

1 **Influence of game crops on the distribution and productivity of red-**
2 **legged partridges *Alectoris rufa* in Mediterranean woodlands**

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4 **Luís Reino^{1,2}, Rui Borralho^{3,4} and Beatriz Arroyo^{5*}**

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6 ¹ CIBIO/InBIO-Centro de Investigação em Biodiversidade e Recursos Genéticos,
7 Universidade do Porto, Campus Agrário de Vairão, Rua Padre Armando Quintas, 4485-661
8 Vairão, Portugal

9 ² Centro de Ecologia Aplicada Prof. Baeta Neves, Instituto Superior de Agronomia/InBIO,
10 Tapada da Ajuda, 1349-017 Lisboa, Portugal

11 ³ ERENA, S.A., Rua do Conde de Redondo, N° 8 - 5° Dt°, 1150-105 Lisboa, Portugal

12 ⁴ Present address: Rua Robalo Gouveia, n°-1ª, 1900-392 Lisboa, Portugal

13 ⁵ Instituto de Investigación en Recursos Cinegéticos (IREC, CSIC-UCLM-JCCM), Ronda
14 de Toledo 12, 13005 Ciudad Real, Spain.

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16 *corresponding author. Beatriz.arroyo@uclm.es

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18

19 **Abstract**

20

21 Red-legged partridges *Alectoris rufa* are one of the most important game species in
22 extensively managed Mediterranean agro-forest systems. Population declines have led to
23 management to increase their populations. This includes the creation of game crops, but
24 their efficacy for red-legged partridges has not been tested. We developed in October 1996
25 an experimental introduction of 32 100x8 m plots in a 6.46 km² mixed agro-forest system
26 area in Portugal. These plots were planted with either lupin *Lupinus* sp., vetch *Vicia* sp. or
27 triticale *Triticum aestivum* x *Secale cereale*. The main goal of this study was the evaluation
28 of the potential effect of game crops on partridge distribution and productivity, after
29 controlling for the effect of habitat or other management actions. Partridge abundance and
30 distribution were assessed during spring and summer 1997 by intensive territory mapping.
31 We compared characteristics of territory centres with those of random points in relation to
32 land uses, game crops, and location of water points or supplementary grain sites. The most
33 important variable explaining partridge's location in spring was the density of
34 supplementary water points. In summer, partridge territories were positively associated
35 with the density of water points and lupin game crops, as well as olive trees. Productivity
36 (number of young per territory in relation to adults observed) increased with the density of
37 lupin game crops, but decreased with density of water points and vetch game crops and
38 proportion of woodland within the territories. Overall, this study suggests that management
39 for partridges in areas of agricultural abandonment, such as those in Mediterranean
40 woodlands, would benefit from the introduction of leguminous game crops and water
41 provision, though more studies are required for a more adequate optimization of these
42 measures of habitat improvement, in particular about the specific cover of the crops and
43 their spatial distribution so they provide adequate resources in summer for nestlings.

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46 **Keywords:** Red-legged partridge; Game crops; habitat management; Mediterranean
47 woodland management.

48 **Introduction**

49 Hunting is an important socioeconomic activity in many areas around the world, and in
50 particular in southern Europe. In these areas, small game is commonly hunted in many
51 areas associated to farmland or agro-forestry systems. As many other species associated to
52 these habitats, farmland small game such as partridges have suffered severe declines in
53 recent decades (e.g. Blanco-Aguilar 2007), through a combination of factors such as
54 agriculture intensification, climate change and overexploitation. As a consequence of these
55 declines, and in order to maintain hunting activities, many management techniques focused
56 on increasing or maintaining post-breeding game populations have been implemented in
57 many areas, particularly where economic interests are strong (Arroyo et al. 2012).

58 The most common management practices applied in Europe to increase small game
59 populations are predator control, population supplementation through the release of
60 captive-reared animals, and habitat management (providing supplementary food or water,
61 or increasing the quantity or quality of habitats used by game species; Arroyo & Beja 2002,
62 Beja et al. 2009, Rios-Saldaña 2010, Mustin et al. 2011). In relation to the latter, game
63 crops (crops established primarily for the benefit of gamebirds) are widely used in certain
64 areas managed for gamebird hunting (Arroyo & Beja 2002). Game crops are usually small
65 blocks sown with mixtures of seeds attractive to game, that provide additional food at
66 critical times of the year, nesting cover, protection from predators or green vegetation as a
67 source of water in areas with dry summers (CTGREF 1975, Peeters & Decamps 1998,
68 Reino et al. 2000, Stoate et al. 2003).

69 Several studies in France or the UK have concluded that winter game crops benefit
70 gamebirds (e.g. Mollot & Granval 1996, Sage et al. 2005). However, other studies have
71 shown contradictory results (e.g. Bro et al. 2004). Additionally, not much information
72 exists about the impact of game crops on game populations in southern Europe, where the
73 general level of intensification of agriculture is relatively low but where land abandonment
74 in less-productive areas is also an issue. In these areas, summer is the most critical period
75 for most species, due to the summer draught typical of Mediterranean type climates, so
76 game crops should aim to provide resources (food, water, cover) for this time of the year,
77 rather than in winter.

