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What traits underpin the successful establishment and spread of the invasive water bug *Trichocorixa verticalis verticalis*?

--Manuscript Draft--

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Abstract:	<p>The introduction of exotic species has a major impact on a wide range of ecosystems, especially in aquatic ecosystems. <i>Trichocorixa v. verticalis</i> (Fieber, 1851), an euryhaline aquatic hemipteran native North America, has occurred as an exotic species in the Iberian Peninsula since at least 1997. In this study, we compared several physiological and biological traits (salinity tolerance of the different developmental stages, thermal tolerance, fecundity and dispersal ability) in the alien species and three native, syntopic corixid species (<i>Sigara lateralis</i>, <i>Sigara scripta</i> and</p>	

	<p>Sigara selecta), to determine which traits may explain its invasion success. Trichocorixa verticalis was the species most resistant to high conductivity at the egg stage, while S. selecta showed the highest halotolerance as adults. The invader had the highest upper thermal limit and a much higher fecundity than Sigara species. Wing morphometry suggested that T. verticalis may be a stronger flier than the native species. Our findings provide an example of how functional and ecological niche interactions among alien and native species can help predict impacts of invasion on aquatic communities.</p>
<p>Response to Reviewers:</p>	<p>Dear Editor,</p> <p>Please find enclosed a revised version of the ms “What traits underpin the successful establishment and spread of the invasive water bug Trichocorixa verticalis verticalis?” (Ref.: HYDR-D-15-00385) by Carbonell et al.</p> <p>We have followed the suggestions of the editor, which we found helpful and insightful. We are now resubmitting a revised version, which we would like you to consider of publication in Hydrobiologia.</p> <p>In the letter which follows, we indicate how we have responded to the comments (the responses have been written between –dashes–). We hope that our ms is now acceptable for publication. If not, we are ready to make whatever further changes are required.</p> <p>Thank you for considering our work to be published in Hydrobiologia.</p> <p>Sincerely, JA Carbonell</p> <p>-----</p> <p>COMMENTS TO THE AUTHOR:</p> <p>All earlier comments have been adequately addressed and I am pleased to inform you that this manuscript may be accepted for publication, pending these minor editorial comments and the final assessment by the EIC.</p> <p>23-24 change “a good example” to “an example” for a more unbiased representation.</p> <p>186-7 Both species are different in colouration, size (S. lateralis is bigger) and form (T. v. verticalis is more parallel shaped) (Günter, 2004; Nieser et al., 1994). Change to “The two species are different in coloration, size (S. lateralis is bigger) and shape, T. v. verticalis being more??? [elongated? “parallel shaped” is not a term used in English] (Günter, 2004; Nieser et al., 1994).</p> <p>336 Differences between T. v. verticalis nymphs and adults in survival time... Change to: “Differences in survival time between T. v. verticalis nymphs and adults...”</p> <p>436 that allowing – change to “that allow”.</p> <p>477 Resistant structures- perhaps ‘resting stages’? It is unclear that resistant structures refers to any biological entity.</p> <p>- In the new version of the manuscript we have included the suggested changes of the editor. Track changes (in yellow) has been used to speed up review process. -</p> <p>252,4 what is ramping programs?</p> <p>- A ramping program performs a dynamic temperature change along time (in our case 1 °C min⁻¹) until the end points (positive or negative). To reduce possible confusion the sentence has been rewritten from “employing a dynamic method using two ramping programs with a rate of temperature change of 1 °C min⁻¹, from acclimation temperature (20 °C) until the upper (70 °C) or lower temperature limit (-45 °C)” to “employing a dynamic method using a ramping rate of ± 1 °C min⁻¹, from acclimation temperature (20 °C) until the upper (70 °C) or lower temperature limit (-45 °C)” (L 252-</p>

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1 **TITLE:** What traits underpin the successful establishment and spread of the invasive
2 water bug *Trichocorixa verticalis verticalis*?

3

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10

11 SUMMARY

12 The introduction of exotic species has a major impact on a wide range of ecosystems,
13 especially in aquatic ecosystems. *Trichocorixa v. verticalis* (Fieber, 1851), an
14 euryhaline aquatic hemipteran native North America, has occurred as an exotic species
15 in the Iberian Peninsula since at least 1997. In this study, we compared several
16 physiological and biological traits (salinity tolerance of the different developmental
17 stages, thermal tolerance, fecundity and dispersal ability) in the alien species and three
18 native, syntopic corixid species (*Sigara lateralis*, *Sigara scripta* and *Sigara selecta*), to
19 determine which traits may explain its invasion success. *Trichocorixa verticalis* was the
20 species most resistant to high conductivity at the egg stage, while *S. selecta* showed the
21 highest halotolerance as adults. The invader had the highest upper thermal limit and a
22 much higher fecundity than *Sigara* species. Wing morphometry suggested that *T.*
23 *verticalis* may be a stronger flier than the native species. Our findings provide **an**
24 **example** of how functional and ecological niche interactions among alien and native
25 species can help predict impacts of invasion on aquatic communities.

26

27 KEYWORDS: corixids, *Trichocorixa*, *Sigara*, alien species, expansion, traits.

28

29 INTRODUCTION

30 The introduction of alien species has a major impact on a wide range of ecosystems and
31 is a major cause of global change (Ricciardi, 2006). Alien invaders are of particular
32 concern in aquatic ecosystems (Dudgeon *et al.*, 2006), especially in protected areas
33 (Guareschi *et al.*, 2013). For example, in the protected Doñana wetlands (south-western
34 Spain), multiple invasions of aquatic organisms have occurred: e.g., the copepod
35 *Acartia tonsa* (Dana, 1849), the crayfish *Procambarus clarkii* (Girard, 1852), the
36 gastropod *Potamopyrgus antipodarum* (Gray, 184), the fishes *Gambusia affinis* (Baird
37 & Girard, 1853) and *Lepomis gibbosa* (Linnaeus, 1758) (Rodríguez-Pérez & Green,
38 2012) and the macrophyte *Azolla filiculoides* (Lam. 1783) (Espinar *et al.*, 2015). A
39 recent addition to this list is the water boatman *Trichocorixa verticalis verticalis*
40 (Fieber, 1851). This species is a small (c. 5 mm) euryhaline corixid (Hemiptera)
41 originally distributed in North America and the Caribbean, where it mainly lives in
42 coastal wetlands (Hutchinson, 1931; Jansson, 2002). It has been recorded as an alien
43 species in South Africa, New Caledonia, Morocco and Portugal, being the only water
44 bug recognized as an alien species in Europe (Rabitsch, 2008). Since its first report in
45 Europe (Cádiz, Spain) by Günther (2004), it has been recorded from various areas of
46 southern Portugal (Sala & Boix, 2005), southwest Spain (Millán *et al.*, 2005;
47 Rodríguez-Pérez *et al.*, 2009; Van de Meutter *et al.*, 2010a) and Morocco (L´Mohdi *et*
48 *al.*, 2010), and it is predicted to spread widely across Europe and the Mediterranean
49 region in future years (Guareschi *et al.*, 2013).

50

51 The distribution of *T. v. verticalis* in the introduced range in south-west Spain has been
52 expanding in recent years (Carbonell *et al.*, 2012a; Guareschi *et al.*, 2013), especially in
53 and around Doñana, where it has become the dominant breeding corixid at several sites,
54 especially in brackish and saline permanent water bodies (Rodríguez-Pérez *et al.*, 2009;
55 Van de Meutter *et al.*, 2010b).

56

57 The invasion process can be divided into a series of stages (transport, establishment,
58 spread and impact) and at each stage there are barriers that need to be overcome for a
59 species or population to pass on to the next stage (Blackburn *et al.*, 2011). In the first
60 phase of the invasion process, the species must travel from its native range to a new
61 ecosystem (transport). In the second phase it must survive, grow and reproduce under
62 the new environmental conditions (establishment). In the third phase (spread), it must
63 acquire a high rate of population growth, invading new regions through active and/or
64 passive dispersal (including possible dispersal by migratory birds as eggs, see Figuerola
65 *et al.*, 2003) and short generation time. Finally, in the last phase, the alien species must
66 alter the structure and functioning of the invaded ecosystem (impact) (Parker *et al.*,
67 1999) through interactions with other species or ecosystem engineering.

68

69 The pathway of entry for *T. v. verticalis* is unknown, but some authors have suggested
70 that it was facilitated by the introduction of American fishes *Gambusia affinis* and/or
71 *Fundulus heteroclitus* (Sala & Boix, 2005; L'Mohdi *et al.*, 2010), while others have
72 suggested it might have been transported in ballast water (Guareschi *et al.*, 2013). Some
73 biological and ecological traits of *T. v. verticalis* relevant to its invasiveness have been
74 studied in its native (Tones, 1977; Kelts, 1979; Simonis, 2013) or invasive range
75 (Coccia *et al.*, 2013; 2014). Salinity tolerance was compared between adults of *T. v.*

76 *verticalis* and three native *Sigara* species (Van de Meutter *et al.*, 2010b), concluding
77 that *S. selecta* outperforms *T. v. verticalis* in terms of salinity tolerance, and that
78 therefore its invasion success cannot simply be explained only by its high halotolerance
79 as an adult. Thermal tolerance has also been studied (Coccia *et al.*, 2013), finding that
80 *T. v. verticalis* presents greater thermal plasticity than the native *Sigara lateralis*,
81 suggesting that this may facilitate its spread in the future, since *T. v. verticalis* may be
82 more able to positively respond to climate change than native species. Furthermore, Van
83 de Meutter *et al.* (2010a) studied interactions between *T. v. verticalis* with the aquatic
84 macroinvertebrate community and found indirect evidence that the native *S. selecta* has
85 been displaced by *T. v. verticalis*, especially in artificial saline water bodies (Rodríguez-
86 Pérez *et al.*, 2009). There is similar evidence for displacement of *S. selecta* by *T. v.*
87 *verticalis* in Moroccan salt-pans (L'Mohdi *et al.*, 2010). Other studies have focused on
88 interactions of *T. v. verticalis* and *Sigara* spp. with predators and parasites in the
89 Doñana area (Coccia *et al.*, 2014; Sánchez *et al.* 2015), finding that in freshwaters the
90 invasive species is more sensitive to predation by Odonata larvae and to parasitism by
91 aquatic mites than the native species.

