



Swansea University
Prifysgol Abertawe



Cronfa - Swansea University Open Access Repository

This is an author produced version of a paper published in :
Philosophical Transactions of the Royal Society B: Biological Sciences

Cronfa URL for this paper:
<http://cronfa.swan.ac.uk/Record/cronfa33127>

Paper:

Doerr, S. & Santín, C. (2016). Global trends in wildfire and its impacts: perceptions versus realities in a changing world. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 371(1696), 20150345
<http://dx.doi.org/10.1098/rstb.2015.0345>

This article is brought to you by Swansea University. Any person downloading material is agreeing to abide by the terms of the repository licence. Authors are personally responsible for adhering to publisher restrictions or conditions. When uploading content they are required to comply with their publisher agreement and the SHERPA RoMEO database to judge whether or not it is copyright safe to add this version of the paper to this repository.
<http://www.swansea.ac.uk/iss/researchsupport/cronfa-support/>

Global trends in wildfire and its impacts: perceptions and realities in a changing world

Stefan H. Doerr^{1*} & Cristina Santín²

Geography Department, Swansea University, Singleton Park, Swansea SA28PP, United Kingdom
(1) <http://orcid.org/0000-0002-8700-9002>, (2) <http://orcid.org/0000-0001-9901-2658>

Keywords: fire occurrence, area burned, fire severity, media, risk, costs, loss of lives, health, climate change

Summary

Wildfire has been an important process affecting Earth's surface and atmosphere for over 350 Mill. years and human societies have co-existed with fire since their emergence. Yet many consider wildfire as an accelerating problem with widely held perceptions, both in the media and scientific papers, of increasing fire occurrence, severity and resulting losses. However, important exceptions aside, the quantitative evidence available does not support these perceived overall trends. Instead, global area burned appears to have overall declined over the last decades and there is increasing evidence that there is less fire in the global landscape today than centuries ago. Regarding fire severity, limited data are available. For the western USA they indicate overall little changes over the last decades, but also that area burned at high severity has overall declined compared to pre-European settlement. Direct fatalities from fire and economic losses also show no clear trends over the last three decades. Trends in indirect impacts, such as health problems from smoke or disruption to social functioning remain insufficiently quantified to be examined. Global predictions for increased fire under a warming climate highlight the already urgent need for a more sustainable co-existence with fire. The data evaluation presented here aims to contribute to this by reducing misconceptions and facilitating a more informed understanding of the realities of global fire.

1. Introduction

Fire has been an important factor in the dynamics of Earth's climate and in the development of biomes since its widespread occurrence began 400-350 Mill. years ago [1,2]. In fire-prone ecosystems, humans have always co-existed with fire in the landscape, and its use can be seen as the first anthropogenic tool that has affected ecosystem dynamics beyond the very local scale [3]. Whether as open biomass burning or as the relatively recent practice of combusting fossil fuels in engines and power stations, fire has been a key factor in the rise of human societies [4,5]. Yet, over the last couple of centuries the traditional European perception of fire has been implemented in many parts of the world (see Box 1), and fire in the landscape (commonly termed wildfire, wildland fire or landscape fire) has been typically considered as 'bad' and our focus on the whole has been on eliminating or at least containing it [6-8]. The 'command and control' attitude of most Western societies neglects the fundamental role that fire has in

*Author for correspondence (s.doerr@swan.ac.uk)

1
2
3
4 sustaining biodiversity and ecosystem health [9,10].

5
6 The media still promote perceptions of wildfire as the enemy even in very fire-prone regions,
7 such as the Western USA or Eastern Australia where managers are attempting to move away
8 from aggressive suppression policies and residents are slowly assimilating the concept of fire as
9 an ecological factor [10–12]. Whilst the vast majority of 30–46 Mill. km² of the global land
10 surface burned per year (~4% the global land surface) [13] have little direct impact on
11 individuals and therefore do not attract wider attention, the media tend to report on the costly
12 and sometimes tragic impacts of some wildfires, with a focus on the fate of individuals [12,14].
13 This is not surprising given the fundamental risk some specific fires pose for human lives,
14 infrastructures and the value of commodities such as forest plantations, yet this type of media
15 coverage can be a barrier to expand the notion of our need of learning to coexist with fire
16 [15,16]. Numerous reports, ranging from popular media through to peer-reviewed scientific
17 literature, have led to a common perception that fires have increased or worsened in recent years
18 around the world (e.g. [10,17–20]. Where these reports are accompanied by quantitative
19 observations, they are often based on short timescales and regional data for fire incidence or area
20 burned, which do not necessarily reflect broader temporal or spatial realities.
21
22

23
24 Unlike other natural hazards such as earthquakes or volcanic eruptions, fire is perceived as an
25 avoidable risk and enormous resources are directed towards fire suppression efforts particularly
26 in the more developed world [21]. Yet the now widely acknowledged consequence that fire
27 suppression often comes at the cost of an increased risk of more severe or extensive future fire
28 within fire-prone landscapes [22] has to date only led to limited changes to fire suppression
29 practice in most regions [10].
30

31 The aim of this paper is to illuminate the discrepancies between the perceptions about global fire
32 against the quantitative realities that have emerged through research on landscape fire
33 occurrence and its impacts on society as a whole. Achieving a more balanced and realistic
34 perspective about fire occurrence, its risks and impacts amongst fire specialists, decisions
35 makers and the wider public is perhaps the most critical step towards regaining a more
36 sustainable coexistence with landscape fires.
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Box 1: A Western-biased perception of fire

In this paper we discuss widely held perceptions of fire and compare them with fire data and statistics available to date. We also highlight that our scientific knowledge and social perceptions are Western-biased because most available data is derived from Western societies in fire-prone countries such as USA, Australia and the Mediterranean Europe. In these countries, current policies and social perceptions share a common starting point: the German forestry school of the 19th Century, which spread the systematic protection of forests against fire across the Old Continent and former colonies [23–25]. This 100% fire exclusion policy has long proven to be impractical unsustainable and ecologically detrimental in fire-prone regions [21,26]. Although fire management is now slowly changing, with prescribed burning also being increasingly used, policies of aggressive wildfire suppression still apply almost everywhere [10,21,24]. For example, in the USA only 0.4% of wildfires, whether ignited by lightning or humans, are allowed to burn [10]. All others are actively suppressed. Regarding social perceptions, it is important to stress that, in many of these regions, intentional burning had been used for very long both by native people and settlers, thus, in rural areas fire was understood as part of the landscape management culture [27]. However, the current general public perception is predominantly different. Until very recently, governments refused to present fire as a potential positive ecological factor out of concern that any admission of a positive role for fire would sound contradicting [21]. Smokey the bear in the United States is the best, but not the only, example of effective public awareness campaigns supporting 100% fire suppression (Fig. a). Nowadays, the perception of fire in Western communities living in high fire risk areas is slowly moving towards the recognition of fire as a valuable natural factor [28], however, in many other regions fire is still perceived by the whole society as a natural hazard with only negative implications. This Western perception of fire currently dominates the world and is thus the focus of this paper. It is, however, not the only one. In this same issue other contributions discuss societies which have long co-existed with fire and remain doing so sustainably, such as the aboriginal people of the Western Dessert of Australia [29] or indigenous communities in Venezuela, Brazil and Guyana [30].



Fig. a. Public awareness campaigns supporting total exclusion of fire from our forests have driven largely our current perceptions of fire. E.g. Left: Smokey the bear has been the American champion against fire since the 1950s. Right: The “All against fire” campaign in Spain during the late 80s and early 90s also had wide national relevance.