78 Portuguese *montados* (equivalent to the Spanish *dehesas*) are cork oak *Quercus*
79 *suber* and holm oak *Q. rotundifolia* extensive agro-forestry systems of savannah-like
80 physiognomy, particularly diverse in both human use and wildlife, being of high economic
81 and conservation relevance in large tracts of the Mediterranean part of Iberia (Meeus 1993,
82 Díaz et al. 1997, Sa-Sousa 2014). Typically, these open woodlands are multipurpose
83 extensively managed areas, generating a high diversity of products (such as cork, firewood,
84 game, domestic stock, cereals), and are frequently pointed out as an example of a
85 sustainable way of land exploitation, with less conservation conflicts than other alternative
86 land use options (Pinto-Correia 1993, Joffre et al. 1999).

87 Game hunting and its management are important activities on these systems, with
88 the red-legged partridge *Alectoris rufa* being one of the most important game species
89 exploited there (Borrvalho et al. 2000). Although red-legged partridges may be common on
90 *montado* (Carvalho & Borrvalho, 1998), they tend to avoid large tracts of forested areas
91 (Lucio & Purroy 1992, Meriggi et al. 1992). In sites experiencing a process of agriculture
92 abandonment, such as in many cork oak *montados* (Pinto-Correia & Mascarenhas 1999),
93 the planting of game crops conceivably may contribute to stopping or reversing population
94 declines associated with the degradation of mixed farmland, potentially increasing partridge
95 density and reproductive success, and reducing dispersion (Pépin & Blayac 1990, Lucio &
96 Purroy 1992). Nevertheless, there is a lack of information to substantiate this supposition.

97 This study assessed whether distribution or productivity of red-legged partridges
98 were affected by a set of experimental plots of game crops, of small size and low cost, in an
99 area of mixed Mediterranean woodland with a high proportion of cork oak *montado*. We
100 discuss results in relation to management for this game species.

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103 **Material and methods**

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105 **Study area**

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107 The study was undertaken in a 6.46-km² hunting estate of mixed Mediterranean woodland,
108 located in Alenquer municipality, central Portugal (at 39°07'N, 8°56'W). The climate in
109 this area is thermo-mediterranean, with hot dry summers and mild winters and an average
110 annual precipitation of 600 mm (Rivas-Martínez et al. 1990). Altitude ranged from 20 to
111 170 m asl. About 8% of the area was an agricultural lowland plain (Figure 1), mostly
112 cultivated with irrigated crops. The remaining area was covered by cork oak *montado*, with
113 and without shrubby understorey (ca. 20% and 38% of the area, respectively); these areas
114 had a relatively high density of trees, so their physiognomy was that of open woodland
115 rather than grassland with a few trees; forest stands of maritime pine *Pinus pinaster*,
116 umbrella pine *P. pinea* and eucalyptus trees *Eucalyptus globulus* (17%); a mixed cork oak
117 and maritime pine woodland with a patchy undergrowth vegetation of different heights
118 (15%); and a small grove of olive trees. During the study period, the partridges were not
119 hunted to allow the build-up of numbers. Red foxes *Vulpes vulpes*, Egyptian mongooses
120 *Herpestes ichneumon*, and feral cats and dogs were controlled, cereal grain was made
121 available at several feeding sites all year round, and water points were implemented.

122

123 **Partridge distribution and game crops**

124

125 In October 1996, 32 100×8-m single species plots were established in the study area. Plots
126 were sown in autumn because autumn sowing is the common farming practice for all crops
127 in this area, except for irrigated crops. These plots were sown with different covers aiming
128 to provide cover and food in spring and, particularly summer: a leguminous crop (either
129 lupin *Lupinus* sp. or vetch *Vicia* sp.) or a cereal crop (triticale *Triticum aestivum* x *Secale*
130 *cereale*). 11 plots of each crop type were planted except for cereal, of which only 10
131 existed. The plots were distributed according to the following method: 30 random points
132 were plotted on the wooded areas of the study site, and the nearest location to each point of
133 easy access to tillage equipment (clearings, firebreaks and high points with sparser
134 vegetation) was chosen to establish a game crop plot. The choice of random location

135 responded to the fact that there were no *a priori* regards to what would be implemented in
136 non-experimental conditions, as game crops were not implemented in the area prior to our
137 study, and game managers are not landowners in most areas (Arroyo et al. 2012).