92

93 However, no previous comparative information exists on reproductive rates, dispersal
94 ability or salinity tolerance of eggs or nymphs. There is a need for an integrative study
95 considering the major traits involved in the spread and establishment phase of *T. v.*
96 *verticalis* and how these traits influence the interactions (competition, coexistence) with
97 native corixids. In the present study, we combined field distribution and experimental
98 data on *T. v. verticalis* and three native corixid species to address the major
99 physiological and biological traits related with the establishment and spread phases of
100 the *T. v. verticalis* invasion. We aimed to identify traits that enhance the potential of this

101 alien species to outcompete native species and that explain its distribution and recent
102 expansion in southern Spain, particularly in mesosaline waters. We compared
103 physiological, life history and morphological traits among *T. v. verticalis* and the
104 coexisting *Sigara* species (corixids with a similar size) along the salinity gradient at
105 which they occur: *Sigara lateralis* (Leach, 1817) and *S. scripta* (Rambur, 1840) in
106 subsaline and hyposaline waters and *S. selecta* (Fieber, 1848) in mesosaline waters. We
107 studied salinity tolerance and thermal tolerance as determinants of the ability to tolerate
108 a wide range of abiotic conditions (Alonso & Camargo, 2003, 2004; Gérard *et al.*,
109 2003). We also studied oviposition rate, hatching time and hatching success as measures
110 of reproductive capacity and the recruitment of new individuals (Richards, 2002), as
111 well as potential flight dispersal ability related to the mechanism for active dispersal and
112 spread (Sakai *et al.*, 2001). We studied thermal tolerance and fecundity traits at two
113 salinity concentrations representing the extremes of the gradient where the alien species
114 coexists with native species, to test if these traits change with salinity, since exposure to
115 different salinities can influence thermal tolerance (Sánchez-Fernández *et al.*, 2010;
116 Arribas *et al.*, 2012; Coccia *et al.*, 2013) and fecundity (Gelin *et al.*, 2001; Castro-
117 Longoria, 2003) in aquatic Coleoptera and Hemiptera. We hypothesised that the
118 establishment success of *T. v. verticalis* in mesosaline waters would be mediated by
119 wide environmental tolerance and a high reproductive rate that exceeds that of the
120 native *S. selecta*. On the other hand, we hypothesized that, in hyposaline waters the
121 reproductive ability of *T. v. verticalis* will be lower than that of native species. Finally,
122 we hypothesised that a high active dispersal capacity could contribute to rapid spread of
123 this alien species.

124

125 MATERIALS AND METHODS

126

127 *Study species*

128 *Trichocorixa v. verticalis* is a euryhaline species which can inhabit low salinity waters,
129 but prefers coastal brackish and saline lentic waterbodies, and is even recorded in the
130 open sea (Hutchinson, 1931). In the invaded range, it breeds in permanent water bodies
131 with salinity exceeding 10 g L⁻¹ (Rodríguez-Pérez *et al.*, 2009).

132

133 The three *Sigara* species show different habitat preferences with regard to salinity,
134 ranging from freshwater to saline (see also our results). *Sigara lateralis*, widely spread
135 in the Palaearctic Region (Aukema & Rieger, 1995), is an opportunistic species
136 frequently inhabiting temporary freshwater pools (Millán *et al.*, 1988; Boda & Csabai,
137 2009; Carbonell *et al.*, 2011). It has been described as a migratory species (Macan,
138 1976; Weigelhofer *et al.*, 1992) and is typically captured by light-trapping more often
139 than any other aquatic insect (Popham, 1964; Weigelhofer *et al.*, 1992). *Sigara scripta*,
140 a mainly circum-Mediterranean species (Aukema & Rieger, 1995), inhabits freshwater
141 or hyposaline ponds and stream pools (Carbonell *et al.*, 2011), as well as coastal
142 brackish waters (Moreno *et al.*, 1997; Van de Meutter *et al.*, 2010b). *Sigara selecta*,
143 distributed in Western Europe and Northern Africa (Aukema & Rieger, 1995), inhabits
144 coastal lentic brackish and saline water bodies (Nieser *et al.*, 1994).

145

146 *Conductivity niche analysis*

147 In the Iberian Peninsula, *T. v. verticalis* coexists with native *Sigara* species throughout
148 its conductivity range. The realised conductivity niches of adults and nymphs of the
149 study species were analysed in order to determine the conductivity range where species
150 can coexist, with overlapping niches. When sampling, we follow a multihabitat protocol

151 to estimate relative abundances for adults (232 localities) and nymphs (92 localities) of
152 the four species in the Iberian Peninsula. The information was gathered from the
153 Aquatic Ecology Research Group's Biodiversity dataset at the University of Murcia,
154 and the Wetland Ecology Department's dataset at the EBD-CSIC, Seville (Coccia *et al.*,
155 unpublished data). The realised conductivity niche of the four species was assessed by
156 an Outlying Mean Index (OMI) analysis (Dolédec *et al.*, 2000), using the niche
157 procedure in ADE-4 package (Thioulouse *et al.*, 1997). The OMI, or species
158 marginality index, measures the distance between the mean habitat conditions used by
159 species (the species centroid), and the mean habitat conditions of the sampling area (the
160 origin of the niche hyperspace). OMI analysis places species along habitat gradients, in
161 this case a conductivity gradient. The position of the species depends on their niche
162 deviation from a reference point, which represents neither the mean nor the most
163 abundant species but rather a theoretical ubiquitous species that tolerates the most
164 general habitat conditions (i.e., that is uniformly distributed among habitat conditions).
165 This analysis also calculates niche breadth (tolerance) as a measure of the amplitude in
166 the distribution of each species along the sampled gradient. The species abundance and
167 conductivity were logarithmically transformed prior to analysis, to reduce data
168 variability. This analysis was performed in R v 3.0.2 for Windows (R Core Team,
169 2014).

170

171 *Evaluating tolerance to abiotic conditions: salinity and thermal tolerance*

172

173 Salinity tolerance tests

174 Salinity tolerance tests for adults, nymphs and eggs of *T. v. verticalis* and *S. lateralis*
175 were performed in the laboratory in the spring of 2013 following the methodology of

176 Carbonell *et al.* (2012b), whose data for *S. scripta* and *S. selecta* from spring 2011 were
177 used in this study. In this earlier study, individuals of both species (*S. scripta* and *S.*
178 *selecta*) were collected in the same localities as individuals used for the current study in
179 2013 (see Table 1). Those sites represent the optimum salinity for the species
180 (Carbonell *et al.*, 2012b) and their natural conditions have not changed significantly in
181 recent years.

182

183 For *S. lateralis* and *T. v. verticalis*, about 200 adults and nymphs were collected from
184 two different localities (Table 1) with a hand net and transported in containers with their
185 original water to the laboratory. These two species are easy to distinguish from each
186 other in the field. Both species are different in colouration, size (*S. lateralis* is bigger)
187 and form, *T. v. verticalis* being more elongated (Günter, 2004; Nieser *et al.*, 1994).

188 Before their use in tolerance tests, adults from each species were kept for two days in a
189 5-L aquarium with their original water, held at 20-22°C to acclimatize them to
190 laboratory conditions, and fed daily with frozen chironomid larvae. Nymphs were kept
191 in the same laboratory conditions, and fed with dried microscopic algae (*Tetraselmis*
192 *chuii*) and frozen chironomid larvae. Finally, to obtain eggs, males and females of each
193 species were placed in aquaria and allowed to lay eggs on a plastic mesh that was
194 removed daily and transferred to test vials.

195

196 Five treatments were used in conductivity tests: 0.6, 16, 37, 74 and 100 mS cm⁻¹ EC,
197 which correspond to 0.3, 10, 25, 50 and 75 g L⁻¹ respectively. The solutions were
198 prepared by dissolving marine salt (Ocean Fish, Prodac®. Citadella Pd, Italy) in
199 distilled water. After acclimatizing, adults were placed individually in 50 mL aerated
200 plastic vials filled with solutions in a controlled-temperature chamber (20 ± 1 °C, 12 h

201 light: 12 h dark; no food) (Carbonell *et al.*, 2012b). A total of 60 adults from each
202 species were used in the experiment (ten individuals x five conductivities + ten control
203 individuals in filtered -1.2 μm vacuum filter - water from sampling sites). Survival was
204 monitored daily for one week or until control mortality rate reached 20 %. The same
205 number of nymphs and eggs from each species were used as for adults. Nymphs (third
206 or fourth instars) and eggs (less than 24 hours old) were placed individually in glass
207 vials (5 mL solution) under the same experimental conditions as adults. The nymph's
208 survival was monitored at 24 h and 48 h, and eggs were monitored until hatching or up
209 to 25 days. Hatching time and success of eggs was quantified. Interspecific differences
210 in survival time of adults and nymphs and the hatching time of eggs, as well as
211 differences in survival after 48 h between adults and nymphs, were analysed for each
212 species using Kruskal – Wallis non-parametric tests because these variables did not
213 meet residual normality assumptions (Shapiro–Wilks test, $p < 0.05$). Fisher's post-hoc
214 LSD t-tests were applied using a Bonferroni correction for multiple comparisons to
215 identify significant differences in the response variables among species and conductivity
216 treatments (Quinn & Keough, 2002). Analyses were performed using SPSS for
217 Windows, Rel. 15.0.1. 2006. Chicago: SPSS Inc.

218

219 Thermal tolerance tests

220 Thermal tolerance of *T. v. verticalis* was compared experimentally with the three *Sigara*
221 species with which it can coexist at hyposaline conditions ($16 \text{ mS cm}^{-1} = 10 \text{ g L}^{-1}$), and
222 only with *S. selecta* at mesosaline conditions ($37 \text{ mS cm}^{-1} = 25 \text{ g L}^{-1}$). Thermal tolerance
223 tests were carried out to determine upper and lower thermal limits of adults acclimated
224 at the two selected conductivity treatments.

225

226 Adults were collected from the field in autumn 2013 at the same localities as previous
227 experiments (Table 1) and were directly transferred to 5-L aquaria at experimental
228 conductivities in the laboratory, then kept for 72 h at 20 °C in a controlled-temperature
229 chamber. Organisms were fed with frozen chironomid larvae during the first 48 hours,
230 but no food was provided for 24 h prior to determination of thermal limits because gut
231 contents can modify freezing temperature (Chown & Nicholson, 2004). Following the
232 acclimation period, 80 individuals (40 males and 40 females) of each species were
233 removed from each conductivity treatment and divided into two sub-groups of 40
234 individuals (20 males and 20 females): one used to measure upper thermal limits (UTL)
235 and the other to measure lower thermal limits (LTL). Although a number of potential
236 end-points exist for tolerance both to heat and cold (Chown & Nicolson, 2004), we
237 considered the *supercooling* point (SCP) or point before the freezing temperature
238 (Wilson *et al.*, 2003) as LTL and the *heat coma* point or point of paralysis prior to death
239 and preceded by spasmodic leg movement as the UTL (Chown & Terblanche, 2007).
240 Although death was easily identifiable in upper thermal tolerance experiments, defining
241 lower lethal limits proved more difficult because individuals exhibiting total paralysis
242 would revive and recover full or partial locomotor abilities shortly after the end of the
243 exposure period. Thus, SPC was selected as the comparable objective variable among
244 species for LTL, since freezing temperature is largely unaffected by cooling rate
245 (Chown & Nicholson, 2004).

