

## **Predicted decrease in global climate suitability masks regional complexity of invasive fruit fly species response to climate change**

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### **Abstract**

Climate change affects the rate of insect invasions as well as the abundance, distribution and impacts of such invasions on a global scale. Among the principal analytical approaches to predicting and understanding future impacts of biological invasions are Species Distribution Models (SDMs), typically in the form of correlative Ecological Niche Models (ENMs). An underlying assumption of ENMs is that species-environment relationships remain preserved during extrapolations in space and time, although this is widely criticised. The semi-mechanistic modelling platform, CLIMEX, employs a top-down approach using species ecophysiological traits and is able to avoid some of the issues of extrapolation, making it highly applicable to investigating biological invasions in the context of climate change. The tephritid fruit flies (Diptera: Tephritidae) comprise some of the most successful invasive species and serious economic pests around the world. Here we project 12 tephritid species CLIMEX models into future climate scenarios to examine overall patterns of climate suitability and forecast potential distributional changes for this group. We further compare the aggregate response of the group against species-specific responses. We then consider additional drivers of biological invasions to examine how invasion potential is influenced by climate, fruit production and trade indices. Considering the group of tephritid species examined here, climate change is predicted to decrease global climate suitability and to shift the cumulative distribution poleward. However, when examining species-level patterns, the predominant directionality of range shifts for 11 of the 12 species is eastward. Most notably, management will need to consider regional changes in fruit fly species invasion potential where high fruit production, trade indices and predicted distributions of these flies overlap.

## Introduction

Global climate change together with elevated volumes of trade, human movement, transport, habitat modification and agricultural production represent key stressors currently impacting ecosystem function, human health and food security (Patz *et al.* 2005; Thuiller 2007; Brook *et al.* 2008; Hellmann *et al.* 2008). Changes including altered climatic constraints and mechanisms of introduction will facilitate the establishment and spread of some organisms outside of their native range (Hellmann *et al.* 2008). For invasive insects, current research predicts expanded geographic distributions and elevated population densities and/or voltinism under various climate change scenarios (Bale *et al.* 2002, Harrington *et al.* 2007; Walther *et al.* 2009). Further, a non-trivial proportion of agricultural pests are alien and/or invasive insects (Ward & Masters, 2007), and thus expected to have potentially serious consequences for sustained agricultural production under climate change (Ziska *et al.* 2011). General patterns of climate change response for agriculturally important pest insects include poleward range expansion and increased herbivory at higher latitudes (e.g. Bale *et al.* 2002; Bebber *et al.* 2013). The search for generality in anticipated climate change impacts is an important component of effective management at a regional or global scale. However, particularly for invasive insects that tend to have disproportionately strong impacts on agricultural production (Ziska *et al.* 2011), it is becoming increasingly clear that species responses are largely idiosyncratic and must also be understood individually (Dukes *et al.* 2009; Gutierrez and Ponti 2014).

Among the key determinants of invasion success for insect species are climatic suitability, propagule pressure and the availability of suitable hosts (for parasitic or phytophagous insects; Ward & Masters, 2007; Bacon *et al.* 2014). Climate change is expected to influence the distribution and abundance of invasive insects both directly (e.g., by altering where species and hosts can occur) and indirectly (e.g., via changes in population growth rates, propagule pressure, and spread), amongst other factors (Lantschner *et al.* 2014). It is important to note that effects of climate change on the distribution and severity of invasive insects may be negative or positive (see Ward & Masters, 2007; Bertelsmeier *et al.* 2015). While each of these key determinants are important, they are often only examined in isolation (Bacon *et al.* 2014) or may be highly correlated and/or intertwined, such that identifying the principal drivers of new invasions is extremely challenging (Yamanaka *et al.* 2015). Nevertheless, in order to help control and limit the spread of invasive insect species, understanding the direct and indirect impacts of climate change on invasions is a critical exercise. Further, combining multiple species into a common analytical framework and considering multiple time scales may reveal important general trends or patterns, including the identification of potential “invasion hotspots” - areas that are likely to hold higher-than-expected suitable climate space as a prerequisite for a number of potentially invasive species (Bertelsmeier *et al.* 2015). To predict how distributions and invasions may change with climate change, Species Distribution Models (SDMs) are widely used (Elith & Leathwick, 2009). Typically these are in the form of correlative Ecological Niche Models (ENMs) that aim to describe statistical relationships between environmental variables and known occurrences of species, and to subsequently project these relationships into novel space or conditions (e.g.,

under different climate change scenarios). Ecological niche models have been used effectively to approximate general patterns of species invasions, including likely introduction points (Fitzpatrick *et al.* 2007), assessments of invasion risk as a function of climate change (Bertelsmeier *et al.* 2013a; 2015), and generating hypotheses concerning invasion range extents that can then be tested experimentally (Hill *et al.*, 2013).