138 We assessed partridge abundance and distribution throughout the study area in
139 spring and summer 1997. That year was similar in temperature and rainfall to others of the
140 same decade, i.e. was not particularly hot or dry (e.g. see data for nearby Lisbon in
141 PORDATA 2016). Partridge abundance and distribution was assessed by intensive territory
142 mapping (Pépin 1983, Borralho et al. 1996). The study area was divided into 1-km² plots,
143 which were intensively and similarly surveyed during the first three hours after dawn and
144 preceding dusk, both by observers using a four-wheel-drive vehicle throughout the
145 extensive track network of the study area, and by walking observers accompanied by
146 pointing dogs in systematic transects in each plot. The locations of partridge sightings,
147 calling birds, tracks and droppings were plotted on 1:5,000 and 1:25,000 topographic maps.
148 Daily location maps were generated through this procedure and from these we compiled
149 composite maps of partridge locations, which were interpreted to delineate territories.
150 Partridge signs one day did not influence searches in day D+1. Counts were stopped when
151 the cumulative number of detected birds plotted against cumulative searching effort
152 levelled off (Borralho et al. 1996). From the composite location maps generated during the
153 partridge counts, we determined the approximate centre of each territory using a
154 Geographic Information System (GIS-ArcCAD). We believe our method is adequate to
155 separate the range of different pairs/coveys. Average home range size of red-legged
156 partridges before hatching is ca. 8 ha per pair (Sumozas 2008), and partridge density at that
157 time was ca. 14 pairs per km² (see results). In summer, we also evaluated for each territory
158 the number of young and adults observed.

159 The following layers of information were incorporated and manipulated in the GIS:
160 the locations of game crops, land uses, locations of supplementary water points and the
161 locations of feeding sites where cereal grain was regularly provided. For each territory
162 centre and for 134 random points distributed throughout the study area, we assessed the
163 following variables in a radius of 250 m around each point: percentage of different land
164 uses (arable land, olive trees, montado, woodland); density of field edges (m/ha); density of

165 game crop plots of each type (number/ha); density of supplementary feeders (number/ha)
166 and density of supplementary water points (density/ha).

167 To assess the influence of game crops on partridge distribution, we compared the
168 characteristics of random points with those of territory centres, both in spring and summer.
169 By using the territory centres, the analyses concerned mostly the scale of habitat selection
170 related to the location of home ranges, i.e., the dispersal scale (*sensu* Morris 1992). For this,
171 we used Generalized Linear Models, using “type of point” as a response variable (with
172 Territory=1, and Random point = 0, binomial error distribution, logit link function). As
173 explanatory variables, we included all the variables mentioned above (land uses and those
174 related to game management). We checked for potential collinearity and redundancy of the
175 explanatory variables by analysing the Variable Inflation Factor (VIF). Land use variables
176 are explained as a combination of the other ones (as they are expressed as percentages), and
177 thus had very high VIF values (> 20). The one with highest VIF value was “*montado*”.
178 When excluding this variable, all explanatory variables had VIF values <2, well below the
179 threshold suggested for eliminating them (Zuur et al. 2010); therefore, all variables except
180 “*montado*” were included in the analysis.

181 Models were implemented with function glm in R (R Development Core Team,
182 2014). Using a multimodel inference approach, model-averaged parameter estimates were
183 derived on the basis of corrected Akaike’s information criteria (AICc) for the subset of
184 models constructed from combinations of these variables that had a $\Delta AICc < 2$. This
185 approach takes into account that when data is inadequate to reach strong inferences from a
186 single best model, and several models may be equally useful to describe variation in data,
187 more robust inferences on effect size and its precision may be made from combining all
188 alternative models in the set (Burnham & Anderson 2002). Information-theoretic
189 approaches are preferable over hypothesis testing approaches particularly in observational
190 studies, where randomization or replication is not achievable (Burnham & Anderson 2002).
191 In this approach, P values and statistical significance are not relevant to address the strength
192 of evidence for the models. Estimates of the relative importance of predictor variables can
193 instead be made by summing the Akaike weights across all the models in the set where this
194 variable occurs (Burnham & Anderson 2002). We thus also calculated the relative

195 importance of each variable as the sum of Akaike weights across all the models in the set
196 where that variable occurred. We only discuss the effects of variables for which the
197 standard errors were smaller than the average coefficient, because otherwise effect size was
198 considered too imprecise. Multimodel inference was implemented in R software by the
199 functions ‘dredge’ and ‘model.avg’ from the ‘MuMIn’ library.

200 Secondly, we evaluated (for summer censuses only) variations in partridge
201 productivity (number of young/number of adults observed) across territories. For this, the
202 response variable was log-transformed (Gaussian error distribution). We included the same
203 explanatory variables and model selection procedure as above.

204

205

206 **Results**

207

208 Our survey rendered 94 red-legged partridge territories in spring, occupied by partridge
209 pairs. In summer, our survey rendered 93 territories, occupied by 1 (n = 53) to 5 (n = 3)
210 adults, and 0 (n = 55) to 13 (n = 1) juveniles. Average productivity (mean \pm SD) was 1.26 \pm
211 1.87 young/adult (n = 93). When juveniles were observed (n = 38), covey size was 3.8 \pm
212 2.7 (median = 4).

213 A number of equivalent models with relatively low explanatory power explained the
214 location of partridge territories in spring (Table 1). According to model-averaged parameter
215 estimates, the most important variable explaining location of partridge territories at that
216 time was the density of supplementary water points (appearing in all of the best models),
217 which increased the likelihood of a point including a partridge territory (Fig. 2). The
218 proportion of arable land appeared in 3 of the 8 best models, but the estimate of the effect
219 size of this variable included 0 (Table 2). The density of cereal game crops appeared in 1 of
220 the 8 best models, but this variable had the lowest relative importance, and its effect was
221 not adequately estimated (Table 2).