2. Has fire increased in many regions around the globe?

Analysis of charcoal records in sediments [31] and isotope-ratio records in ice cores [32] suggest that global biomass burning during the past century has been lower than at any time in the last 2000 years. Although the magnitude of the actual differences between pre-industrial and current biomass burning rates may not be as pronounced as suggested by those studies [33], modelling approaches agree with a general decrease of global fire activity at least in the last centuries [34].

1
2
3
4 In spite of this, fire is often being quoted as an increasing issue around the globe (e.g. [10,17–
5 20]). One reason for this apparent contradiction may be that the global extent of fire is not
6 necessarily correlated to impacts on human society as explored in the subsequent section.
7 Another reason may be that our wider perception of fire is shaped by some widely publicised
8 regional trends and a lack of discrimination between reported fire activity parameters. An
9 important distinction regarding the latter is that between area burned (i.e. total ha or km²) and
10 fire occurrence (i.e. the number of fires for a given area and period). Recent trends in area
11 burned can now be derived from satellite observations and national records with reasonable
12 accuracy at regional and global scales [35,36]. Trends in occurrence, however, are less reliable
13 as recording efforts and methods vary between administrative regions. A striking example where
14 the lack of discrimination has led to contrasting perceptions is that of fire occurrence and
15 associated area burned in the Mediterranean region in the last three decades (Fig. 1). There was
16 indeed an increase in the number of fires from the early 1980's to the late 1990's. However, the
17 last three decades have been characterised by an overall decrease in area burned, and also a
18 decrease in the number of fires from mid-2000 (Fig. 1). This is often not recognised even within
19 the scientific community, with some authors continuing to underpin the importance of their fire-
20 related research with an increase of fire in this region [6,37].
21
22

23
24 Area burned is perhaps the most commonly used parameter when fire trends are being examined.
25 It is a relatively simple and globally relevant parameter and it underpins estimations for carbon
26 emissions by wildfire [13]. A summary of global trends in area burned during the 20th century is
27 given in Flannigan et al. [38]. During the first half century the global average area burned
28 decreased somewhat by about 7% [39]. This was largely attributed to human factors, such as
29 increased fire prevention, detection and fire fighting efficiency, abandonment of slash-and-burn
30 cultivation in some areas, and permanent agricultural practice in others. During the second half
31 of the past century, this trend reportedly reversed with a 10% increase in global area burned.
32 However, this trend was not reflected everywhere and there are regional variations and
33 substantial uncertainties [38]. Overall this increase in the latter half of the last century has been
34 attributed to land management changes including increases in deforestation fires in the tropics
35 [39], but it may also partially reflect a 'return' to a more 'normal' fire regime in areas where fire
36 had been suppressed [38].
37
38

39
40 The availability of satellite data now allows a more consistent evaluation of temporal patterns in
41 area burned. Thus, from an analysis based on MODIS burned area maps between 1996-2012,
42 Giglio et al. [35] present some rather notable outcomes. In contrast to what is widely perceived,
43 the detected global area burned has actually decreased slightly over this period (by 1 % yr⁻¹). A
44 more recent global analysis by van Lierop et al. [36], based primarily on nationally reported fire
45 data supplemented by burned area estimates from satellite observations, shows an overall decline
46 in global area burned of 2 % yr⁻¹ for the period of 2003-2012.
47

48
49 At coarse regional scales, overall trends for the period 1996-2012 are rather contrasting [35]. For
50 example, data for Europe and Australia/New Zealand show a strong decline in area burned of 5
51 % yr⁻¹, despite the latter region experiencing the largest annual area burned in the final year of
52 the observation period. In contrast, for SE-Asia, The Middle East and Boreal North America the
53 estimated area burned increased by 3-4 %. For Temperate North America the very small increase
54 in area burned (0.1 % yr⁻¹) estimated by Giglio et al. [35] over this period may seem surprising
55 when compared to the widely reported increase in area burned for the USA [40] and particularly
56 the western USA in recent decades [41–44]. This discrepancy may be at least in part be due to (i)
57 the region used in Giglio et al.'s analysis excludes the Boreal and drier south-eastern zones of
58 the USA and (ii) area burned in the studies focused on the USA [40–44] is based on national and
59 regional fire statistics produced using a variety of methods. These need to be viewed with some
60 caution when examining trends as they have undergone changes in annual reporting methods and

1
2
3
4 biases over time [45]. Indeed, according to national statistics for the USA, whilst area burned by
5 prescribed fire has changed little overall since reporting began in 1998 (10-year average: 8,853
6 km²), area burned by wildfires has seen an overall strong trend of increase by over 5 % yr⁻¹ over
7 the period 1991 to 2015, with 2015 exceeding 40,000 km² burned for the first time during the
8 last 25 years (Fig. 2). This increase has been accompanied by an overall decline in the number of
9 fires (Fig. 2). This suggests a general trend of fewer, but larger wildfires, which is also
10 highlighted for forests in the Western USA by Westerling for the period 1983-2012 [44].
11 However, caution is advised when considering the relative rates of change for area burned. The
12 comparatively brief periods of observation discussed here are strongly influenced by regional
13 inter-annual variability and are too short to be indicative of longer-term trends. For example, if
14 only the last 16 full reporting years for the USA are considered (2000-2015), where annual area
15 burned ranged between 14,284 km² (2001) and 40,975 km² (2015), the overall annual increase
16 has been < 1 % [46]. Longer-term records can indeed reveal rather different perspectives. For
17 example, for the Californian Cascades and Sierra Nevada, Mallek et al. [47] suggest that
18 'modern' (1984-2009) annual area burned was only 14% of that burned annually prior to
19 European settlement (~1500-1850). In addition to climate, changes in vegetation patterns and fire
20 regimes also play an important role here and are discussed in the context of fire severity in the
21 following section.
22
23

24
25 Thus, whilst there are clearly some noteworthy trends in area burned for specific recent periods
26 and regions, the general perception of increasing fire around the world is not supported by the
27 data available to date. This does not withstand the observation of increasing fire season length in
28 some areas [48], which is an important contributor to the increase in area burned during this
29 century in the Northwestern USA [41,44], boreal Canada and Alaska [49,50]. A future
30 lengthening of the fire season is also anticipated for other many regions of the globe, with a
31 potential associated increase of fire activity [9,51-54]. It is, however, important to recognise that
32 in addition to direct climatic factors other factors, such as fuel availability and human influence,
33 will also strongly affect future fire activity [55,56].
34
35

36 Thus, the widespread use of limited datasets or excessive extrapolation of short-term regional
37 trends may go some way in explaining the widely held view of generally increasing fire around
38 the world. The wider impacts of fire examined in the following section, however, may be even
39 more relevant in driving the overall perceptions of fire trends.
40
41

42 3. Have fire impacts increased in many regions around the globe?

