



Article (refereed) - postprint

Van der Sande, Masha T.; Poorter, Lourens; Balvanera, Patricia; Kooistra, Lammert; Thonicke, Kirsten; Boit, Alice; Dutrieux, Loic P.; Equihua, Julian; Gerard, France; Herold, Martin; Kolb, Melanie; Simoes, Margareth; Pena-Claros, Marielos. 2017. **The integration of empirical, remote sensing and modelling approaches enhances insight in the role of biodiversity in climate change mitigation by tropical forests.**

© 2017 Elsevier B.V.

This manuscript version is made available under the CC-BY-NC-ND 4.0 license <http://creativecommons.org/licenses/by-nc-nd/4.0/>



This version available <http://nora.nerc.ac.uk/515990/>

NERC has developed NORA to enable users to access research outputs wholly or partially funded by NERC. Copyright and other rights for material on this site are retained by the rights owners. Users should read the terms and conditions of use of this material at <http://nora.nerc.ac.uk/policies.html#access>

NOTICE: this is the author's version of a work that was accepted for publication in *Current Opinion in Environmental Sustainability*. Changes resulting from the publishing process, such as peer review, editing, corrections, structural formatting, and other quality control mechanisms may not be reflected in this document. Changes may have been made to this work since it was submitted for publication. A definitive version was subsequently published in *Current Opinion in Environmental Sustainability* (2017), 26-27. 69-76. [10.1016/j.cosust.2017.01.016](https://doi.org/10.1016/j.cosust.2017.01.016)

www.elsevier.com/

Contact CEH NORA team at
noraceh@ceh.ac.uk

Highlights

- The role of biodiversity on climate change mitigation by tropical forests remains poorly understood.
- Empirical, remote sensing and modelling studies provide complementary information.
- In more than 75% of the studies, biodiversity significantly affected carbon storage or sequestration.
- Biodiversity underpins short-term ecosystem functioning and assures long-term carbon storage and sequestration in tropical forests.
- Integrating approaches by using ‘boundary objects’ will lead to a comprehensive understanding.

1 **Title:** The integration of empirical, remote sensing and modelling approaches enhances insight in the
2 role of biodiversity in climate change mitigation by tropical forests

3
4
5
6 **Authors:**

7 Masha T. van der Sande^{1,2*}, Lourens Poorter¹, Patricia Balvanera³, Lammert Kooistra⁴, Kirsten
8 Thonicke⁵, Alice Boit⁵, Loïc P. Dutrieux³, Julian Equihua⁶, France Gerard⁷, Martin Herold³, Melanie
9 Kolb⁶, Margareth Simões^{8,9} and Marielos Peña-Claros¹

10
11
12
13 ¹ Forest Ecology and Forest Management Group, Wageningen University and Research, PO Box 47,
14 6700 AA Wageningen, The Netherlands

15
16
17 ² Environmental Research (Alterra), Wageningen University and Research, PO Box 47, 6700 AA
18 Wageningen, The Netherlands

19
20 ³ Instituto de Investigaciones en Ecosistemas y Sustentabilidad, Universidad Nacional Autónoma de
21 México, Morelia, Mexico.

22
23 ⁴ Laboratory of Geo-Information Science and Remote Sensing, Wageningen University and Research,
24 P.O. Box 47, 6700 AA Wageningen, The Netherlands

25
26 ⁵ Earth System Analysis, Potsdam Institute for Climate Impact Research e.V. (PIK), Telegrafenberg
27 A31, D-14473 Potsdam, Germany

28
29 ⁶ CONABIO, National Commission for the Knowledge and Use of Biodiversity, Mexico DF, Mexico

30
31 ⁷ Centre for Ecology and Hydrology, Maclean Building, Wallingford, OX10 8BB, UK

32
33 ⁸ Embrapa Solos - Brazilian Agriculture Research Corporation, Rio de Janeiro, Brasil

34
35 ⁹ UERJ/DESC/PPGMA Rio de Janeiro State University, Rio de Janeiro, Brasil

36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

* Corresponding author:

Masha T. van der Sande

Deutscher Platz 5e

04103 Leipzig

Germany

29

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

Email: masha.vandersande@idiv.de

Abstract

Tropical forests store and sequester high amounts of carbon and are the most diverse terrestrial ecosystem. Studies show potentially important effects of biodiversity on carbon storage and sequestration, but a complete understanding of this relationship across spatiotemporal scales relevant for climate change mitigation needs three approaches: empirical, remote sensing and ecosystem modelling. Here, we review the contribution of these *individual approaches* to the understanding of the relationship of biodiversity with carbon storage and sequestration, and find short-term and long-term benefits of biodiversity at both broad and fine spatial scales. We argue that enhanced understanding is obtained by *combining approaches*, i.e., by using output from one approach to improve another approach and thus results in better input, validation and comparison between approaches. This can be further improved by *integrating approaches* through using ‘boundary objects’ (i.e., variables) that can be understood and measured by all approaches, such as the diversity of leaf traits of the upper canopy and forest structure indices. Combining and especially integrating approaches will therefore lead to a better understanding of biodiversity effects on climate change mitigation. This is crucial for making sound policy decisions.

Keywords: biodiversity-ecosystem functioning, carbon sequestration, carbon storage, forest structure, functional diversity, REDD+, species diversity

Introduction

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
66
67
68
69
70
71
72
73
74
75
76
77
78
79
80
81
82
83

Tropical forests play a major role in the global carbon cycle and are therefore important for climate change mitigation [1]. Tropical forests are also biodiversity hotspots and thus relevant for biodiversity conservation [2]. With ‘biodiversity’ we refer here to different vegetation properties: species and trait diversity, community-mean trait values and forest structure. Biodiversity is important for human wellbeing; it provides essential supporting, provisioning, cultural and regulating ecosystem services [3]. For example, biodiversity can potentially increase the capacity for carbon storage and sequestration, not only in temperate systems [e.g., 4,5] but also in highly diverse tropical forests [6]. In turn, this increased carbon uptake capacity may lead to a higher potential for climate mitigation. By evaluating how biodiversity affects carbon storage and sequestration, we can underscore the importance of tropical forests not only for nature conservation but also for climate mitigation.