Using correlative ENMs for extrapolation of species-environment relationships assumes that they remain stable, both temporally and in novel environments. This assumption is often violated (Jiménez-Valverde *et al.* 2011; Webber *et al.* 2011), particularly when using ENMs to examine synergies between climate change and invasions (see Bertelsmeier *et al.* 2013a,b). Thus, SDMs that rely on at least a partial mechanistic understanding of the species-environment relationship are perhaps more appropriate tools in predicting distributions under climate change and biological invasion scenarios. One such SDM is CLIMEX (Hearne Scientific Software Pty Ltd, Australia) (Sutherst & Maywald 1985; Sutherst *et al.* 2007) which employs a semi-mechanistic approach to examine the relationship between climate, species distributions and patterns of growth (Macfadyen and Kriticos, 2012). CLIMEX models are typically fitted using a combination of empirically measured life history parameters (e.g. ecophysiological traits and developmental rates), abundance data and point distribution records. The CLIMEX software calculates a weekly Growth Index (GI) based on species physiological response curves for temperature and soil moisture availability. This GI is then combined with stress indices (e.g. hot/cold and wet/dry) to calculate an Ecoclimatic Index (EI) (Sutherst *et al.* 2007; Macfadyen and Kriticos 2012). The EI can then be determined across spatially explicit climate data to estimate the suitability of each grid cell (ranging from completely unsuitable to optimal) (see methods; Sutherst *et al.* 2007). CLIMEX is thus able to avoid some of the issues involved with transferability associated with ENMs, as the model is rebuilt *de novo* for climate change scenarios rather than being extrapolated from current distributional estimates (Webber *et al.* 2011). Thus, CLIMEX is well suited for investigating the interactions between biological invasions and potential responses to climatic change (e.g. Mika and Newman 2010; Lozier and Mills 2011; Hill *et al.* 2014).

The tephritid flies (Diptera: Tephritidae), or “true fruit flies”, are some of the most successful invaders and are serious economic pests around the world (Aluja and Mangan 2008; Papadopoulos *et al.* 2013; Karsten *et al.* 2015). The pest flies in this family typically cause major damage to fruits, especially via larval feeding, though damage from oviposition also occurs (Aluja and Mangan 2008). Strict phytosanitary and trade regulations enacted to manage invasive spread can result in indirect economic losses (Duyck *et al.* 2004; De Meyer *et al.* 2008). In addition there is increasingly well-documented history of invasions of various tephritid species globally, which have continued to spread rapidly despite major efforts to control their movement (see Duyck *et al.* 2004; 2007; Hill and Terblanche, 2014; Papadopoulos *et al.* 2013). Tephritid invasions are thought to be primarily associated with the global transport of fruit. Recent analyses of invasion pathways for *Ceratitis capitata* support this hypothesis, for at least the last two centuries (Karsten *et al.* 2015). More recently, passenger baggage for air travel has revealed itself as a major pathway (Liebhold *et*

al. 2006; Ma *et al.* 2012), and that these introductions are not isolated events but rather occur with some regularity. This is seen in examples which include a number of pest tephritids that are continually reappearing in California (Papadopoulos *et al.* 2013), a large increase in the number of *B. latifrons* interceptions in China in recent years (Ma *et al.* 2012) and *Bactrocera dorsalis* (previously *B. invadens*) intercepted repeatedly in traps in northern South Africa after spreading from recent introduction in Kenya (Manrakhan *et al.* 2011).

Tephritids are an excellent system for examining how climate change might influence biological invasions of insects. Many of the pest species within the genera *Anastrepha*, *Bactrocera* and *Ceratitis* spp. occupy largely overlapping ecological niches including broadly similar life-histories. Whilst many of these tephritid species have polyphagous diets, others such as those within the genus *Rhagoletis* are specialists, although this may be on a widely planted fruit crop such as apples (*R. pomonella*) (Aluja and Mangan 2008). Thus, the general ubiquity of host plants for these species has facilitated the global invasion of many of these species. Many of the pest tephritids also have tropical origins (although temperate for *Rhagoletis* spp.), and thus any poleward range expansion associated with changing climate opens an abundance of new habitat for these species, in many cases into regions with high fruit production (Stephens *et al.* 2007; Ni *et al.* 2012; Fu *et al.* 2014). Specifically, climate warming in temperate regions improve conditions for the flies to establish, through fewer frost days, a longer growing season and greater frequency of warm nights (Papadopoulos *et al.* 2014). Tephritids often have high dispersal ability and rapid growth and reproductive rates, as is characteristic of many invasive species and agricultural pests (Tschardt *et al.* 2005). Competition between invasive tephritids is a possibility when there is a shared pathway resulting in multiple species introduction, and resident host plants are limited (e.g. Duyck *et al.* 2004; 2007). In such cases it may be that the successful invaders are the better competitors than resident species, and better colonists than the other invaders (Duyck *et al.* 2007). Overall however, local biotic resistance from endemic species and natural enemies appears to be low for the highly pestiferous species. Finally, due to their important pest status and the amount of research that is directed at this group, there have been numerous attempts to model current and potential distributions for members of this fly family (Yonow and Sutherst 1998, Vera *et al.* 2002), several using CLIMEX. Although there are a few individual assessments (Stephens *et al.* 2007; Ma *et al.* 2012; Ni *et al.* 2012; Fu *et al.* 2014), a comprehensive assessment of their responses to global climate change has not yet been explored.