222 In summer, location of partridge territories was positively associated with the
223 density of supplementary water points, the density of lupin game crops and the proportion
224 of olive tree fields (appearing in all of the best models, Tables 1 and 2, Fig. 3).

225 Additionally, partridges at this time appeared to select areas with more field edges, arable
226 land and cereal game crops, and avoid areas with more supplementary feeders and more
227 woodland (Table 2, Fig. 3).

228 Productivity (number of young per territory in relation to adults observed)
229 decreased with density of water points but increased with density of lupin game crops.
230 Additionally, productivity declined with density of vetch game crops and proportion of
231 woodland within the territories, although the relative importance of these variables was
232 lower (Table 2, Fig. 4).

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234

235 **Discussion**

236

237 According to our results, game crops sown with lupin appear to have some positive
238 influence on the location of partridges in summer, particularly of larger coveys, in our study
239 area. In contrast, the influence of game covers sown with cereal was less important, and
240 vetch game crops were apparently avoided. This suggests that game crops sown with lupin
241 provide cover or food resources necessary for partridge chicks (Buenestado et al. 2008,
242 Sumozas 2008), Average productivity observed in the study area (1.26 young per adult)
243 was in the lower range compared with that observed in hunting estates in central Spain
244 (Díaz-Fernández et al. 2013). This may indicate, among other things, that food supply
245 during chick the rearing period is low (as also supported by the relatively small covey sizes
246 observed), or that predation is high (as also supported by the large proportion of adults
247 without juveniles). The fact that productivity was positively associated with lupin game
248 crops, but negatively with vetch game crops, suggests that cover may be more important
249 than food resources in these crops (lupin is higher and denser than vetch, but both are rich
250 in nutrients being leguminous), and that the resource sought when using game crops is
251 protection from predators. However, these assumptions should need to be tested in future
252 studies.

253 The fact that location of partridge territories in spring was apparently not associated
254 with game crops suggests that they are not selected for nest cover, possibly because they

255 were too small and partially fenced, or because habitats providing nest cover were not
256 limiting for the study population. To our knowledge, this is the first study evaluating the
257 effect of game crops on farmland game in the Iberian Peninsula, so it is not possible to
258 compare our results with others. Additionally, our study was made in only one year, and
259 results may be biased if environmental conditions were not usual for the area. However,
260 1997 was quite similar regarding weather conditions to other years of the same decade
261 (PORDATA 2016), and we thus believe that our data can be representative of game crop
262 use by partridges in general conditions. Our results overall emphasize that environmental
263 conditions are more limiting in summer than spring in Mediterranean climates, and indicate
264 that management aiming to improve environmental conditions during the summer period
265 are likely to be beneficial for red-legged partridges. Nevertheless, it would be important to
266 further study effectiveness of game crops in extreme weather conditions (e.g. in dry years),
267 and develop studies to optimize these measures of habitat improvement, in particular about
268 the specific cover of the crops so they provide adequate resources (either cover, food, or
269 both) in summer for nestlings.

270 Other management variables were also important, in particular the provision of
271 supplementary water points. Location of partridge territories was strongly associated to
272 areas with high density of supplementary water points both in spring and summer. The
273 importance of water on partridge distribution in Southern Portugal has already been
274 described (Borrallho et al. 1998). Together, these results highlight the importance of water
275 areas with dry climates such as southern Iberian Peninsula, and indicate that water
276 availability may be a limiting factor there. Thus, they support that supplementation of water
277 is a useful management tool for this species in these areas. However, it has been suggested
278 that artificial water points may act as areas for transmission of diseases and infections
279 (Villanua et al. 2006), so it would be important to further assess this in future studies, if
280 using this management tool.

281 Surprisingly, and at odds with the above-mentioned results, young/adult ratio was
282 negatively related to the density of supplementary water points. This may reflect density-
283 dependence effects (if more partridges locate their territories close to water points, local
284 density may be higher and this may have negative effects through competition for food, Bro

285 et al. 2003), or else also indicate, as suggested above, higher disease transmission to
286 nestlings around water points. More research should look at these aspects in the future.

287 In contrast to what was found for supplementary water points, we found a negative
288 effect of availability of supplementary food devices on location of partridge territories in
289 summer, and no effect on productivity. In other areas of the Iberian Peninsula (central
290 Spain), supplementary food has been associated to higher productivity and density of red-
291 legged partridges (Díaz-Fernández et al. 2013). However, in central Spain, supplementary
292 food and water were usually placed together, so it was difficult to separate the influence of
293 both management tools (Díaz-Fernández et al. 2013). This could however indicate that food
294 is more limiting in central Spain than in *montado* areas in Portugal at the time of the study,
295 or else that other factors are more influential than grain provision. One way or other, given
296 that grain is one of the highest expenses in game management in farmland areas in the
297 Iberian Peninsula (Díaz-Fernández 2012), this raises questions about the cost-efficiency of
298 using supplementary grain for partridges in other areas with similar land cover, as in central
299 and southern Portugal. Further studies should try to quantify this in order to optimize
300 management.