246

247 Prior to testing, individuals were removed from their acclimation aquaria, washed with
248 distilled water, dried on absorbent paper, then placed upside down tied by their wings
249 onto a clean, dry rectangular piece of pottery using a small piece of adhesive tape. This
250 procedure keeps individuals motionless during the experiment. Tests were carried out in

251 air in a controlled-temperature chamber (BINDER MK53. BINDER GmbH, Tuttlingen,
252 Germany), employing a dynamic method using a ramping rate of $\pm 1 \text{ }^\circ\text{C min}^{-1}$, from
253 acclimation temperature (20 $^\circ\text{C}$) until the upper (70 $^\circ\text{C}$) or lower temperature limit (-45
254 $^\circ\text{C}$). An identical ramping rate was employed in previous studies, thus allowing
255 comparison amongst treatments and taxa (Calosi *et al.*, 2008; Sánchez-Fernández *et al.*,
256 2010; Arribas *et al.*, 2012; Coccia *et al.*, 2013). To determine UTL and LTL, the body
257 surface temperature was measured throughout the experiment using an infrared video
258 camera (FLIR A325sc. FLIR Systems Inc., Wilsonville, USA) coupled to the climatic
259 chamber, at a distance of c. 0.4 m from the insects. The infrared camera converted body
260 surface radiation into a thermal pattern and was capable of detecting temperature
261 differentials of 0.1 $^\circ\text{C}$. As a supplement, the insects were recorded throughout the
262 experiment using a video camera (Sony DCR-DVD110E. Sony Co., Tokyo, Japan)
263 synchronised with the infrared camera, From video images, the moment at which
264 movement of legs ceased was determined and the body surface temperature at that
265 moment was obtained from infrared thermal images. Infrared recordings were analysed
266 with ThermoCAM™ Researcher Pro 2.10 software. Differences in UTL and LTL were
267 examined by ANOVA considering acclimation conductivity, species and sex as fixed
268 factors, together with their interactions. As preliminary analyses showed no differences
269 between sexes (UTL: $F = 0.01$, $P = 0.932$; LTL: $F = 1.21$, $P = 0.273$), sex was removed
270 from further analysis.

271

272 Evaluating the reproductive capacity: species fecundity tests

273 Fecundity tests were carried out on adults of the four species in autumn 2013. The
274 oviposition rate of *T. v. verticalis* was compared experimentally with the three *Sigara*

275 species with which it can coexist at hyposaline conditions (16 mS cm^{-1}), and only with
276 *S. selecta* at mesosaline conditions (37 mS cm^{-1}).

277

278 The oviposition rate of species (expressed in eggs day⁻¹) was determined under
279 laboratory conditions ($20 \pm 2 \text{ }^\circ\text{C}$, 12 h light: 12 h dark). Adults were collected from the
280 field (Table 1) and 10 pairs (one female and one male) of each species were directly
281 transferred to 500 ml aquaria (one pair per aquarium) filled with salt solutions and fed
282 daily with frozen chironomid larvae. A piece of plastic mesh was placed in each
283 aquarium as an oviposition site. Egg production was monitored daily until the female
284 died. If the male died before the female, it was replaced. Eggs laid during the first two
285 days were excluded before calculating oviposition rate, to minimise the effect of any
286 females carrying eggs when collected in the field. Differences in oviposition rate were
287 examined by ANOVA, considering acclimation conductivity and species as fixed
288 factors.

289

290 Evaluating dispersal ability: wing morphometry measures

291 Morphometric measurements of metathoracic wings (flight wings) were taken in order
292 to compare active dispersal ability among study species. Measurements were taken from
293 individuals collected in the field and preserved in 100% alcohol at $-20 \text{ }^\circ\text{C}$ (Arribas *et al.*,
294 2012). Forty individuals (20 males and 20 females) of each species were dried at $60 \text{ }^\circ\text{C}$
295 in an oven (JP Selecta Prebatem. JP Selecta, Barcelona, Spain) for 24 h, and dry weight
296 was measured using a digital balance (Sartorius BP221S. Sartorius AG, Goettingen,
297 Germany) to $\pm 0.0001 \text{ g}$. Measured individuals were kept in hot water (about $80 \text{ }^\circ\text{C}$) for
298 1 minute to rehydrate them and facilitate manipulation of the wings. After rehydration,
299 the left metathoracic wing was removed, teased open, stained with chlorazol black

300 (Martínez, 2002), and mounted in 50% dimethyl hydantoin formaldehyde (DMHF) on a
301 microscope slide restoring their natural size and shape. Wings and individuals were
302 photographed under a Zeiss Stemi 2000C Trinocular Zoom Stereomicroscope (Zeiss,
303 Thornwood, NY, USA), and subsequently length, width and area of the wing were
304 measured using a Spot Insight Firewire digital camera (Spot Imaging Solutions, Sterling
305 Heights, MI, USA) and associated software. Raw measures of body mass and wing
306 were used to derive wing loading (body mass/wing area) and wing aspect ratio (wing
307 length/wing width) (Gibb *et al.*, 2006; Arribas *et al.*, 2012). For both indices, species
308 and sex, together with their interaction, were considered as fixed factors in ANOVA.
309 All ANOVA analyses were conducted using SPSS for Windows, Rel. 15.0.1. 2006.
310 Chicago: SPSS Inc.

311

312 RESULTS

313 *Realised conductivity niche*

314 OMI results revealed an overlapping conductivity niche among adults of the study
315 species (Fig. 1a). The exotic species showed a wide niche, similar to that of *S. selecta*
316 although with a lower conductivity optimum, occurring from subsaline to hypersaline
317 waters (1.03-120 mS cm⁻¹), but was more frequent in mesosaline waters. *Sigara selecta*
318 had a niche in the upper part of the conductivity gradient, being an exclusive inhabitant
319 of brackish and saline waters (16.6-200 mS cm⁻¹). *Sigara scripta* adults occurred along
320 a broad conductivity gradient (0.25-41.6 mS cm⁻¹), with a niche midway in the overall
321 conductivity range. *Sigara lateralis* presented the lowest conductivity tolerance,
322 occurring from fresh to hyposaline waters (0.15-33.6 mS cm⁻¹). Nymphal stages
323 presented less conductivity tolerance and lower niche breadth than conspecific adults,

324 although with similar overlap in niches among species along the conductivity gradient
325 (Fig. 1b).

326

327 *Salinity tolerance*

328 There were significant differences in salinity tolerance among species ($H = 53.49$, $p <$
329 0.001) (Figure 2 and Table S1.1). Conductivity had significant effects on adult and
330 nymph survival times, and egg hatching time. *Trichocorixa v. verticalis* adults survived
331 until 37 mS cm^{-1} , but did not show significant differences in survival time with *S.*
332 *lateralis* or *S. scripta* at lower conductivities (0.6 and 16 mS cm^{-1}). Adults of the most
333 tolerant species, *S. selecta*, were able to survive along the entire conductivity gradient
334 with a higher survival time than other species (Figure 2a).

335

336 Differences between *T. v. verticalis* nymphs and adults in survival time were not
337 significant ($H = 3.34$, $p = 0.068$). The same pattern was found for *S. lateralis* ($H = 2.12$,
338 $p = 0.146$) and *S. scripta* ($H = 0.042$, $p = 0.838$). Only *S. selecta* ($H = 16.33$, $p < 0.001$)
339 showed significant differences between nymphs and adults in survival time. Nymphs of
340 *T. v. verticalis* and *S. selecta* survived up to 37 mS cm^{-1} and did not differ significantly
341 in survival time at that conductivity (Figure 2b). Nymphs of *S. lateralis* and *S. scripta*
342 survived until 16 mS cm^{-1} , although *S. lateralis* was significantly less tolerant than other
343 species at that conductivity.

344

345 *Trichocorixa v. verticalis* eggs showed the highest conductivity tolerance, being the
346 only species hatching up to 74 mS cm^{-1} , when mean hatching time was 20.9 days
347 (Figure 2c). Their hatching success was high (80%) at lower conductivities, decreasing
348 with increasing conductivity (Table 2). Eggs of *S. selecta* hatched with 100% success at

349 conductivities up to 37 mS cm⁻¹, showing no significant differences in hatching time
350 with *T. v. verticalis* eggs (Table 2). Eggs of *S. lateralis* and *S. scripta* hatched until 16
351 mS cm⁻¹, *S. lateralis* showing the shortest hatching time and the highest hatching
352 success at lower conductivities (0.6 and 16 mS cm⁻¹).

353

354 *Thermal tolerance*

355 Upper and lower thermal limits showed significant differences among species (UTL: F
356 = 10575.54, $p < 0.001$; LTL: $F = 84.37$, $p < 0.001$) but not between sexes, and both
357 limits were significantly dependent on acclimation conductivity (UTL: $F = 29014.84$, p
358 < 0.001 ; LTL: $F = 491.48$, $p < 0.001$) (Table S1.2). Significant conductivity x species
359 interactions were found for both thermal limits (UTL: $F = 9249.39$, $p < 0.001$; LTL: F
360 = 69.09, $p < 0.001$). At 16 mS cm⁻¹ *T. v. verticalis* presented the highest UTL (51.5 °C ±
361 0.05) with a similar value to *S. scripta* (50.4 °C ± 0.13), and significantly higher than *S.*
362 *selecta* (49.6 °C ± 0.07) and *S. lateralis* (47.5 °C ± 0.14) (Figure 3a). However, at 37 mS
363 cm⁻¹ UTL did not differ between *T. v. verticalis* and *S. selecta* (Figure 3a), but
364 decreased in *T. v. verticalis* when conductivity increased. LTL for *T. v. verticalis* (-8.9
365 °C ± 0.17) was weaker than for *S. selecta* (-10.8 °C ± 0.14), and similar to that of *S.*
366 *lateralis* and *S. scripta* at 16 mS cm⁻¹ (Figure 3b). As for UTL, LTL decreased as
367 conductivity increased for both *T. v. verticalis* and *S. selecta*.

368

369 *Oviposition rates*

370 Oviposition rates differed significantly among species ($F = 3.40$, $p = 0.024$) (Table
371 S1.3). The oviposition rate of *T. v. verticalis* was significantly higher than that of *Sigara*
372 species at 37 mS cm⁻¹ (Figure 4) but not at 16 mS cm⁻¹. *Trichocorixa v. verticalis*
373 oviposition rate increased from 1.6 eggs day⁻¹ at 16 mS cm⁻¹ to 2.2 eggs day⁻¹ at 37 mS

374 cm^{-1} , in contrast to what was observed for *S. selecta* (0.2 eggs day^{-1} at both
375 conductivities). *Trichocorixa v. verticalis* females were able to lay up to 29 eggs at 16
376 mS cm^{-1} in one day, whilst the maximum number of eggs laid by a *Sigara* female in
377 one day was six (for *S. scripta* and *S. selecta*).

378

379 *Wing morphology*

380 Differences among species (WL: $F = 12.81, p < 0.001$; WAR: $F = 28.98, p < 0.001$) and
381 between sexes (WL: $F = 40.52, p < 0.001$; WAR: $F = 6.07, p = 0.015$) were significant
382 for both wing loading and wing aspect ratio (Figure 5 and Table S1.4). *Trichocorixa v.*
383 *verticalis* (particularly males) had the lowest wing loading values, although significant
384 differences were only found with *S. scripta* which had the highest value (Figure 5a).
385 *Trichocorixa v. verticalis* also presented the highest wing aspect ratio (Figure 5b),
386 indicating a longer and narrower wing shape compared to the broader and more rounded
387 wings for *Sigara* species (Figure 5c). Females presented significantly higher wing
388 loading than males except for *S. lateralis*, where no differences between sexes were
389 found. For wing aspect ratio, significant differences between sexes were only found for
390 *S. selecta*, males showing a higher ratio than females.