Here we examine how climate change may alter patterns of geographic distributions through global climatic suitability for a number of tephritid species, both individually and across this group. We first collate SDMs that have been built in CLIMEX, and project them into both baseline and future climate scenarios. We next identify regions of overlap in climatic suitability of these key invasive species in an attempt to highlight areas where additive or synergistic effects of invasion may occur, allowing for the identification of putative invasion hotspots. We further extend these climatic models and combine outputs with information on trade (correlated with invasion risk; Bacon *et al.* 2014; Lantschner *et al.* 2014) and fruit production (a reasonable proxy for agricultural host availability; Bacon *et al.* 2014) to enhance the predictability of invasion potential and impacts in light of three core drivers of invasion success: climate, host availability and dispersal barriers.

We contrast our results to what is generally expected for climate change impacts on invasive insect species; namely poleward expansion, increased distributional range and elevated invasive potential.

## Methods

### *Model parameters*

There are a number of tephritid CLIMEX models present in the literature and included with the CLIMEX software. CLIMEX (Version 4.0 beta) currently ships with model parameters for the following tephritid fruit fly species: *Bactrocera tryoni* (Yonow & Sutherst, 1998), *Bactrocera dorsalis* (Stephens *et al.* 2007) and *Ceratitis capitata* (Vera *et al.* 2002). These parameters are populated from the literature which has established key physiological and population growth parameters such as thermal tolerance, desiccation tolerance, growth rates and phenology (see Table 1, and references therein). This combination of data from different sources allows for testing and validation of the model using both empirical data and field observations (e.g. Macfadyen and Kriticos 2012). Some modelling attempts have incorporated using statistical tests of goodness of fit with known distribution points (e.g. Area Under the receiver-operator characteristic (ROC) Curve (AUC); Ni *et al.* 2011). The focus of this study was to look at general patterns of invasion potential with climate change, rather than provide fine scale risk assessments of establishment for any particular species or region. We collated 9 additional models to those included with CLIMEX (Table 1) to give a total of 12 parameterised tephritid models. The different genera represented were *Anastrepha* (*A. ludens*, *A. obliqua*), *Bactrocera* (*B. dorsalis*, *B. correcta*, *B. latifrons*, *B. tryoni*, *B. zonata*), *Ceratitis* (*C. capitata*, *C. rosa*), *Rhagoletis* (*R. pomonella*, *R. indifferens*) and *Zeugodacus cucurbitae* (Previously *B. cucurbitae*; Virgilio *et al.* 2015). We assumed that models built on subsets of their range (e.g. *Z. cucurbitae* in China; Lingbin *et al.* 2008) would generally perform well when projected to a global surface, though perhaps conservative in their predictions. We used CLIMEX 4.0 (beta version) to rebuild these models and produce spatial outputs of Ecoclimatic Index (EI). At each grid cell the EI scale ranges 0–100, with 0 representing an unfavourable environment, values over 20 considered to be ideal conditions, between 10 and 20 suitable, and between 0.01 and 10 marginal (Sutherst and Maywald, 2005; Ni *et al.* 2012), though interpretations of EI are species-specific to some extent.

### *Environmental data*

All climatic data were obtained from the CliMond dataset (Version 1.2; Kriticos *et al.* 2012). The baseline climate reflects average conditions between 1961 and 1990 with 1975 as a midpoint of these data. CLIMEX models were initially predicted to the baseline data for each species. For climate change projections, we used the CSIRO Mk. 3.0 and Miroc-H models (available from CliMond) for the time periods 2030, 2050 and 2070 under the A2 SRES for future climate change. The A2 SRES represents one of the higher rates, though not the highest, of warming in the Fourth Assessment Report (AR4; IPCC 2007). We then predicted the CLIMEX models on the same climate change scenarios and time frames globally, and then averaged the EI across the CSIRO and Miroc-H models for each time frame (as in Hill *et al.* 2014). As we are most interested in broad scale patterns across countries, and as most of the models were initially parameterised at this

resolution, we employed a resolution of 30 minutes, approximately 50×50 km at the equator.

