 interactions between membranes and uranium during the water purification process. 83 Therefore, the aims of this study were to determine (i) the effect of pH and speciation 84 on uranium-membrane interactions by NF and RO membranes; (ii) the effect of 85 membrane pore size and (iii) the effect of pressure on the uranium-membrane 86 interactions.

88 Materials and Methods. The behaviour of uranium was investigated using an experimental feed solution containing 0.5 mg L⁻¹ uranium (uranyl nitrate, TAAB, 89 90 UK), background electrolyte and buffer (20 mM NaCl and 1 mM NaHCO₃, Fisher 91 Scientific, UK). The uranium concentration chosen is comparable to concentrations found naturally.^{4,29,37,38} MilliQ water was used to prepare all solutions (Elga Purelab 92 93 Ultra, High Wycombe, UK). The pH of the experimental solution was adjusted (1 M 94 HNO₃, VWR Aristar, or 1 M NaOH, Fisher Scientific) according to the conditions to 95 be tested (described later), prior to adding the solution to the stainless steel cross-flow 96 system (MMS, Switzerland). Two membranes with different molecular-weight cut-off 97 (MWCO) values but of similar materials (polyamide active layer on a polysulphone 98 support layer) were selected: TRC-SR2, a NF membrane (Koch Membrane systems, 99 USA) and BW30, a RO membrane (Dow Filmtec). Cross-flow experiments were 100 carried out with and without applied pressure to investigate the effects of pressure on 101 uranium-membrane interactions. Two types of cross-flow experiments were 102 conducted: 1) across the pH range 3-10 (in units of one, with one pH value per 103 experiment) but with no applied pressure and 2) at two selected pH values (pH 6 and 104 8.5), across the pressure range 5-15 bar (in steps of 2.5 bar, one pressure and pH value 105 per experiment).

106

Filtration set-up and procedure. New rectangular membrane coupons (membrane area of 0.0046 m²) were cut for each experiment from a membrane sheet. The coupons were washed and soaked in MilliQ water at 4°C overnight, washed again, before being mounted in the stainless steel cross-flow system, operated in total recirculation mode. Membranes were compacted using MilliQ water at 25 bar prior to

112 experiments and a solution sample was collected as a quality check. A detailed 113 description of the experimental set-up is given in the supporting information (SI) and a schematic is published in Semião and Schäfer.³⁹ The system was drained and 114 115 experimental solution added (feed volume of 2 L). The flow-rate was set to 0.6 L min⁻ 116 ¹ and, for *pressure experiments*, pressure was applied at this point. Temperature, 117 pressure, feed flow, feed and permeate pH and conductivity were monitored 118 throughout the experiments which lasted up to seven hours. Feed and permeate 119 samples were regularly collected. The pH value of the feed solution was adjusted (1 120 M HNO₃, VWR Aristar, or 1 M NaOH) during experiments to maintain the pH value 121 within \pm 0.2. The pH value of permeate samples deviated from that of the feed with 122 up to a value of ± 0.5 .

123

124 The mass of uranium sorbed by the membrane (mg U m⁻² of membrane) was simply 125 calculated as the mass uptake of uranium by the membrane divided by the membrane 126 area. To report the mass sorbed as percentage uptake, the uranium mass sorbed by the 127 membrane was calculated in relation to the initial mass of uranium available in the 128 feed (*i.e.* 1 mg U):

129

$$Uptake (\%) = \frac{V_{b0}C_{b0} - V_{bf}C_{bf} - \sum_{i=0}^{i=n} V_S C_{si}}{V_{b0}C_{b0}} \times 100$$
 Eq. 1

where, V = volume (L), C = uranium concentration (mg L^{-1}), s = permeate or feed sample taken, b = bulk (or feed), f = final, 0 = initial and n indicates the number of samples collected during the experiment.

137

Following completion of each experiment and the removal of the membrane, the system was thoroughly washed using dilute HNO₃ (2%, v/v Analar, VWR) followed by consecutive washes with MilliQ water. For quality control purposes a sample was collected of the final system washes and analysed for uranium content (below detection in washes).

