

Investigation of Hybrid Plasma-Catalytic Removal of Acetone over CuO/ γ -Al₂O₃ Catalysts Using Response Surface Method

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30 **Abstract**

31 In this work, plasma-catalytic removal of low concentrations of acetone over CuO/ γ -Al₂O₃
32 catalysts was carried out in a cylindrical dielectric barrier discharge (DBD) reactor. The
33 combination of plasma and the CuO/ γ -Al₂O₃ catalysts significantly enhanced the removal
34 efficiency of acetone compared to the plasma process using the pure γ -Al₂O₃ support, with the
35 5.0 wt.% CuO/ γ -Al₂O₃ catalyst exhibiting the best acetone removal efficiency of 67.9%.
36 Catalyst characterization was carried out to understand the effect the catalyst properties had
37 on the activity of the CuO/ γ -Al₂O₃ catalysts in the plasma-catalytic reaction. The results
38 indicated that the formation of surface oxygen species on the surface of the catalysts was
39 crucial for the oxidation of acetone in the plasma-catalytic reaction. The effects that various
40 operating parameters (discharge power, flow rate and initial concentration of acetone) and the
41 interactions between these parameters had on the performance of the plasma-catalytic removal
42 of acetone over the 5.0 wt.% CuO/ γ -Al₂O₃ catalyst were investigated using central composite
43 design (CCD). The significance of the independent variables and their interactions were
44 evaluated by means of the Analysis of Variance (ANOVA). The results showed that the gas
45 flow rate was the most significant factor affecting the removal efficiency of acetone, whilst
46 the initial concentration of acetone played the most important role in determining the energy
47 efficiency of the plasma-catalytic process.

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49 **Keywords:** Plasma-catalysis; Dielectric barrier discharge; Acetone removal; CuO/ γ -Al₂O₃;
50 Response surface method

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55 **1. Introduction**

56 Acetone, one of the most abundant oxygenates in air, has been widely used as paint
57 thinner, solvent and raw material in chemical industry. The emission of acetone has negative
58 effects on both the global environment and human health (Koppmann, 2008). Exposure to
59 acetone can cause dizziness, unconsciousness and nausea (Flowers et al., 2003). Great efforts
60 have been devoted to technology research and development to meet the stringent regulations
61 for air pollution control. However, conventional technologies including catalytic combustion,
62 regenerative oxidation, photo-catalytic oxidation, adsorption and condensation are not
63 cost-effective for the removal of low concentrations of acetone in high volume waste gas
64 streams (Schnelle Jr and Brown, 2001).

65 For the last two decades, non-thermal plasma (NTP) has been regarded as a promising
66 gas cleaning technology for the abatement of low concentration volatile organic compounds
67 (VOCs) in high volume waste gas streams (Chen et al., 2009; Tu and Whitehead, 2012).
68 Using air as a carrier gas, energetic electrons and a large number of highly reactive species
69 including O, O₃, N and metastable N₂ can be generated in the plasma even at room
70 temperature. Both high energy electrons and reactive species are capable of initiating a
71 cascade of physical and chemical reactions, which contribute to the removal of gas pollutants.
72 The main challenges in the industrial application of NTP for waste gas clean-up are the
73 formation of unwanted by-products and the low energy efficiency of the plasma process
74 (Kogelschatz, 2003; Kim, 2004). CO, CH₄, HCOOH and HCHO were found to be the major
75 organic by-products in plasma decomposition of acetone (Lyulyukin et al., 2010; Narengerile
76 and Watanabe, 2012; Zheng et al., 2014).

77 Recently, the combination of plasma and heterogeneous catalysis, namely
78 plasma-catalysis, has been considered as a promising solution for waste gas clean-up. The
79 presence of a catalyst in the plasma has great potential to generate a synergistic effect, which

80 can reduce the activation energy of the reaction, enhance the removal of the gas pollutant and
81 the selectivity of the desired final products, and minimize the formation of unwanted
82 by-products. All of these contribute in different ways to increasing the energy efficiency of
83 the plasma-catalytic process (Van Durme et al., 2008; Chen et al., 2009; Vandenbroucke et al.,
84 2011). Chang and Lin (2005) reported the acetone decomposition efficiency of a plasma
85 process to be 25% higher in the presence of TiO₂ compared to that using NTP alone. Trinh
86 and Mok (2014) found that placing ceramic supported MnO₂ catalysts in a dielectric barrier
87 discharge (DBD) significantly improved the removal efficiency of acetone, by 37%, at a
88 specific energy density (SED) of 600 J L⁻¹. In our previous work, we showed that the energy
89 yield of acetone removal (3.72 g kWh⁻¹) was 51.0% higher in the presence of γ -Al₂O₃ than
90 when using plasma alone (Zheng et al., 2014).

91 Catalysts are of great significance in a plasma-catalysis system. Various catalysts have
92 been reported for plasma-catalytic oxidation of VOCs, among which Cu-based catalysts
93 showed their advantages over other transition metal oxide catalysts due to their low cost and
94 comparative reaction performance (Guo et al., 2007; An et al., 2011; Wu et al., 2013; Zhu et
95 al., 2015a). Our previous work showed that the addition of 10 wt% transition metal oxides
96 (Ce, Co, Cu, Mn and Ni) on γ -Al₂O₃ support enhanced the removal of acetone, with the
97 supported copper oxide catalyst exhibiting the best performance among the tested transition
98 metals (Zhu et al., 2015b).

99 Although plasma-catalytic removal of acetone has been reported before, far less has been
100 done for the optimization of the plasma-catalytic process since its reaction performance is
101 largely affected by various operation parameters (Vandenbroucke et al., 2011; Samukawa et
102 al., 2012). The optimization of plasma-catalytic systems in previous work has been mostly
103 carried out via experimental approaches. The traditional univariate method fails to consider
104 and represent the interactions between different input variables. Moreover, this method

105 requires a large amount of experimental data to obtain the favorable sets of operating
106 parameters for the optimization of the plasma process, which makes it time consuming and
107 labor intensive (Aerts et al., 2013; Thevenet et al., 2014; Xu et al., 2014). Recently, response
108 surface methodology (RSM) has drawn attention for the investigation and optimization of
109 processes. RSM is a statistical model considering the non-linear relationships between the
110 multiple input and output variables based on design of experiments (DoE), which aims to
111 predict and optimize the performance of complex systems via experiment design, model
112 building, and evaluation of the significance of independent variables and the interactions
113 between them. Until now, only limited work has been focused on the investigation of plasma
114 processes using the DoE method (Butron-Garcia et al., 2015; Mei et al., 2015), while the use
115 of DoE for the optimization of plasma-catalytic gas clean-up has not been reported before.