43 a) Fire intensity and severity

44 Whilst the trends in area burned explored above have implications for the effects of fire on
45 global carbon emissions, ecosystems and society, the spatial extent of burning is not always
46 closely linked to the impacts of a fire. From a perspective of fire ecology or risk to
47 infrastructures, the intensity of a fire (i.e. its rate of energy output), its severity (its ecosystem
48 impacts) and its spatial patterns (degree of patchiness) may be more important than the total area
49 burned. For example, the degree of vegetation consumption, the depth of burning into the
50 organic and mineral soil, and the proximity of areas less- or not affected by fire are important in
51
52
53
54
55
56
57
58
59
60

1
2
3
4 determining the length of time for a burned area to ‘recover¹’ [3,57–59]. The notion that fire
5 intensity and severity have increased in recent years pervades media reports and some of the
6 literature [10,60–62]. Whether or not this is the case is not easy to ascertain given that these
7 parameters and associated trends are much more difficult to determine compared to area burned.
8 All else being equal, fire intensity can indeed be expected to increase with air temperature [63],
9 and it can be deduced that areas that are experiencing higher atmospheric temperatures in the fire
10 season associated with global warming would experience more intense fires. For example, the
11 catastrophic 2009 Black Saturday fires of Victoria (Australia) were reportedly associated,
12 amongst other factors, with unprecedented high atmospheric temperatures (since measurements
13 began) and fire intensity [64]. Whether or not this extreme event signifies a trend or may simply
14 be the result of longer-term natural variability in fire behaviour remains open. Indeed, it has
15 subsequently been suggested that the fire weather potential witnessed during Black Saturday and
16 the associated level of fire intensity was not unprecedented in south-eastern Australia [65].
17
18

19 Few studies exist that have explicitly examined trends in fire severity. These have focused on the
20 western USA, an area where there are particular concerns about increased fire activity [40,66].
21 Examining trends from 1984-2006 for large ecoregions in the North- and Southwest USA,
22 Dillon et al. [67] found no significant increase in the proportion of annual area burned at high
23 severity for five of the six regions considered, with the Southern Rockies being the exception.
24 For the Sierra Nevada region (California), which was not covered the previous study [67],
25 Hanson and Odion [68,69] found no general increase in fire severity within the period of 1984-
26 2010. However, considering ten national forests in California for the same period, Miller and
27 Safford [70] found a significant increase in burn severity for yellow pine-mixed conifer forests.
28 They attribute this largely to decades of fire suppression and other management practices rather
29 than climate, which have led to major changes in forest composition and structure, and increases
30 in density and fuel-loading, and hence fire behavior. Covering the much larger area of the dry
31 forest landscapes of the western USA, including large parts of those examined in the
32 aforementioned studies, Baker [71] found that the rate of high-severity fire in the period 1984-
33 2012 was within or below that of historical century- to millennial-scale estimates.
34
35

36
37 Thus, whilst there is evidence of a recent increase in proportional fire severity for a specific
38 forest type in California, these independent studies do not support the notion of an overall
39 increase in fire severity over the last few decades in the fire-adapted forested landscapes in the
40 western USA. Indeed a longer term perspective focused on the Californian Sierra Nevada and
41 Cascades by Mallek et al. [47] suggests that the annual area burned at high severity between
42 1984-2009 was only half that prior to European settlement (~1500-1850), associated with an
43 overall smaller area burned compared to pre-European times. Whether or not the overall lack of
44 change in burn severity applies also to other regions where perceptions of increases in fire
45 severity exist too have to remain open until robust data emerge to test this notion.
46
47
48
49
50

51
52
53
54
55
56
57 ¹ The concept of the ‘post-fire recovery window’ or ‘window of disturbance’ can be viewed as the time it takes for
58 ecosystem properties such as biomass, biodiversity, soil characteristics or the hydrological balance to return to a
59 pre-fire status [93]. This assumes that fire is an episodic or even rare disturbance event. A more appropriate view in
60 fire-adapted ecosystems is that fire is a natural process that is part of a natural cycle between fire and a post-fire
recovery conditions with varying recurrence [94].

b) Impacts on society: direct effects on people

Whilst the ecological impacts of fire or their interactions with climate are of concern to scientists, natural resource managers, policy makers and the public, policy and public perception regarding fire in the landscape is primarily shaped by the impacts of fire on people and society as a whole (see Box 2). Lives lost, together with direct damage to homes and other infrastructures create wide media attention and are probably of greatest importance here. For example the Black Saturday fires of 2009, in which 173 people lost their lives shook Australian society and led to major reconsideration of landscape fire related policy [64]. These and other tragic losses to lives from fire may or may not have been preventable, but should be also seen in perspective to other risks to lives. When considering some of the extreme landscape fires as a form of natural disaster, the number of deaths is actually relatively low compared to other natural disaster types. For example, data by the Emergency Events Database (EM-DAT)² suggest that over the period 1901-2014 3,753 people have been killed by wildfire, compared to over 2.5 Million from earthquakes and nearly 7 Million from floods [72]. These figures are likely to be inaccurate and substantial underestimations of direct deaths from fire. For example the EM-DAT reports 21, 35 and 17 deaths for 2012, 2013 and 2014, whereas data collected for recent years by the Global Fire Monitoring Centre report 215, 209 and 217 fatalities from landscape fires for the same years [73]. Considering that ca 4% of the global land vegetated land surface burns every year, annual direct deaths, whether they number in the 10's or the low 100's, indicate a comparatively low risk of death as a direct result of fire compared to that from other natural disaster types (Table 1).

It is also worth noting that many of the deaths recorded as a result of landscape fires have indirect 'medical' or operational causes. For example of the 26 total landscape fire deaths recorded in the USA in 1999 [74], only one was a direct fire death (burnover), nine were due to heart attack and other causes included crushing by engines and electrocution. Unsurprisingly, fire fighters are at greatest risk from fires, particularly in regions where fire suppression involves the use of personnel on the ground in topographically complex terrain. The death of 19 wildland firefighters in Arizona in 2013, who became entrapped in steep terrain under changing fire behaviour [75], serve as a recent tragic example. Data from the USA show a total of 338 firefighter fatalities between 1977 and 2006 [76]. Additional deaths occur in training, road and aircraft accidents. Amongst these there is no clear temporal trends in wildland fire deaths, except when considering those from aircraft crashes, which have risen probably due to the increased use of aircraft in wildland firefighting over this period [76]. A study examining all recorded wildland fire fatalities in Spain between 1980 and 2010 reported 241 fatalities of which 169 were firefighters and with no increasing or decreasing temporal trend [77]. Considering the reported global direct death toll from landscape fire 'disasters' between 1977 through to 2014, no clear trend emerges either with large fluctuations between years ranging from 0 in 1990 to a maximum of 266 in 1997 [72].

² EM-DAT is a global database on natural and technological disasters which fulfil one or more of the following four criteria: (i) 10 or more people dead, (ii) 100 or more people affected, (iii) declaration of a state of emergency, (iv) call for international assistance [72]. It therefore excludes some landscape fire events where fatalities have occurred or less than 100 people have been affected. Lives lost and economic damage based on EM-DAT reported here are therefore likely to be an underrepresentation of actual global values.

Box 2: Good fire, bad fire?

Fire has long been a natural factor in many ecosystems around the world, from boreal forests to tropical savannas [95,96]. In these systems fire is a necessary perturbation to preserve ecosystem health and stimulate rejuvenation [97,98]. Each ecosystem is adapted to a specific fire regime (i.e. fire type and recurrence), which could be understood as “good fire”. However, when the fire regime moves away from the established one (e.g. due to human influence), ecosystem resilience to fire may be surpassed [98]. The resulting long-lasting damage to the ecosystem would thus be caused by “bad fire”. From an ecosystem perspective it is therefore relatively easy to distinguish between “good” and “bad fire”. Although this is an oversimplification as ecosystems are dynamic entities, which evolve and change also without human influence [94]. Notwithstanding this, a more complicated picture arises when considering the human perspective. An ecologically “good” stand-replacing fire in a fire-dependent forest, essential for forest regeneration, will be viewed as a “bad fire” when it results in losses of homes or lives, or perhaps even by it resulting, in the short-term, in a black and desolate landscape. Equally, an ecologically “bad” fire in a heathland, occurring too soon after the last one for full ecosystem recovery, can indeed be perceived as a “good” fire for the landowner whose intention is to convert the heather into grass. Often a range of different perceptions comes into play, complicating even more the full picture, as highlighted in this issue by Davies et al. [99] in relation to the role of fire in U.K. peat- and moorland management. Prescribed burning there is strongly supported by land managers whereas opposition from the general public is a growing trend. Another example, which is of global concern, is the recurring problem of peat fires in Southeast Asia. These are a consequence of land use changes and have enormous impacts on air and water quality, human health, ecosystem resilience and the global carbon cycle [100]. In September 2015, Indonesia’s peat fires emitted carbon at a rate of 15-20 million tons per day, well over the daily carbon emissions of the whole American economy [101]. These human-caused tropical peat fires are amongst the few examples of unequivocally “bad fires”. In most cases, however, whether a fire is considered “good” or “bad” will depend on its context, which can be ecological, social, economic or a combination of all. It is the role of the scientific community to provide an objective basis for society to understand and judge the consequences of the choices we make in how we manage, modify and co-exist with fire.