391

392 DISCUSSION

393 Of the studied species, the invasive *T. v. verticalis* had higher values for traits related
394 with establishment (i.e. the highest oviposition rate at mesosaline conditions, and
395 highest conductivity tolerance of eggs) and spread (i.e. the highest wing aspect ratio and
396 the lowest wing loading in males) phases of invasion. This combination of traits
397 together with wide physiological tolerance to salinity and temperature (Table 2) may
398 largely explain the high rates of survival and reproduction of the introduced species in

399 Doñana, and the recent geographical expansion of this species (Carbonell *et al.*, 2012a;
400 Guareschi *et al.*, 2015). These results are consistent with the characteristics attributed to
401 successful invaders, considered to be species which have some combination of high
402 dispersal ability, high reproductive output, generalist feeding, and broad environmental
403 tolerance (Havel *et al.*, 2015).

404

405 In subsaline and hyposaline conditions (0.6 and 16 mS cm⁻¹, respectively), *T. v.*
406 *verticalis* showed similar survival times to native species for adults and nymphs, but
407 showed the highest tolerance to elevated temperature (at 16 mS cm⁻¹ *T. v. verticalis*
408 presented the highest UTL). This finding is consistent with the thermal tolerance
409 previously described for *T. v. verticalis* at different acclimation temperatures and
410 salinities (Coccia *et al.* 2013). However, *S. lateralis* is probably the most competitive
411 species at low conductivities, because it had a similar oviposition rate but the shortest
412 hatching time and the highest hatching success (100%).

413

414 In mesosaline conditions, *T. v. verticalis* may have a competitive advantage over *S.*
415 *selecta* because of its higher oviposition rate, although these two species showed similar
416 salinity tolerance and upper thermal limits at both adult and nymph stages. However,
417 the invasive species was less tolerant to low temperature than *S. selecta*. According to
418 our results, the ability of *T. v. verticalis* to deal with high and low temperatures
419 decreases at high salinities, reducing its thermal range. Climatic models of its potential
420 distribution (Guareschi *et al.*, 2013) appear to support our results. These models indicate
421 that *T. v. verticalis* prefers areas with low variability in climatic conditions, such as
422 coastal areas, in its potential invasive distribution range. This could be due to the

423 limited capacity of adults to withstand extreme cold in continental areas and at high
424 altitudes (Guareschi *et al.*, 2013).

425

426 At the highest conductivity (74 mS cm^{-1}), *T. v. verticalis* out-performed *S. selecta* due to
427 higher salinity tolerance and hatching success of eggs in chloride-rich waters, despite
428 the greater adult salinity tolerance of the native species. Previous laboratory studies
429 (Carbonell *et al.*, 2012b and unpublished data) found that both species showed wide
430 tolerance to different anionic composition, although the invasive species is restricted to
431 coastal wetlands where salts, derived from seawater, are dominated by NaCl. The native
432 species (*S. selecta*) is also capable of occupying inland waters with higher sulphate
433 concentrations than coastal wetlands (Carbonell *et al.*, 2012b), and is apparently more
434 sulphate tolerant.

435

436 Many aquatic species have adaptations that allow persistence in temporary
437 environments, such as dormant stages and spores in prokaryotes, seeds of aquatic plants
438 or resting eggs in zooplankton (Havel *et al.*, 2015). Therefore, production of eggs
439 resistant to high salinities could be advantageous in temporary saline pools. In these
440 habitats, evaporation produces a marked increase in salinity, thus the production of
441 resistant eggs to drought and high salinity conditions is a useful strategy that permits
442 hatching when water levels increase and salinity decreases (Williams, 2006). However,
443 there is no evidence of drought tolerance of eggs of *T. v. verticalis* in the invaded area
444 (unpublished data). In their native range, two other subspecies of *T. verticalis* (*T. v.*
445 *interiores* and *T. v. sellaris*) produce diapausing eggs that allow them to survive in ice,
446 hypersaline waters or temporary pools (Tones, 1977; Kelts, 1979).

447

448 Resistant eggs coupled with the higher egg production observed in our study and the
449 ability to breed throughout the year in the invaded area (Rodríguez-Pérez *et al.*, 2009)
450 are biological traits common in invasive species (Havel *et al.* 2015) that may promote
451 the establishment and dispersal of *T. v. verticalis* in new areas and the displacement of
452 *S. selecta* from mesosaline to hypersaline wetlands, as apparently observed in Morocco
453 (L'Modhi *et al.*, 2010). In Doñana wetlands, the invasive species attains the highest
454 population densities in waters of intermediate conductivities of 17 - 55 mS cm⁻¹
455 (Rodríguez-Pérez *et al.*, 2009), overlapping with the conductivity range of *S. selecta*,
456 except in extreme salinities where *T. v. verticalis* could be excluded by the lower
457 salinity tolerance of the adult stage. However, the lower part of the *T. v. verticalis*
458 salinity niche overlaps with the upper part of the *S. scripta* and *S. lateralis* niches,
459 where these native species could be less competitive.

460

461 Although the distribution of *T. v. verticalis* in the invaded area can be explained largely
462 by its wide physiological tolerance to salinity, biological interactions with native
463 hemiptera and other macroinvertebrates (competition, predation or even parasitism) may
464 also be important determinants of its abundance in specific habitats, as in other corixid
465 species (Scudder, 1983). The predation rate on *T. verticalis* by Odonata larvae was
466 higher than that of *S. lateralis*, probably because of the smaller size of the alien (Coccia
467 *et al.*, 2014). This might help to explain the rarity of *T. v. verticalis* in Doñana wetlands
468 of lower conductivity, where Odonata larvae are abundant. In addition, *T. v. verticalis*
469 showed higher parasitism by the water mites *Hydrachna skorikowi* and *Eylais*
470 *infundibulifera* than *S. lateralis* and *S. scripta* in low salinity Doñana wetlands (Sánchez
471 *et al.* 2015). Thus, *T. v. verticalis* is under higher parasitic and predation pressure than
472 native *Sigara* species at low salinities, which may help to explain why *T. v. verticalis* is

473 particularly successful in saline habitats where parasitic mites and Odonata larvae are
474 absent.

475

476 The survival of invasive species under harsh environmental conditions or biological
477 interactions is only half of the invasion story. **Resting stages** and/or adults must also
478 move from one habitat to others. With the exception of flying insects and mobile
479 vertebrates, most aquatic species lack the means for active dispersal into isolated
480 hydrological catchments and instead use a variety of transport vectors for passive
481 dispersal, such as transport by wind or vertebrates such as birds (Havel & Shurin, 2004).

482 In the case of *T. v. verticalis*, its rapid spread to nearby wetlands may be promoted by
483 its high capacity for active and passive dispersal. *Trichocorixa v. verticalis* presents the
484 lowest wing loading of males, related to superior flying ability (Berwaerts *et al.*, 2002)
485 and the highest wing aspect ratio, related with fast-flapping and energy efficient active
486 flight (Hoffsten, 2004). Thus, wing morphometrics suggest that *T. v. verticalis* may be a
487 stronger flier than native species, which would facilitate its ability to move between
488 water bodies to breed or to avoid unfavorable environmental conditions, especially
489 when temporary habitats dry out. In general, corixids disperse actively to suitable
490 aquatic habitats mainly between early spring, when individuals fly to new potential
491 breeding areas, and late fall when the population density peaks and habitats begin to dry
492 out (Pajunen & Jansson, 1969; Boda & Csabai, 2009). The most suitable periods for
493 dispersal are dawn and late evening hours, which are usually free of wind and rain, with
494 a higher relative air humidity and lower air temperature (Boda & Csabai, 2009).

495 Besides, males of all the studied species, except *S. lateralis*, show greater flight ability
496 than females so as to actively move among water bodies, probably to search for females.

497 In general, corixid females, larger than males, invest more in reproduction and less in
498 dispersal (i.e. oogenesis flight syndrome) when environmental conditions are favorable,

499 owing to a trade-off between the high energy costs of egg production and of flight
500 (Velasco *et al.*, 1990). In contrast, males invest more energy in developing flight
501 muscles to actively search for females.

502

503 In conclusion, the establishment and spread success of the alien *T. v. verticalis* could be
504 principally due to its high egg production, high resistance to salinity, heat and harsh
505 environmental conditions, and its better flight ability. Further studies of other biological
506 traits, such as voltinism, development rate, diet and body size differences should
507 contribute to define its functional and ecological niche and a better understanding of the
508 establishment success and impact of the species in the invaded area.

509

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523

524 REFERENCES

- 525 Alonso, A. & J. A. Camargo, 2003. Short-term toxicity of ammonia, nitrite and nitrate
526 to the aquatic snail *Potamopyrgus antipodarum* (Hydrobiidae, Mollusca). *Bulletin of*
527 *Environmental Contamination and Toxicology* 70: 1006–1012.
- 528 Alonso, A. & J. A. Camargo, 2004. Sub-lethal responses of the aquatic snail
529 *Potamopyrgus antipodarum* (Hydrobiidae, Mollusca) to unionised ammonia: a tolerant
530 invading species. *Fresenius Environmental Bulletin* 13: 607–615.
- 531 Arribas, P., Velasco, J., Abellán, P., Sánchez-Fernández, D., Andújar, C., Calosi, P.,
532 Millán, A., Ribera, I. & D. Bilton, 2012. Dispersal ability rather than ecological
533 tolerance drives differences in range size between lentic and lotic water beetles
534 (Coleoptera: Hydrophilidae). *Journal of Biogeography* 39: 984–994.
- 535 Aukema, B. & C. Rieger, 1995. *Catalogue of the Heteroptera of the Palearctic Region.*
536 Vol. 1: Enicocephalomorpha, Dipsocoromorpha, Nepomorpha, Gerromorpha and
537 Leptopodomorpha. Netherlands Entomological Society. Amsterdam.
- 538 Berwaerts, K., Van Dyck, H. & P. Aerts, 2002. Does flight morphology relate to flight
539 performance? An experimental test with the butterfly *Pararge aegeria*. *Functional*
540 *Ecology* 16: 484–491.
- 541 Blackburn, T. M., Pyšek, P., Bacher, S., Carlton, J. T., Duncan, R. P., Jarošík, V.,
542 Wilson, J. R. U. & D. M. Richardson, 2011. A proposed unified framework for
543 biological invasions. *Trends in Ecology and Evolution* 26: 333–339.
- 544 Boda, P. & Z. Csabai, 2009. Seasonal and diel dispersal activity characteristics of
545 *Sigara lateralis* (Leach, 1817) (Heteroptera: Corixidae) with special emphasis on
546 possible environmental factors and breeding state. *Aquatic Insects* 31: 301–314.
- 547 Calosi, P., Bilton, D. T. & J. I. Spicer, 2008. Thermal tolerance, acclimatory capacity
548 and vulnerability to global climate change. *Biology Letters* 4: 99–102.