To fully understand the effect of biodiversity on carbon storage and sequestration (hereafter referred to as “carbon”), we first need to develop a mechanistic understanding of the short-term and local-scale effects of biodiversity on carbon. For this relationship to be relevant for global climate change mitigation, it should also occur at broader spatial and longer temporal scales that will have most impact and long-term benefit on climate change mitigation. Three approaches are needed to cover these different aspects of the relationship between biodiversity and climate change mitigation: an empirical approach to understand the short-term and local-scale relationship (Figure 1, Arrow 1), a remote sensing approach to scale up to broad spatial scales (Figure 1, Arrow 2) and an ecosystem modelling approach to scale up to long temporal scales (Figure 1, Arrow 3). These approaches are complementary in their ecological realism, spatial and temporal scale and contribute differently to the understanding of the biodiversity-carbon relationship and its consequences for global climate change mitigation (Table 1, Figure 1).

In this paper, we advocate that combining and integrating empirical, remote sensing and ecosystem modelling approaches is needed to understand biodiversity effects on carbon across spatiotemporal scales. To show this, we perform a literature review to bring together evidence from the **individual approaches** to evaluate their contribution to the understanding of the biodiversity-carbon relationship. We then discuss how we can **combine approaches** to improve the assumptions, cross-validation and output of studies evaluating the biodiversity-carbon relationship. Finally, our study moves beyond the

84 concept of combining approaches to **integration of approaches**. This is essential to link, scale and
85 translate among the approaches, and therefore to provide the best understanding of the biodiversity-
86 carbon relationship across spatiotemporal scales that are most relevant for climate change mitigation.

87

88

89 **Contribution of individual approaches – a review**

90

91 *Empirical studies* – Empirical studies and experiments in the field that control for confounding factors
92 provide insight into mechanisms underlying the biodiversity-carbon relationship, identify important
93 aspects of biodiversity and provide evidence whether this relationship is strong enough (i.e., detectable
94 and significant) to have a substantial effect on the functioning of natural systems. Empirical evidence
95 for the biodiversity-carbon relationship is increasing rapidly, starting from conceptual ideas [7] to
96 testing this relationship for different ecosystems [e.g., 4,5]. For tropical forests, however, the evidence
97 is still emerging and scattered among local studies [e.g., 8,9] and regional to continental studies [e.g.,
98 6,10].

99 A recent review [11] evaluated 38 empirical studies on the role of different biodiversity indicators
100 for carbon storage and dynamics (i.e., fluxes of carbon over time such as growth and mortality) in
101 tropical forests. This review provided three main results. First, carbon dynamics increased most often
102 with taxonomic diversity [e.g., 12], whereas carbon storage depended most on the average trait values
103 of the tree community (i.e., community-mean traits) [13]. These results indicate that biodiversity is of
104 major importance, but that different biodiversity indicators represent different mechanisms by which
105 they contribute to carbon storage or dynamics: i) taxonomic (or functional) diversity can increase
106 complementarity among species in their strategies to acquire and use resources, and as such increase
107 overall carbon storage and sequestration [14] and ii) community-mean traits mainly represent the most
108 dominant species in a community, which may most strongly influence carbon storage and
109 sequestration [15]. A thorough understanding of the role of different biodiversity indicators on
110 multiple carbon-related variables is therefore necessary to guide climate change mitigation policies.

111 Second, the review [11] showed that this relationship is stronger in mature forests than in disturbed or

112 plantation forests, perhaps because of stronger competition and thus higher importance of biodiversity
113 for carbon in denser forests. Third, the biodiversity-carbon relationship was stronger at broader spatial
114 scales across sites (e.g., across Neotropical forests [6,10]), possibly because of stronger variation in
115 biodiversity across sites at broader spatial scale. However, since empirical studies mostly capture
116 processes at the plot or landscape scale, the role of spatial scale in the biodiversity-carbon relationship
117 remains unclear.

Remote sensing studies – Remote sensing allows to assess the biodiversity-carbon relationship at
118 continuous and broader (i.e., regional to global) spatial scales relevant for policy. Remote sensing
119 monitors changes in carbon and biodiversity over time, which is important for, among others, the
120 measurement, verification and reporting of countries' efforts to Reduce Emissions from Deforestation
121 and forest Degradation (REDD+). However, remote sensing is based on indirect proxies for ecosystem
122 processes and properties and is limited in analysing site-specific conditions such as soil fertility that
123 can co-determine carbon.

Several studies reviewed the potential and limitations of remote sensing based methods for
125 measuring and monitoring carbon [16] and biodiversity [17,18] of tropical forests (for relevant
126 advances in this field see Appendix S1). For forest carbon, wall-to-wall pan-tropical benchmark maps
127 based on different techniques and resolutions have been developed [19–21]. However, remote sensing
128 based maps of biodiversity are still rare [22], thus limiting the number of studies, especially broad
129 scale, that evaluate biodiversity-carbon relationships. We identified and qualitatively assessed 10
130 studies that evaluated this relationship (Appendix S2a-c). Nine of the ten studies show a positive
131 relationship between biodiversity and carbon storage (no studies evaluated carbon sequestration), for
132 different biodiversity indicators: plant species diversity (7 studies), fauna species diversity (2) and
133 plant trait diversity (1). The strength of the biodiversity-carbon relationship varied considerably
134 among studies ($r = -0.01 - 0.83$) but seems to be scale-independent: both the strongest and the weakest
135 correlations were found at the fine scale (Appendix S2a). At least three possible reasons may explain
136 the variation in correlation strength. First, differences in environmental conditions may explain this
137 variation. Spatial variation in rainfall seasonality and species richness was significantly positively
138 related to the strength of the correlation between species richness and carbon storage (Figure 2,

140 Appendix S2), indicating that species richness effects on carbon storage increase towards more
141 seasonal and towards more diverse forests. Second, the strength may depend on the method used to
142 derive biodiversity and carbon variables. When biodiversity and carbon storage are derived using the
143 same method (e.g., LiDAR), they are not independent and may show a stronger correlation compared
144 to when the variables are obtained from independent remote sensing sources. Third, the strength of
145 this correlation may depend on the prediction accuracy of remote sensing indicators for biodiversity
146 and carbon. A range of retrieval methods is used to estimate carbon storage and biodiversity indicators
147 by relating remote sensing data sources to field observations (Appendix S2b), but an optimal method
148 is still to be agreed on [23]. Although the small number of studies does not yet allow formal testing of
149 the biodiversity-carbon relationship, the studies indicate that hotspots for carbon storage are related to
150 hotspots for biodiversity.