### *Distribution changes*

We built spatial raster files from the CLIMEX output using the raster package (Hijmans 2015) in R (R Core Team (2015), version 3.2.1). Using the EI values, we calculated the net change in climate suitability (area expressed in grid cells) for each species and time point. To examine the tephritids as a whole, grid cells were selected at a binary suitability threshold ( $EI > 10$ ) for each species and each time point (baseline, 2030, 2050, 2070), and then combined into single maps to determine current and future climatically suitable areas and potential “invasion hotspots”. Beta diversity values including species turnover and nestedness were then calculated by comparing the combined predictions for all 12 species between the baseline and 2070 climates, using R and following the same approach as used in Bishop *et al.* (2015). Species turnover is a measure of non-overlap among the species considered and in this case is based on suitable climate space for each grid cell at baseline versus 2070. Nestedness maps display absolute species losses relative to the baseline group, ignoring species gains (Baselga 2010).

To investigate species individually, the baseline time period was compared to the model projections for 2070 to compare suitable range increases and decreases (see Delean *et al.* 2013). For each species the grid cells that changed with projected climate change were classed as either “lost”, “gained” or remained as “stable range”. Global net change was then calculated by  $(\text{Gains-Losses})/\text{stable range}$ . We also measured the amount of overlap between the baseline conditions and the three future time periods across the whole global surface for each individual species using Schoener's *D* (dismo package in R; Hijmans *et al.* 2015). In addition to losses and gains, we measured the amount of change and calculated the movement of the range margins of each species (range shifts). Range shifts were calculated using the *sp* (Pebesma and Bivand 2005) and raster packages in R, at 15% of the range margin in each cardinal direction (North, South, West, East) and the vector connecting the range margin in 1975 and the range margin in 2070. In addition, the vector of the movement of each distribution's centroid was calculated by connecting the centers of gravity in 1975 and 2070.

### *Additional drivers*

We employed a simple measure of risk by relating climate suitability to fruit production and total agricultural trade data. Data for total fruit production, including melons, was obtained for all countries for the time period 2009 to 2013 from the FAO (FAO 2015). While total fruit production does not take into account the relative proportions and suitabilities of hosts for each tephritid species, nor for the group as a whole, it is a reasonable and considerably more tractable parameter that we hypothesize will broadly correlate with the overall availability of hosts at a landscape scale. For each country, the quantity of fruit production (kilotons) was calculated across these years and mapped by country, but most islands were excluded from the analysis. As a proxy for propagule pressure, we investigated trade data. Data for trade indices, including an import and an export quantity index for agricultural production, was obtained for the years 2007 to 2011 (FAO). Average

values across these years were determined and then this was mapped by country. As there are no data on projected trade or fruit production rates, this analysis was carried out for the baseline time period only. In this way we are assuming that high levels of trade will correlate with increased pathways for these fruit flies, but not necessarily trade of specific fruit products. This method does not specifically account for movement via other pathways such as air passenger baggage (Liebhold *et al.* 2006).

We then examined how indices of import, export and fruit production related to the countries for which there was suitable climate space (determined through the CLIMEX models) and whether the species had been recorded as present or absent in that country (CABI Invasive species compendium; <http://www.cabi.org/isc/> accessed July 2015). For example, suitable climate space, high fruit production and a high import index with a species not recorded may correspond to higher risk of invasion. Conversely, high export index, high level of fruit production and the presence of a tephritid species may indicate a higher risk of contributing to invasions. For each of the 11 species (*R. indifferens* was excluded from this analysis due to lack of data), we selected all countries that hold a 50 km grid cell with an EI >10 and either a presence or absence for that species. We determined two separate indices, first using the presences, we multiplied the export index with the fruit production index, and secondly, using the absences we multiplied the import index with the fruit production index (fruit production was log<sub>10</sub> transformed prior in both cases). Here we assume that fruit production will correlate with available hosts and that trade indices will correlate with propagule pressure, ignoring differences in quarantine and detection programmes across these different countries. These indices provide a simple risk analysis of invasion or export potential for each country, and when summed up across all species allows invasion or escape potential for each species/country to be ranked. Our method also makes the assumption that all fruit production and trade will include suitable hosts for a given species. Due to recent taxonomic changes (Schutze *et al.* 2015), we combined data for *B. invadens* with *B. dorsalis*.

## Results

### *Distributions and climate change*

The CLIMEX model parameters for the 12 species of tephritid fruit fly are shown in Table 1. The global projections of all the species demonstrate that they have very wide potential niches, with suitable climate space available for most species throughout tropical, subtropical and temperate regions (see Supplementary Material 1 for individual species maps). The current climate suitability for the group of all species considered is very broad, distributed throughout equatorial and subtropical through to temperate regions, and particularly in the Southern Hemisphere (Fig. 1a). Sub-Saharan Africa and South America especially are highly suitable (the highest number of species in any grid cell is 12), as are South-east Asia and northeastern Australia. North America, Europe and Russia are marginal from climate suitability standpoint for this group of species (Fig. 1a). Under the climate change scenarios investigated here the climate suitability pattern remains similar to current conditions, though there are noticeable contractions in central (equatorial) South

America and Africa, as well as a general decrease in suitable climate space for the tephritid group in Southeast Asia, Australia and India. Marginal suitability observed in North America and Russia under current climate expands under future climate change scenarios (Fig. 1, b-d), largely driven by *R. indifferens* and *R. pomonella* (Supplementary Material 1.12-13).