116 In this work, the effect of discharge power, gas flow rate, initial acetone concentration
117 and Cu loading amount on the performance of the plasma-catalytic removal of acetone were
118 investigated. Initial experiments were carried out to find an optimal Cu loading amount for
119 the highest removal efficiency of acetone. A series of catalyst characterization techniques
120 were performed to establish the relationships between catalyst properties and reaction
121 performance. A central composite design (CCD) method was applied to investigate the
122 interactions between the main process variables and to optimize the plasma-catalytic process
123 in terms of removal efficiency and energy efficiency.

124

125 **2. Experimental**

126 **2.1 Catalyst preparation and characterization**

127 The x wt% CuO/ γ -Al₂O₃ catalysts (x=2.5, 5.0, 7.5 and 10.0) were prepared by incipient
128 wetness impregnation using copper nitrate (Alfa Aesar, 99.5%) as the precursor. The
129 appropriate weight of support (γ -Al₂O₃) was added to the copper nitrate solution with a

130 concentration of 0.1 M and continuously stirred at 80 °C for 4 h. The resulting slurry was
131 dried in an oven at 110 °C overnight, followed by calcination at 500 °C for 5 h. Pure γ -Al₂O₃
132 support was treated in the same way for comparison in this study. All the catalysts were
133 sieved to 40-60 meshes prior to use.

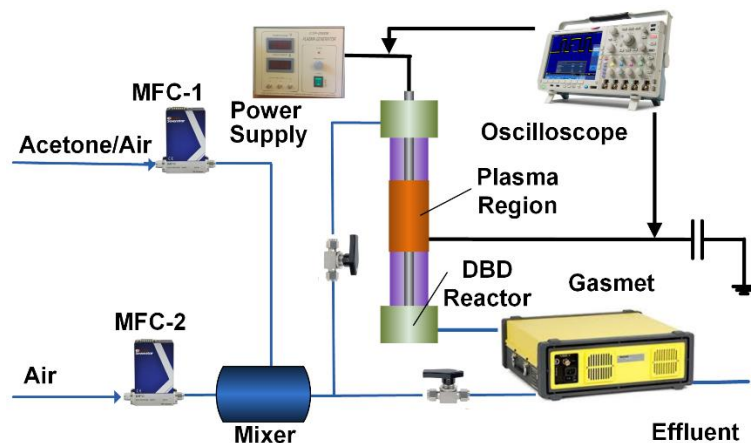
134 The structural properties of the CuO/ γ -Al₂O₃ catalysts, including specific surface area,
135 average pore size and pore volume, were acquired via N₂ adsorption-desorption experiments
136 using a Quantachrome Autosorb-1 instrument at -196 °C. The X-ray diffraction (XRD)
137 patterns of the catalyst samples were analyzed by a Rigaku D/max-2000 X-ray diffractometer.
138 The instrument was equipped with a Cu-K α radiation source, with the scan conducted in the
139 2 θ range from 10 ° to 80 ° with a scanning rate of 4 ° min⁻¹ and a step size of 0.02°. The
140 reducibility of the CuO/ γ -Al₂O₃ catalysts was evaluated by temperature-programmed
141 reduction with hydrogen (H₂-TPR) using a gas chromatograph (GC-1690). Each catalyst (50
142 mg) was pre-treated at 200 °C in a N₂ flow for 1 h before the test. The samples were then
143 heated from room temperature to 800 °C at a heating rate of 10 °C min⁻¹. A 5% H₂/Ar flow
144 with a flow rate of 40 mL min⁻¹ was used. The amount of consumed H₂ was calculated by the
145 integration of the H₂-TPR signals.

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147 **2.2 Experimental set-up**

148 The schematic diagram of the experimental set-up is shown in Fig. 1. A 60 mm-long
149 aluminum foil (ground electrode) was wrapped over a quartz tube with an inner diameter of 8
150 mm and wall thickness of 1 mm. A stainless steel rod with an outer diameter of 4 mm was
151 placed in the axis of the quartz tube and acted as a high voltage electrode. The length of the
152 discharge zone was 60 mm with a discharge gap of 2 mm. Zero grade air (99.999%) was used
153 as carrier gas in this work. Gaseous acetone was generated from a gas cylinder (0.5% acetone,
154 balanced air). All gas streams were controlled by mass flow controllers and premixed prior to

155 the DBD reactor. In each experiment, catalyst samples with a dielectric constant of around
156 12.6 were placed in the discharge region, held in place by glass wool. The reactor was
157 powered by an AC power supply with a frequency of 10.2 kHz, while the maximum peak
158 voltage was 30 kV.
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Fig. 1. Schematic diagram of the experimental setup

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The applied voltage was measured by a Tektronix 6015A high voltage probe (1000:1), while the voltage across the external capacitor (0.47 μF) was monitored by a Tektronix TPP500 probe. All electrical signals were sampled by a digital oscilloscope (Tektronix 3034B). The discharge power was calculated using Q-U Lissajous method.

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Gas products were measured by an online multi-component analyzer (Gasetm Dx4000, Finland) with a resolution of 8 cm^{-1} . The Gasetm was calibrated with a standard acetone gas cylinder (1%, air balanced). The effective path length of the gas analyzer was 5 m, while the volume of the gas cell was 0.4 L. Measurements were carried out after running the plasma reaction for about 40 min, when a steady-state of the process was reached. All experimental data were obtained by repeating 3 times, with the average value of the three measurements being presented. The removal efficiency of acetone (η_{acetone}) and energy efficiency (EE) of the plasma-catalytic process can be defined as:

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$$\eta_{\text{acetone}} = \frac{c_{\text{in}} - c_{\text{out}}}{c_{\text{in}}} \times 100\% \quad (1)$$

$$EE (\text{g kWh}^{-1}) = \frac{M_{\text{acetone}} \times \eta_{\text{acetone}} \times c_{\text{in}} \times Q}{P \times V_m} \times 3.6 \times 10^6 \quad (2)$$

177

178 where c_{in} and c_{out} are the inlet and outlet acetone concentration (ppm); M_{acetone} is the molar
179 weight of acetone (g mol^{-1}); Q is the total flow rate (L min^{-1}), P is the discharge power (W)
180 and V_m is the gas molar volume.