143

144 Chemical analysis and quality control. The uranium concentrations in the feed samples was determined by Inductively Coupled Plasma - Optical Emission 145 146 Spectroscopy (ICP-OES, Optima 5300 DV, Perkin Elmer, USA) whilst those in the 147 permeate samples were determined by Inductively Coupled Plasma - Mass 148 Spectrometry (ICP-MS, Agilent 7500ce). Calibration standards were prepared using a uranium stock solution (1000 mg L⁻¹ U, Merck) and dilute HNO₃ (2% v/v, Aristar, 149 150 VWR International, UK). For both ICP-OES and ICP-MS analysis, check standards 151 and blank samples (2% v/v HNO₃) were inserted after every 5-10 samples during the 152 run. The accuracy of the calibration was asserted for both techniques by analyzing a 153 standard reference solution (ICP Multi Element Standards Solution VI, Certipur). The 154 average value obtained for the standard reference material during ICP-OES analysis was $0.978 \pm 0.140 \text{ mg L}^{-1}$ (expected: $1.0 \pm 0.05 \text{ mg L}^{-1}$). The average value for the 155 ICP-MS analysis was $0.979 \pm 0.59 \ \mu g \ L^{-1}$ (expected: $1.0 \pm 0.05 \ \mu g \ L^{-1}$). 156

158 Micro-X-ray Fluorescence spectroscopy. The penetration of uranium into 159 membranes was explored for selected experiments using micro-X-ray Fluorescence spectroscopy (µ-XRF, XGT-7000 microscope from Horiba Jobin Yvon). Cross-160 161 sections of the dry membranes were cut using scissors and placed between paperboard 162 support for analysis in a line over multiple points (1000s/point) along the membrane 163 cross-section with an incident X-ray beam size of 10 µm (Rh target, accelerating 164 voltage of 50 kV, current 1 mA). More detailed information on this method is 165 provided in the Supporting Information.

166

Speciation modeling. The modeling of the uranium and other ions present in the
aqueous solution was carried out across the pH range 3-10 using visual Minteq 2.53
(KTH, Stockholm, Sweden), as described in Rossiter et al.²⁹

170

171 Membrane characterization. NF and RO membranes have been shown by other 172 studies to have different surface chemistries. The BW30 membrane consists of a fully 173 aromatic polyamide active layer coated with an aliphatic layer rich in alcoholic 174 groups, whilst other membranes have a fully or semi aromatic polyamide active layer with no coating.⁴⁰ As these different functional groups may affect membrane surface 175 176 charge⁴¹ and hence potentially, the interaction with different uranium species, the 177 membrane zeta potential was measured in background electrolyte solution (1 mM 178 NaHCO₃ and 20 mM NaCl) using an electro-kinetic analyzer (EKA, Anton Paar, 179 Austria). MWCO and pore-size was determined experimentally using a range of 180 neutral organic molecules (dioxane, dextrose (Fisher, UK), xylose (Acros Organics, 181 UK) and polyethylene glycol of different molecular weight (400, 600 and 1000 g mol⁻

¹, Fisher, UK)) at a concentration of 25 mg carbon L^{-1} applied to the same membrane 182 sample, following methods described in Hilal et al. and Nghiem et al.^{42,43} For the 183 determination of pore radius, film theory was used⁴⁴ and a correlation was used to 184 estimate the mass transfer coefficient of UO2^{2+,45,46} Organic carbon concentrations 185 186 were measured in non-purgeable organic carbon mode using a TOC Analyser 187 (Shimadzu TOC-VCPH, UK) with an ASI autosampler and high-sensitivity catalyst. 188 Salt flux was calculated using conductivity measurements. Average permeability and 189 standard deviation of the membrane coupons was calculated using the stabilized flux 190 measurements during compaction.

191

192 **Results and Discussion.**

Membrane characterisation. The zeta potential measurements showed that the net charge of TFC-SR2 and BW30 were similar to each other and varied with pH: the overall charge was positive at pH 3, the iso-electric points for TFC-SR2 and BW30 were at pH 4.25 and pH 4.19 (Table 1), respectively, after which the magnitude of the negative charge increased with increasing pH value. The nominal MWCO (90% retention), permeability measurements and pore radius calculation showed that TFC-SR2 has a more open structure compared with BW30 (Table 1).

200

Uranium sorption to NF and RO membrane across the pH range 3-10. The objective of these experiments was to determine the influence of pH on the extent of uranium sorption by the membrane without the application of pressure. The results for uranium sorption by the membrane are shown in Figure 1 together with the dominant uranium species across the pH range. Sorption increased from < 5.5% at pH 3 for both 206 membranes, to a maximum of 31% for BW30 and 50% for TFC-SR2 at pH 6. 207 Thereafter sorption decreased to < 10% at pH 10 for both membranes. Notably, the 208 sorption was similar for both membranes at the pH extremes, whereas, although 209 following the same pattern, the sorption was at least 20% lower for BW30 at peak 210 sorption than for TFC-SR2.