Maps of beta diversity are useful for understanding how species composition is likely to change between baseline and future projections. Overall, tephritid community composition is predicted to change most dramatically in northern and eastern Europe, parts of North America, interior India, South America and Australia, and in tropical and temperate sub-Saharan Africa (areas of blue in Fig. 2a) according to the 2070 climate projections. Regions of large potential differences in composition occur on all continents and include tropical regions of South America and Africa, parts of southern Africa, parts of Europe and North America ( $\beta_{sor}$ ; Fig. 2a). Compositional changes due to species turnover ( $\beta_{sim}$ ; Fig. 2b) appear to be high in the regions where large overall differences in composition are predicted (Fig. 2a) suggesting that turnover appears to be largely responsible for the differences in overall composition. Compositional changes due to nestedness ( $\beta_{sne}$ ; Fig. 2c), where high values indicate areas of greatest potential loss of species in future, were at their peak in tropical Africa, southern Africa, tropical regions South America, Southeast Asia and India. A large region of moderate nestedness occurs across most of central Africa, indicating a potential loss of species in this region in future. There are also large areas of moderate nestedness across most of Europe.

#### *Range shifts, niche overlap and climate change*

When examined individually, all species display a decrease in climate suitability under the A2 SRES for all time periods, except for *R. pomonella* that shows increases under all time periods, and *B. correcta* that has slight increases in climate suitability for 2030 and 2050, but a loss at 2070 (Fig. 3a). Niche overlap between the baseline conditions and the different time frames for climate change under the A2 scenario, as measured using Schoener's *D* (Fig. 3b), reflects that not only is climate suitability generally decreasing, but that the potential niches for each species are becoming increasingly dissimilar. For *R. pomonella*, the overlap also decreases indicating that the gains in climate suitability under climate change are also driving elevated dissimilarity from baseline conditions (Fig. 3b). The change in climate suitability maps for each species, indicating which grid cells are “lost”, “gained” or maintain a “stable range” between the baseline conditions and 2070 are presented in Supplementary Material 2 and data underlying the shift metrics are available in Supplementary Material 3. The range shifts by species (‘shift vectors’) reveal a more complex pattern of distributional changes across latitudes (Fig. 4) relative to simple poleward expansion. For most species (11 of 13), the greatest change is eastward, with lesser shifts towards the southern or northern range margins (Fig. 4). In combination with these range margin changes there were also centroid shifts (see Supplementary Material 2 and 3).



## *Additional drivers*

From our simple classification scheme, we ranked potential invasion risk countries by determining which hold suitable climate space, the presence or absence of each species, and the fruit production and trade (import and export) indices. The top 15 countries that are most at risk of exporting a fruit fly species include India, Thailand, the United States of America and countries in eastern sub-Saharan Africa (Fig. 5, red shading). Countries throughout South America, Spain, Turkey and New Zealand have the highest risk of import of a new invasive fruit fly species (blue shading). For some countries, such as Indonesia and Vietnam, both the import and export risk of invasive tephritids is high (purple shading). Typically, the countries at risk of importing fruit flies under baseline conditions also maintain suitable climate space under future climate change. Likewise, the potential sources are countries that have many endemic tephritid species and are also mostly developing countries that are increasing trade (and see Supplementary Material 4).

## **Discussion**

Our approach of simultaneously considering species separately and as a group demonstrated that overall patterns of climate change response may mask the complexity of individual range shifts and species turnover. Climate change is predicted to alter global climate suitability for this group of tephritid species, and the combined shift in distributions is expected to be poleward, mainly through contraction from equatorial regions. Within regions with preserved or enhanced climate suitability, species display varying levels of expansion and contraction, driving species turnover at the grid cell (50 km<sup>2</sup>) level. Such patterns are difficult to generalise across species. Indeed, when considered individually, there is global reduction in climate suitability for each species, except for *R. pomonella*. However, the principal direction of range shifts is eastward, rather than poleward shifts anticipated by most studies predicting distributional range changes under warming (Stephens *et al.* 2007; Ni *et al.* 2011; Fu *et al.* 2014). An eastward range shift may be associated with complex interactions between temperature and precipitation (see Lenoir and Svenning 2015). Such complexity and idiosyncratic species responses suggest that predicting and managing future invasions of these fruit flies and other pest species will be a dynamic and enduring challenge.