181

182 **2.3 Response surface exploration**

183 In this work, a three-factor, five-level central composite design was used to investigate
184 the effects of the independent variables and the interactions of these factors on
185 plasma-catalytic removal of acetone using the trial version of Design-Expert[®] 8.05b
186 (Stat-Ease Inc., Minneapolis, USA). Three plasma process parameters, discharge power (X_1),
187 gas flow rate (X_2), and the initial concentration of acetone (X_3), were chosen as the input
188 factors for the design, while the removal efficiency (Y_1) and energy efficiency (Y_2) of the
189 plasma-catalytic process were employed as the responses based on our previous work [19]. A
190 total of 20 experiments including 6 axial points, 8 factorial points and 6 replicates at the
191 central point were designed using the CCD method (Table 1). Each input parameter was
192 coded into five levels as -2, -1, 0, +1 and +2 according to Eq. (3):

$$x_i = (X_i - X_0) / \Delta X_i \quad (3)$$

194 where x_i is the coded value of the i^{th} variable, X_i is the original value of the i^{th} variable, X_0 is
195 the value of X_i at the centre point of the tested data range and ΔX_i is the step size. The levels

196 of the selected plasma processing parameters were given in both coded and real values (**Table**
 197 **1**).

198

199 **Table 1.** Independent variables and their levels used in the CCD method.

Symbols	Variables	Ranges and levels				
		-2	-1	0	+1	+2
x_1	Discharge power (W)	15	17.5	20	22.5	25
x_2	Gas flow rate (L·min ⁻¹)	0.5	0.75	1	1.25	1.5
x_3	Initial concentration of acetone (ppm)	100	150	200	250	300

200

201 In the CCD design, a quadratic polynomial response equation was used to correlate
 202 and describe the relationship between the independent plasma processing parameters and the
 203 responses:

$$204 \quad Y = \beta_0 + \sum_{i=1}^k \beta_i x_i + \sum_{i=1}^k \beta_{ii} x_i^2 + \sum_{i < j} \beta_{ij} x_i x_j + \varepsilon \quad (4)$$

205 where Y, k, x_i and ε are the response, the number of variables, the coded values of
 206 independent variables and the residual value, respectively. β_0 is a constant coefficient, whilst
 207 β_i , β_{ii} and β_{ij} are linear, quadratic and interaction coefficients, respectively. The quality of fit
 208 and the significance of the polynomial model can be identified by the coefficient of
 209 determination (R^2) and the *F*-test, which were completely analyzed by the analysis of variance
 210 (ANOVA). The interactions of the independent variables were investigated by constructing
 211 the response surfaces and contour plots based on the model (Montgomery et al., 1984).

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213 3. Results and discussions

214 3.1 Catalysts characterizations

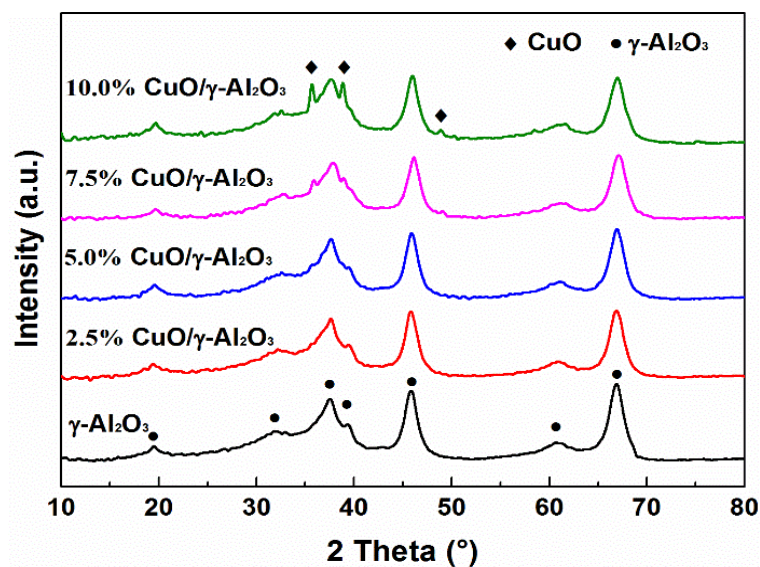
215 The physicochemical properties of the CuO/ γ -Al₂O₃ catalysts were analyzed by N₂
216 adsorption-desorption experiments. The isotherms of all the catalysts are of typical type V,
217 while the hysteresis loops exhibit type H4, indicating the formation of narrow slit-like pores
218 in the catalysts (Lippens and De Boer, 1965; Sing, 1985). The γ -Al₂O₃ support has a large
219 specific surface area (241.6 m² g⁻¹) and a well-developed total pore volume (0.377 cm³ g⁻¹).
220 The specific surface area and total pore volume of the CuO/ γ -Al₂O₃ catalysts decreases from
221 209.6 to 187.8 m² g⁻¹ and from 0.338 to 0.299 cm³ g⁻¹ in the Cu loading amount range of 2.5%
222 to 10%, which can be attributed to the partial coverage of the γ -Al₂O₃ surface by Cu species.
223 In contrast, the average pore diameter slightly increases from 5.01 to 5.19 nm, indicating the
224 clogging of micro-pores in the presence of Cu species (Zakaria et al., 2012).

225 **Table 2.** Physicochemical properties of CuO/ γ -Al₂O₃ catalysts.