211

The results of these experiments confirm the observations by Rossiter et al.²⁹ that 212 213 there is a strong interaction between uranium and NF/RO membranes, especially at 214 pH values 5-7. The speciation modelling showed that the dominant uranium species 215 vary greatly with pH and the valence of the uranium species also changed from being 216 positive at acidic pH (pH 3-6), to either neutral or carrying single negative charge 217 under near-neutral conditions to highly negative at alkaline pH (pH 8-10). Since the 218 overall membrane charge also varies with pH, going from weakly positive (pH 2-4) to 219 highly negative at pH values above pH 5, charge interactions are likely to play an 220 important role in uranium sorption. Electrostatic repulsion can explain the low 221 interaction between uranium and the membranes at pH 3-4 and pH 8-10, where 222 uranium species and membrane carry the same charge. Electrostatic attraction is a 223 likely contributor to the greater sorption of uranium to the membrane at pH 5 and 6, 224 which was 49% for the TFC-SR2 and varied from 25 to 31% for the BW30 225 membrane. Charge interactions between uranium species and NF/RO membranes 226 hence govern uranium sorption to the membranes.

227

228 The uranium species also vary in molecular weight (Table 2), and hence size 229 exclusion might contribute to the difference in sorption for both membranes studied. 230 At pH 5-6, where high uranium sorption takes place, the MW of the main uranium species (UO₂OH⁺, UO₂²⁺ and UO₂CO₃, Figure 1) is considerably lower (270-330 g 231 mol⁻¹) than the MWCO of the TFC-SR2 membrane (MWCO: 486 g mol⁻¹). The size 232 233 difference between the membrane MWCO and the uranium species, allied to charge 234 attraction between the negatively charged membrane and the positively charged or neutral uranium species, suggests ease in penetrating the TFC-SR2 membrane⁴⁷ in the 235 236 absence of pressure and with access to the active and support layers for sorption. 237 Charge attraction will also occur between the uranium species and the BW30 membrane, as both TFC-SR2 and BW30 membranes have similar negative surface 238 charge.⁴⁸ However, the BW30 membrane has a MWCO of 88 g mol⁻¹, hence based on 239 240 size exclusion, a much lower uranium penetration into the membrane active layer and 241 support layer would be expected. In fact, the sorption by TFC-SR2 reached 242 equilibrium more slowly (generally 30-60 minutes longer) than for BW30, which 243 would be consistent with slower diffusion of uranium, followed by sorption into the 244 porous active and support layer structure. The higher sorption onto the TFC-SR2 245 membrane as opposed to the BW30 membrane could hence be caused by a higher 246 surface area available for the uranium species to sorb onto the membrane.

To provide further evidence for the penetration of uranium into the TFC-SR2 membrane, μ -XRF analysis was performed on four selected membrane samples after experiments at pH 6 and pH 8.5 for both BW30 and TFC-SR2. These pH values were selected in order to investigate a point of high sorption of uranium at pH 6 and one of lower uranium sorption at pH 8.5. The μ -XRF analysis showed significant differences with regards to uranium distribution for both membranes studied (Figure 2). As the spatial resolution of this method is relatively low (the incident X-ray beam size is 10 254 µm and penetrates through the sample so the lateral resolution is very low *i.e.* mm 255 range), the exact location of uranium cannot be conclusively determined, *i.e.* whether 256 uranium is present in the active layer, support layer or both, since the NF/RO membrane active layers have a thickness of around 200 nm.⁴⁹ Neither the thickness of 257 258 the active layer nor that of the membrane can be accurately determined with this 259 method as the method picks up the sulfur signal of the polysulphone support layer but 260 not the signal corresponding to the polyester support layer. However, the sulfur 261 signals indicate the presence of the polysulfone support layer, whereas the calcium 262 signals indicate the *approximate* boundaries of the membrane since the calcium 263 signals originate from the mounting material (see Supporting Information). The XRF 264 analysis presented in Figure 2 a confirms that uranium entered into the more open 265 membrane structure of the TFC-SR2 at pH 6, as the uranium peak for this membrane 266 overlapped with the sulfur peak. In contrast, no uranium could be detected for the 267 BW30 membrane (Figure 2 c), showing low or no penetration into this membrane. 268 Size exclusion hence plays an important role in uranium sorption. At pH 8.5, albeit 269 lower compared to pH 6 due to charge repulsion, uranium penetration and internal 270 sorption occurred for the TFC-SR2 membrane (Figure 1 and Figure 2 b). This 271 occurred independently of conditions of charge repulsion between the negatively 272 charged membrane and the negatively charged uranium species, showing the effect of 273 membrane pore size and hence size exclusion in uranium penetration and sorption into 274 the membrane structure. Uranium sorption onto the more opened TFC-SR2 membrane 275 is hence not only governed by charge interactions but also by access to internal 276 surface area governed by membrane pore size. In contrast, no uranium could be 277 detected for the BW30 membrane for pH 8.5, as can be seen in Figure 2 d, showing that any sorption observed in Figure 1 by this membrane was not likely to occur deep inside the membrane structure but mainly on the surface: uranium sorption for the dense RO membrane is hence governed by charge interactions. The lower penetration into BW30 compared to TFC-SR2 hence indicates that membrane pore size acts as a limiting factor to sorption of uranium by the membranes. Pore size has similarly been determined as an important factor in membrane sorption of hormones.⁵⁰