The response of tephritid distributions to climate change has been predicted to be primarily poleward, linked to a release from cold stress conditions in higher latitudes under climate change. Climate change is also expected to benefit certain species, such as *A. obliqua*, which was predicted to expand its range on all continents (Fu *et al.* 2014). We did not observe such a benefit for the majority of species investigated here, including *A. obliqua*. This discrepancy may be due to the differences in climate data used (Fu *et al.* applied a uniform increase in temperature and precipitation instead of an SRES), the time frame for future projections (i.e., 2020 in Fu *et al.* and 2030, 2050, 2070 in this study) or at the spatial scale employed. Another potential source of variability across our results is the use of a different climate dataset, CliMond, to what most of the species were calibrated on. As CLIMEX parameters are sensitive to both changes in spatial resolution

(Kriticos and Leriche 2010; Taylor and Kumar 2012) and differences in the underlying climate data, there is potential that use of a different climate dataset has contributed to slight under- or over-predictions. These factors, among others are all known to contribute to differences between SDM studies and such errors in prediction need to be taken into account when interpreting results (Elith and Leathwick 2009; Synes and Osborne 2011). We therefore recommend that any study interested in determining a single-species risk assessment for any of these species assess the possible influence of model training data quality and spatial scale for their suitability prior to management or risk forecasting.

The models herein also describe an increase in predicted climate suitability for *Rhagoletis pomonella* under climate change. This may be at least partially due to *Rhagoletis* species having a narrower thermal niche than the other species considered here (Table 1) and is primarily a temperate species. Thus climate change predictions generally forecast a range expansion into higher latitudes. In terms of host range, *Rhagoletis* species are also specialist exploiters rather than opportunistic broad range exploiters (e.g. *Anastrepha*, *Bactrocera*, *Ceratitis*; Aluja and Mangan 2008), which must be considered when making more targeted assessments of response to climate change for this genus. Additionally, species like *B. latifrons* and *Z. cucurbitae* are restricted to host plants in the solanaceae and cucurbitaceae, respectively, and our approach therefore generalises this aspect of the species biology. However, the ubiquity of such host plants is not likely to be restrictive at a global scale, and correlating distributions with total fruit production should also capture respective hosts. Our results present an interesting comparison to a similar study that examined 15 species of invasive ants and climate change (Bertelsmeier *et al.* 2015). Overall, a decline in global climate suitability was demonstrated for the ant species investigated, although five of the species did display potential for range expansion (Bertelsmeier *et al.* 2015). Thus for two major groups of invasive insects, ants and fruit flies, climate change is not expected to broadly favour invasions, though both groups hold some notable exceptions.

While there are some significant benefits to applying CLIMEX to a group of species such as the tephritid flies examined here, caveats apply to the interpretation of these models. First, broad scale climatic variables (~50 km) are estimated and forecast at a vastly different scale relative to most organisms including fruit flies. Effective regional management of the individual species will require that finer temporal and spatial scales be employed. Tephritid flies typically use habitats on the scales of orchards and the trees within them, representing a range of microclimates that can buffer against the broad scale climatic conditions (see Barton and Terblanche 2014), although capable of long-distance dispersal in a small fraction of the population. Microclimates that provide refuges from climatic change may also be facilitated through local irrigation regimes that simulate rainfall and increase available moisture to buffer against desiccation (Stephens *et al.* 2007). Additionally, a 50 km grid cell is the product of much statistical downscaling and interpolation (Austin and Van Niel 2011), and thus the value it holds represents a long term average across a large area for a particular variable, typically not incorporating much variation. The physiological parameters used in CLIMEX models are typically grounded on laboratory data and seldom incorporate environmental

heterogeneity, nor diurnal ranges of temperature or microclimatic effects (Geng *et al.* 2011). To the degree that especially native distributions may be constrained by biotic interactions (e.g., competition, natural enemies), may lead to underestimates of potential distributions in introduced regions (Stephens *et al.* 2007). Future studies attempting to predict the future impacts of these invasive insects and response to climate change should also move beyond static parameter estimates and investigate adaptive and plasticity of traits that are involved with resistance to the cold and heat stresses that define range limits and invasiveness (e.g. Chown *et al.* 2007) including establishment success. Rapid adaptation in invasive insects may facilitate persistence in a novel environment (Gibert *et al.* THIS ISSUE), including climate change. CLIMEX models thus serve as an important tool to detect patterns of invasion and highlight differences of these patterns across species, rather than specific risk assessments.

Although climate suitability is a prerequisite for invasion and can be used as a guide to help predict and hopefully prevent future invasions, propagule pressure and host availability are also critical to invasion success (Ward and Masters 2007; Bacon *et al.* 2014). To comprehensively combine host information and propagule pressure is a challenging task, as future trade and fruit production projections are largely unavailable in public repositories. Insect introductions are typically accidental and driven through hitchhiking on plant material, soil and wood products (Desurmont and Pearse, 2014; Liebhold *et al.* THIS ISSUE) and many interceptions of some species of tephritid are from individual people carrying fruit across borders (Ma *et al.* 2012). Thus while our use of trade data and fruit production as proxies for propagule pressure and host availability are large assumptions, we anticipate there will be correlation with the actual processes underlying tephritid invasions. To forecast agricultural production and trade links accurately will require that ecologists work with economists in a pest risk analysis framework (Baker *et al.* 2000) to determine how trade partners are likely to change over the next century, and how new trading opportunities into foreign markets may alter invasion risk of fruit flies and other invasive species (see Nuñez and Pauchard 2010).