Sample	S _{BET} (m ² g ⁻¹)	Total pore volume (cm ³ g ⁻¹)	Average pore diameter (nm)	Amount of H ₂ consumed (μ mol g ⁻¹)
γ -Al ₂ O ₃	241.6	0.377	4.99	-
2.5 wt % CuO/ γ -Al ₂ O ₃	209.6	0.338	5.01	233.4
5.0 wt % CuO/ γ -Al ₂ O ₃	206.7	0.324	5.07	422.7
7.5 wt % CuO/ γ -Al ₂ O ₃	192.8	0.315	5.14	594.0
10.0 wt % CuO/ γ -Al ₂ O ₃	187.8	0.299	5.19	791.9

226
227 The XRD patterns of the CuO/ γ -Al₂O₃ catalysts and γ -Al₂O₃ support are shown in Fig. 2.
228 All the catalysts show diffraction peaks that correspond to the typical cubic structure of
229 γ -Al₂O₃ crystalline (JCPDS 00-010-0425). No obvious diffraction peaks ascribed to the
230 crystalline phase of copper oxides are observed at low Cu loading (2.5 wt % and 5.0 wt%),
231 which suggests the Cu species are well dispersed on the γ -Al₂O₃. The diffraction peaks of

232 crystalline CuO (JCPDS 01-089-5899) located at $2\theta=35.5^\circ$ and 38.8° are clearly seen when
233 increasing the Cu loading amount, indicating the formation of bulk CuO at high Cu loading
234 (7.5 wt % and 10.0 wt%).



235

236

Fig. 2. XRD patterns of the CuO/γ-Al₂O₃ catalysts.

237

238 Fig. 3 shows the H₂-TPR profiles of all the CuO/γ-Al₂O₃ catalysts used in this study as
239 the γ-Al₂O₃ support cannot be reduced within the tested temperature range (Zhu et al., 2015b).

240 The amounts of consumed H₂ were calculated based on the H₂-TPR profiles. As expected, the
241 intensities of the reduction peaks increase significantly with increasing Cu loading amount. It

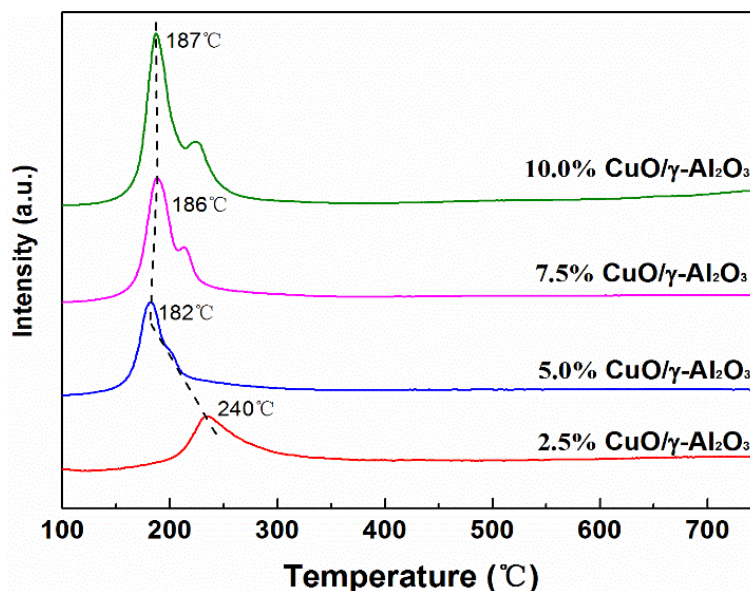
242 can be seen that there exists a single reduction peak located at around 240 °C at the Cu
243 loading amount of 2.5 wt%. At higher Cu loading amounts, the reduction peaks are shifted to

244 lower temperatures. At low loading amount, the existence of isolated Cu species was
245 dominant, leading to a higher reduction temperature (Yamamoto et al., 2002). For the other

246 catalysts, the H₂-TPR profiles show two distinct peaks. The first peak can be attributed to the
247 reduction of highly dispersed CuO species, while the second peak is associated with the

248 reduction of bulk CuO (Águila et al., 2008). The XRD spectra also confirm the existence of
249 bulk CuO at the Cu loading amounts of 7.5 wt% and 10 wt%. The lowest reduction

250 temperature of 182 °C can be observed for the 5.0 wt% CuO/ γ -Al₂O₃ catalyst, indicating that
251 it is easy to activate oxygen species on the surface of the 5.0 wt% CuO/ γ -Al₂O₃ catalyst. At
252 the Cu loading amounts of 7.5 wt% and 10 wt%, the reduction peaks shift to higher
253 temperatures (López-Suárez et al., 2008).



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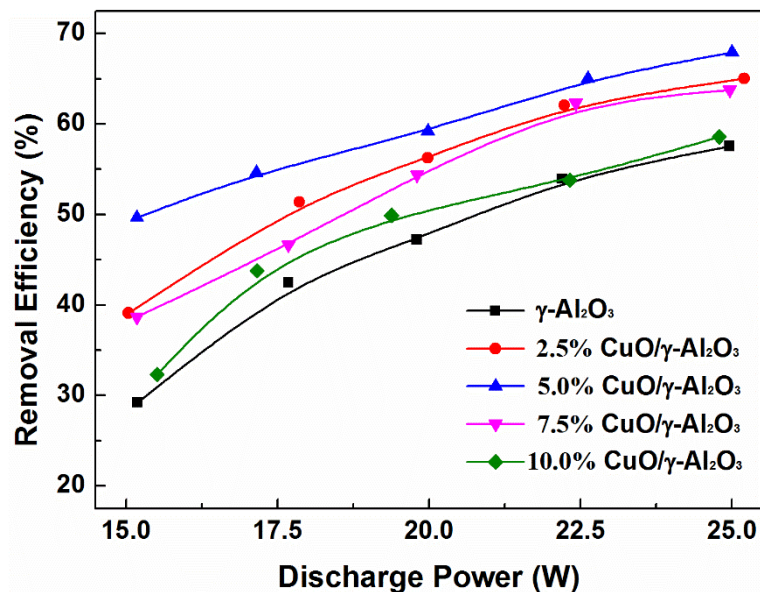
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Fig. 3. H₂-TPR profiles of CuO/ γ -Al₂O₃ catalysts.

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257 3.2 Plasma-catalytic removal of acetone

258 Fig. 4 shows the effect of Cu loading on the plasma-catalytic removal of acetone. The
259 removal of acetone increases monotonically with increasing discharge power regardless of the
260 Cu loading amount. The maximum acetone removal efficiency of 67.9% was achieved at a
261 discharge power of 25 W in the presence of the 5.0 wt% CuO/ γ -Al₂O₃. It is widely recognized
262 that higher discharge power could lead to the formation of more microdischarges in the DBD,
263 which generates more reaction channels and chemical reactive species (e.g., O, O₃, N and
264 metastable N₂) for chemical reactions. These reactive species collide and react with acetone
265 and intermediates, forming reaction products including organic fragments, CO, CO₂ and H₂O.
266 Hence, higher discharge power improves the removal efficiency of acetone in the
267 plasma-catalytic process.