549 Carbonell, J. A., Guareschi, S., Coccia, C., Sánchez-Fernández, D., Velasco, J., Boyero,
550 L., Green, A. J. & A. Millán, 2012a. Distribución de *Trichocorixa verticalis verticalis*
551 (Fieber, 1851) (Heteroptera: Corixidae) a nivel mundial y su expansión en la Península
552 Ibérica. (2012) En: GEIB Grupo Especialista en Invasiones Biológicas (ed) EEI 2012
553 Notas Científicas. pp. 148-152. 4º Congreso Nacional sobre Especies Exóticas
554 Invasoras “EEI 2012”. GEIB, Serie Técnica Nº 5. León, 218 pp.

555 Carbonell, J. A., Gutiérrez-Cánovas, C., Bruno, D., Abellán, P., Velasco, J. & A.
556 Millán, 2011. Ecological factors determining the distribution and assemblages of the
557 aquatic hemiptera (Gerromorpha & Nepomorpha) in the Segura river basin (Spain).
558 *Limnetica* 30: 59–70.

559 Carbonell, J. A., Millán, A. & J. Velasco, 2012b. Concordance between realised and
560 fundamental niches in three Iberian *Sigara* species (Hemiptera: Corixidae) along a
561 gradient of salinity and anionic composition. *Freshwater biology* 57: 2580–2590.

562 Castro-Longoria, E. 2003. Egg production and hatching success of four *Acartia* species
563 under different temperature and salinity regimes. *Journal of Crustacean Biology* 23:
564 289–299.

565 Chown S. L. & W. N. Nicolson, 2004. *Insect Physiological Ecology: Mechanisms and*
566 *Patterns*. Oxford University Press, U.K.

567 Chown, S. L. & J. S. Terblanche, 2007. Physiological diversity in insects: ecological
568 and evolutionary contexts. *Advances in Insect Physiology* 33: 50–152.

569 Coccia, C., Boyero, L. & A. J. Green, 2014. Can differential predation of native and
570 alien corixids explain the success of *Trichocorixa verticalis verticalis* in the Iberian
571 Peninsula? *Hydrobiologia* 734: 115–123.

572 Coccia, C., Calosi, P., Boyero, L., Green, A. J. & D. T. Bilton, 2013. Does
573 Ecophysiology Determine Invasion Success? A Comparison between the Invasive

574 Boatman *Trichocorixa verticalis verticalis* and the Native *Sigara lateralis* (Hemiptera,
575 Corixidae) in South-West Spain. PLoS ONE, 8(5): e63105.
576 doi:10.1371/journal.pone.0063105.

577 Dolédec, S., Chessel, D. & C. Gimaret-Carpentier, 2000. Niche separation in
578 community analysis: a new method. Ecology 81: 2914–2927.

579 Dudgeon, D., Arthington, A. H., Gessner, M. O., Kawabata, Z., Knowler, D. J.,
580 Lévêque, C., Naiman, R. J., Prieur-Richard, A. H., Soto, D., Stiassny, M. L. J. & C. A.
581 Sullivan, 2006. Freshwater biodiversity: Importance, threats, status and conservation
582 challenges. Biological reviews of the Cambridge Philosophical Society 81: 163–182.

583 Espinar, J. L., Díaz-Delgado, R., Bravo-Utrera, M. A. & M. Vilà, 2015. Linking *Azolla*
584 *filioloides* invasion to increased winter temperatures in the Doñana marshland (SW
585 Spain). Aquatic Invasions 10. doi: <http://dx.doi.org/10.3391/ai.2015.10.1.02>

586 Figuerola, J., Green, A. J. & L. Santamaria, 2003. Passive internal transport of aquatic
587 organisms by waterfowl in Doñana, south–west Spain. Global Ecology and
588 Biogeography 12: 427–436.

589 Gelin, A., Crivelli, A. J., Rosecchi, E. & P. Kerambrun, 2001. Can salinity changes
590 affect reproductive success in the brown shrimp *Crangon crangon*? Journal of
591 Crustacean Biology 21: 905–911.

592 Gérard, C., Blanc, A. & K. Costil, 2003. *Potamopyrgus antipodarum*
593 (Mollusca:Hydrobiidae) in continental aquatic gastropod communities: impact of
594 salinity and trematode parasitism. Hydrobiologia 493: 167–172.

595 Gibb, H., Hjalten, J., Ball, J. P., Pettersson, R. B., Landin, J., Alvini, O. & K. Danell,
596 2006. Wing loading and habitat selection in forest beetles: are red-listed species poorer
597 dispersers or more habitat-specific than common congeners? Biological Conservation
598 132: 250–260.

599 Guareschi, S., Abellán, P., Laini, A., Green, A. J., Sánchez-Zapata, J. A., Velasco, J. &
600 A. Millán, 2015. Cross-taxon congruence in wetlands: Assessing the value of waterbirds
601 as surrogates of macroinvertebrate biodiversity in Mediterranean Ramsar sites.
602 Ecological Indicators 49: 204–215.

603 Guareschi, S., Coccia, C., Sánchez-Fernández, D., Carbonell, J. A., Velasco, J., Boyero,
604 L., Green, A. J. & A. Millán, 2013. How Far Could the Alien Boatman *Trichocorixa*
605 *verticalis verticalis* Spread? Worldwide Estimation of Its Current and Future Potential
606 Distribution. PLoS ONE 8(3): e59757. doi:10.1371/journal.pone.0059757.

607 Günther, H. 2004. *Trichocorixa verticalis verticalis* (Fieber), eine nearktische
608 Ruderwanze in Europa (Heteroptera: Corixidae). Mitteilungen des Internationalen
609 Entomologischen Verereins 29: 45–49.

610 Havel, J.E., Kovalenko, K.E., Thomaz, S.M., Amalfitano, S. & L. B. Kats, 2015.
611 Aquatic invasive species: challenges for the future. Hydrobiologia 750: 147–170.

612 Havel, J. E. & J. B. Shurin, 2004. Mechanisms, effects, and scales of dispersal in
613 freshwater zooplankton. Limnology and Oceanography 49: 1229–1238.

614 Hoffsten, P. 2004. Site-occupancy in relation to flight-morphology in caddisflies.
615 Freshwater Biology 49: 810–817.

616 Hutchinson, G. E. 1931. On the occurrence of *Trichocorixa kirkaldy* (Corixidae,
617 Hemiptera-Heteroptera) in salt water and its zoo-geo-graphical significance. American
618 Naturalist 65: 573–5749.

619 Jansson, A. 2002. New records of Corixidae (Heteroptera) from northeastern USA and
620 eastern Canada, with one new synonymy. Entomologica Fennica 13: 85–88.

621 Kelts, L. J. 1979. Ecology of a tidal marsh corixid, *Trichocorixa verticalis* (Insecta,
622 Hemiptera). Hydrobiologia 64: 37–57.

623 L'Mohdi, O., Bennis, N., Himmi, O., Hajji, K., El Haissoufi, M., Hernando, C.,
624 Carbonell, J. A. & A. Millán, 2010. *Trichocorixa verticalis verticalis* (Fieber 1851)
625 (Hemiptera, Corixidae): une nouvelle espeeie exotique au Maroc. Boletín de la
626 Sociedad Entomológica Aragonesa 46: 395–400.

627 Macan, T. T. 1976. A revised key to the British Water Bugs (Hemiptera-Heteroptera)
628 with notes on their Ecology. Freshwater Biological Association Scientific Publication,
629 16, Ambleside: Freshwater Biological Association, 77p.

630 Martínez, I. 2002. Técnicas básicas de anatomía microscópica y de morfometría para
631 estudiar los insectos. Boletín de la Sociedad Entomológica Aragonesa 30: 187–195.

632 Millán, A., Hernando, C., Aguilera, P., Castro, A. & I. Ribera, 2005. Los coleópteros
633 acuáticos y semiacuáticos de Doñana: Reconocimiento de su biodiversidad y
634 prioridades de conservación. Boletín de la Sociedad Entomológica Aragonesa 36: 157–
635 164.

636 Millán, A., Velasco, J., Nieser, N. & C. Montes, 1988. Heterópteros acuáticos
637 (Gerromorpha & Nepomorpha) de la Cuenca del Río Segura. S.E. de la Península
638 Ibérica. Anales de Biología 11: 71–80.

639 Moreno, J. L., Millán, A., Suarez, M. L., Vidal-Abarca, R. & J. Velasco, 1997. Aquatic
640 Coleoptera and Heteroptera assemblages in waterbodies from ephemeral coastal streams
641 (“ramblas”) of south-eastern Spain. Archiv für Hydrobiologie 141: 93–107.

642 Nieser, N., Baena, M., Martínez-Avilés, J. & A. Millán, 1994. Claves para la
643 identificación de los heterópteros acuáticos (nepomorpha & gerromorpha) de la
644 Península Ibérica –Con notas sobre las especies de las Islas Azores, Baleares, Canarias
645 y Madeira. Asociación Española de Limnología. Madrid.

646 Pajunen, V. I. & A. Jansson, 1969. Dispersal of the rock pool corixids *Arctocorixa*
647 *carinata* (Sahlb.) and *Callicorixa producta* (Reut.) (Heteroptera, Corixidae). *Annales*
648 *Zoologici Fennici* 6: 391–427.

649 Parker, I. M., Simberloff, D., Lonsdale, W. M., Goodell, K., Wonham, M., Kareiva, P.
650 M., Williamson, M. H., Von Holle, B., Moyle, P. B., Byers, J. E. & L. Goldwasser,
651 1999. Impact: toward a framework for understanding the ecological effects of invaders.
652 *Biological Invasions* 1: 3–9.

653 Popham, E. J. 1964. The migration of aquatic bugs with special reference to the
654 Corixidae (Hemiptera Heteroptera). *Archiv für Hydrobiologie* 60: 450–496.

655 Rabitsch, W. 2008. Alien true bugs of Europe (Insecta: Hemiptera: Heteroptera).
656 *Zootaxa* 1827: 1–44.

657 Quinn, G. & Keough, 2002. *Experimental Design and Data Analysis for Biologists*.
658 Cambridge University Press. Cambridge. 537 pp.

659 R Core Team. 2014. R: A language and environment for statistical computing. R
660 Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0, URL
661 <http://www.R-project.org>.

662 Ricciardi, A. 2006. Are modern biological invasions an unprecedented form of global
663 change? *Conservation Biology* 21: 329–336.

664 Richards, D. C. 2002. The New Zealand mudsnail invades the Western United States.
665 *Aquatic Nuisance Species* 4: 42–44.

666 Rodríguez-Pérez, H., Florencio, M., Gómez-Rodríguez, C., Green, A.J., DÍa-Paniagua,
667 C. & L. Serrano, 2009. Monitoring the invasion of the aquatic bug *Trichocorixa*
668 *verticalis verticalis* (Hemiptera: Corixidae) in the wetlands of Doñana National Park
669 (SW Spain). *Hydrobiologia* 634: 209–217.

670 Rodríguez-Pérez, H. & A. J. Green, 2012. Strong seasonal effects of waterbirds on
671 benthic communities in shallow lakes. *Freshwater Science* 31: 1273–1288.