301 Various land use variables influenced partridge distribution (at least in summer) and
302 productivity, with a general positive effect of arable land and olive trees, and negative
303 effect of woodland agreeing with previous results (Lucio & Purroy 1992, Borralho et al.
304 1999). Among landscape features, field edge density was positively related to location of
305 partridges in summer. The importance of field edges for this species has been mentioned in
306 many studies (Lucio & Purroy 1992, Peiró et al. 1993, Vargas et al. 2006), influencing
307 nesting selection and success (Casas & Viñuela 2010, Villanúa et al. 2011) and survival
308 (Buenestado et al. 2009). In this study, we found no influence between this variable and
309 productivity, but this may be related to the small variation in field edge density found
310 among the different study territories. Additionally, results could be influenced by the spatial
311 distribution of different habitats and partridge territories, so future studies should evaluate
312 this.

313 Overall, our results indicate that management for partridges in areas of agricultural
314 abandonment, such as those in Mediterranean woodlands, would benefit by including the

315 use of game crops and the provision of water, although it would be useful to think about the
316 spatial design of devices to minimize negative effects of the latter, and about the specific
317 cover of the crops so they provide the most needed resources in summer (either cover
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333 **References**

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- 335 Arroyo B, & Beja P (2002) Impact of hunting management practices on biodiversity.
336 Report to EC within REGHAB Concerted Action.
- 337 Arroyo B, Delibes-Mateos M, Díaz-Fernández S, Viñuela J (2012) Hunting management in
338 relation to profitability aims: red-legged partridge hunting in central Spain. *Eur J*
339 *Wildlife Res* 58:847–855.
- 340 Beja P, Gordinho L, Reino L, Loureiro F, Santos-Reis M, Borralho R (2009) Predator
341 abundance in relation to small game management in southern Portugal: conservation
342 implications. *Eur J Wildlife Res* 55:227–238.
- 343 Blanco-Aguilar JA (2007) Variación espacial en la biología de la perdiz roja (*Alectoris*
344 *rufa*): una aproximación multidisciplinar. PhD thesis, Universidad Complutense de

345 Madrid, España.

346 Borralho R, Rego F, Vaz Pinto P (1996) Is driven transect sampling suitable for estimating
347 red-legged partridge *Alectoris rufa* densities? *Wildlife Biology* 2: 259-268.

348 Borralho R, Rito A, Rego F, Simões H, Vaz Pinto P (1998) Summer distribution of red-
349 legged partridges *Alectoris rufa* in relation to water availability on Mediterranean
350 farmland. *Ibis* 140: 620-625.

351 Borralho R, Carvalho J, Vaz Pinto P, Bugalho J, Castro Pereira D, Capelo M (2000) Gestão
352 cinegética e evolução demográfica de duas populações de Perdiz-vermelha (*Alectoris*
353 *rufa*) do Centro e Sul de Portugal. *Silva Lusitana*, 8: 49-59.

354 Bro E, Deldalle B, Massot M, Reitz FO, Selmi F (2003) Density dependence of
355 reproductive success in grey partridge *Perdix perdix* populations in France:
356 management implications. *Wildlife Biology* 9: 93-102

357 Bro E, Mayot P, Corda E, Reitz F (2004) Impact of habitat management on grey partridge
358 populations: assessing wildlife cover using a multisite BACI experiment. *Journal of*
359 *Applied Ecology* 41: 846-857

360 Buenestado FJ, Ferreras P, Delibes-Mateos M, Tortosa FS, Blanco-Aguilar JA, Villafuerte
361 R (2008) Habitat selection and home range size of red-legged partridges in Spain.
362 *Agriculture Ecosystems and Environment* 126: 158-162.

363 Buenestado FJ, Ferreras P, Blanco-Aguilar JA, Tortosa FS, Villafuerte R (2009) Survival
364 and causes of mortality among wild Red-legged Partridges *Alectoris rufa* in southern
365 Spain: Implications for conservation. *Ibis* 151: 720-730.

366 Casas F, Viñuela J (2010) Agricultural practices or game management: which is the key to
367 improve red-legged partridge nesting success in agricultural landscapes?
368 *Environmental Conservation* 37:177-186.

369 Carvalho J, Borralho R (1998) Produtividade e sucesso reprodutivo de duas populações de
370 Perdiz-vermelha (*Alectoris rufa*) em diferentes habitats. *Silva Lusitana*, 6: 215-226.

371 CTGREF (1988) Aménagement des territoires de chasse/Petit Gibier. Note Technique n°
372 28, Nogent-sur-Vernisson.

373 Díaz M, Campos P, Pulido FJ (1997) The Spanish dehesas: a diversity in land-use and
374 wildlife. - In: Pain, D.J. & Pienkowski, M.W. (Eds.); *Farming and Birds in Europe*.

375 Academic Press, Cambridge, Great Britain, pp. 178-209.

376 Díaz-Fernández S. (2012) Relationships between red-legged partridge hunting
377 management, red-legged partridge populations, and human populations. PhD thesis,
378 Universidad de Castilla- La Mancha, España.