c) Impacts on society: direct economic impacts

Human losses aside, the direct financial costs, such as the damage to homes and other infrastructures, often dominate the perception of the fire impacts and an increase in these is often highlighted in the media [78–80] or scientific papers and reports [81–83] (see also Box 2). The data on fire disasters with continuous annual records of economic damage (1987-2014; [72]) gives annual global values (adjusted to 2015 \$ value) ranging from \$4.6 Mill. to \$12,318 Mill. (annual average \$2,677 Mill.), and showing no apparent trend. These estimates of losses, however, only include damage to property, crops and livestock and do not reflect losses from fire events not classified as disasters (see footnote 2). Other important economic parameters not included here are the costs arising from human losses, injuries and longer-term health implications [84]. Furthermore, fire suppression costs are not considered in these figures. These can be very substantial (see Fig 2). For example, Greece, France, Italy, Portugal and Spain together invest €2,500 Mill. each year in fire management, with most of this budget dedicated to fire detection and suppression [6]. This is similar to the estimated global average annual losses from fires reported by EM-DAT for 1987 to 2004. Canada spends an average of \$US 531 Mill. annually on fire prevention and suppression (2000-2010; [85]). There is limited data available from most countries to examine any global temporal trends.

Fig. 2 shows suppression costs (adjusted to 2016 \$ rates) in relation to the number of fires and area burned for the USA during the last 25 years. Whilst the area burned has seen an overall

1
2
3
4 increase of $\sim 5\% \text{ yr}^{-1}$ (see also section 2), suppression costs have overall increased by $\sim 1.5 \times$ that
5 rate. It is not clear to what degree this trend is (i) representative of any trends elsewhere in the
6 world and (ii) has resulted in a concomitant reduction in the actual area burned. The fact the
7 period of 2000–2016 has seen an increase of $< 1\% \text{ yr}^{-1}$ in inflation adjusted suppression costs,
8 which is similar to the rate of increase in area burned over the same period, indicates that the
9 preceding period of a relative increase in resources allocated to suppression in the 1990s was
10 followed by a levelling off of suppression expenditure per unit area affected. That said, area
11 burned is perhaps not the most important factor to consider when examining suppression cost in
12 the USA. Of greater relevance may be the increasing population density and hence need for fire
13 suppression in the wildland-urban interface (WUI). For example, in the western US states of
14 California, Oregon and Washington, housing in the WUI comprised 61 % of all new homes built
15 during the 1990s, and 43 % of the total housing in the region [86]. Given that 2.9 Mill. American
16 homes are in areas with fire return intervals ≤ 100 years [86], an increase in suppression need
17 would be expected even if the area burned had remained unchanged. This may not only be one
18 of the reasons why the American continent is leading the global ‘league table’ (Table 2) in terms
19 of total economic damage over the period of 1984–2014. It will have also have resulted in more
20 people experiencing fire, which may be associated with greater media coverage of fire from
21 these areas.
22
23
24
25

26 d) Impacts on society: indirect impacts

27 In addition to direct impacts on people health and economic losses, fires also have other
28 substantial effects on society through indirect impacts. Post-fire environmental effects such as
29 accelerated flooding, soil erosion, mass movement and pollution of water bodies are amongst the
30 most costly impacts on society [3,58,59]. Other important indirect effects are the longer-term
31 health implications [84]. A notable example of this is how smoke from landscape fires has
32 historically, and is currently, contributing to premature deaths amongst the world population
33 [87]. Estimates for the period 1997–2006 suggest these to be in the region of 340,000 per year
34 [88]. These figures are orders of magnitudes greater than more direct deaths from fires (Section
35 3b). Other indirect social impacts include disruptions to social processes and functioning such as
36 disruptions to road and air traffic, and businesses closure during and immediately after the fire,
37 or even long-term reduction of tourism, aesthetic value of the landscape or home values [89].
38 Catastrophic fires can even change social dynamics and the way people interact with each other
39 and with the landscape [89]. Efforts are increasing to examine these indirect impacts more
40 closely as they are currently only poorly understood and quantified [89]. It is therefore not
41 possible here to explore any trends or their potential effects on people’s perceptions.
42
43
44
45
46

47 4. Synthesis and Conclusion

49 We have shown here that the widely-held perception of increasing fire and fire impacts at the
50 global and some regional scales is not well supported by the realities that the available data
51 show. We do not question that fire season length and area burned has increased in some regions
52 over the last decades, as documented for parts of North America, or that climate and land use
53 change could lead to major shifts in future fire with potential increases in area burned, severity
54 and impacts over large regions [9,48,51]. The data available to date, however, do not support a
55 general increase in area burned or in fire severity for many regions of the world. Indeed there is
56 increasing evidence suggesting that there is overall less fire in the landscape today than there has
57 been centuries ago [34,90] although the magnitude of this reduction still needs to be examined in
58 more detail [33].
59
60

1
2
3
4
5 The data evaluated here do not support either the perception of increasing direct losses from fire.
6 Over the last decades there is no clear trend of increasing direct losses such as losses of life or
7 infrastructures. Whilst any fire-related death can be seen as one too many, at least the risk of
8 direct death from fire for the population as a whole is low compared to other natural hazards.
9 From the data available for the USA covering the last 25 years, it is clear that suppression costs
10 have increased substantially, and during the 1990s at a greater rate than increases in area burned
11 (Fig. 2). This increased expenditure and effort in the USA will most likely have saved many
12 lives whilst it also led to the loss of others. Increases in suppression expenditure may, at least in
13 part, be driven by a concern of worsening fire situation. The media are dominated by reports
14 from fires where lives are lost or at risk and these are typically from fire-prone regions
15 exhibiting high population densities (Fig. 3). The increased population density in the WUI over
16 the past decades, for example, may itself have resulted in increased media reports. It is important
17 to highlight that there is likely to be a bias in reporting of losses for western countries given that
18 the largest number of people affected by fire and losses of life appears to be elsewhere (i.e. Asia,
19 see Table 2 and Box 1).
20
21

22 Perhaps rather than a ‘wildfire problem’ that has worsened globally in recent decades, the
23 negative, and sometimes tragic, consequences of fire themselves may be gaining wider public
24 attention and, therefore, recognition. The fact that nowadays the latest news reports about
25 disasters from around the world are readily available to large parts of the population may be a
26 contributing factor. What is not spreading equally well is the recognition that fire is a
27 fundamental natural ecological agent in many of our ecosystems and only a ‘problem’ where we
28 choose to inhabit these fire-prone regions or we humans introduce it to non fire-adapted
29 ecosystems [3]. The ‘wildfire problem’ is essentially more a social than a natural one.
30
31

32 The warming climate, which is predicted to result in more severe fire weather in many regions of
33 the globe in this century [51] will probably contribute further to both perceived and actual risks
34 to lives, health and infrastructures. Therefore, the need for human societies to co-exist with fire
35 will continue, and may increase in the future [9]. We thus need to move towards a more
36 sustainable co-existence with fire. This requires a balanced and informed understanding of the
37 realities of wildfire occurrence and its effects. It is hoped that the data and discussion presented
38 here, together with the other contributions in this special issue, reduce misconceptions about fire
39 and assist in providing this understanding.
40
41
42

43 **Acknowledgments**

44 We sincerely thank A. Scott, C. Belcher, C. Ross and W. Chaloner for organizing the Royal Society Discussion
45 Meeting “The Interaction of Fire and Mankind” and for inviting us to participate. Our fruitful discussions with
46 experts from a wide range of fire-related disciplines greatly helped to shape this manuscript.