151 *Modelling studies* – Modelling studies allow assessment of the biodiversity-carbon relationship at
152 temporal scales of up to centuries, and evaluate impacts of alternative future climate change scenarios
153 and selected policy interventions. However, modelling is a simplification of the real world and
154 therefore the representation of multiple interacting processes may miss relevant processes.

155 Testing biodiversity-carbon relationships using ecosystem models requires a modelling framework
156 that simulates physiological and morphological processes, plant competition and mortality, and
157 functional and structural diversity. We found only three models that studied biodiversity-carbon
158 relationships (Appendix S4). First, a dynamic plant functional trait model was applied to Australian
159 forests [24]. This study found that, with modest climate change, plant trait diversity increased carbon
160 sequestration in lowland forests, but this effect decreased with strong climate change (under SRES
161 A1FI scenario). Second, species diversity weakly increased forest productivity in northern India
162 (simulated by the remote-sensing based Carnegie-Ames-Stanford Approach (CASA) model) under
163 current climate conditions [25]. Third, functional trait diversity increased forest recovery of carbon
164 stocks, and hence forest resilience, after climate change in a dynamic global vegetation model
165 (DGVM) that accounts for competition and plant trait diversity (Lund-Potsdam-Jena managed Lands
166 with Flexible Individual; LPJmL-FIT, [26]).

167 One reason for the limited amount of studies is the lack of a realistic representation of biodiversity
168 in ecosystem models. A potentially useful modelling approach is the use of DGVMs. Initially,
169 DGVMs had a very simplified representation of biodiversity, using several plant functional types [e.g.,
170 27], but recent model developments focussed on implementing functional diversity or plant trait
171 diversity in the DGVM framework. DGVMs can now include variation in some plant traits, adaptive
172 responses, and trade-offs between traits [e.g., 28] (see Appendix S5 for more details on the models).
173 These model developments will allow testing the biodiversity-carbon relationship at various temporal
174 scales, including the effect of biodiversity on forest resilience.

175 176 177 **The biodiversity–carbon relationship: state of the art**

178
179 We evaluated the biodiversity-carbon relationship using three complementary approaches, and found
180 that biodiversity significantly and positively affected carbon storage and/or sequestration in the
181 majority of the empirical studies (75%) and remote sensing studies (90%) and a weak positive effect
182 on long-term carbon in the most recent models. These results extend the well-known findings from
183 experimental studies and temperate systems that *biodiversity matters for ecosystem functioning in*
184 *tropical forests*.

185 The different approaches provided complementary information on the role of spatial scale. Among
186 empirical studies, the biodiversity-carbon relationship was stronger at large spatial scale (e.g., across
187 Neotropical forests) than at fine spatial scale (e.g., within one forest type). In contrast, remote sensing
188 studies found that the strength of the biodiversity-carbon relationship did not vary with spatial scale,
189 perhaps because of the indirect way in which they assess both biodiversity and carbon. Modelling
190 studies showed that biodiversity is important for carbon not only at short, but also at long temporal
191 scales where it serves as an ‘insurance’ against environmental hazards. Hence, although scale seems to
192 affect the strength of the biodiversity-carbon relationship, *biodiversity underpins short-term*
193 *ecosystem functioning and assures long-term carbon storage and sequestration in tropical forests,*
194 *at both fine and broad spatial scales*. These results indicate that biodiversity conservation is not a

195 mere co-benefit of management for REDD+, but should be considered as a requirement for long-term
1
2 196 effective REDD+ activities [29].
3

4 197

5
6 198

7 8 9 **Combining approaches**

10
11 200

12
13 201 To go beyond individual contributions (Figure 3a), we advocate to *combine approaches* to improve
14
15 202 the quality of input data, refine assumptions, facilitate cross-validation and evaluate the robustness of
16
17 203 relationships across approaches (Figure 3b). We here discuss opportunities to combine the three
18
19 204 approaches. First, empirical and remote sensing approaches can be combined (Figure 3b, Arrow 1) to
20
21 205 validate remote sensing results, e.g. by evaluating the detection algorithm, and to facilitate accurate
22
23 206 upscaling of local observations to broad spatial scales. Second, empirical and ecosystem modelling
24
25 207 approaches can be combined (Figure 3b, Arrow 2) in several ways. For example, the mechanisms
26
27 208 underlying biodiversity-carbon relationships found in empirical studies can be included in modelling
28
29 209 approaches and used to refine model assumptions for more accurate long-term predictions.
30
31
32 210 Furthermore, findings from fine-scale empirical studies can be tested in models over longer temporal
33
34 211 scales, thus facilitating the generalisation of the mechanisms. Third, remote sensing and ecosystem
35
36 212 modelling approaches can be combined (Figure 3b, Arrow 3) by using remote sensing data as an input
37
38 213 for ecosystem models [30], or to validate modelled patterns and processes [31].
39
40
41

42 214 The combination of the three approaches would thus allow better exploration of the mechanisms
43
44 215 behind the biodiversity-carbon relationship at broad spatiotemporal scales. Hence, combining
45
46 216 approaches in such ways – by using output from one approach to improve another approach – leads to
47
48 217 opportunities for better input, validation and scaling.
49
50