268
269 **Fig. 4.** Effect of Cu loadings on the plasma-catalytic removal of acetone.
270

271 The Cu loading amount significantly affects the reaction performance of the
272 plasma-catalytic oxidation of acetone in the tested discharge power range. The acetone
273 removal efficiency increases with the Cu loading amount up to 5%, while further increasing
274 the Cu loading decreases the reaction performance of the plasma-catalytic process. The
275 catalysts play an important role in the plasma-catalytic system for VOC removal. In the
276 presence of the catalysts, acetone molecules and the organic fragments could be adsorbed and
277 oxidized to CO_2 and H_2O via surface reactions on the surface of the catalysts (Zhu et al.,
278 2015a). The different reaction performances might be attributed to the different
279 physicochemical properties of the $\text{CuO}/\gamma\text{-Al}_2\text{O}_3$ catalysts.

280 As presented in section 3.1, all the $\text{CuO}/\gamma\text{-Al}_2\text{O}_3$ catalysts possessed comparable specific
281 surface area, with no obvious changes found in the structure and crystallites of the catalysts.
282 The large specific surface area of the catalysts could offer many adsorption sites for acetone
283 molecules and intermediates, resulting in a longer residence time of the pollutants in the
284 plasma region and benefitting the removal of acetone. The adsorbed species are further
285 converted via surface reactions driven by surface oxygen species. The H_2 -TPR profiles show

286 that the most easily reducible Cu species (corresponding to the first peak in H₂-TPR profiles)
287 increase significantly from 2.5 wt% CuO/ γ -Al₂O₃ to 5.0 wt% CuO/ γ -Al₂O₃, and continue to
288 slightly increase for higher Cu loadings, indicating the abundance of surface oxygen species
289 with high mobility which could participate in the plasma-induced surface reactions.
290 López-Suárez et al. (2008) reported that a maximum surface Cu loading amount on Al₂O₃ can
291 be achieved at around 5.0 wt% Cu loading, with further addition of Cu decreasing the surface
292 Cu amount. This is in line with the formation of bulk CuO at a high Cu loading amount
293 derived from the XRD results, which could in turn inhibit the surface reactions (Luo et al.,
294 2005). Among the tested catalysts the lowest reduction temperature is found when using the
295 5.0 wt% CuO/ γ -Al₂O₃ catalyst, indicating easier activation of the catalyst. This is closely
296 related to the redox properties of the catalyst and consequently affects the reaction
297 performance.

298 The main gaseous products in the effluent were CO, CO₂ and H₂O, while small amounts
299 of HCOOH, HCHO, NO₂ and N₂O were also detected. The removal of acetone in the
300 plasma-catalytic process could be attributed to the combination of plasma-induced gas phase
301 reactions and plasma-assisted surface reactions on the CuO/ γ -Al₂O₃ catalysts. The plasma gas
302 phase reactions for acetone removal were initiated by direct electron impact dissociation of
303 the carrier gas (air) to form chemically reactive species such as O, OH, N and metastable N₂
304 for the stepwise decomposition and oxidation of acetone and/or intermediates into CO, CO₂,
305 H₂O and other by-products (Fridman, 2008). Acetone molecules can be decomposed by the
306 rupture of C-C and C-H bonds, forming methyl groups (CH₃•) and acetone groups
307 (CH₃COCH₂•). Consequently, acetone radicals can be oxidized by O and OH to acetyl
308 radicals, methyl groups and ketenes. The further oxidation of acetyl radicals leads to the
309 formation of methanol and acetic acid (Magne et al., 2009). Methyl groups can be further
310 decomposed to CH and CH₂ by energetic electrons (Huang et al., 2011). These species can

311 react with O and OH radicals, forming CO, CO₂, HCHO and HCOOH. In the plasma-catalytic
 312 process, catalysts were placed in the plasma region in direct contact with the discharge. Both
 313 short-lived radicals and acetone/intermediates can be adsorbed on the catalyst surfaces to
 314 initiate a series of surface oxidation reactions, forming CO, CO₂, H₂O and by-products.

315

316 3.3 DoE analysis

317 3.3.1 Regression models and data analysis

318 In this work, optimization of the plasma-catalytic removal of acetone was carried out
 319 using the CCD method in the presence of the most active catalyst (5.0 wt% CuO/ γ -Al₂O₃).
 320 The designed experiments and corresponding results of the CCD method are summarized in
 321 **Table 3**. The removal efficiency of plasma-catalytic removal of acetone varies from 44.3% to
 322 81.3%, while the energy efficiency of the plasma-catalytic process is in the range of 0.60 g
 323 kWh⁻¹ to 1.30 g kWh⁻¹. The obtained responses were correlated to the aforementioned
 324 independent plasma processing parameters using the polynomial equation (4). The best-fit
 325 models of removal efficiency and energy efficiency in terms of coded factors are as follows:

Removal Efficiency(%):

$$326 \quad Y_1 = 59.69 + 4.75x_1 - 8.43x_2 - 4.99x_3 + 0.74x_1x_2 - 0.67x_1x_3 + 1.86x_2x_3 \quad (5)$$

$$\quad + 0.075x_1^2 + 0.72x_2^2 + 0.79x_3^2$$

Energy Efficiency(g kWh⁻¹):

$$327 \quad Y_2 = 0.92 - 0.043x_1 + 0.11x_2 + 0.16x_3 + 0.012x_1x_2 - 0.015x_1x_3 + 0.04x_2x_3 \quad (6)$$

$$\quad + 7.222 \times 10^{-3}x_1^2 - 0.021x_2^2 - 7.736 \times 10^{-3}x_3^2$$

328

329 **Table 3.** Experimental design matrix and experimental results of the CCD

Run order	Coded values (X)			Responses (Y)	
	Discharge power (x_1)	Gas flow rate (x_2)	Initial concentration (x_3)	Y ₁ : Removal efficiency (%)	Y ₂ : Energy efficiency (g kWh ⁻¹)