47 **Authors' Contributions**

48 Both authors contributed to the ideas presented here, to drafting the article and approved the final version.
49

50 **Competing Interests**

51 We have no competing interests in relation to the article's content.
52

53 **Funding**

54 CS has been supported by a Leverhulme Trust Grant (RPG-2014-095).
55

56 **References**

- 57
58
59 1. Belcher, C. M., Hadden, R. M., intrinsic flammability
60 Yearsley, J. M., McElwain, J. C. & of Earth's ecosystems
Rein, G. 2010 Baseline estimated from

- paleoatmospheric oxygen over the past 350 million years. *Proc. Natl. Acad. Sci. U. S. A.* **107**, 22448–22453. (doi:10.1073/pnas.1011974107)
2. He, T., Belcher, C. M., Lamont, B. B. & Lim, S. L. 2015 A 350-million-year legacy of fire adaptation among conifers. *J. Ecol.*, 1–12. (doi:10.1111/1365-2745.12513)
3. Santin, C. & Doerr, S. H. In press. Fire effects on soils: the human dimension. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*
4. Scott, A. C., Bowman, D. M. J. S., Bond, W. J., Pyne, Stephen J & Alexander, M. E. 2014 *Fire on Earth: An introduction*. John Wiley & Sons, Ltd.
5. Gowlett, J. A. J. 2016 The discovery of fire by humans: a long and convoluted process. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, This issue.
6. Raftoyannis, Y. et al. 2014 Perceptions of forest experts on climate change and fire management in European Mediterranean forests. *iForest - Biogeosciences For.* **7**, 33–41. (doi:10.3832/ifor0817-006)
7. Fabra-crespo, M. & Rojas-briales, E. 2015 media news on forest issues : a case study of Spain. **24**.
8. Kyriazopoulos, A. P., Arabatzis, G., Abraham, E. M. & Parissi, Z. M. 2013 Threats to Mediterranean rangelands: A case study based on the views of citizens in the Viotia prefecture, Greece. *J. Environ. Manage.* **129**, 615–620. (doi:10.1016/j.jenvma.2013.08.035)
9. Moritz, M. a et al. 2014 Learning to coexist with wildfire. (doi:10.1038/nature13946)
10. North, B. M. P., Stephens, S. L., Collins, B. M., Agee, J. K., Aplet, G., Franklin, J. F. & Fulé, P. Z. 2015 Reform forest fire management. *Science (80-)*. **349**, 1280–1281.
11. McCaffrey, S. 2015 Community Wildfire Preparedness: a Global State-of-the-Knowledge Summary of Social Science Research. *Curr. For. Reports* **1**, 81–90. (doi:10.1007/s40725-015-0015-7)
12. Yell, S. 2010 “Breakfast is now tea, toast and tissues”: affect and the media coverage of bushfires. *Media Int. Aust. Inc. Cult. Policy* **137**, 109–119.
13. Randerson, J. T., Chen, Y., Van Der Werf, G. R., Rogers, B. M. & Morton, D. C. 2012 Global burned area and biomass burning emissions from small fires. *J. Geophys. Res. Biogeosciences* **117**. (doi:10.1029/2012JG002128)
14. Graham, A. 2015 How Journalists Fan the Flames of Wildfire in the West. *Mont. Journal. Rev. Mag.*
15. Paveglio, T., Norton, T. & Carroll, M. S. 2011 Fanning the Flames? Media Coverage during Wildfire Events and its Relation to Broader Societal Understandings of the Hazard. *Hum. Ecol. Rev.* **18**, 41–52.
16. Varela, E., Jacobsen, J. B. & Soliño, M. 2014 Understanding the heterogeneity of social preferences for fire prevention management. *Ecol. Econ.* **106**, 91–104. (doi:10.1016/j.ecolecon.2014.07.014)
17. Moreira, N. In press. Study Links Increase in Wildfires to Global Warming. *Boston Globe*.
18. Scientists, U. of C. In press. Is Global Warming Fueling Increased Wildfire Risks?
19. Almagro, C. 2009 El futuro en llamas.

20. Northoff, E. 2003 Fire are increasingly damaging the world's forests. *Fire. An Environmental History, Told Through Fire, of Europe and Europe's Encounter with the World*. Washington: University of Washington Press.
21. Donovan, G. H. & Brown, T. C. 2007 Be careful what you wish for : the legacy of Smokey Bear.
22. Stephens, S. L. et al. 2014 Temperate and boreal forest mega-fires: characteristics and challenges. *Front. Ecol. Environ.* **12**, 115–122. (doi:10.1890/120332)
23. Fernandes, P. M., Davies, G. M., Ascoli, D., Fernández, C., Moreira, F., Rigolot, E., Stoof, C. R., Vega, J. A. & Molina, D. 2013 Prescribed burning in southern Europe: Developing fire management in a dynamic landscape. *Front. Ecol. Environ.* **11**, e4–e14. (doi:10.1890/120298)
24. Burrows, N. & McCaw, L. 2013 Prescribed burning in southwestern Australian forests. *Front. Ecol. Environ.* **11**. (doi:10.1890/120356)
25. Pyne, S. J. 2016 Fire in the mind: changing understandings of fire in Western civilization. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, this issue.
26. Wallace, W. 1965 Fire in the jarrah forest environment. *J. R. Soc. West. Aust.* **49**, 33–44.
27. Pyne, S. J. 1997 *Fire. An Environmental History, Told Through Fire, of Europe and Europe's Encounter with the World*. Washington: University of Washington Press.
28. McCaffrey, S., Toman, E., Stidham, M. & Shindler, B. 2015 Chapter 2 - Social Science Findings in the United States. *Wildfire Hazards, Risks and Disasters*, 15–34. (doi:http://dx.doi.org/10.1016/B978-0-12-410434-1.00002-6)
29. Bliege-Bird, R., Bird, D. W. & Codding, B. F. 2016 People, ENSO, and fire in Australia: fire regimes and climate controls in hummock grasslands. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, this issue.
30. Mistry, J., Bilbao, B. & Berardi, A. 2016 Community owned solutions for fire management in tropical ecosystems: case studies from Indigenous communities of South America. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, this issue.
31. Marlon, J. R., Bartlein, P. J., Carcaillet, C., Gavin, D. G., Harrison, S. P., Higuera, P. E., Joos, F., Power, M. J. & Prentice, I. C. 2008 Climate and human influences on global biomass burning over the past two millennia. *Nat. Geosci.* **1**, 697–702. (doi:10.1038/ngeo313)
32. Wang, Z., Chappellaz, J., Park, K. & Mak, J. E. 2010 Large Variations in Southern Hemisphere Biomass Burning During the Last 650 Years. *Science (80-.)*. **330**, 1663–1666. (doi:10.1126/science.1197257)
33. Van der Werf, G. R., Peters, W., van Leeuwen, T. T. & Giglio, L. 2013 What could have caused pre-industrial biomass burning emissions to exceed current rates? *Clim. Past* **9**, 289–306. (doi:10.5194/cp-9-289-2013)
34. Knorr, W., Kaminski, T., Arneeth, a. & Weber, U. 2014 Impact of human population density on fire frequency at the global scale. *Biogeosciences* **11**, 1085–1102. (doi:10.5194/bg-11-1085-2014)
35. Giglio, L., Randerson, J. T. & Van Der Werf, G. R. 2013 Analysis of daily, monthly, and annual burned area using the fourth-generation global fire emissions database (GFED4). *J. Geophys. Res. Biogeosciences* **118**, 317–328. (doi:10.1002/jgrg.20042)
36. Van Lierop, P., Lindquist, E.,