In examining how the potential additive effects of climate suitability, host availability (fruit production by country) and propagule pressure (trade) contribute to invasion success, the tephritid species were weighted equally and no competition was incorporated into our models. Interaction with endemic species (particularly other tephritids), and other invasive species that occupy similar niches to these (Duyck *et al.* 2004; 2007) is an important factor that contributes to the success of invasions, particularly at a local scale and on islands. For example, *Bactrocera dorsalis* outcompetes a range of other invasive tephritids when they have come into contact (see Duyck *et al.* 2004 for review) and *B. tryoni* competitively displaced *C. capitata* when their distributions overlapped in Australia (Duyck *et al.* 2004). We also assumed that the species are not limited by seasonal fruit availability, particularly for winter, which is not only likely to influence the range of some of these species (Stephens *et al.* 2007), but also their competitive ability. Competitive interactions are, however, unlikely to strongly influence the broad scale distribution assessment employed here (Pearson and Dawson 2003).

Fruit fly invasions under climate change will challenge global, regional and local food security, and the high variability in potential tephritid invasions from multiple sources will complicate future management issues regionally. Regions such as South America, and sub-Saharan Africa (i.e., those with economically important fruit production and developing markets) will likely be forced to maintain strict quarantine, particularly where export to the Northern Hemisphere or when another climatically suitable, uninvaded region predominates (e.g. Stephens *et al.* 2007). Due to the economic importance of tephritid species, understanding how climate change will influence the probability and severity of future invasions will be paramount for developing trade relations and economic security.

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## Figures and Tables

**Table 1.** CLIMEX parameters for 12 species of tephritid fruit flies. **Temperature threshold** (°C) (DV0 = lower temperature threshold, DV1 = lower optimum temperature, DV2 = upper optimum temperature, DV3 = upper temperature threshold); **Soil moisture threshold** (SM0 = lower soil moisture threshold, SM1 = lower optimum soil moisture, SM2 = upper optimum soil moisture, SM3 = upper soil moisture threshold); **Heat stress** (TTHS = Heat stress temperature threshold (°C), THHS = Heat stress temperature rate, DTHS = Heat stress degree-day threshold, DHHS = Heat stress degree-day rate); **Cold stress** (TTCS = Cold stress temp threshold (°C), THCS = Cold stress Temperature rate, DTCS = Cold Stress degree-day threshold, DHCS = Cold stress degree-day rate); **Dry Stress** (SMDS = Dry stress threshold, HDS = Dry stress rate); **Wet stress** (SMWS = Wet stress threshold, HWS = Wet stress rate); PDD is the number of degree-days per generation. **Diapause index** (*Rhagoletis pomonella* only; DPD0 = Weekly day length that induces diapause [12], DPT0 = Weekly minimum temperature that induces diapause [20], DPT1 = weekly minimum temperature that terminates diapause [5], DPD = Minimum number of days for diapause development to be completed[83]). Refer to the CLIMEX manual for further explanation of the units and derivations of these parameters.

**Figure 1.** Suitable climate space for 12 species of tephritid fruit flies for (a) baseline climatic conditions (1961-1990 centered on 1975) and future climate change (A2 SRES for (b) 2030, (c) 2050 and (d) 2070 time periods). Each species is equally weighted at the threshold value of the Ecoclimatic Index (EI) >10. The shading indicates number of species that have suitable climate space for a given grid cell, with a value of 0 indicating the grid cell is not suitable for any species, through to 13 indicating that all species have the potential to occupy that grid cell.

**Figure 2.** Global species turnover and nestedness of tephritid species under climate change. Ecoclimatic Index values of 10 or more were taken to represent suitable climate space per species for each map. (a) Total compositional variation between a group of tephritid fruit flies for 1975 and 2070, calculated using  $\beta_{\text{sor}}$ . Darker areas show highest variation. (b) Compositional changes between a group of tephritid fruit flies for 1975 and 2070 that are due to turnover, calculated using  $\beta_{\text{sim}}$ . Darker areas show greatest turnover in species composition. (c) Compositional changes between a group of tephritid fruit flies for 1975 and 2070 due to nestedness, calculated using  $\beta_{\text{sne}}$ . Darker areas show greatest loss in species from 1975 to 2070.

**Figure 3.** Range size and niche changes under climate change for 12 species of tephritid fruit flies. (a) Potential geographic range size changes measured as losses or gains of grid cells under climate change (A2 SRES for 2030, 2050 and 2070 time periods). (b) Temporal fidelity in niche overlap (calculated as Schoener's  $D$ ) between baseline conditions and three future time points (0 = complete dissimilarity across time points, 1=no change).