51 218

52
53 219

54 55 56 **Integrating approaches**

57
58 221

59
60
61
62
63
64
65

222 ***Integrating approaches*** moves beyond combining them by using similar indicators as input and/or
1
2 223 output. Using similar indicators allows direct comparison among, and scaling between, approaches to
3
4 224 better understand biodiversity-carbon relationships. To avoid translation problems of indicators across
5
6 225 approaches, we propose to use ‘*boundary objects*’, which are indicators that “are both adaptable to
7
8 226 different viewpoints [in our case approaches] and robust enough to maintain identity across them”
9
10 227 [32]. Boundary objects are frequently used in interdisciplinary studies to communicate across
11
12 228 scientific disciplines (such as ‘resilience’ and ‘ecosystem services’ [33,34]). Using boundary objects
13
14 229 that can be measured by the three research approaches could greatly facilitate scaling among them and
15
16 230 advance our understanding of biodiversity effects on climate change mitigation.
17
18

19
20 231 Several potential boundary objects can be used for carbon and biodiversity (see examples in Table
21
22 232 2). Indicators for aboveground carbon storage are relatively easy to quantify by all approaches and are
23
24 233 already being used. Aboveground net carbon change (i.e. net carbon uptake or net biomass growth at
25
26 234 the ecosystem level) can serve as a boundary object for carbon sequestration as it can be measured by
27
28 235 all approaches albeit using different methodologies. Boundary objects for biodiversity are more
29
30 236 complicated to define as the concept of biodiversity is broadly defined, ranging from genetic to
31
32 237 ecosystem diversity (Convention on Biological Diversity). In this review, we separated biodiversity
33
34 238 into three important groups of vegetation properties: species and trait diversity, community-mean trait
35
36 239 values and forest structure (Table 2 [cf. 11]). A useful boundary object for biodiversity is the diversity
37
38 240 and the mean of leaf traits of the upper canopy, such as specific leaf area [35] and leaf nutrient
39
40 241 concentrations [36]. Leaf trait diversity can be easily measured in the field [37] by empirical studies,
41
42 242 can be seen from space for the upper canopy by new hyperspectral remote sensing techniques [e.g.,
43
44 243 35] and are explicitly included in recently developed dynamic global vegetation models [e.g., 38].
45
46 244 Indicators for forest structure, such as crown size distribution of the upper canopy, can also serve as
47
48 245 boundary object, as they can be captured by all three approaches (Table 1). These example boundary
49
50 246 objects can be similarly measured by all approaches and therefore directly used to scale between
51
52 247 approaches. Such boundary objects may thus allow for integration of empirical, remote sensing and
53
54 248 modelling approaches. This, in turn, will help advancing our understanding of biodiversity effects on
55
56 249 carbon across spatiotemporal scales, and thus on climate change mitigation (Figure 1).
57
58
59
60
61
62
63
64
65

250

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

251

252 Concluding remarks

253 Empirical, remote sensing and ecosystem modelling approaches each have their complementary
254 strengths in evaluating biodiversity effects on carbon storage and sequestration. These individual
255 approaches show that biodiversity is generally important for short-term and long-term carbon storage
256 and sequestration, indicating that biodiversity conservation is not only a co-benefit of REDD+
257 activities, but is an integral and crucial component of effective REDD+ implementation [29].
258 However, we advocate that combining, and especially integrating these three approaches will provide
259 an enhanced understanding of how biodiversity contributes to climate change mitigation. We propose
260 the use of boundary objects as a means of integrating all three approaches and span across spatial and
261 temporal scales relevant for climate change mitigations. Such integration of approaches can provide
262 input to guide society and policies such as REDD+ to reach the goals of the UNFCCC Paris
263 Agreement.

264






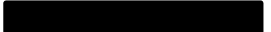
265

266 Acknowledgements

267 The research leading to these results has received partial funding from the European Union Seventh
268 Framework Programme (FP7/2007-2013) under grant agreement n° 283093 – The Role Of
269 Biodiversity In climate change mitigatioN (ROBIN). MvdS received partial funding through the
270 strategic research program KBIV (KB-14) “Sustainable spatial development of ecosystems,
271 landscapes, seas and regions”, funded by the Dutch Ministry of Economic Affairs and carried out by
272 Wageningen University & Research (project code KB-14-003-030). We thank Nashieli Garcia Alaniz
273 and three anonymous reviewers for comments on an earlier version of this manuscript, and Sarah
274 Passonneau for English editing.




1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

Table 1: Overview of the differences among empirical, remote sensing and ecosystem modelling approaches in terms of: spatial scale, temporal scale, biodiversity variables that can be accurately estimated, how likely it is that carbon estimates are correct, the workload per unit area or time evaluated and their main strength. Please note that this overview highlights the main aspects per approach that is relevant for this manuscript, rather than that it provides an exhaustive overview of the properties of the approaches.

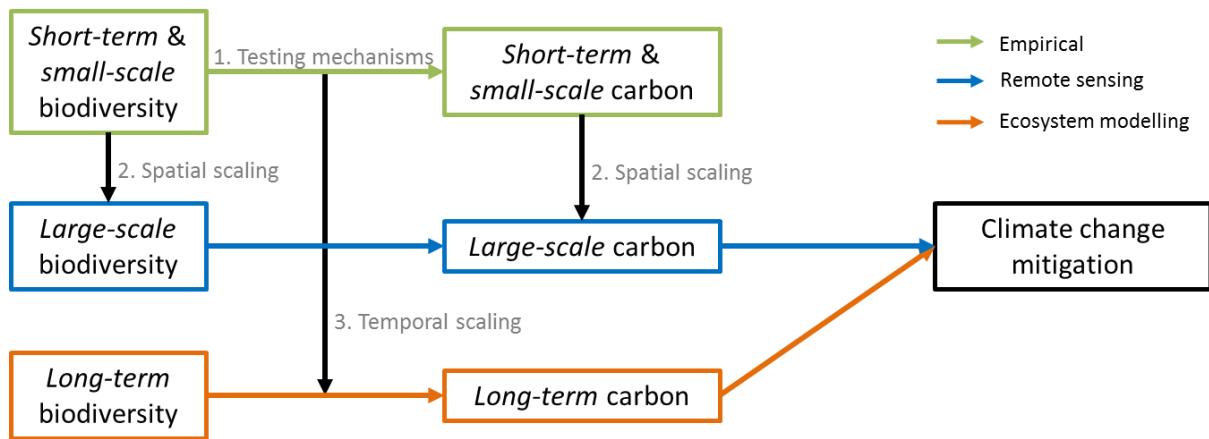
	Spatial scale		Temporal scale		Biodiversity variables that can accurately be estimated	How likely carbon estimate is correct?	Workload per unit area or time evaluated	Main strength
	Small	Large	Short	Long				
Empirical					Species, functional traits, forest structure	Very likely	High	Underpinning mechanisms
Remote sensing					Forest structure	Likely	Medium	Spatial scaling
Ecosystem modelling					Functional groups, forest structure	Likely	Medium	Temporal scaling