- Sathyapala, S. & Franceschini, G. 2015 Global forest area disturbance from fire, insect pests, diseases and severe weather events. *For. Ecol. Manage.* **352**, 78–88. (doi:10.1016/j.foreco.2015.06.010)
37. Caon, L., Vallejo, V. R., Ritsema, C. J. & Geissen, V. 2014 Effects of wildfire on soil nutrients in Mediterranean ecosystems. *Earth Sci. Rev.* **139**, 47–58. (doi:10.1016/j.earscirev.2014.09.001)
38. Flannigan, M., Krawchuk, M., de Groot, W., Wotton, B. & Gowman, L. 2009 b: Implications of changing climate for global wildland fire. *Int. J. Wildl. Fire* **18**, 483–507.
39. Mouillot, F. & Field, C. B. 2005 Fire history and the global carbon budget: A fire history reconstruction for the 20th century. *Glob. Chang. Biol.* **11**, 398–420. (doi:10.1111/j.1365-2486.2005.00920.x)
40. Rocca, M. E., Miniati, C. F. & Mitchell, R. J. 2014 Introduction to the regional assessments: Climate change, wildfire, and forest ecosystem services in the USA. *For. Ecol. Manage.* **327**, 265–268. (doi:10.1016/j.foreco.2014.06.007)
- Higuera, P. E., Abatzoglou, J. T., Littell, J. S. & Morgan, P. 2015 The Changing Strength and Nature of Fire-Climate Relationships in the Northern Rocky Mountains, U.S.A., 1902-2008. *PLoS One* **10**, e0127563. (doi:10.1371/journal.pone.0127563)
42. Dennison, P. E., Brewer, S. C., Arnold, J. D. & Moritz, M. A. 2014 Large wildfire trends in the western United States, 1984-2011. *Geophys. Res. Lett.* **41**, 2928–2933. (doi:10.1002/2014GL061184.Received)
43. Westerling, a. L., Hidalgo, H. G., Cayan, D. R. & Swetnam, T. W. 2006 Warming and earlier spring increase western U.S. forest wildfire activity. *Science (80-.)*. **313**, 940–3. (doi:10.1126/science.1128834)
44. Westerling, a. L. 2016 Increasing western US forest wildfire activity: sensitivity to changes in the timing of Spring. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*
45. Short, K. C. 2015 Sources and implications of bias and uncertainty in a century of US wildfire activity data. *Int. J. Wildl. Fire* **24**, 883–891. (doi:10.1071/WF14190)
- NIFC 2016 National Interagency Fire Center Statistics.
47. Mallek, C. M., Safford, H., Viers, J. & Miller, J. 2013 Modern departures in fire severity and area vary by forest type , Sierra Nevada and southern Cascades , California , USA. *Ecosphere* **4**, 1–28. (doi:10.1890/ES13-00217)
48. Jolly, W. M., Cochrane, M. a., Freeborn, P. H., Holden, Z. a., Brown, T. J., Williamson, G. J. & Bowman, D. M. J. S. 2015 Climate-induced variations in global wildfire danger from 1979 to 2013. *Nat. Commun.* **6**, 7537. (doi:10.1038/ncomms8537)
49. De Groot, W. J., Cantin, A. S., Flannigan, M. D., Soja, A. J., Gowman, L. M. & Newbery, A. 2013 A comparison of Canadian and Russian boreal forest fire regimes. *For. Ecol. Manage.* **294**, 23–34. (doi:10.1016/j.foreco.2012.07.033)
50. Kelly, R., Chipman, M. L., Higuera, P. E., Stefanova, I., Brubaker, L. B. & Hu, F. S. 2013 Recent burning of boreal forests exceeds fire regime limits of the past 10,000 years. *Proc. Natl. Acad. Sci. U. S. A.* **110**, 13055–60.

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
51. Flannigan, M., Cantin, A. S., De Groot, W. J., Wotton, M., Newbery, A. & Gowman, L. M. 2013 Global wildland fire season severity in the 21st century. *For. Ecol. Manage.* **294**, 54–61. (doi:10.1016/j.foreco.2012.10.022)
52. De Groot, W. J., Flannigan, M. D. & Cantin, A. S. 2013 Climate change impacts on future boreal fire regimes. *For. Ecol. Manage.* **294**, 35–44. (doi:10.1016/j.foreco.2012.09.027)
53. Fox-Hughes, P., Harris, R., Lee, G., Grose, M. & Bindoff, N. 2014 Future fire danger climatology for Tasmania, Australia, using a dynamically downscaled regional climate model. *Int. J. Wildl. Fire* **23**, 309–321. (doi:10.1071/Wf13126)
54. Liu, Y., Stanturf, J. & Goodrick, S. 2010 Trends in global wildfire potential in a changing climate. *For. Ecol. Manage.* **259**, 685–697. (doi:10.1016/j.foreco.2009.09.002)
55. Liu, Z., Yang, J., Chang, Y., Weisberg, P. J. & He, H. S. 2012 Spatial patterns and drivers of fire occurrence and its future trend under climate change in a
56. Fox, D. M. et al. 2015 Increases in fire risk due to warmer summer temperatures and wildland urban interface changes do not necessarily lead to more fires. *Appl. Geogr.* **56**, 1–12. (doi:10.1016/j.apgeog.2014.10.001)
57. Keeley, J. E. 2009 Fire intensity, fire severity and burn severity: A brief review and suggested usage. *Int. J. Wildl. Fire* **18**, 116–126. (doi:10.1071/WF07049)
58. Shakesby, R. a. & Doerr, S. H. 2006 Wildfire as a hydrological and geomorphological agent. *Earth-Science Rev.* **74**, 269–307. (doi:10.1016/j.earscirev.2005.10.006)
59. Martin, D. A. In press. Fire impacts on water resources. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*
60. Stephens, S. L., Agee, J. K., Fulé, P. Z., North, M. P., Romme, W. H., Swetnam, T. W. & Turner, M. G. 2013 Managing Forests and Fire in Changing Climates. *Science (80-.)*. **342**, 41–2. (doi:10.1126/science.1240294)
61. Kramer, M. In press. Why Big, Intense Wildfires Are the New Normal. *Natl. Geogr. Mag.*
62. Parker, L. In press. How Megafires Are Remaking American Forests. *Natl. Geogr. Mag.*
63. Schroeder, M. J. & Buck, C. C. 1970 Fire Weather : A Guide for Application of Meteorological Information to Forest Fire Control Operations. *USDA For. Serv. Agric. Handb.*
64. Victorian Government 2010 *Final report*.
65. Cruz, M. G., Sullivan, a. L., Gould, J. S., Sims, N. C., Bannister, a. J., Hollis, J. J. & Hurley, R. J. 2012 Anatomy of a catastrophic wildfire: The Black Saturday Kilmore East fire in Victoria, Australia. *For. Ecol. Manage.* **284**, 269–285. (doi:10.1016/j.foreco.2012.02.035)
66. Marlon, J. R. et al. 2012 PNAS Plus: Long-term perspective on wildfires in the western USA. *Proc. Natl. Acad. Sci.* **109**, E535–E543. (doi:10.1073/pnas.1112839109)
67. Dillon, G. K., Holden, Z. a., Morgan, P., Crimmins, M. a., Heyerdahl, E. K. & Luce, C. H. 2011 Both