379 Díaz-Fernández S, Arroyo B, Casas F, Haro M, Viñuela J (2013) Effect of management
380 techniques on red-legged partridge abundance. PLOS One. 8(6), e66671

381 Joffre R, Rambal S, Ratte JP (1999) The dehesa system of southern Spain and Portugal as a
382 natural ecosystem mimic. Agroforestry systems 45 : 57-79.

383 Lucio AJ, Purroy FJ (1992) Red-legged partridge (*Alectoris rufa*) habitat selection in
384 Northwest Spain. Gibier Faune Sauvage 9: 417-430.

385 Meeus JH (1993) The transformation of agricultural landscapes in Western Europe. - The
386 Science of the Total Environment 129: 171-190.

387 Meriggi A, Saino N, Montagna D, Zacchetti D (1992) Influence of habitat on density and
388 breeding success of grey and red-legged partridges. Bolletino di Zoologia 59: 289-
389 295.

390 Mollot B, Granval P (1996). Efficacité de la jachère faune sauvage. - Bulletin Mensuel de
391 l'Office National de la Chasse 214: 24-29.

392 Morris DW (1992). Scales and costs of habitat selection in heterogeneous landscapes.
393 Evolutionary Ecology, 6: 412-432.

394 Mustin K, Newey S, Irvine J, Arroyo B, Redpath S (2011) Biodiversity impacts of game
395 bird hunting and associated management practices in Europe and North America.
396 Contract report, James Hutton Institute, Aberdeen, UK.

397 Peeters A, Decamps C (1998) Choix et gestion de couverts herbacés dans les jachères et les
398 tournières faunistiques - Gibier Faune Sauvage, Game and Wildlife 15 (Hors série
399 Tome 1): 117-129.

400 Peiró V, Seva E, Almiñana N (1993) Selección de hábitat en una población de perdiz roja
401 (*Alectoris rufa*) en una zona de sierra del sur de la provincia de Alicante.
402 Mediterránea Serie de Biología 14: 5-22.

403 Pépin D (1983) Utilisation et valeur de diverses méthodes d'estimation de la densité de la
404 perdrix rouge (*Alectoris rufa*) au printemps. - In: J. Castroviejo (Ed.); Actas del XV

405 Congreso Internacional de Fauna Cinegética y Silvestre: 725-735. Trujillo.

406 Pépin D, Blayac J (1990) Impacts d'un aménagement de la garrigue et de l'instauration
407 d'un plan de chasse sur la démographie de la perdrix rouge (*Alectoris rufa*) en milieu
408 méditerranéen. *Gibier Faune Sauvage* 7: 145-158.

409 Pinto-Correia T (1993). Threatened landscape in Alentejo, Portugal: the 'montado' and
410 other 'agro-silvo-pastoral' systems. *Landscape and Urban Planning* 24: 43-48.

411 Pinto-Correia T, Mascarenhas J (1999) Contribution to the extensification/intensification
412 debate: new trends in the Portuguese montado. *Landscape and Urban Planning* 46:
413 125-131.

414 PORDATA (2016). <http://www.pordata.pt/en/>, accessed in April 2016.

415 R Development Core Team (2014) R: A Language and Environment for Statistical
416 Computing. R Foundation for Statistical Computing. Available at: [http://www.r-](http://www.r-project.org)
417 [project.org](http://www.r-project.org).

418 Reino LM, Borralho R, Bugalho JFF (2000) Revisão da utilização de culturas para a fauna
419 na gestão de espécies cinegéticas. *Revista de Ciências Agrárias* 23: 48-71.

420 Ríos-Saldaña CA (2010) Los planes técnicos de caza de Castilla-La Mancha y su aplicación
421 en la gestión y conservación de las especies cinegéticas. PhD Thesis, University of
422 Castilla-La Mancha, Ciudad Real.

423 Rito A, Borralho R (1997) Importância da disponibilidade de água para galiformes bravios
424 em situações de carência. *Revista de Ciências Agrárias* 20: 35-50.

425 Rivas-Martínez S, Lousã M, Díaz TE, Fernandez-González F, Costa JC (1990) La
426 vegetación del sur de Portugal (Sado, Alentejo y Algarve). *Itenera Geobotánica* 3: 5
427 126.

428 Sage, RB, Parish DMB, Woodburn MIA Thompson PGL. (2005). Songbirds using crops
429 planted on farmland as cover for game birds. *European Journal of Wildlife Research*
430 51: 248-253

431 Sa-Sousa P (2014) The Portuguese montado : conciliating ecological values with human
432 demands within a dynamic agroforestry system. *Annals of Forest Science* 71: SI: 1-3

433 Sumozas N (2008) Estudio de la selección de hábitat y el dominio vital de la perdiz roja en
434 pseudoestepas agrarias. Diploma de Estudios Avanzados. Universidad de Castilla-La

435 Mancha.

436 Stoate C, Szczur J, Aebischer NJ (2003) Winter use of wild bird cover crops by passerines
437 on farmland in northeast England. *Bird study* 50, 15-21.

438 Vargas JM, Guerrero JC, Farfán MA, Barbosa AM, Real R (2006) Land use and
439 environmental factors affecting Red-legged Partridge (*Alectoris rufa*) hunting yields
440 in southern Spain. *European Journal of Wildlife Research* 52: 188-195.