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285 Uranium sorption by membrane at pressures 5-15 bar. The previous section 286 demonstrated the higher penetration of uranium species into the membrane of greater 287 porosity (TFC-SR2), across the pH range and irrespective of uranium species present. 288 The objective of these experiments was to investigate the effect of pressure on the 289 uranium sorption to the membranes. Pressure is likely to enhance the permeation of 290 solutes inside the membrane and hence facilitate access to the internal membrane 291 surface area, where sorption may occur. It may also lead to a higher concentration of 292 uranium at the membrane surface (through concentration polarisation) and, as a 293 consequence, precipitation might occur. To investigate this, a pressure range of 5-15 294 bar was selected based on the typical range for spiral wound membranes (3-20 bar)⁵¹ and manufacturer recommendations.^{52,53} Again, pH values 6 and 8.5 were selected. 295

296

The resulting membrane sorption of uranium as a function of pressure is presented in Figure 3. The sorption varied significantly between membrane types and also between pH values and thus uranium species present. For TFC-SR2 at pH 6, a constant sorption of $50 \pm 5\%$ of uranium (equivalent to a range of 109-125 mg U m⁻² of the membrane surface) was observed across the pressure range (5-12.5 bar) (Figure 3a); 302 only at 15 bar was there a significant increase in uranium sorption to 69%. The 303 sorption for the same membrane but at pH 8.5 was different: sorption increased with 304 pressure, from < 20% at 5 bar up to 61% at 12.5 and 15 bar (Figure 3b). Conversely, 305 for BW30 the sorption of uranium by the membrane remained unaffected by pressure 306 at both pH 6 and pH 8.5 (Figure 3c and d). The results show that irrespectively of 307 pressure and pH, uranium sorption is higher in TRC-SR2 than in BW30. TFC-SR2 308 also gave lower retention for uranium, across the pressure range of 5-15 bar: 90% \pm 309 6% at pH 6, 94% \pm 5% at pH 8.5 while for BW30 uranium retention was 99.7% \pm 310 0.3% for both pH values and across the entire pressure range. The results are 311 consistent with the larger pore size of TFC-SR2 compared to BW30 which is related 312 with permeability. The permeability of the two membranes were compared using the 313 pure water flux before and after completed experiments, and the permeate flux during 314 experiments (Figure 4). Pure water flux for BW30 was approximately half of that for 315 TFC-SR2 at the same pressure, reflecting the difference in permeability between the 316 two membranes. As expected, BW30 experienced a flux decline during the uranium 317 experiments, consistent with effects of concentration polarisation and osmotic pressure difference between the feed and permeate side.⁵⁴ Once the pure water flux 318 319 was again measured after the uranium experiments, it was restored to its original 320 value. By contrast, the flux of TFC-SR2 unexpectedly increased with the addition of 321 experimental solution (and also with addition of salt solution not containing uranium; 322 results not included) and remained high even when the experimental solution was 323 drained and pure water was filtered. Although unusual, a similar effect has been reported by several studies.^{55–57} Nilsson et al. linked the flux increase to pore 324 expansion caused by salt ions reducing the strength of the membrane cross-links.⁵⁷ 325

326 Such pore expansion within the TFC-SR2 membrane would further enhance the 327 penetration of uranium into the membrane.