- topography and climate affected forest and woodland burn severity in two regions of the western US, 1984 to 2006. *Ecosphere* **2**, art130. (doi:10.1890/ES11-00271.1)
68. Hanson, C. T. & Odion, D. C. 2014 Is fire severity increasing in the Sierra Nevada, California, USA? *Int. J. Wildl. Fire* **23**, 1–8.
69. Hanson, C. T. & Odion, D. C. 2015 Sierra Nevada fire severity conclusions are robust to further analysis: a reply to Safford et al. *Int. J. Wildl. Fire* **24**, 294–295. (doi:10.1071/WF13016)
70. Miller, J. D. & Safford, H. 2012 Trends in Wildfire Severity: 1984 To2010 in the Sierra Nevada, Modoc Plateau, and Southern Cascades, California, Usa. *Fire Ecol.* **8**, 41–57. (doi:10.4996/fireecology.0803041)
71. Baker, W. L. 2015 Are High-Severity Fires Burning at Much Higher Rates Recently than Historically in Dry-Forest Landscapes of the Western USA? *PLoS One* **10**, e0136147. (doi:10.1371/journal.pone.0136147)
72. Guha-Sapir, D., Below, R. & Hoyois, P. In press. EM-DAT: International Disaster
- www.emdat.be. *Univ. Cathol. Louvain – Brussels – Belgium.*
73. (Gfmc), T. G. F. M. C. 2012 IFFN-GFMC UNISDR Global Wildland Fire Network Bulletin. **17**.
74. NIFC 2015 National Interagency Fire Center. , 22pp.
75. Mutch, R. W. 2013 Just leave the line. *Wildfire Mag.*
76. Fahy, R. F., Leblanc, P. R. & Molis, J. L. 2007 What ' S Changed Over the Past 30 Years ? *Analysis*
77. Cardil, a. & Molina, D. M. 2014 Factors Causing Victims of Wildland Fires in Spain (1980–2010). *Hum. Ecol. Risk Assess. An Int. J.* **21**, 67–80. (doi:10.1080/10807039.2013.871995)
78. González, D. 2012 Los incendios forestales en España generan gastos y pérdidas que superan anualmente los 1.000 millones. *ARN Digit.*
79. Gorman, S. & Simpson, I. 2015 Property losses from northern California wildfire nearly double. *Reuters, U.S. Ed.*
80. Futuro, V. 2014 Conaf estima en US\$ 100 millones el costo directo por incendios forestales. *Emol.*
81. Badger, S. G. 2014 Large loss fires in the United States, 2013.
82. Stephenson, C., Handmer, J. & Robyn, B. 2013 Estimating the economic, social and environmental impacts of wildfires in Australia. *Environ. Hazards* **12**, 93–111.
83. Rahn, M. 2009 Wildfire Impact Analysis. *San Diego State Univ.* , 15pp.
84. Kochi, I., Donovan, G. H., Champ, P. a & Loomis, J. B. 2010 The economic cost of adverse health effects from wildfire-smoke exposure: a review. *Int. J. Wildl. Fire* **19**, 803–817. (doi:10.1071/WF09077)
85. Gonzalez-Caban, a 2013 The economic dimension of wildland fires. *Notes* , 229–237.
86. Hammer, R. B., Radeloff, V. C., Fried, J. S. & Stewart, S. I. 2007 Wildlandurban interface housing growth during the 1990s in California, Oregon, and Washington. *Int. J. Wildl. Fire* **16**, 255–265. (doi:10.1071/WF05077)
87. Johnston, F. H., Shannon, M. & Bowman, D. M. J. S. In press. The pyrohealth transition – how fire emissions have influenced human

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
88. Johnston, F. H., Henderson, S. B., Chen, Y., Randerson, J. T., Marlier, M., DeFries, R. S., Kinney, P., Bowman, D. M. J. S. & Brauer, M. 2012 Estimated global mortality attributable to smoke from landscape fires. *Environ. Health Perspect.* **120**, 695–701. (doi:10.1289/ehp.1104422)
89. Paveglio, T. B., Brenkert-Smith, H., Hall, T. & Smith, A. M. S. 2015 Understanding social impact from wildfires: advancing means for assessment. *Int. J. Wildl. Fire* **24**, 212–224. (doi:10.1071/WF14091)
90. Prentice, I. C. 2010 The Burning Issue. *Science (80-.)*. **330**, 1636–1637. (doi:10.1126/science.1199809)
91. Thomas, D. S. & Butry, D. T. 2013 Areas of the U.S. wildland–urban interface threatened by wildfire during the 2001–2010 decade. *Nat. Hazards* **71**, 1561–1585.
92. San-Miguel-Ayanz, J., Moreno, J. M. & Camia, A. 2013 Analysis of large fires in European Mediterranean landscapes: Lessons learned and perspectives. *For. Ecol. Manage.* **294**, 11–22. (doi:10.1016/j.foreco.2012.10.050)
93. Prosser, I. P. & Williams, L. 1998 The effect of wildfire on runoff and erosion in native Eucalyptus forest. *Hydrol. Process.* **12**, 251–265.
94. Millar, C. I. & Stephenson, N. L. 2015 Temperate forest health in an era of emerging megadisturbance. *Science (80-.)*. **349**, 823–826.
95. Bond, W. J. & Zaloumis, N. P. 2016 The deforestation story: testing for anthropogenic origins of Africa’s flammable grasslands. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, this issue.
96. He, T., Pausas, J. G., Belcher, C. M., Schwilk, D. W. & Lamont, B. B. 2012 Fire-adapted traits of *Pinus* arose in the fiery Cretaceous. *New Phytol.* **194**, 751–759.
97. Bowman, D. M. J. S., Perry, G. L. W., Higgins, S. I., Johnson, C. N., Fuhlendorf, S. D. & Murphy, B. P. 2016 Pyrodiversity is the coupling of biodiversity and fire regimes in food-webs. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, this issue.
98. Keeley, J. E., Pausas, J. G., Rundel, P. W., Bond, W. J. & Bradstock, R. a. 2011 Fire as an evolutionary pressure shaping plant traits. *Trends Plant Sci.* **16**, 406–411. (doi:10.1016/j.tplants.2011.04.002)
99. Davies, G. M. 2016 The role of fire in U.K peatland and moorland management; the need for informed, unbiased debate. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, this issue.
100. Page, S. E. & Hooijer, A. 2016 In the line of fire: the peatlands of Southeast Asia. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*
101. Laurance, S. G., William, F., Rose, M. & National, C. P. 2015 Peat fires: emissions likely to worsen. *Nature* **527**, 305. (doi:10.1038/527305e)

Figures

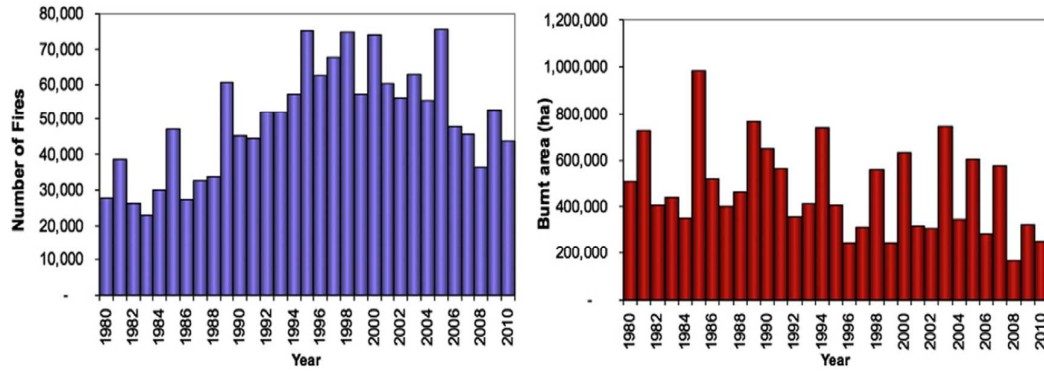


Figure 1: Wildfire occurrence (left) and corresponding area burnt (right) in the European Mediterranean region for the period 1980-2010. Source: San-Miguel-Ayanz et al. [92].