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

Table 2: Potential ‘boundary objects’ that can be used to integrate empirical, remote sensing and ecosystem modelling approaches. We identify three potential boundary objects for ‘biodiversity’ (species, functional trait and structural diversity) and two potential boundary objects for ‘carbon’ (storage and net change).

		Empirical studies	Remote sensing studies	Ecosystem modelling studies
				
Biodiversity	Species and trait diversity	Number of species Functional diversity (based on leaf traits)	Number of species, obtained from imaging spectroscopy. Variation in specific leaf area and leaf nutrient concentrations from imaging spectroscopy and hyperspectral imaging	Number of functional groups. Distribution of specific leaf area and other trait values in the modelled tree community
	Trait mean	Community-weighted mean leaf traits	Leaf trait values of tree canopy averaged by area	Average trait values of the modelled tree community.
	Forest structure	Variation in crown size (e.g. diameter)	Variation in crown shape and diameter from LiDAR	Variation in crown size
Carbon	Storage	Standing stocks per unit area	Standing stocks per unit area	Standing stocks per unit area
	Sequestration	Aboveground biomass growth or net change	Aboveground net biomass change	Aboveground gross or net primary productivity

283



284

285 **Figure 1:** Conceptual framework showing how three different research approaches (empirical, remote
 286 sensing and ecosystem modelling) contribute to the understanding of biodiversity effects on carbon
 287 sequestration and storage (“carbon”) and, hence, on climate change mitigation. Their main advantage
 288 and contribution to assessing the role of biodiversity for climate change mitigation is displayed in
 289 boxes, although not being exhaustive. Empirical studies (green, Arrow 1) provide a mechanistic
 290 understanding of biodiversity effects on carbon, both measured at fine spatial scales (e.g., local) and
 291 short temporal scales (e.g., a decade). Remote sensing studies (blue, Arrow 2) scale up to broader
 292 spatial scales (e.g., continental), and ecosystem modelling (orange, Arrow 3) scale up to longer
 293 temporal scales (e.g., centuries). Remote sensing scales up variables (biodiversity and carbon),
 294 whereas ecosystem models generally use the relationship to scale up.