672 Sakai, A. K., Allendorf, F. W., Holt, J. S., Lodge, D. M., Molofsky, J., With, K. A.,
673 Baughman, S., Cabin, R. J., Cohen, J. E., Ellstrand, N. C., McCauley, D. E., O'Neil, P.,
674 Parker, I. M., Thompson, J. N. & S. G. Weller, 2001. The population biology of
675 invasive species. *Annual Review of Ecology and Systematics* 32: 305–332.

676 Sala, J. & D. Boix, 2005. Presence of the nearctic water boatman *Trichocorixa*
677 *verticalis verticalis* (Fieber, 1951) (Heteroptera, Corixidae) in the Algarve region (S
678 Portugal). *Graellsia* 61: 31–36.

679 Sánchez-Fernández, D., Calosi, P., Atfield, A., Arribas, P., Velasco, J., Spicer, J. I.,
680 Millán, A. & D. T. Bilton, 2010. Reduced salinities compromise the thermal tolerance
681 of hypersaline specialist diving beetles. *Physiological Entomology* 35: 265–273.

682 Sánchez, M., Coccia, C., Valdecasas, A. G., Boyero, L. & A. J. Green, 2015. Parasitism
683 by water mites in native and exotic Corixidae: Are mites limiting the invasion of the
684 water boatman *Trichocorixa verticalis* (Fieber, 1851)? *Journal of Insect Conservation*.
685 Doi: 10.1007/s10841-015-9764-7

686 Scudder, G. G. E. 1983. A review of factors governing the distribution of two closely
687 related corixids in the saline lakes of British Columbia. *Hydrobiologia* 105: 143–154.

688 Simonis, J. L. 2013. Predator ontogeny determines trophic cascade strength in
689 freshwater rock pools. *Ecosphere* 4: 1–25.

690 Thioulouse, J., Chessel, D., Dolédec, S. & J. M. Olivier, 1997. ADE-4: a multivariate
691 analysis and graphical display software. *Statistics and Computing* 7: 75–83.

692 Tones, P. I., 1977. Life-cycle of *Trichocorixa verticalis interiores* Sailer (Hemiptera:
693 Corixidae) with special reference to diapause. *Freshwater Biology* 7: 31–36.

694 Van de Meutter, F., Trekels, H. & A. Green, 2010a. The impact of the North American
695 waterbug *Trichocorixa verticalis* (Fieber) on aquatic macroinvertebrate communities in
696 southern Europe. *Fundamental and Applied Limnology* 177: 283–292.

697 Van de Meutter, F., Trekels, H., Green, A. & R. Stoks, 2010b. Is salinity tolerance the
698 key to success for the invasive water bug *Trichocorixa verticalis*? *Hydrobiologia* 649:
699 231–238.

700 Velasco, J., Millán, A. & N. Nieser, 1990. Observaciones sobre la colonización y el
701 ciclo de vida de *Heliocorisa vermiculata* (puton, 1874) (Heteroptera, Corixidae) en
702 pequeños estanques del se español. *Limnetica* 6: 101–108.

703 Weigelhofer G., Weissmair, W., & J. Waringer, 1992. Night migration activity and the
704 influence of meteorological parameters on light-trapping for aquatic Heteroptera.
705 *Zoologischer Anzeiger* 229: 209–218.

706 Williams, D. D., 2006. *The biology of temporary waters*. Oxford, Biology, 337 pp.

707 Wilson, P. W., Heneghan, A. F. & A. D. J. Haymet, 2003. Ice nucleation in nature:
708 supercooling point (SCP) measurements and the role of heterogeneous nucleation.
709 *Cryobiology* 46: 88–98.

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715 FIGURE LEGENDS

716

717 **Fig. 1.** Realised niche breadths of four study species in relation to conductivity for a)
718 adults and b) nymphs. X-axis represents the range of conductivity in the study area.
719 Tick marks represent conductivities of sampling sites included in the analysis. *Mean
720 conductivity of sampling sites.

721

722 **Fig. 2.** Mean \pm SE survival time of adults, nymphs and eggs in conductivity treatments.
723 According to post hoc analysis with Bonferroni correction, different numbers above
724 bars indicate significant differences ($p \leq 0.05$) in survival time among species in a given
725 conductivity treatment, whereas different letters above bars indicate significant
726 differences in survival time among conductivity treatments for a given species.

727

728 **Fig. 3.** Mean \pm SE a) upper thermal limits and b) lower thermal limits in conductivity
729 treatments. According to post hoc analysis with Bonferroni correction, different
730 numbers above bars indicate significant differences ($p \leq 0.05$) among species in a given
731 conductivity treatment, whereas different letters above bars indicate significant
732 differences among conductivity treatments for a given species. Experiments were not
733 carried out on *S. lateralis* and *S. scripta* at 37 mS cm⁻¹.

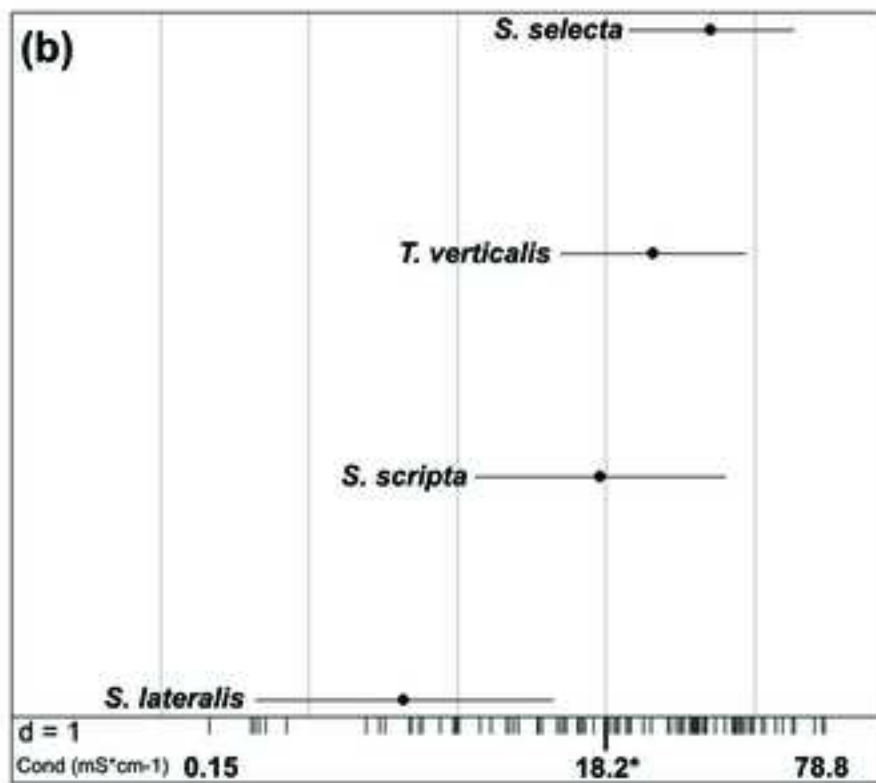
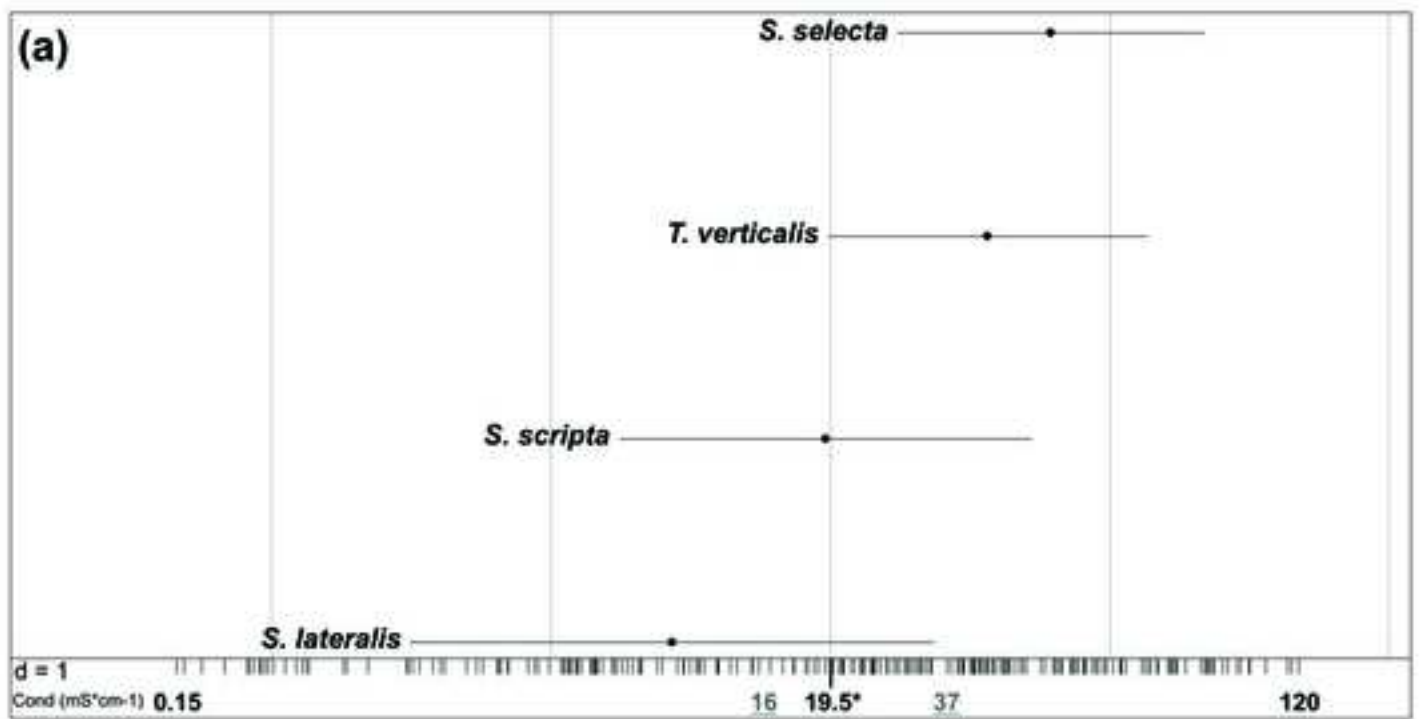
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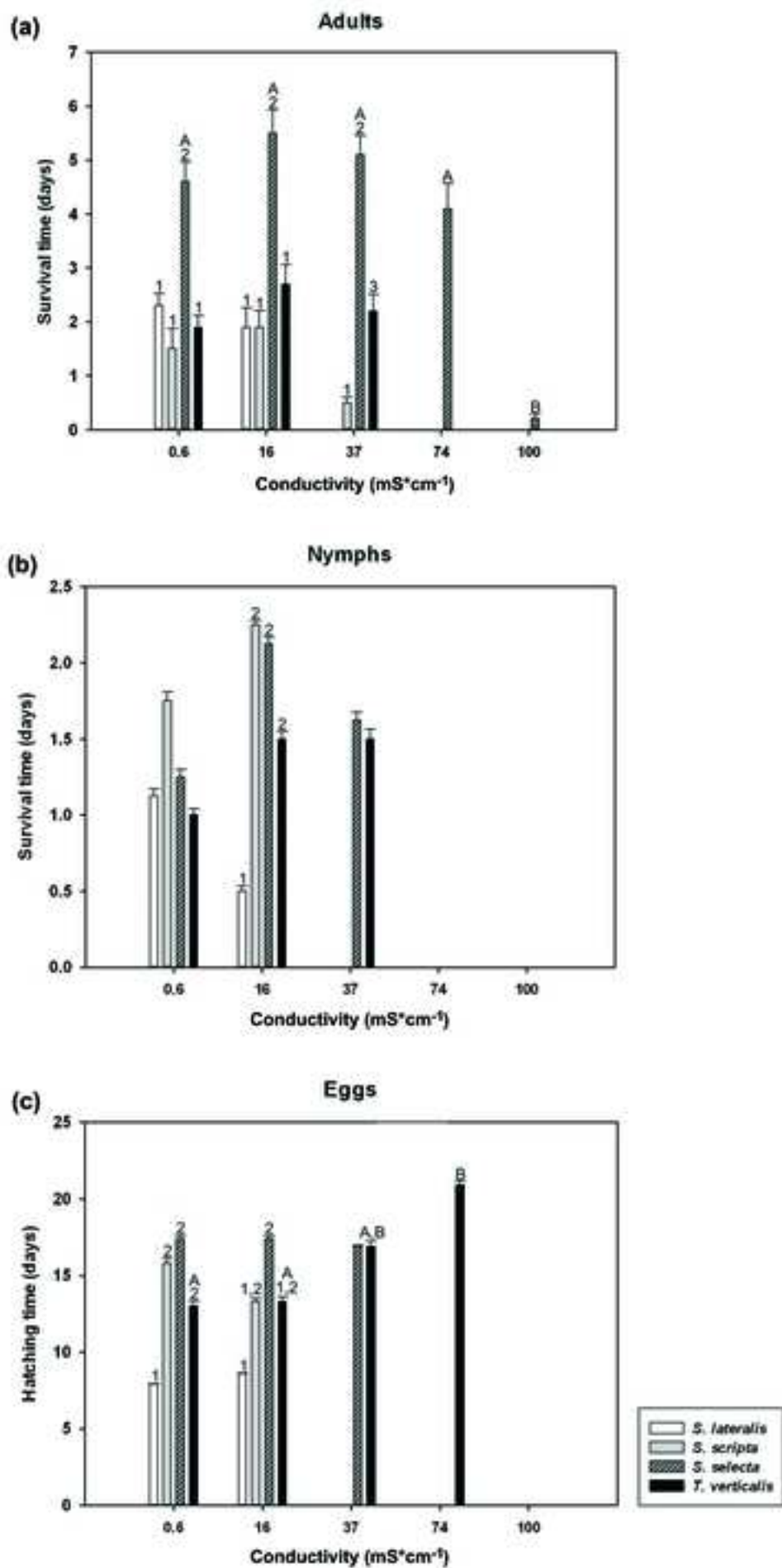
735 **Fig. 4.** Mean \pm SE oviposition rate of species in conductivity treatments. According to
736 post hoc analysis with Bonferroni correction, different numbers above bars indicate
737 significant differences ($p \leq 0.05$) among species in a given conductivity treatment.
738 Experiments were not carried out on *S. lateralis* and *S. scripta* at 37 mS cm⁻¹.