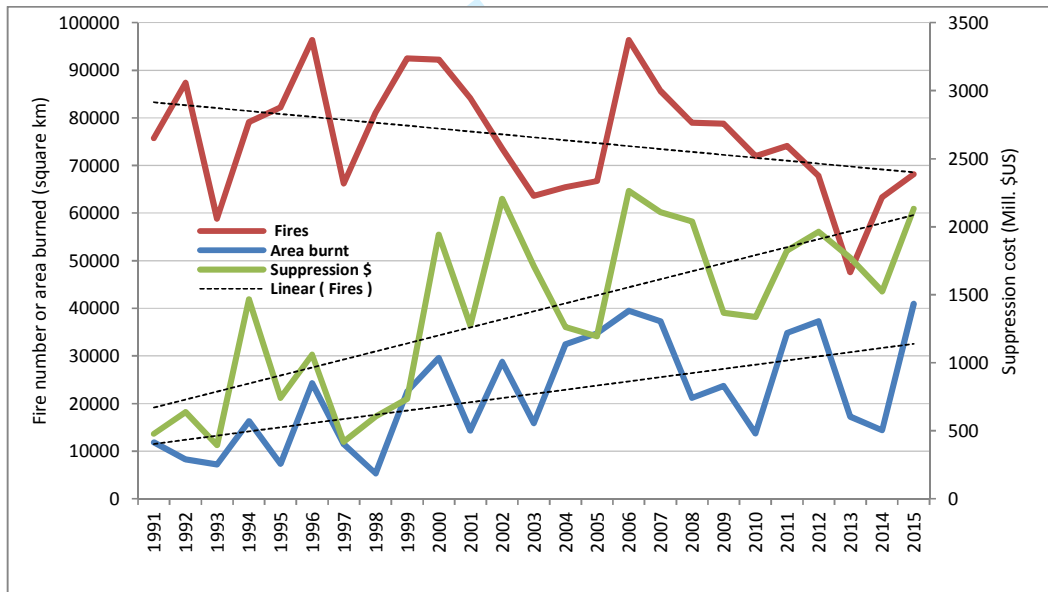


Figure 2: Area burned, number of fires and suppression costs (inflation adjusted to 2016 equivalent) for the USA with linear trend lines (1991-2015). Data: National Interagency Fire Center [46].

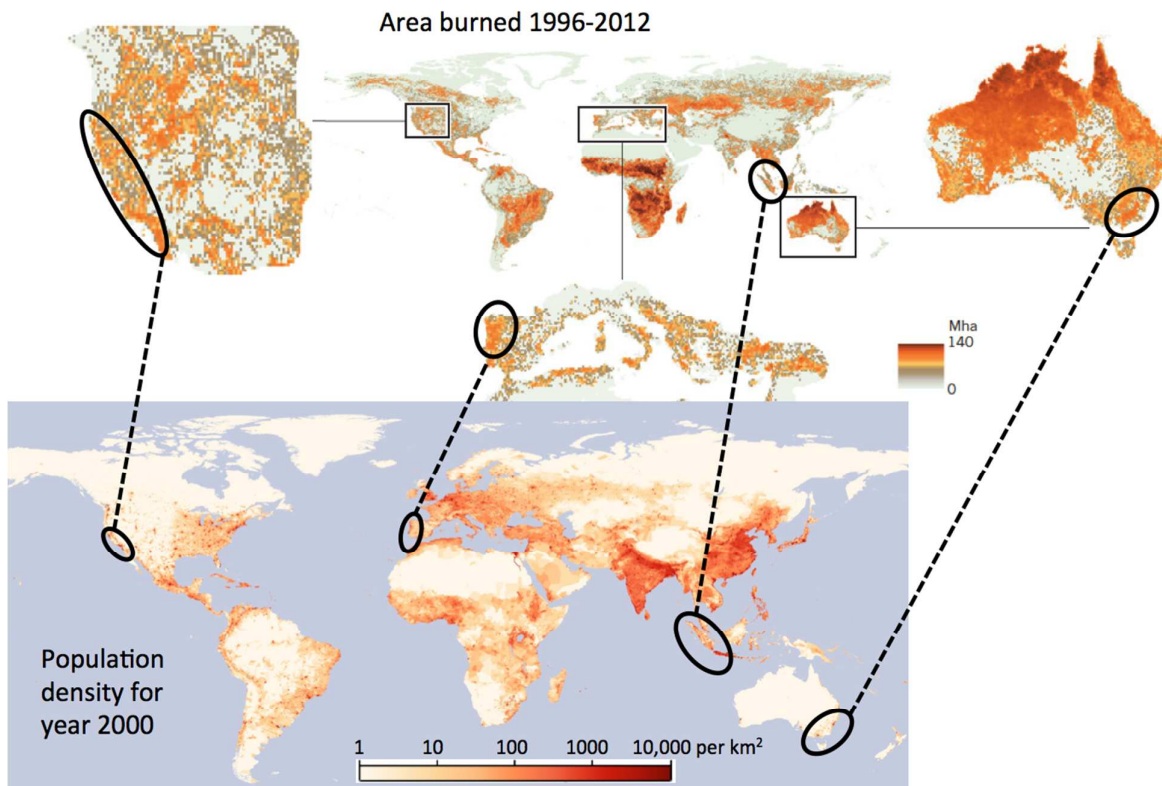


Figure 3: Global area burned with enlarged sections of the globe (1996-2012) and global population density with examples where regions with high proportions of area burned coincide with high population densities. Based and modified from Moritz et al. [9] and NASA (http://neo.sci.gsfc.nasa.gov/view.php?datasetId=SEDAC_POP).

Tables

Table 1: Global comparison of human and economic losses derived from wildfire, earthquakes and flood disasters from 1901 to 2014. (Source: EM-DAT 2015 [72]).

	Wildfires	Earthquakes	Floods
N. of Events	387	1291	4,481
People killed	3,753	2,574,627	6,947,908
People injured	6,812	2,614,875	1,329,923
People affected (Mill.)	6	190	3,604
Risk of death (%)*	0.06	1.4	0.02
Total direct damage (Mill. \$)	54,828	774,771	681,427
Cost per event (Mill. \$)	142	600	152
Cost per person affected (\$)	9,138	4,078	189

*N. of fatalities per N. people affected (%)

Table 2: Human and economic losses from wildfire 'disasters' by global region from 1984 to 2013. Costs are based on the actual value of \$ in a given reporting year. Data: EM-DAT 2013 [72].

	N. of events	People killed	Total people affected	Death rate/event	Economic costs (\$Mill)
Africa	25	272	21,672	11	440
America	118	234	1,229,175	9	25,229
Asia	50	748	3,188,257	30	11,892
Europe	89	462	1,295,562	18	12,619
Oceania	21	224	74,320	9	2,121
Total	303	1940	5808986	78	52,301