**Figure 4.** Comparison of the magnitude and direction of range shifts calculated for 12 tephritid species.

Range shift distance was calculated as shift vectors of the range margins and the movement of the centroid vector between the predicted distributions for baseline and 2070 climates. Values are unitless as they are centered on the mean and divided by the standard deviation. White: range contraction (negative values), Black: range expansion (positive values. represents by small and large squares respectively). To see the direction of these shifts, refer to individual species maps in Supplementary Material 2.

**Figure 5.** Map of countries ranked based on holding suitable climate space, high fruit production and suitable trade indices for 10 different tephritid species. Red shading indicates that these are the top 15 countries that pose an export risk of an invasive tephritid. Blue shading indicates that these 15 countries have a risk of importing an invasive tephritid. Countries that hold both risks are shaded purple.

Table 1

Species	Temperature Index				Soil Moisture Index				Heat Stress			
	DV0	DV1	DV2	DV3	SM0	SM1	SM2	SM3	TTHS	THHS	DTHS	DHHS
<i>Anastrepha ludens</i>	7,8	14,0	25,0	35,0	0.05	0.2	0.8	1.2	37,0	0.00015	NA	NA
<i>Anastrepha obliqua</i>	10,0	20,0	30,0	36,0	0.05	0.1	1.1	1.6	39,0	0.00015	NA	NA
<i>Ceratitis capitata</i>	12,0	22,0	30,0	35,0	0.1	0.3	1	1.5	39,0	0.01	100	0
<i>Ceratitis rosa</i>	8,6	24,0	28,0	33,0	0.105	0.4	1	1.5	NA	NA	NA	NA
<i>Bactrocera dorsalis</i>	13,0	25,0	33,0	36,0	0.1	0.25	1	1.5	36,0	0.005	NA	NA
<i>Bactrocera correcta</i>	15,7	20,0	34,0	35,0	0.1	0.15	1.4	1.5	36,0	.00008	NA	NA
<i>Bactrocera latifrons</i>	15,7	18,0	33,0	36,0	0.1	0.5	1	1.8	36,0	0.005	NA	NA
<i>Bactrocera tryoni</i>	12,0	25,0	33,0	36,0	0.1	0.5	1.75	2	36,0	0.005	0.4375	0.01
<i>Bactrocera zonata</i>	12.6	20,0	30,0	36,0	0.1	0.2	1	1.6	36,0	0.0005	NA	NA
<i>Rhagoletis pomonella</i>	8,3	15,0	25,0	31,0	0.2	0.6	1	1.5	33,0	0.0005	NA	NA
<i>Rhagoletis indifferens</i>	3,0	5,0	25,0	28,0	0.0	0.1	0.6	0.76	28,0	0.01	NA	NA
<i>Zeugodacus cucurbitae</i>	10,4	20,0	30,0	35,0	0.1	0.2	1	1.4	36,0	0.001	NA	NA
Species	Cold Stress				Dry Stress		Wet Stress		Degree Days			
	TTCS	THCS	DTCS	DHCS	SMDS	HDS	SMWS	HWS	PDD	Reference		
<i>Anastrepha ludens</i>	9,0	-0.003	NA	NA	0.1	-0.0002	1.1	0.001	607	Jian et al. 2008		
<i>Anastrepha obliqua</i>	7.6	-0.003	NA	NA	0.05	-0.0001	1.65	0.001	642	Fu et al. 2014		
<i>Ceratitis capitata</i>	10,0	0	5	0.0015	0.02	0.05	1.6	0.0015	622	Vera et al. 2002		
<i>Ceratitis rosa</i>	0,0	0	15	-0.001	0.105	-0.021	1.6	0.0015	400	de Villiers et al. 2012		
<i>Bactrocera dorsalis</i>	2.5	-0.012	8	-0.002	0.1	-0.024	1.5	0.007	470	Stephens et al. 2007		
<i>Bactrocera correcta</i>	9,0	-0.0005	NA	NA	0.1	-0.00036	1.5	0.0006	726.3	Lu et al 2010		
<i>Bactrocera latifrons</i>	2,0	-0.10	NA	NA	0.10	-0.005	1.8	0.002	415.40	Ma et al. 2012		
<i>Bactrocera tryoni</i>	2,0	0.1	20	0.00025	0.1	0.005	2	0.002	380	Yonow & Sutherst 1998		
<i>Bactrocera zonata</i>	2,0	-0.008	NA	NA	0.08	-0.0007	1.60	0.005	380	Ni et al. 2012		
<i>Rhagoletis pomonella</i>	NA	NA	NA	NA	0.2	-0.001	1.5	0.02	1065	Geng et al. 2011		
<i>Rhagoletis indifferens</i>	-6,0	-0.001	NA	NA	0.0	0.0	0.76	0.001	1800	Kumar et al 2014		
<i>Zeugodacus cucurbitae</i>	5,0	-0.005	NA	NA	0.02	-0.05	1.4	0.0015	5599999999	Lingbin et al 2008		