1	20	1	200	59.3	0.91
2	20	1	200	59.2	0.91
3	20	0.5	200	78.5	0.60
4	15	1	200	49.7	1.02
5	20	1	200	59.5	0.92
6	17.5	0.75	250	59.7	0.98
7	20	1	300	51.1	1.18
8	22.5	0.75	250	68.7	0.88
9	22.5	1.25	250	55.4	1.18
10	20	1	200	59.3	0.91
11	17.5	1.25	250	47.2	1.30
12	25	1	200	67.9	0.84
13	22.5	0.75	150	81.3	0.63
14	20	1.5	200	44.3	1.02
15	20	1	100	72.2	0.56
16	22.5	1.25	150	64.3	0.82
17	20	1	200	59.2	0.91
18	17.5	0.75	150	73.4	0.73
19	20	1	200	59.3	0.91
20	17.5	1.25	150	49.7	0.82

330

331 **Table 4** shows the ANOVA of the generated regression models. The results confirm that
332 the models are highly significant since the F-values for both Y_1 and Y_2 are found to be 72.33
333 and 83.20, both of which are greater than the critical value of 3.02 in our case (Montgomery
334 et al., 1984). Moreover, the ultimate low probability value (p -value < 0.0001) indicates the
335 significance of both models at a confidence level greater than 95%. It can be confirmed that
336 most variations in the response can be explained by the generated models considering the high

337 F -values and low p -values. The obtained regression correction coefficients (R^2) (0.9849 for Y_1
338 and 0.9868 for Y_2) are close to unity, indicating the regression models are well fitted to the
339 experimental results. The adequate precision presents the signal-to-noise ratio of the models,
340 while values greater than 4 are desirable. In this study, the adequate precisions are 29.645 and
341 32.610 for the removal efficiency and energy efficiency of the plasma-catalytic process,
342 respectively, which indicate adequate intensities of the signals. The coefficients of variations
343 (C.V.), as the ratio of the standard error of the estimations to the mean value of the responses,
344 could be used to measure the reproducibility of the regression models. The obtained C.V. are
345 2.84% for Y_1 and 3.33% for Y_2 , which are less than the critical value of 10%, indicating the
346 reliability and reproducibility of the models (Mousavi et al., 2014).

347

348 **Table 4.** ANOVA of magnitude and significance of factor effects on the responses

Response	Model terms	Sum of square	Degree of freedom	Mean square	F -value	p -value (Prob.> F)
Removal Efficiency	Model	1957.54	9	217.50	72.33	< 0.0001
	x_1	361.50	1	361.50	120.22	< 0.0001
	x_2	1137.46	1	1137.46	378.28	< 0.0001
	x_3	398.40	1	398.40	132.49	< 0.0001
	x_1x_2	4.43	1	4.43	1.47	0.2527
	x_1x_3	3.58	1	3.58	1.19	0.3005
	x_2x_3	27.58	1	27.58	9.17	0.0127
	x_1^2	0.14	1	0.14	0.047	0.8333
	x_2^2	13.01	1	13.01	4.33	0.0642
	x_3^2	15.62	1	15.62	5.19	0.0459
	Residual	30.07	10	3.01		
	Total	1987.61	19			

R²=0.9849, Adequate precision=29.645, C.V.=2.84%

	Model	0.67	9	0.075	83.20	< 0.0001
	x_1	0.029	1	0.029	32.14	0.0002
	x_2	0.19	1	0.19	211.41	< 0.0001
	x_3	0.42	1	0.42	469.64	< 0.0001
	x_1x_2	1.152E-3	1	1.152E-3	1.28	0.2843
	x_1x_3	1.872E-3	1	1.872E-3	2.08	0.1797
Energy Efficiency	x_2x_3	0.013	1	0.013	14.46	0.0035
	x_1^2	1.311E-3	1	1.311E-3	1.46	0.2551
	x_2^2	0.012	1	0.012	12.88	0.0049
	x_3^2	1.505E-3	1	1.505E-3	1.67	0.2250
	Residual	8.997E-3	10	8.997E-4		
	Total	0.68	19			

R²=0.9868, Adequate precision=32.610, C.V.=3.33%

349

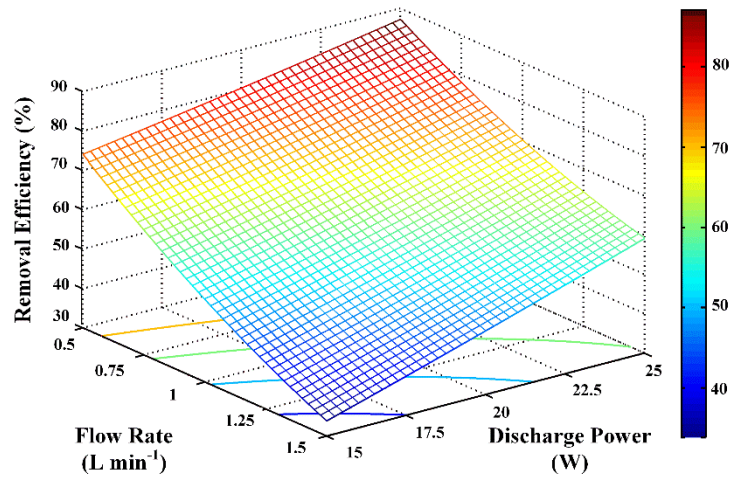
350 3.3.2 Effect of variables on removal efficiency

351 A model term is considered to play an important role in the plasma-catalytic process
 352 once its p-value is below the level of significance (0.05 in this work). In the plasma-catalytic
 353 removal of low concentrations of acetone, x_1 , x_2 , x_3 , x_2x_3 and x_3^2 are identified as the
 354 significant terms for the removal efficiency of acetone, while x_1 , x_2 , x_3 , x_2x_3 and x_2^2 are
 355 important for the energy efficiency of the plasma-catalytic process. Considering the highest
 356 F-value of 378.22, the air flow rate is believed to be the most important factor affecting the
 357 removal efficiency of the plasma-catalytic process. Similarly, the largest F -value of initial
 358 concentration confirms its role in determining the energy efficiency of the process.