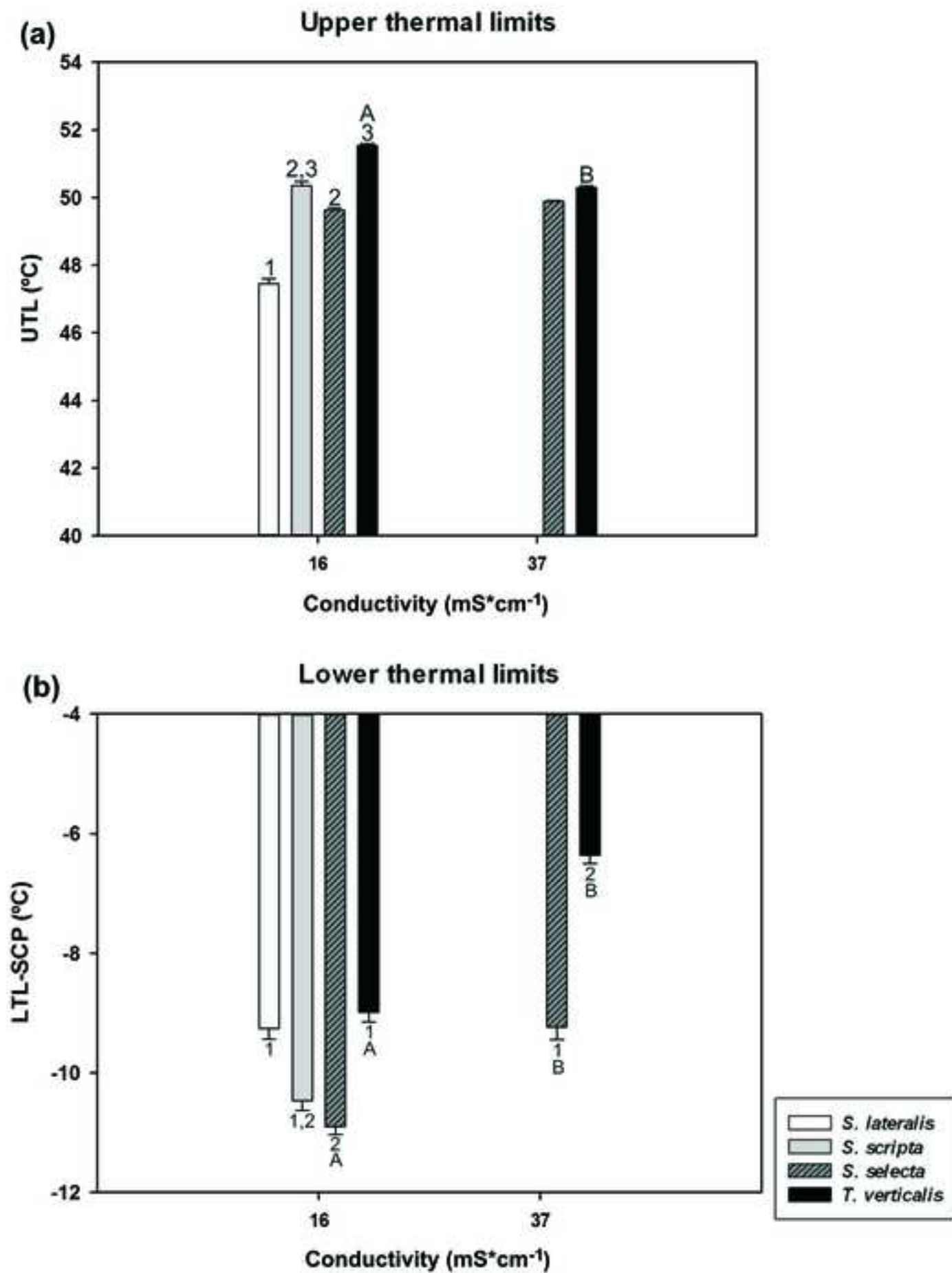
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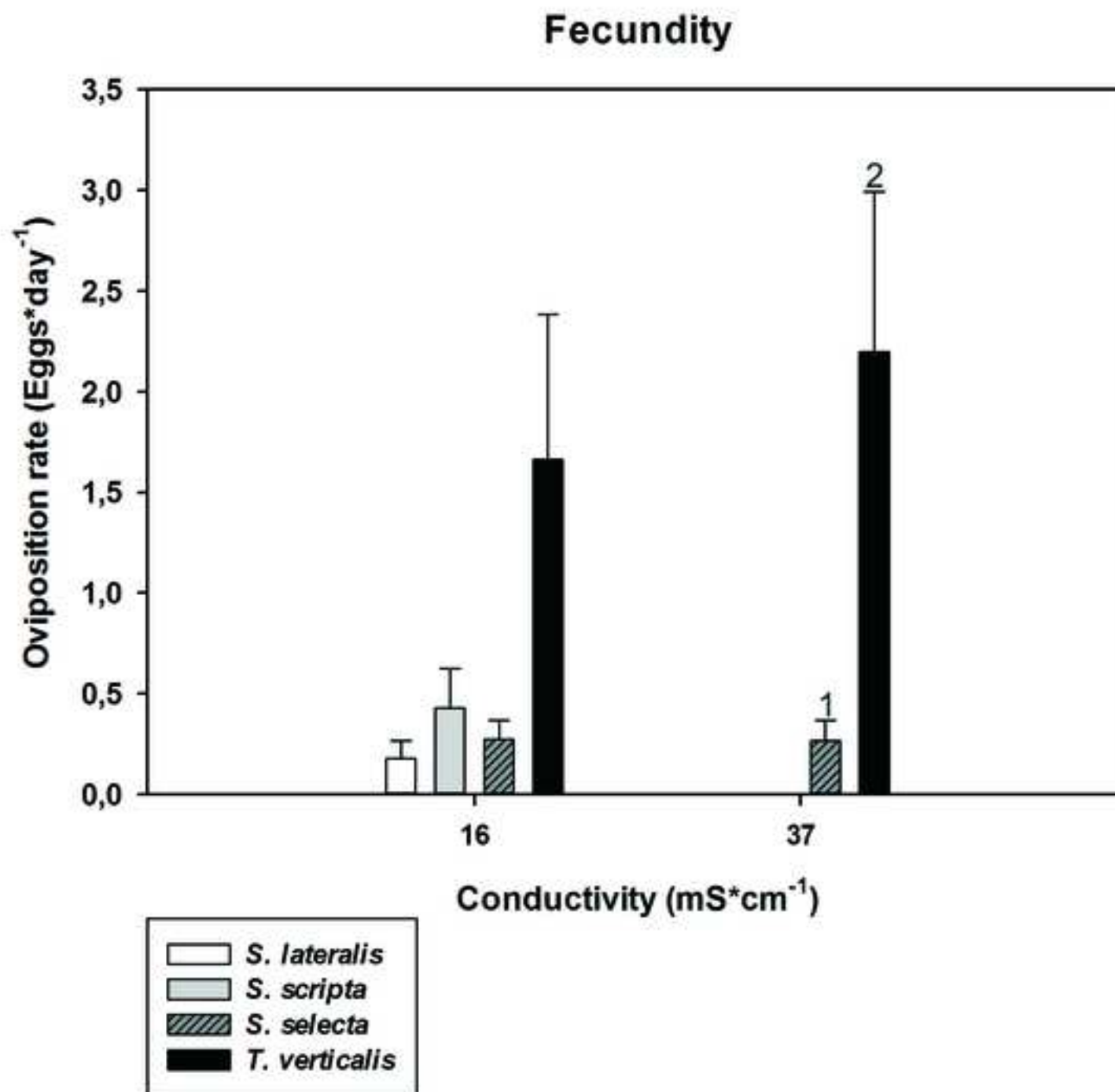
740 **Fig. 5.** Wing characteristics of the studied species: (a) wing loading, (b) wing aspect
741 ratio and (c) examples of wings of the four species. Data points represent mean values \pm
742 SE for each species (black circles, males; white circles, females). According to post hoc
743 analysis with Bonferroni correction, different numbers indicate significant differences
744 ($p \leq 0.05$) among species, whereas different letters indicate significant differences
745 among sexes for a given species.