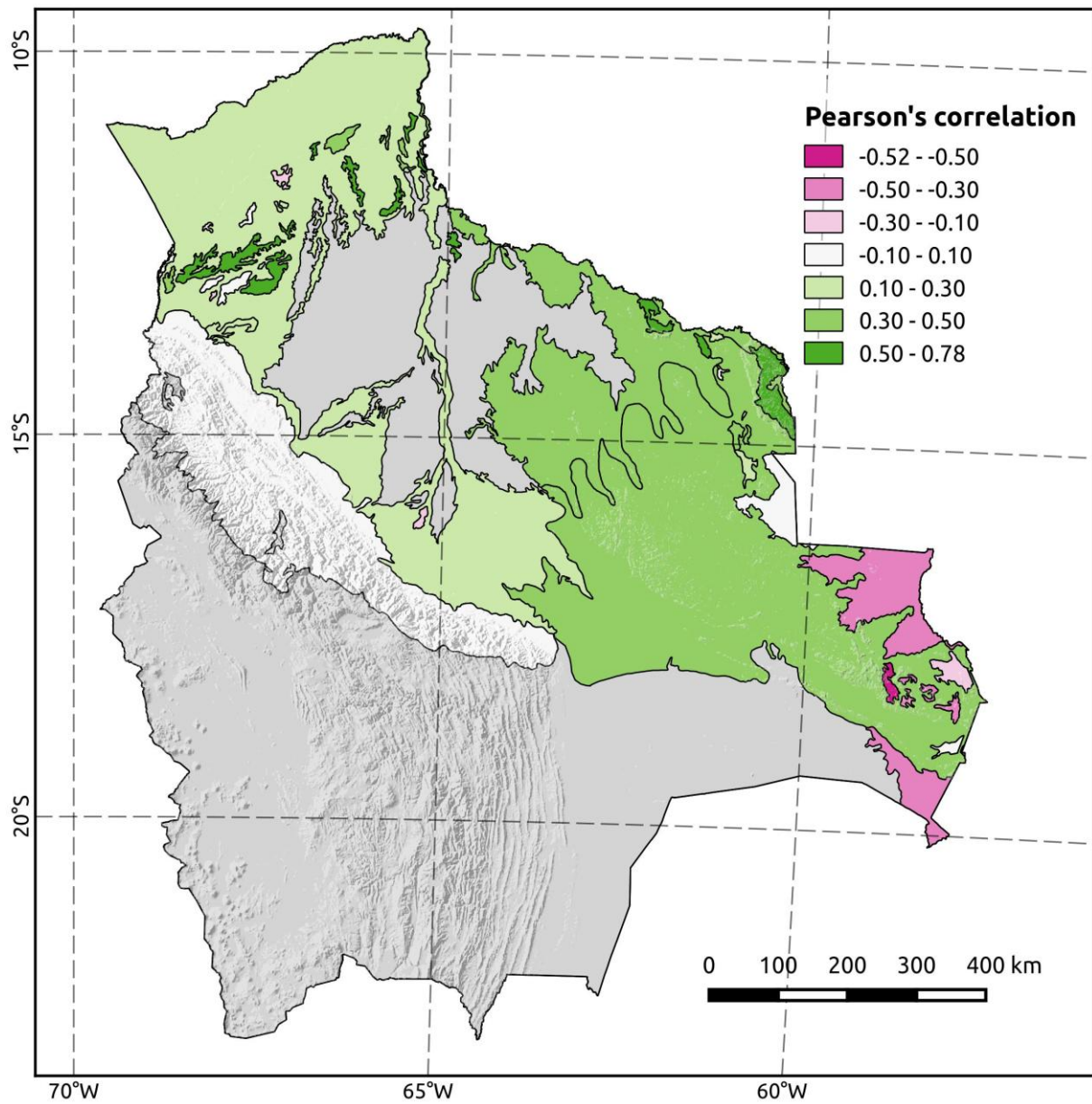


Figure 2: Spatial correlation between remote sensing-derived tree species richness and aboveground biomass for tropical forest in different biogeographic zones in lowland Bolivia (see Appendix S3 for methods). The correlation strength increased with rainfall seasonality (i.e., the coefficient of variation of monthly rainfall; $P < 0.001$, $t = 4.3$, $N = 53$) and with predicted species richness ($P < 0.001$, $t = 5.4$, $N = 53$). In both regression analyses, we included the size of the area as a variable to correct for possible effects of differences in pixel number on which the correlation coefficient was based. Rainfall seasonality and predicted species richness were not significantly correlated ($r = 0.20$, $P = 0.12$, $t = 1.55$). Data were obtained from Kooistra *et al.* [39].

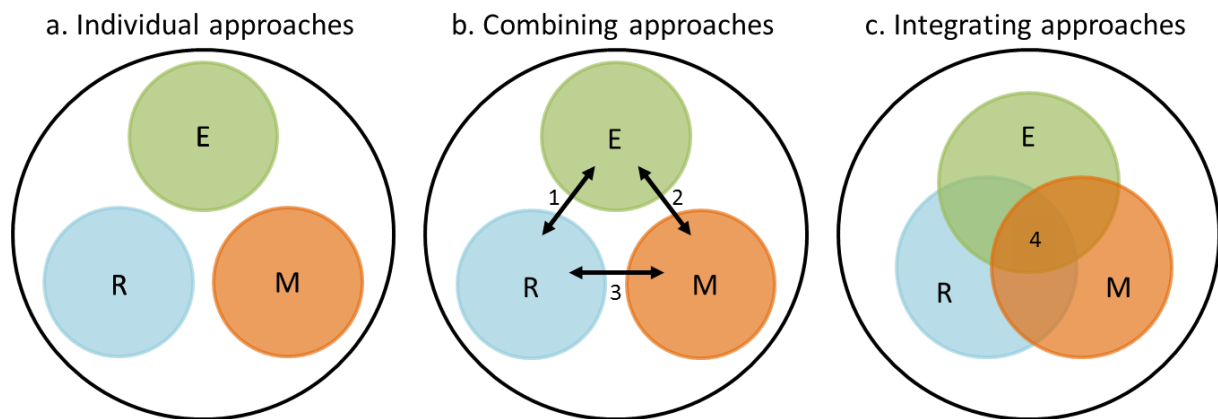


Figure 3: Representation of the differences between a) using individual approaches (E and green: empirical; R and blue: remote sensing; M and orange: ecosystem modelling) to compare results, b) combining approaches (e.g., validations and spatial and temporal upscaling) and c) integrating approaches through the use of ‘boundary objects’, for example by using diversity in leaf traits or indices of forest structure, which can be measured in empirical field studies, scale up over larger areas using remote sensing and included in modelling studies. Possible combinations are: empirical and remote sensing approaches to scale the biodiversity-carbon relationship to broader spatial scales (Arrow 1), empirical and modelling approaches to scale this relationship to larger temporal scales (Arrow 2) and remote sensing and modelling approaches for further validation and improvements of the approaches (Arrow 3). Integrating approaches seeks for boundary objects, i.e. indicators that can be quantified by each approach (number 4 in the figure). For examples of boundary objects, see Table 2.

317 **References:**

- 1
2 318
3
4 319 1. Bonan GB: **Forests and climate change: forcings, feedbacks, and the climate benefits of**
5
6 320 **forests.** *Science*. 2008, **320**:1444–1449.
7
8
9 321 2. Slik JWF, Arroyo-Rodríguez V, Aiba S-I, Alvarez-Loayza P, Alves LF, Ashton P, Balvanera P,
10
11 322 Bastian ML, Bellingham PJ, van den Berg E, et al.: **An estimate of the number of tropical**
12
13 **tree species.** *Proc. Natl. Acad. Sci.* 2015, **111**:7472–7477.
14 323
15
16
17 324 3. Mace GM, Norris K, Fitter AH: **Biodiversity and ecosystem services: a multilayered**
18
19 325 **relationship.** *Trends Ecol. Evol.* 2012, **27**:19–25.
20
21
22 326 4. Tilman D, Reich PB, Knops J, Wedin D, Mielke T, Lehman C: **Diversity and productivity in**
23
24 **a long-term grassland experiment.** *Science*. 2001, **294**:843–5.
25 327
26
27
28 328 5. Morin X, Fahse L, Scherer-Lorenzen M, Bugmann H: **Tree species richness promotes**
29
30 329 **productivity in temperate forests through strong complementarity between species.** *Ecol.*
31
32 330 *Lett.* 2011, **14**:1211–1219.
33
34
35 331 6.* Poorter L, van der Sande MT, Thompson J, Arets EJMM, Alarcón A, Álvarez-Sánchez J,
36
37 332 Ascarrunz N, Balvanera P, Barajas-Guzmán G, Boit A, et al.: **Diversity enhances carbon**
38
39 **storage in tropical forests.** *Glob. Ecol. Biogeogr.* 2015, **24**:1314–1328.
40 333
41
42
43 334 The authors evaluate the empirical relationship between species richness and biomass storage
44
45 335 across Neotropical forests. The study demonstrates that species richness has a positive effect on
46
47 336 biomass storage which is independent of the effects of environmental conditions.
48
49
50
51 337 7. Hooper DU, Chapin III FS, Ewel JJ, Hector A, Inchausti P, Lavorel S, Lawton JH, Lodge DM,
52
53 338 Loreau M, Naeem S, et al.: **Effects of biodiversity on ecosystem functioning: a consensus of**
54
55 339 **current knowledge.** *Ecol. Monogr.* 2005, **75**:3–35.
56
57
58 340 8. Prado-Junior JA, Schiavini I, Vale VS, Arantes CS, van der Sande MT, Lohbeck M, Poorter L:
59
60 341 **Conservative species drive biomass productivity in tropical dry forests.** *J. Ecol.* 2016,
61
62
63
64
65