441 Villanua D, Hofle U, Perez-Rodriguez L, Gortazar G (2006) *Trichomonas gallinae* in
442 wintering common wood pigeons *Columba palumbus* in Spain. *Ibis* 148: 641-648.

443 Villanúa D, Torres J, Ardaiz J, Alzaga V, Ros F, Cormenzana A, Castián E (2011)
444 Relationship between landscape heterogeneity loss and red-legged partridge
445 (*Alectoris rufa*) populations survival. XXXth Congress of the International Union of
446 Game Biologist (UIGB) and PERDIX XIII. Barcelona, España.

447 Zuur AF, Ieno EN, Elphick CS (2010) A protocol for data exploration to avoid common
448 statistical problems. *MEE* 1: 3-14.

449

Table 1. Variables included in the best models ($\Delta AIC_c < 2$) explaining location of partridge territories (comparison between territories and random points) or partridge productivity (young/adult ratio). Edges: density of field edges; GC1: density of game crops sown with lupin within territory; GC2: density of game crops sown with vetch within territory; GC3: density of game crops sown with cereal within territory; Water: density of water points; Feeder: density of feeders; Arable: % of arable land; Wood: % of forested area; Olive: % of olive trees.

Model	Edges	GC1	GC2	GC3	Water	Feeder	Arable	Wood	Olive	Df	delta	weight
<u>Spring location</u>												
1					X					2	0.00	0.211
2					X		X			3	0.25	0.187
3	X				X					3	0.65	0.152
4					X	X	X			4	1.46	0.102
5					X	X				3	1.49	0.100
6	X				X		X			4	1.79	0.086
7					X					3	1.87	0.083
8				X	X				X	3	1.97	0.079
<u>Summer location</u>												
1	X	X			X	X			X	6	0.00	0.133
2	X	X			X	X		X	X	7	0.29	0.115
3	X	X			X	X			X	7	0.67	0.095
4		X			X	X			X	6	0.75	0.092
5	X	X			X				X	5	0.80	0.089
6	X	X	X		X	X			X	7	0.94	0.083
7		X			X				X	5	1.13	0.076
8	X	X			X				X	6	1.14	0.075
9	X	X	X		X	X		X	X	8	1.31	0.069
10		X			X	X			X	5	1.54	0.062
11	X	X	X		X	X			X	8	1.54	0.062
12		X	X		X	X			X	7	1.99	0.049
<u>Productivity</u>												
1		X			X					3	0.00	0.364
2		X			X			X		4	1.37	0.184
3		X	X		X					4	1.73	0.153
4		X	X		X			X		4	1.76	0.151
5		X			X				X	4	1.81	0.147

Table 2. Model-averaged coefficients for variables included in models with delta AICc<2. Also indicated is the Relative Variable Importance (RVI) of those variables. Highlighted in bold those where parameter estimates do not encompass 0 taking into account SE. Edges: density of field edges (m/ha); GC1: density of game crops sown with lupin within territory; GC2: density of game crops sown with vetch within territory; GC3: density of game crops sown with cereal within territory; Water: density of water points; Feeders: density of feeders; Arable: % of arable land; Woods: % of forested area; Olive: % of olive trees. In bold, variables with RVI>0.3

	Spring			Summer			Productivity		
	β	SE	RVI	β	SE	RVI	β	SE	RVI
Water	5.44	2.23	1	7.81	2.89	1	-0.97	0.48	1
GC1				14.60	6.25	1	2.85	1.13	1
GC2							-1.46	1.35	0.32
GC3	0.17	1.48	0.08	6.18	5.82	0.26			
Arable	0.003	0.005	0.37	0.01	0.007	0.45			
Edges	6.75	18.45	0.20	45.49	41.48	0.72			
Feeders				-3.68	3.20	0.76			
Woods				-0.009	0.007	0.18	-0.002	0.001	0.36
Olive	-0.0005	0.007	0.08	0.06	0.02	1	0.002	0.003	0.13

Figure 1. Map of the distribution of land uses in the study area.

Figure 2. Modelled probability (and 95% CI) that a given spot would hold a partridge territory in spring in relation to game management in *montado* woodland in Portugal.

Figure 3. Modelled probability (and 95% CI) that a given spot would hold a partridge territory in summer in relation to game management in *montado* woodland in Portugal.

Figure 4. Modelled relationship between partridge productivity (young/adult ratio, and 95% CI) and game management in *montado* woodland in Portugal.