328

329 Furthermore, concentration polarisation, which increases with increasing 330 permeability, is likely to affect the filtration, leading to an accumulation of solutes 331 and consequently higher concentration adjacent to the membrane surface. Taking 332 concentration polarisation into account, the solute concentration at the membrane surface was calculated using Equation 2,⁴⁴ 333

334

$$\frac{Cm - Cp}{Cb - Cp} = \exp\left(\frac{Jv}{k}\right)$$
Eq. 2

where Cm = concentration at membrane surface (mg L⁻¹), Cp = concentration inpermeate (mg L⁻¹), Cb = concentration in bulk solution (mg L⁻¹), Jv = permeate flux(m s⁻¹) and k = the mass transfer coefficient (m s⁻¹). Whereas Cp, Cb and Jv aredetermined experimentally, k had to be calculated using correlations relevant to theexperimental conditions (slit channel and laminar flow). For the experimentalconditions of the system used, the Sherwood number can be related to the Reynoldsand Schmidt number⁵⁸ as described in Equation 3:

$$Sh = 1.85 \times \text{Re}^{0.33} Sc^{0.33} (d_h / L)^{0.33}$$
 Eq. 3

where Re = the Reynolds number, Sc = the Schmidt number, dh = channel hydraulic diameter, L = the length of the membrane cell. The Reynolds, Schmidt and Sherwood number were calculated as described in Semião et al.³⁷

351

352 The extent of concentration polarisation experienced by a membrane can be reported 353 as the concentration polarisation modulus, giving the ratio of initial concentration at 354 the membrane surface (Cm) to that in the bulk solution (Cb) at the start of the 355 experiment. The concentration polarisation modulus for each pressure experiment are 356 displayed together with the uranium uptake in Figure 3. During filtration with BW30, 357 uranium uptake (around 30 and 20% for pH 6 and 8.5, respectively) for both pH 358 values remained unaffected by pressure and concentration polarisation (Figure 3c and 359 d). Due to its high permeability the TFC-SR2 membrane was more affected than 360 BW30 by concentration polarisation. There were some important differences between 361 the results for TFC-SR2 at the two pH values. At pH 8.5, uranium sorption to the 362 membrane clearly followed the polarisation modulus trend. This trend was similar to a 363 previous study with hormones and a NF270 membrane, where higher polarisation 364 modulus resulted in higher concentration at the membrane surface translating into a higher adsorption.³⁷ In contrast, at pH 6 the uranium uptake to the TFC-SR2 365 366 membrane remained constant despite pressure and concentration polarisation increase, 367 with the exception of the highest pressure point. It appears that concentration 368 polarisation does not affect the interaction between the uranium species and 369 membrane at pH 6, suggesting that variations in uranium concentration only have a 370 small effect on the amount of uranium sorbed by the membranes. To confirm this, an 371 adsorption isotherm for uranium at pH 6 was plotted (Figure S2). Using linear fit of 372 the sorption isotherm, the uranium sorption, based on the concentration at the membrane surface Cm of 0.62-1.03 mg L⁻¹ for pressures 5 to 15, was estimated to 373 around 107-123 mg U m⁻² of membrane surface, i.e. a sorption of 50% to 57% based 374 375 on mass balance, showing that pressure and hence concentration polarisation had an 376 insignificant effect on uranium sorption at pH 6. Precipitation as an uptake 377 mechanism could be excluded, as even at high concentration polarisation the uranium 378 concentration at the membrane surface was calculated to be a maximum of 1.03 mg L⁻ 379 ¹ and still remained below maximum solubility of uranium, thus confirming sorption 380 to be the main mechanism governing uranium-membrane interactions.

381

382 This study confirmed that the uranium-membrane interactions were highly speciation and pH dependent, with affinity determined by the charge of both membrane and 383 384 uranium species (e.g. UO_2OH^+ or $UO_2(CO_3)_3^{4-}$) as well as species size relative to 385 membrane pore size. Pore size and subsequent permeability of the membranes 386 governed uranium sorption, where TFC-SR2 was subject to higher uranium sorption 387 than BW30 under all experimental conditions. Concentration polarisation affected only one of the uranium species, $UO_2(CO_3)_3^{4-}$, a species which generally tends to 388 display low sorption and high mobility,^{59,60} but its sorption to TFC-SR2 increased 389 390 with increasing pressure. UO₂OH⁺, whose sorption to TFC-SR2 was initially higher, 391 remained largely unaffected by concentration polarisation. This study has provided a 392 first insight into the nature of the interactions of uranium with NF and RO membranes 393 and the clear effects of pH and charge interactions, membrane pore size and 394 concentration polarisation on uranium sorption. Uranium sorption might be further 395 affected by the presence of different functional groups on the membrane active layer. Tang et al,⁶¹ for example, showed that some RO membranes possess a surface layer coating rich in -COH groups in addition to the aromatic or semi-aromatic polyamide active layer. Hence, future work focused on a more in-depth analysis to determine the chemical nature and spatial distribution of the uranium sorption to the membranes, as well as the effect of hardness on the removal of uranium by NF and RO membranes is needed.