342 **104:817–827.**

1
2
3 343 9. Ruiz-Jaen MC, Potvin C: **Tree diversity explains variation in ecosystem function in a**
4
5 344 **neotropical forest in Panama.** *Biotropica* 2010, **42:638–646.**

6
7
8 345 10.* Chisholm RA, Muller-Landau HC, Abdul Rahman K, Bebbber DP, Bin Y, Bohlman SA, Bourg
9
10 346 NA, Brinks J, Bunyavejchewin S, Butt N, et al.: **Scale-dependent relationships between tree**
11
12 347 **species richness and ecosystem function in forests.** *J. Ecol.* 2013, **101:1214–1224.**

13
14
15
16 348 This paper evaluates the effect of species richness on biomass storage and productivity in
17
18 349 tropical forests across the globe. The authors find that species richness effects are positive and
19
20 350 strongest at small grain sizes (0.04 ha) and weaker at larger grain sizes (0.25 ha).

21
22
23 351 11. van der Sande MT, Poorter L, Kooistra L, Balvanera P, Thonicke K, Thompson J, Arets
24
25 352 EJMM, Garcia Alaniz N, Jones L, Mora F, et al.: **Biodiversity enhances climate change**
26
27 353 **mitigation by tropical forests.** *In revision.*

28
29
30
31 354 12. Balvanera P, Aguirre E: **Tree diversity, environmental heterogeneity, and productivity in a**
32
33 355 **Mexican tropical dry forest.** *Biotropica* 2006, **38:479–491.**

34
35
36 356 13. Becknell JM, Powers JS: **Stand age and soils as drivers of plant functional traits and**
37
38 357 **aboveground biomass in secondary tropical dry forest.** *Can. J. For. Res.* 2014, **613:604–**
39
40 358 **613.**

41
42
43
44 359 14. Tilman D: **The ecological consequences of changes in biodiversity: a search for general**
45
46 360 **principles.** *Ecology* 1999, **80:1455–1474.**

47
48
49 361 15. Grime JP: **Benefits of plant diversity to ecosystems: immediate, filter and founder effects.**
50
51 362 *J. Ecol.* 1998, **86:902–910.**

52
53
54 363 16. De Sy V, Herold M, Achard F, Asner GP, Held A, Kelldorfer J, Verbesselt J: **Synergies of**
55
56 364 **multiple remote sensing data sources for REDD+ monitoring.** *Curr. Opin. Environ. Sustain.*
57
58 365 *2012, 4:696–706.*

- 366 17. Duro DC, Coops NC, Wulder MA, Han T: **Development of a large area biodiversity**
1
2 367 **monitoring system driven by remote sensing.** *Prog. Phys. Geogr.* 2007, **31**:235–260.
3
4
- 5 368 18. Kuenzer C, Ottinger M, Wegmann M, Guo H, Wang C, Zhang J, Dech S, Wikelski M: **Earth**
6
7 369 **observation satellite sensors for biodiversity monitoring: potentials and bottlenecks.** *Int. J.*
8
9 370 *Remote Sens.* 2014, **35**:6599–6647.
10
11
- 12 371 19.* Saatchi SS, Harris NL, Brown S, Lefsky M, Mitchard ETA, Salas W, Zutta BR, Buermann W,
13
14 372 Lewis SL, Hagen S, et al.: **Benchmark map of forest carbon stocks in tropical regions**
15
16 373 **across three continents.** *Proc. Natl. Acad. Sci. U. S. A.* 2011, **108**:9899–904.
17
18
19
20 374 A "benchmark" map of biomass carbon stocks over 2.5 billion ha of forests on three continents
21
22 375 is presented encompassing all tropical forests, for the early 2000s, which will be invaluable for
23
24 376 REDD assessments at both project and national scales.
25
26
- 27 377 20. Baccini A, Goetz SJ, Walker WS, Laporte NT, Sun M, Sulla-Menashe D, Hackler J, Beck
28
29 378 PSA, Dubayah R, Friedl MA, et al.: **Estimated carbon dioxide emissions from tropical**
30
31 379 **deforestation improved by carbon-density maps.** *Nat. Clim. Chang.* 2012, **2**:182–185.
32
33
34
- 35 380 21.* Avitabile V, Herold M, Heuvelink GBM, Lewis SL, Phillips OL, Asner GP, Armston J, Asthon
36
37 381 P, Banin LF, Bayol N, et al.: **An integrated pan-tropical biomass map using multiple**
38
39 382 **reference datasets.** *Glob. Chang. Biol.* 2016, **22**:1406–1420.
40
41
42
43 383 A scale-independent fusion method is presented which provides improved forest biomass
44
45 384 estimates by integrating existing regional biomass maps and country-specific reference datasets
46
47 385 as input maps.
48
49
- 50 386 22. Asner GP: **Organismic remote sensing for tropical forest ecology and conservation.** *Ann.*
51
52 387 *Missouri Bot. Gard.* 2015, **100**:127–140.
53
54
55
- 56 388 23. Skidmore AK, Pettorelli N, Coops NC, Geller GN, Hansen M, Lucas R, Múcher CA,
57
58 389 O'Connor B, Paganini M, Miguel Pereira H, et al.: **Agree on biodiversity metrics to track**
59
60 390 **from space.** *Nature* 2015, **523**:403–405.
61
62
63
64
65