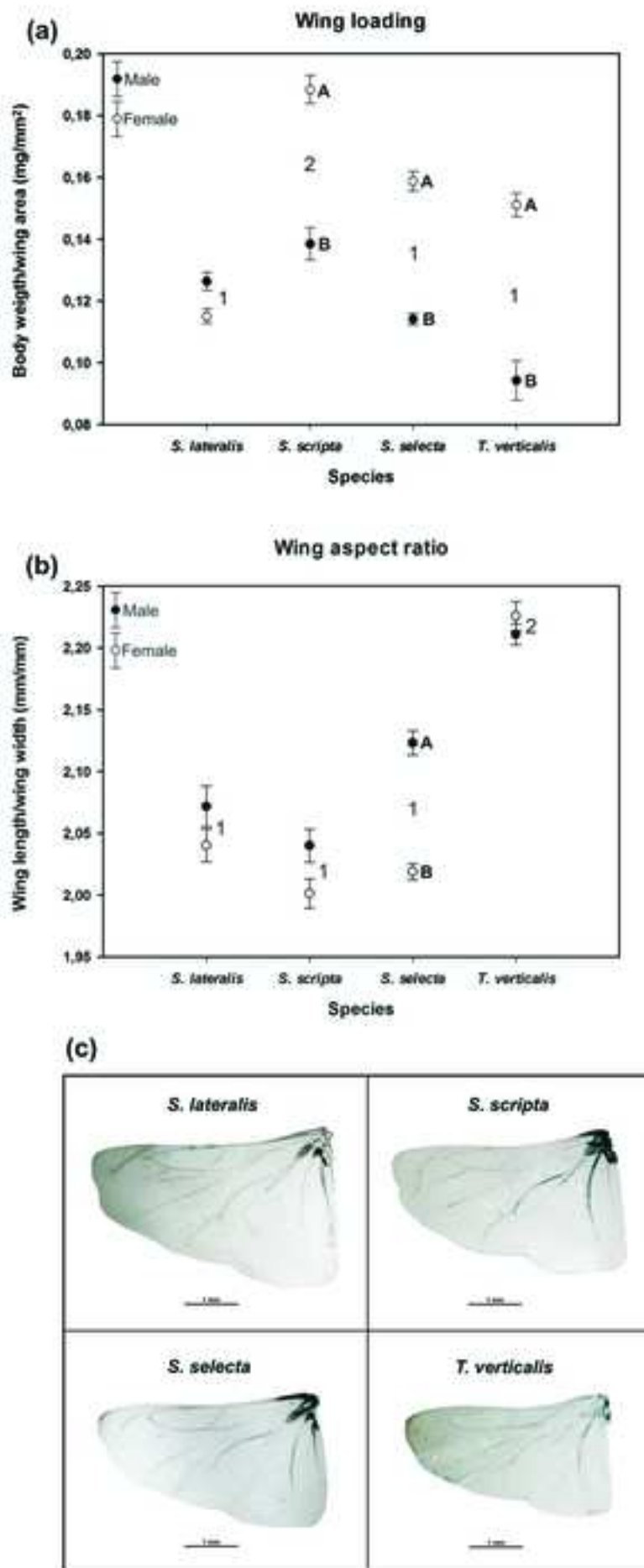
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TABLES

Table 1. Collection site information (dates of collection, geographical coordinates and mean conductivity).

Species	Sample location	Date	Latitude	Longitude	Altitude (m)	Conductivity (mS/cm)
<i>S. lateralis</i>	FAO, Doñana National Park, Sevilla	spring (2013)	37° 04' N	6° 22' O	2	2.08
<i>S. lateralis</i>	Charca de Las Moreras, Mazarrón, Murcia	autumn (2013)	37° 34' N	1° 17' W	10	0.55
<i>S. scripta</i>	Chícamo stream, Abanilla, Murcia	spring (2011)	38° 11' N	1° 03' W	150	21.9
<i>S. scripta</i>	Chícamo stream, Abanilla, Murcia	autumn (2013)	38° 11' N	1° 03' W	150	22.2
<i>S. selecta</i>	Barranco del Diablo, Molina de Segura, Murcia	spring (2011)	38° 07' N	1° 08' W	140	48.3
<i>S. selecta</i>	Barranco del Diablo, Molina de Segura, Murcia	autumn (2013)	38° 07' N	1° 08' W	140	53.7
<i>T. v. verticalis</i>	Veta la Palma, Doñana Natural Park, Sevilla	spring & autumn (2013)	36° 54' N	6° 15' O	1	33.7

Table 2. Summary of the studied traits expressed as average \pm SE, except for salinity tolerance expressed as LC₅₀ (95% confidence interval). * traits at which *T. v. verticalis* outperformed the other species.

		N	<i>S. lateralis</i>	<i>S. scripta</i>	<i>S. selecta</i>	<i>T. verticalis</i>
Sal. Tol. (LC₅₀-48h) adults		50	12.85 (5.27-31.30)	17.18	78.65 (67.01-92.31)	43.41 (33.81-55.74)
Sal. Tol. (LC₅₀-48h) nymphs		50	1.36	23.35	35.58 (12.75-99.29)	46.13*
Sal. Tol. (LC₅₀-25d) eggs		50	52.33	24.33	24.33	55.40 (42.63-72.00)*
Fecundity (eggs/day)	16 mS/cm	10	0.17 \pm 0.07	0.55 \pm 0.18	0.66 \pm 0.09	1.20 \pm 0.43
	37 mS/cm	10	-	-	0.41 \pm 0.07	6.06 \pm 2.04*
	0.6 mS/cm	10	7.90 \pm 0.04	15.80 \pm 0.32	17.40 \pm 0.21	13.00 \pm 0.32
Hatching time (days)	16 mS/cm	10	8.60 \pm 0.03	13.30 \pm 0.21	17.40 \pm 0.20	13.30 \pm 0.31
	37 mS/cm	10	0	0	16.90 \pm 0.05	16.90 \pm 0.35
	74 mS/cm	10	0	0	0	20.90 \pm 0.27*
	0.6 mS/cm	10	100	70	80	80
Hatching success (%)	16 mS/cm	10	100	90	80	80
	37 mS/cm	10	0	0	100	60
	74 mS/cm	10	0	0	0	40*
UTL-HC (°C)	16 mS/cm	40	47.46 \pm 0.14	50.36 \pm 0.13	49.63 \pm 0.07	51.54 \pm 0.06*
	37 mS/cm	40	-	-	49.89 \pm 0.03	50.30 \pm 0.06*
LTL-SCP (°C)	16 mS/cm	40	-9.25 \pm 0.18	-10.47 \pm 0.16	-10.90 \pm 0.14	-8.98 \pm 0.17
	37 mS/cm	40	-	-	-9.23 \pm 0.21	-6.36 \pm 0.14
Wing loading (mg/mm²)		40	0.11 \pm 0.00	0.15 \pm 0.03	0.13 \pm 0.00	0.11 \pm 0.00
Wing aspect ratio (mm/mm)		40	2.06 \pm 0.01	2.02 \pm 0.01	2.07 \pm 0.01	2.22 \pm 0.01*

SUPPORTING INFORMATION

What traits underpin the establishment and spread success of the invasive water bug

Trichocorixa verticalis verticalis?

J.A. Carbonell, A. Millán, A.J. Green, V. Céspedes, C. Coccia & J. Velasco

Hydrobiologia

Appendix S1 Analysis results

S1.1. Significant differences (Kruskal-Wallis test) among species and conductivities for survival times of adults and nymphs, and hatching time of eggs.

Species						Conductivity					
Phase	Species	N	Mean rank	Chi-square	<i>p</i> -value	Phase	Conductivity	N	Mean rank	Chi-square	<i>p</i> -value
Adults	<i>S. lateralis</i>	50	78.12	53.49	<0.001	Adults	0.6	40	130.15	68.82	<0.001
	<i>S. scripta</i>	50	78.92				16	40	135.28		
	<i>S. selecta</i>	50	147.13				37	40	108.64		
	<i>T. v. verticalis</i>	50	97.83				74	40	75.96		
	Total	200	100				40	52.48			
				Total	200						
Nymphs	<i>S. lateralis</i>	50	82.62	10.18	0.017	Nymphs	0.6	40	130.46	87.37	<0.001
	<i>S. scripta</i>	50	101.47				16	40	142.70		
	<i>S. selecta</i>	50	113.12				37	40	101.34		
	<i>T. v. verticalis</i>	50	104.79				74	40	64.00		
	Total	200	100				40	64.00			
				Total	200						
Eggs	<i>S. lateralis</i>	50	67.50	39.93	<0.001	Eggs	0.6	40	137.38	82.14	<0.001
	<i>S. scripta</i>	50	86.67				16	40	137.05		
	<i>S. selecta</i>	50	118.99				37	40	104.44		
	<i>T. v. verticalis</i>	50	128.84				74	40	78.14		
	Total	200	100				40	45.50			
				Total	200						

S1.2. Effects of species, conductivity and their interaction on upper thermal limit (UTL) and lower thermal limit (LTL) of different corixid species.

Effect	SS	df	F	p
UTL (Heat coma)				
Full model	104562.421	7	12641.375	(<0.001)
Intercept	312937.138	1	264834.0	(<0.001)
Conductivity	34284.940	1	29014.835	(<0.001)
Species	37489.286	3	10575.542	(<0.001)
Conductivity x species	32788.195	3	9249.387	(<0.001)
Error	255.233	216		
LTL (Supercooling point)				
Full model	3906.711	7	135.980	(<0.001)
Intercept	10661.880	1	2597.733	(<0.001)
Conductivity	2017.200	1	491.484	(<0.001)
Species	1038.862	3	84.372	(<0.001)
Conductivity x species	850.690	3	69.086	(<0.001)
Error	886.529	216		

S1.3. Effects of species, conductivity and their interaction on the oviposition rate of different corixid species.

Effect	SS	df	F	p
Full model	37.772	5	2.386	(0.050)
Intercept	36.538	1	11.539	(0.001)
Conductivity	0.693	1	0.219	(0.642)
Species	32.303	3	3.400	(0.024)
Conductivity x species	0.739	1	0.234	(0.631)
Error	682.341	71		

S1.4. Effects of species, sex, and their interaction on wing loading (WL) and wing aspect ratio (WAR).

Test	Effect	SS	df	F	p
WL	Full model	0.065	7	14.734	(<0.001)
	Intercept	1.920	1	3036.395	(<0.001)
	Species	0.024	3	12.813	(<0.001)
	Sex	0.026	1	40.516	(<0.001)
	Species x sex	0.015	3	8.062	(<0.001)
	Error	0.071	112		
WAR	Full model	0.787	7	14.267	(<0.001)
	Intercept	524.908	1	66647.400	(<0.001)
	Species	0.685	1	28.978	(<0.001)
	Sex	0.048	3	6.074	(0.015)
	Species x sex	0.054	3	2.288	(0.082)
	Error	0.882	216		