402

403 The results are of significance in the wider membrane application context since it 404 illustrates the importance of taking sorption of contaminants into account. The 405 retention *observed* in experimental and applied water treatment settings may not be 406 the actual or real retention, and the long-term consequences of sorption to the 407 membranes remains unknown. One possible consequence is the risk of uranium de-408 sorption from the membrane into the permeate line during operation, especially at 409 acidic pH values, which could pose a health risk to water consumers. There is great 410 variability in membrane life-time and performance from location to location and 411 contaminant sorption (not necessarily picked up since it may not cause obvious 412 fouling and consequent flux decline) may be one of the determining factors.

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414

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417 FIGURES

418



Figure 1. Uranium speciation (lines) and uranium uptake or sorption (columns) by membranes TFC-SR2 and BW30 across the pH range 3-10. Experimental solution: 0.5 mg L⁻¹ uranium, 20 mM NaCl and 1 mM NaHCO₃. Experimental conditions: flow-rate = 0.6 L min⁻¹, temperature = 24°C, no applied pressure. The variation in uptake was within \pm 4% for repeated experiments for TFC-SR2 and \pm 1% for BW30.



Figure 2. Cross-sections of TFC-SR2 and BW30 showing the elemental distribution of U, S and Ca for experiments performed without pressure for TFC-SR2 at pH 6 (a) and pH 8.5 (b) and BW30 at pH 6 (c) and pH 8.5 (d) determined by μ -XRF. The *approximate* top and bottom edge of the membrane is indicated by the shaded area based on the detection of the calcium mounting material. Note the different intensity scales for Ca and S compared to U.



Figure 3. The percentage uptake (sorption) of uranium (columns) during the experiments is shown on the right y-axis, while the calculated polarisation modulus (Cm/Cb) for uranium is indicated by points and displayed on the left y-axis. Repeatability of U uptake for selected experiments was within \pm 4% for TFC-SR2 and \pm 1% for BW30.

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Figure 4. Pure water flux (PWF) before and after selected pressure experiments, and permeate flux during pressure experiments (5-15 bar) for TFC-SR2 and BW30 at pH 6 and 8.5. Permeability variability for TFC-SR2 was \pm 13% and BW30 was \pm 3% based on pure water flux experiments. Variability of permeate flux during experiments was within \pm 10% for TFC-SR2 and \pm 2% for the BW30. Please note different flux scales for TFC-SR2 and BW30.

- 452
- 453 TABLES

454 Table 1. Membrane characteristics

Parameter	TFC-SR2	BW30
Iso-electric point/pH	4.25	4.19
Nominal MWCO/g mol ⁻¹	486 ^a	88
Pore radius/nm	0.52 ± 0.03	$0.32 \pm 0.01^{b,43}$
Permeability/L m ⁻² h ⁻¹ bar ⁻¹	10.97 ± 1.51	4.84 ± 0.15

455 ^aAbsolute MWCO (100% retention) for TFC-SR2 was determined as 1033 g mol⁻¹.

456 ^bNote that RO membranes are considered to have dense, non-porous structures and so

- 457 "pore-radius" for BW30 was determined only as a comparison with the more open
- 458 structure of TFC-SR2.
- 459
- 460

461 Table 2. Main uranium species present for pH range 3-10^a

pH value	Uranium species	Molecular weight (g mol ⁻¹)
3-4	UO_2^{2+}	270
5-6	$\rm UO_2OH^+$	287

6-7	UO ₂ CO ₃	330
7	(UO ₂) ₂ CO ₃ (OH) ₃ ⁻	651
8	$UO_2(CO_3)_2^{2-}$	390
8-10	$UO_2(CO_3)_3^{4-}$	450

^{462 &}lt;sup>a</sup>Note that each pH value may contain a mixture of several species.

464

465

- 466 ASSOCIATED CONTENT
- 467 Supporting information. Details of experimental set-up and filtration procedure,
- 468 ICP-OES, ICP-MS and XRF analysis, salt retention and isotherm experiment. This
- 469 material is available free of charge via the internet at http://pubs.acs.org.

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