- 391 24. Pichancourt J-B, Firn J, Chadès I, Martin TG: **Growing biodiverse carbon-rich forests.** *Glob.*
1
2 392 *Chang. Biol.* 2014, **20**:382–93.
3
4
- 5 393 25. Chitale VS, Tripathi P, Behera MD, Behera SK, Tuli R: **On the relationships among**
6
7 394 **diversity, productivity and climate from an Indian tropical ecosystem: A preliminary**
8
9 395 **investigation.** *Biodivers. Conserv.* 2012, **21**:1177–1197.
10
11
- 12 396 26.** Sakschewski B, von Bloh W, Boit A, Poorter L, Peña-Claros M, Heinke J, Joshi J, Thonicke K:
13
14 **Resilience of Amazon forests emerges from plant trait diversity.** *Nat. Clim. Chang.* 2016,
15 397 **6**:1032–1036.
16
17 398
18
19
- 20 399 This paper uses a global dynamic vegetation model that can simulate tropical forests with high
21
22 400 levels of functional diversity, and shows that trait diversity increases forest recovery capacity
23
24 401 after climate change.
25
26
- 27 402 27. Sitch S, Huntingford C, Gedney N, Levy PE, Lomas M, Piao SL, Betts R, Ciais P, Cox P,
28
29 Friedlingstein P, et al.: **Evaluation of the terrestrial carbon cycle, future plant geography**
30 403 **and climate-carbon cycle feedbacks using five Dynamic Global Vegetation Models**
31
32 404 **(DGVMs).** *Glob. Chang. Biol.* 2008, **14**:2015–2039,.
33
34 405
35
36
- 37 406 28.** Fyllas NM, Gloor E, Mercado LM, Sitch S, Quesada C a., Domingues TF, Galbraith DR,
38
39 Torre-Lezama a., Vilanova E, Ramírez-Angulo H, et al.: **Analysing Amazonian forest**
40 407 **productivity using a new individual and trait-based model (TFS v.1).** *Geosci. Model Dev.*
41
42 408 *Discuss.* 2014, **7**:1413–1452.
43
44 409
45
46
- 47 410 The authors develop a model to simulate tropical forest productivity by using empirically
48
49 411 measured species composition and functional diversity as measured.
50
51
- 52 412 29. Díaz S, Hector A, Wardle DA: **Biodiversity in forest carbon sequestration initiatives: not**
53
54 **just a side benefit.** *Curr. Opin. Environ. Sustain.* 2009, **1**:55–60.
55 413
56
57
- 58 414 30. Forkel M, Migliavacca M, Thonicke K, Reichstein M, Schaphoff S, Weber U, Carvalhais N:
59
60 415 **Codominant water control on global interannual variability and trends in land surface**
61
62
63
64
65

416 **phenology and greenness.** *Glob. Chang. Biol.* 2015, **21**:3414–3435.

1
2
3 417 31. Forkel M, Carvalhais N, Schaphoff S, Bloh W V., Migliavacca M, Thurner M, Thonicke K:

4
5 418 **Identifying environmental controls on vegetation greenness phenology through model-**
6
7 419 **data integration.** *Biogeosciences* 2014, **11**:7025–7050.

8
9
10 420 32.* Star SL, Griesemer JR: **Institutional ecology, “translations” and boundary objects:**

11
12 421 **amateurs and professionals in Berkeley s museum of vertebrate zoology, 1907-39.** *Soc.*

13 422 *Stud. Sci.* 1989, **19**:387–420.

14
15
16
17
18 423 This paper develops the idea of 'boundary objects' to make actors with different viewpoints and

19
20 424 backgrounds succesfully colaborate in scientific work.

21
22
23 425 33. Abson DJ, von Wehrden H, Baumgärtner S, Fischer J, Hanspach J, Härdtle W, Heinrichs H,

24
25 426 Klein AM, Lang DJ, Martens P, et al.: **Ecosystem services as a boundary object for**

26
27 427 **sustainability.** *Ecol. Econ.* 2014, **103**:29–37.

28
29
30
31 428 34. Brand FS, Jax K: **Focusing the meaning(s) of resilience: resilience as a descriptive concept**

32
33 429 **and a boundary object.** *Ecol. Soc.* 2007, **12**:23.

34
35
36 430 35.** Asner GP, Martin RE, Tupayachi R, Emerson R, Martinez P, Sinca F, Powell GVN, Wright SJ,

37
38 431 Lugo AE: **Taxonomy and remote sensing of leaf mass per area (LMA) in humid tropical**

39
40 432 **forests.** *Ecol. Appl.* 2011, **21**:85–98.

41
42
43
44 433 This paper shows that estimations of leaf mass per area (LMA) from spectroscopic remote

45
46 434 sensing provides comparable results as empirical LMA measurements of the top canopy.

47
48
49 435 36.** Asner GP, Anderson CB, Martin RE, Tupayachi R, Knapp DE, Sinca F: **Landscape**

50
51 436 **biogeochemistry reflected in shifting distributions of chemical traits in the Amazon forest**

52
53 437 **canopy.** *Nat. Geosci.* 2015, **8**:567–573.

54
55
56
57 438 This paper shows that multiple leaf traits at the community-level (i.e., community-average trait

58
59 439 values and spatial variation) can be estimated using remote sensing techniques.

- 440 37. Pérez-Harguindeguy N, Díaz S, Garnier E, Lavorel S, Poorter H, Jaureguiberry P, Cornwell
1 WK, Craine JM, Gurvich DE, Urcelay C, et al.: **New handbook for standardised**
2 441 **measurement of plant functional traits worldwide.** *Aust. Journal Bot.* 2013, **61**:167–234.
3
4 442
5
6
7 443 38. Scheiter S, Langan L, Higgins SI: **Next-generation dynamic global vegetation models:**
8
9 444 **learning from community ecology.** *New Phytol.* 2013, **198**:957–69.
10
11
12 445 39. Kooistra L, Dutrieux L, Equihua J, Kolb M, Roerink G, Gerard F, Peña Claros M: **Current**
13 **contributions of biodiversity and ecosystems to climate change mitigation - an analysis**
14 **using remote sensing datasets.** *Report ROBIN project D113.* 2015.
15 446
16
17 447
18
19
20 448

21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65