Figure 1

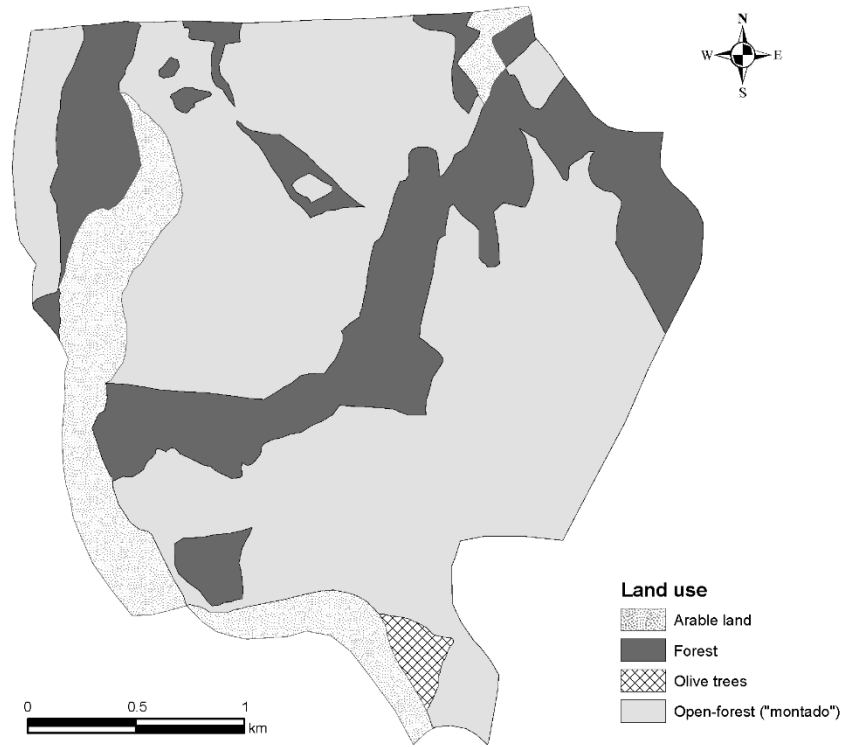


Figure 2.

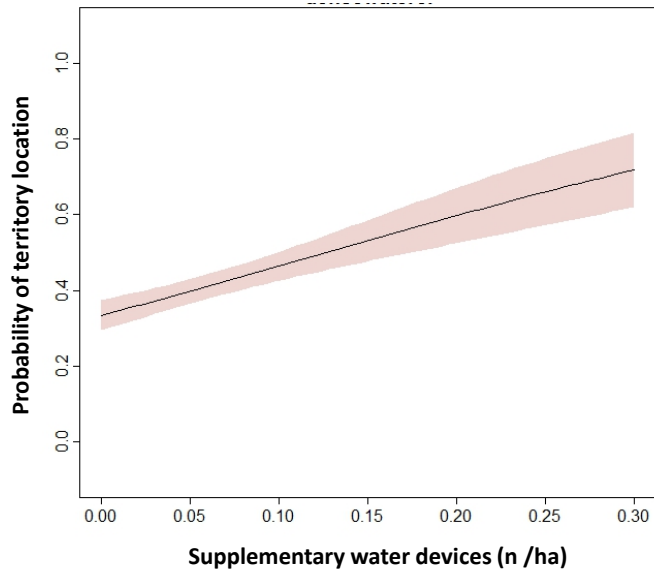


Figure 3.

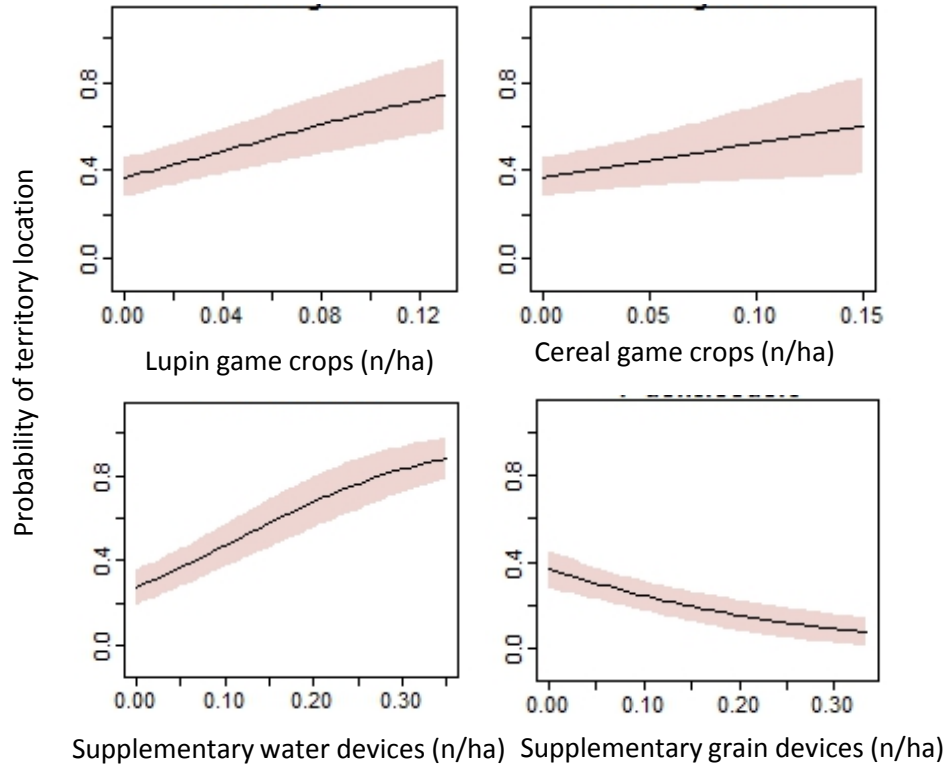


Figure 4.