359 Three-dimensional (3D) response surfaces and two-dimensional (2D) contours are
 360 presented (**Fig. 5 - Fig. 7**) based on the quadratic polynomial regression models to gain new

361 insights into the effects of each individual factor and their interactions on the plasma-catalytic
362 process. **Fig. 5** shows the combined effect of discharge power and flow rate on the removal
363 and energy efficiency at the initial acetone concentration of 200 ppm (the center level). The
364 acetone removal efficiency increases significantly with an increase in the discharge power and
365 flow rate (shown in **Fig. 5a**). The maximum acetone removal efficiency of 86.2% is achieved
366 at a discharge power of 25 W and a flow rate of 0.5 L min⁻¹. As discussed earlier, the number
367 of micro-discharges increases with increasing discharge power, which could contribute to the
368 generation of more reaction channels and reactive species, and consequently enhance the
369 reaction performance. Significant decreases in removal efficiency are observed with
370 increasing flow rate. The residence time of pollutants at 0.5 L min⁻¹ is 3 times that at 1.5 L
371 min⁻¹. Longer residence time is beneficial for the removal of acetone as the possibility of
372 collisions between the reactive species and the pollutants is much higher than at shorter
373 residence times. The highest energy efficiency of 1.12 g kWh⁻¹ is obtained at a discharge
374 power of 15 W and flow rate of 1.5 L min⁻¹, which may be attributed to heating and excitation
375 of the carrier gas by the dissipated discharge power. Similar observations have been reported
376 elsewhere in cases of VOC removal using either DBD reactors or packed-bed reactors (Zheng
377 et al., 2014). The interactions between the two terms on the reaction performance are regarded
378 as insignificant as the gradients are almost the same at varied flow rates and discharge powers,
379 while the contours are almost linear (Mei et al., 2015). The *p*-values of 0.2527 and 0.2843
380 (greater than the critical value of 0.05) also support this conclusion.

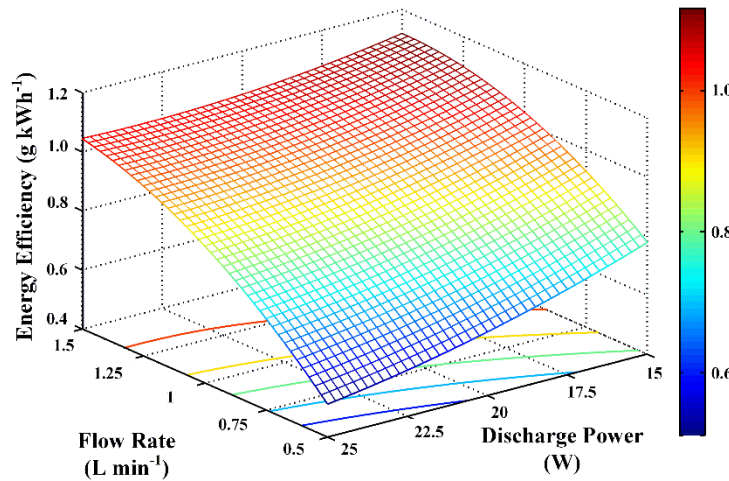
381



382

383

(a)



384

385

(b)

386 **Fig. 5.** Effect of discharge power and flow rate on plasma-catalytic removal of acetone at the

387 initial concentration of 200 ppm: (a) removal efficiency; (b) energy efficiency.

388

389 **Fig. 6** illustrates the effect of discharge power and initial concentration on

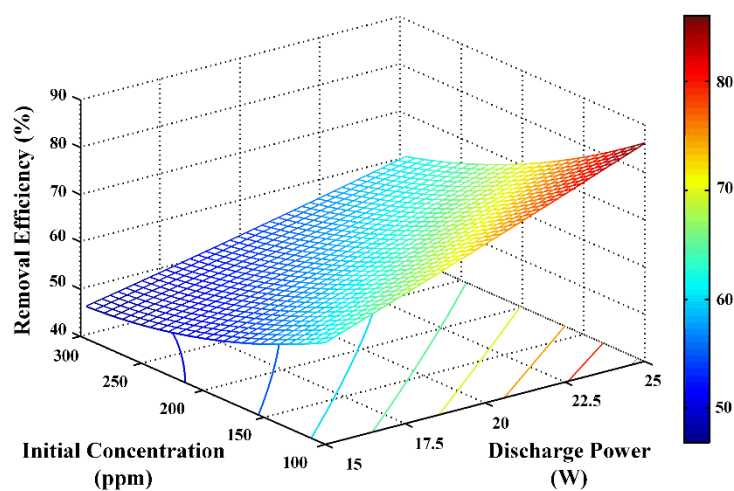
390 plasma-catalytic removal of acetone. The maximum acetone removal efficiency of 85.2% is

391 obtained at a discharge power of 25 W and an initial acetone concentration of 100 ppm, while

392 the highest energy efficiency of 1.51 g kWh⁻¹ is reached at an initial concentration of 300 ppm

393 and a discharge power of 15 W. The removal efficiency of acetone is doubled when the

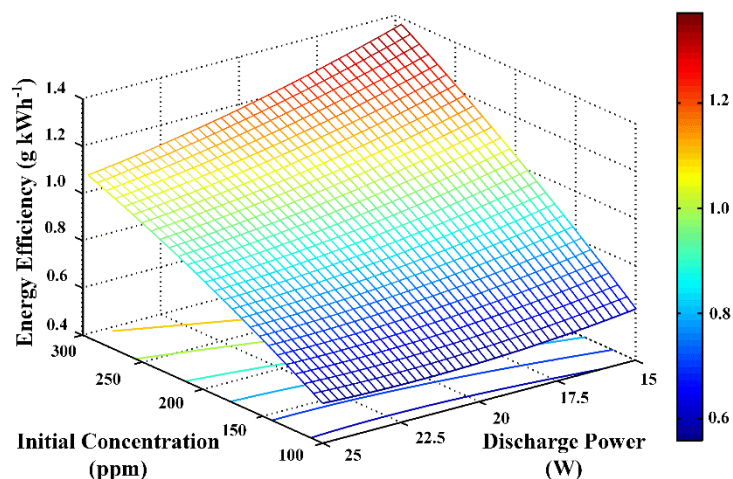
394 discharge power is increased from 15 W to 25 W at 100 ppm, but only increases by 34.8% at an
395 initial concentration of 300 ppm. Higher initial concentration of acetone exhibits a negative
396 effect on acetone removal regardless of the discharge power. For constant reactor and
397 operation parameters, the generation of reactive radicals in the plasma process is almost the
398 same (Nie et al., 2013). At higher initial concentration, more acetone molecules are
399 introduced into the plasma-catalytic system, whilst the concentration of reactive species has
400 been diluted, which lowers the probability of acetone molecules reacting with these reactive
401 species. Consequently, the removal efficiency of acetone decreases with increasing initial
402 concentration. On the other hand, higher initial concentration enhanced the chance of reactions
403 occurring between reactive species and methanol molecules, which led to better utilization of
404 the reactive species. At this point, more acetone molecules can be converted and the energy
405 efficiency of the plasma process is increased at higher initial concentration. The interaction
406 between the discharge power and initial concentration is not significant as the contour lines are
407 linear and the p -values are greater than 0.05, namely 0.3005 for removal efficiency and 0.1797
408 for energy efficiency.
409



(a)

410

411



(b)

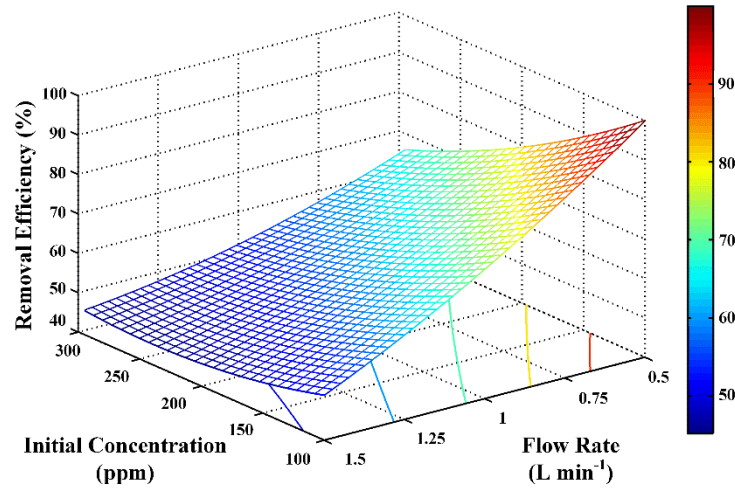
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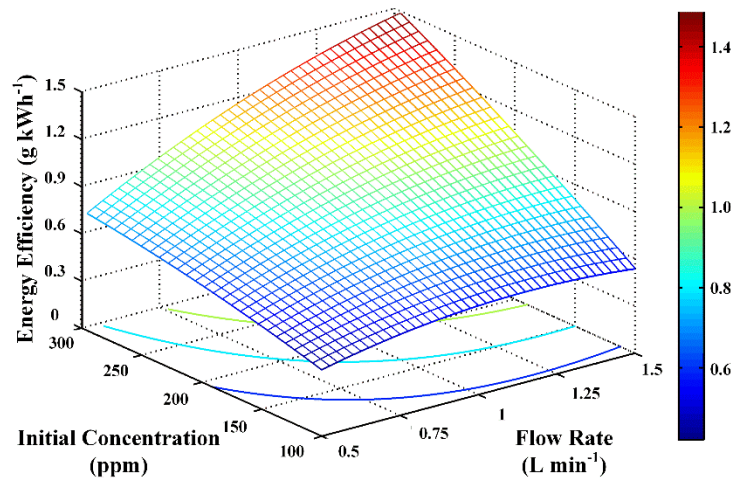
414 **Fig. 6.** Effect of discharge power and initial concentration on plasma-catalytic removal of
 415 acetone at the flow rate of 1 L min⁻¹: (a) removal efficiency; (b) energy efficiency.

416

417 The effect of flow rate and initial concentration on the performance of plasma-catalytic
 418 removal of acetone is plotted in **Fig. 7**. The two terms exhibit a similar effect on the process
 419 performance as discussed before. The highest removal efficiency of 99.8% is obtained at a gas
 420 flow rate of 0.5 L min⁻¹ and an initial concentration of 100 ppm, while the maximum energy
 421 efficiency is reached at a flow rate of 1.5 L min⁻¹ and an initial concentration of 300 ppm. The
 422 removal efficiency of acetone appears to be more sensitive to the flow rate as the gradient of
 423 the removal efficiency is much larger at 100 ppm compared to that at 300 ppm, while the
 424 initial concentration of acetone is more important for improving the energy efficiency,
 425 considering the energy efficiency of the plasma process is almost independent at 100 ppm.
 426 This can also be confirmed by the *F*-values of each term for *Y*₁ and *Y*₂ in **Table 4**. Moreover,
 427 the low *p*-value (<0.001) of the term *x*₂*x*₃ also confirms the strong interactions between initial
 428 concentration and flow rate and the effect these have on the reaction performance of
 429 plasma-catalytic removal of acetone.



(a)



(b)

430

431

432

433

434 **Fig. 7.** Effect of flow rate and initial concentration on plasma-catalytic removal of acetone at

435 a discharge power of 15 W: (a) removal efficiency; (b) energy efficiency.

436

437 4. Conclusions

438 In this work, the plasma-catalytic removal of acetone was investigated using a series of

439 CuO/ γ -Al₂O₃ catalysts. The integration of plasma with the CuO/ γ -Al₂O₃ catalysts significantly

440 improves the removal efficiency of the plasma-catalytic gas cleaning process by 15% to 20%

441 in the tested discharge power range compared to the plasma process using pure γ -Al₂O₃

442 support. The 5.0 wt% CuO/ γ -Al₂O₃ catalyst exhibits the best activity with the maximum
443 acetone removal efficiency of 67.9% at a discharge power of 25 W. Catalyst characterization
444 including BET, XRD and H₂-TPR demonstrates that the activation of surface oxygen species
445 were crucial for the oxidation of acetone molecules and organic by-products on the catalyst
446 surface, which in turn enhances the reaction performance.

447 The effects of various plasma operating parameters, including discharge power, flow rate
448 and initial concentration of acetone, on the plasma-catalytic process and the interactions
449 between these parameters were investigated using CCD method in the presence of the 5.0
450 wt% CuO/ γ -Al₂O₃ catalyst. The generated regression models fits very well with the actual
451 data considering the high coefficient of determination ($R^2=0.9849$ for removal efficiency and
452 0.9868 for energy efficiency). The ANOVA results show that the flow rate was the most
453 significant factor affecting the removal efficiency of acetone and the initial acetone
454 concentration was the most important parameter in determining energy efficiency of the
455 plasma-catalytic process. Moreover, the interactions between flow rate and initial
456 concentration impose a significant effect on the plasma-catalytic process for acetone removal.

457

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462

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